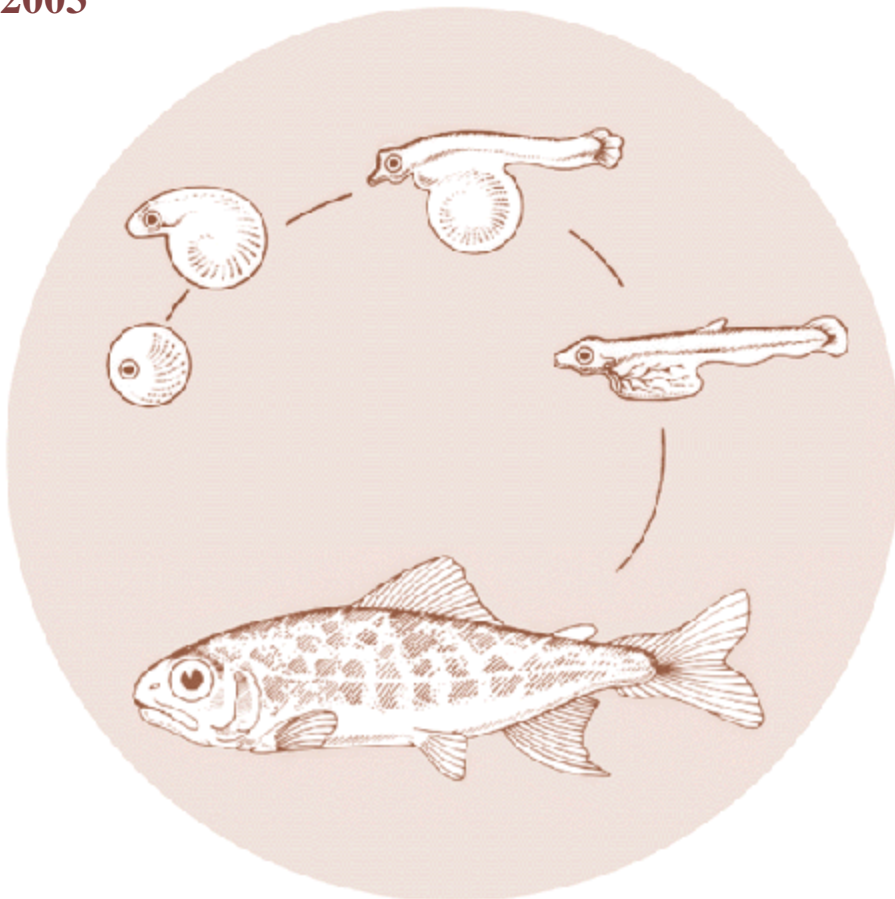


# Coeur d'Alene Tribal Production Facility, Volume III of III

## Submittal to the Northwest Power Planning Council Appendices 2

Technical Report  
2003



This Document should be cited as follows:

*Anders, Paul, John Cussigh, David Smith, Jason Scott, William Towey, Dale Ralston, Ronald Peters, Raymond Beamesderfer, Jeffery Jordan, Ernest Brannon, Douglas Ensor, "Coeur d'Alene Tribal Production Facility, Volume III of III", Project No. 1990-04402, 682 electronic pages, (BPA Report DOE/BP-00006340-4)*

Bonneville Power Administration  
P.O. Box 3621  
Portland, Oregon 97208

This report was funded by the Bonneville Power Administration (BPA), U.S. Department of Energy, as part of BPA's program to protect, mitigate, and enhance fish and wildlife affected by the development and operation of hydroelectric facilities on the Columbia River and its tributaries. The views in this report are the author's and do not necessarily represent the views of BPA.

## **Table of Contents**

- 1. Fisheries Habitat Evaluation on Tributaries of the Coeur d'Alene Indian Reservation, 1993-94 annual report**
- 2. Coeur d'Alene Tribe Fish, Water, and Wildlife Program, Supplementation Feasibility Report on the Coeur d'Alene Indian Reservation, 1998 technical report**
- 3. Fisheries Habitat Evaluation on Tributaries of the Coeur d'Alene Indian Reservation, 1998 annual report**

**FISHERIES HABITAT EVALUATION ON TRIBUTARIES OF  
THE COEUR D'ALENE INDIAN RESERVATION**

**1993,1994 ANNUAL REPORT**

Prepared by:

Kelly Lillengreen  
Angelo Vitale  
Ron Peters

Coeur d' Alene Tribe  
Fish and Wildlife Program

Prepared for:

U.S. Department of Energy  
Bonneville Power Administration  
Environment, Fish and Wildlife  
P.O. Box 3621  
Portland, OR 97208-362 1

Project Number. 90-044  
Contract Number DE-BI79-90BP 10544

September, 1996

---

## **ABSTRACT**

**Bull trout and cutthroat trout are two salmonid species native to the Lake Coeur d'Alene drainage. Historically these species were a critical component of the Coeur d'Alene Tribe's annual subsistence requirements. Since 1932, the cutthroat trout population has declined significantly in the Coeur d'Alene system. The present ecosystem bears little resemblance to habitat composition, diversity and structure of the historic ecosystem.**

**The purpose of this study was to conduct baseline stream and biological surveys of four drainages located within the Coeur d'Alene Reservation and make recommendations on ways to increase the westslope cutthroat and bull trout populations on the Reservation.**

**Data indicated that habitat degradation, specifically, water quantity and lack of habitat complexity, was limiting westslope cutthroat and bull trout populations on the Reservation.**

**Population data indicated that cutthroat trout populations were low when compared to other similar drainages. Surveys revealed a conspicuous absence of bull trout.**

**Recommendations included: Conducting extensive habitat restoration in the study drainages; developing alternate harvest opportunities to reduce pressure on wild stocks; purchasing critical watershed areas for fisheries habitat protection; constructing and operating a trout production facility; and, implementing a five-year monitoring program to evaluate the program effectiveness.**

## **ACKNOWLEDGMENTS**

The authors gratefully acknowledge the following individuals for their assistance in field data collection and laboratory analysis; Deanna Harper, George Aripa, Richard Allen, A.C. Sanchez, Ron Goddard, Jeff Jordan, Cody SiJohn, David SiJohn, Lyle Meshell, Sam Sanchez. Special thanks to Chris Hardy and David St.Amand, for their technical assistance in field work and in editing the final report. Carla Moore provided secretarial skills for formatting and proofing the final report. Funding was provided by the US. Department of Energy, Bonneville Power Administration, Contract Number 99-044-00. Special thank you goes to Tom Morse (BPA Project Manager), Ernie Stensgar (Chairman, Coeur d'Alene Tribal Council).

## TABLE OF CONTENTS

|  |    |
|--|----|
| <b>ABSTRACT</b>  | i  |
| <b>ACKNOWLEDGEMENTS</b>  | ii |
| <b>1.0. INTRODUCTION</b>                                       | 1  |
| 1.1. Fishery Management History of the<br>Coeur d'Alene System | 1  |
| 1.2. Summary of 1990,1991 & 1992 Findings                      | 1  |
| 1.2.1. 1990  | 1  |
| 1.2.2. 1991  | 2  |
| 1.2.3. 1992  | 3  |
| 1.3. Study Objectives  | 6  |
| <b>2.0. METHODS AND MATERIALS</b>                              | 7  |
| 2.1. Description of the Project Area                           | 7  |
| 2.2. Physical Investigations                                   | 7  |
| 2.2.1. Channel Typing  | 12 |
| 2.2.2. Riffle Armor Stability Index                            | 12 |
| 2.2.3. Stream Reach Inventory/Channel Stability Evaluation     | 14 |
| 2.2.4. Habitat Evaluations                                     | 15 |
| 2.2.5. Stream Discharge Measurements                           | 16 |
| 2.2.6. Water Quality Analysis                                  | 16 |
| 2.3. Fisheries Surveys   | 16 |
| 2.3.1. Relative Abundance                                      | 16 |
| 2.3.2. Trout Population/ Biomass Estimates                     | 16 |
| 2.3.3. Habitat Use Evaluation                                  | 18 |
| 2.3.4. Spawning Surveys  | 18 |
| 2.3.5. Migration Trap data                                     | 18 |
| 2.3.6. Age, Growth and Condition                               | 19 |
| 2.4. Macroinvertebrate Surveys                                 | 19 |
| 2.4.1. Benthic Macroinvertebrate                               | 19 |
| 2.4.2. Drift Macroinvertebrate                                 | 20 |
| 2.4.3. Shannon-Weiner Diversity Index                          | 20 |
| 2.5. Analysis of Migration Barriers                            | 21 |
| <b>3.0. RESULTS</b>  | 23 |
| 3.1. Physical Investigations Summary                           | 23 |
| 3.1.1. Lake Creek  | 23 |
| 3.1.2. Benewah Creek   | 35 |
| 3.1.3. Evans Creek   | 56 |
| 3.1.4. Alder Creek   | 68 |
| 3.2. Biological Evaluation                                     | 82 |
| 3.2.1. Relative Abundance                                      | 82 |
| 3.2.1.1. 1993 Lake, Benewah, Evans, and Alder creeks           | 82 |
| 3.2.1.2. 1994 Lake, Benewah, Evans, and Alder creeks           | 89 |
| 3.2.2. Trout Population and Biomass Estimates by Habitat Type  | 96 |

|  |     |
|--|-----|
| 3.2.2.1. Lake Creek                      | 96  |
| 3.2.2.2. Benewah Creek                   | 96  |
| 3.2.2.3. Evans Creek                     | 100 |
| 3.2.2.4. Alder Creek                     | 100 |
| 3.2.3. Habitat Use Evaluation            | 104 |
| 3.2.4. Spawning Surveys                  | 105 |
| 3.2.5. Trout Migration Analysis          | 105 |
| 3.2.6. Age, Growth and Condition         | 113 |
| 3.2.6.1. 1993                            | 113 |
| 3.2.6.2. 1994                            | 120 |
| 3.3. Water Quality and Quantity Analysis | 128 |
| 3.3.1. 1993                              | 128 |
| 3.3.2. 1994                              | 135 |
| 3.4. Macroinvertebrate Surveys           | 135 |
| 3.4.1. Hess and Drift Samples            | 135 |
| 3.4.2. Biological Assessment             | 141 |
| 3.5. Analysis of Migration Barriers      | 141 |
| 4.0. DISCUSSION                          | 143 |
| 4.1. Physical Characteristics            | 143 |
| 4.1.1. Water Quantity and Quality        | 143 |
| 4.1.2. Habitat                           | 144 |
| 4.2. Biological Assessment               | 146 |
| 4.3. Conclusion                          | 147 |
| LITERATURE CITED                         | 150 |
| APPENDICES                               |     |
| Appendix A                               | 154 |
| Appendix B                               | 158 |
| Appendix C                               | 161 |
| Appendix D                               | 165 |
| Appendix E                               | 244 |
| Appendix F                               | 254 |



## 1.0. Introduction

In 1987 the Northwest Power Planning Council amended the Columbia River Basin Fish and Wildlife Program, directing the Bonneville Power Administration (BPA) to fund “A *baseline stream survey of tributaries located on the Coeur d’Alene Indian Reservation to compile information on improving spawning habitat, rearing habitat, and access to spawning tributaries for bull trout (Salvelinus confluentus) cutthroat trout (Oncorhynchus clarki), and to evaluate the existing fish stocks. If justified by the results of the survey, fund the design, construction and operation of a cutthroat and bull trout hatchery on the Coeur d’Alene Indian Reservation; necessary habitat improvement projects; and a three year monitoring program to evaluate the effectiveness of the hatchery and habitat improvement projects. If the baseline survey indicates a better alternative than construction of a fish hatchery, the Coeur d’Alene Tribe will submit an alternative plan for consideration in program amendment proceeding.* In 1990, BPA contracted the Coeur d’Alene Tribe to perform this study. This report contains the results of fiscal year 1993 and 1994 data and the Coeur d’Alene Indian Tribes’ final recommendations for enhancing the cutthroat and bull trout fishery on the Coeur d’Alene Indian Reservation.

### 1.1. Fisheries Management History of the Coeur d’Alene System.

See Graves et al. (1990) for a discussion of the past history of the study area.

### 1.2. Summary of 1990,1991, and 1992 Findings

#### 1.2.1. 1990

Twenty-one creeks, flowing into Lake Coeur d’Alene, and the St. Joe and St. Maries rivers, were initially identified within the study area as having habitat potentially suitable for trout species. Data obtained from an aerial survey further determined that only ten of the original twenty one creeks located within the Coeur d’Alene Indian Reservation contained potential trout habitat (Graves et al. 1990). These tributaries included:

|                |              |
|----------------|--------------|
| Fighting       | Plummer      |
| Bellgrove      | Benewah      |
| Lake           | Hell’s Gulch |
| Squaw          | Evans        |
| Little Plummer | Alder        |

The Missouri method of evaluating stream reaches was modified and used to rank the ten tributaries (Fajen and Wehnes 1981). This ranking, in combination with biological information collected, was used to identify the four

streams with the best potential cutthroat and bull trout habitat (Lillengreen, et. al. 1993).

#### 1.2.2. 1991

During 1991, biological data collected on the ten streams included; relative abundance data, trout population estimates, growth rates, and benthic macroinvertebrate densities (Lillengreen, et. al. 1993). Relative abundance data resulted in the capture of 6,138 fish from June, August and October, 1991. A total of 427 cutthroat trout were collected from all sampled tributaries. Relative abundance of cutthroat trout for all tributaries was 6.7 percent. Fighting Creek had the highest relative abundance of cutthroat trout (93.1 percent). Evans Creek, Lake Creek, Hells Gulch, Alder Creek, Benewah Creek, and Plummer/Little Plummer creeks had relative abundances of 30.8, 12.1, 11 .1, 3.3, 2.1 and 0.5 percent, respectively. No bull trout were captured in any of the surveyed tributaries. For a complete breakdown of fish relative abundance see Lillengreen, et al. (1993).

Population estimates were conducted in only four of the ten tributaries due to intermittent stream conditions found during the summer on the other six selected streams. The four streams in which population estimates were conducted included Benewah, Alder, Evans and Lake creeks. Density estimates for cutthroat trout were 1.2 fish/100 m<sup>2</sup> in Benewah Creek, 1.5 fish/100 m<sup>2</sup> in Alder Creek, 8.1 fish/100 m<sup>2</sup> in Lake Creek and 18.9 fish/100 m<sup>2</sup> in Evans Creek. Density estimates were also determined for eastern brook trout in Alder Creek (11.8 fish/100 m<sup>2</sup>) No bull trout were captured in any surveyed section and are assumed to be absent from the study areas.

Growth rates and condition factors for cutthroat trout captured in each stream tended to be low in comparison to other streams in the region except for Benewah Creek (Lillengreen et al. 1993). Growth rates for cutthroat trout existing in Benewah Creek were comparable to other streams in the region. Eastern brook trout growth and condition factors were also comparable to those found in other streams in the region. Bull trout growth rates and condition factors could not be assessed because no bull trout were captured during the study.

Mean annual invertebrate densities in the tributaries ranged from 1,206 organisms/m<sup>2</sup> in Alder Creek to 2,886 organisms/m<sup>2</sup> in Evans Creek. Mean annual densities in the drift ranged from 21 organisms/m<sup>2</sup> in Alder Creek to 266 organism/m<sup>2</sup> in Evans Creek. Invertebrate densities were similar to other streams of the same size in the region For a more detailed breakdown of invertebrate densities, see Lillengreen et al. (1993).

Land use practices within each selected watershed have contributed to the degradation of the fishery resources on the Coeur d'Alene Indian Reservation. Major habitat problems associated with the area included high

sediment input from non-point sources which included agricultural (grazing and farming) and silvacultural (timber) practices. Stream systems located in low elevation drainages received their primary sources of water from snow melt runoff and rain events. Due to flow constraints (zero flow in summer) and adverse land use practices within the basins, these drainages, have limited potential for resident fish production. However, perennial drainages could have existing land-use practices modified to enhance the habitat quality and quantity for cutthroat and bull trout.

Four out of the ten tributaries, Lake, Benewah, Evans and Alder creeks were chosen for further study based on their relatively high quality fisheries habitat and potential habitat enhancement opportunities.

### 1.2.3. 1992

During 1992, habitat surveys, fish densities, substrate sampling, migratory patterns, discharge measurements, water quality, and predicted cutthroat trout survival were assessed in Lake, Benewah, Evans and Alder creeks.

To estimate relative species abundance, 1,881 fish were captured in May, July and October, 1992. A total of 349 cutthroat trout were collected from all sampled tributaries. Evans Creek had the highest relative abundance of cutthroat trout at 98.8 percent. Twenty-three percent of the fish collected in Alder Creek were cutthroat trout while 44 percent were eastern brook trout. Benewah and Lake creeks had relative abundance values of five percent and two percent, respectively. For a complete breakdown of fish relative abundance see Lillengreen et al. (1993).

Population estimates were conducted in September, 1992. Density estimates for cutthroat trout were 1.4 fish/100 m<sup>2</sup> in Benewah Creek, 11.8 fish/100 m<sup>2</sup> in Alder Creek, 1.5 fish/100m<sup>2</sup> in Lake Creek and 33.0 fish/100 m<sup>2</sup> in Evans Creek. Density estimates were also determined for eastern brook trout in Alder Creek with densities of 6.1 fish/100 m<sup>2</sup>. No bull trout were captured in any surveyed section and were assumed to be absent from the study areas.

Growth rates and condition factors for cutthroat trout captured in each stream tended to be comparable to other streams in north Idaho. Eastern brook trout growth and condition factors were also comparable to those found in other streams in the region. Bull trout growth rates and condition factors could not be assessed since no bull trout were collected in the study area.

Migration trap data in combination with age and growth analyses indicated that Lake and Benewah Creeks had remnant populations of adfluvial westslope cutthroat trout and a resident population of westslope cutthroat trout. Stocks in Alder Creek could not be determined from the data collected and Evans Creek retained only a resident population of cutthroat trout.

Habitat surveys were conducted on each of the four streams. Surveys showed that habitat was a limiting factor for cutthroat and bull trout survival in most of the watersheds.

### Lake Creek

Major factors limiting habitat in Lake Creek were the lack of pool habitat, high water temperatures and cumulative silt loading over time resulting in decreased number of pools and loss of overwintering and rearing habitat. Pools accounted for only eight percent of the total habitat in Lake Creek. Another potentially limiting factor in Lake Creek was the absence of riparian vegetation along most of the stream corridor. Lack of shade from riparian vegetation may be responsible for high summer water temperatures above the acceptable range (>21 °C) for cutthroat trout. High water temperatures create a trout thermal barrier. Substrate sampling indicates that cumulative silt loading over time, has decreased the availability of overwintering and rearing habitat. Average cutthroat trout survival, as estimated from analysis of percent fines is 66 percent.

### Benewah Creek

Factors limiting production in Benewah Creek include, lack of riparian vegetation, unstable stream banks, and high siltation rates. Cattle graze 62 percent of the riparian area, leaving unstable stream banks and little riparian vegetation. Compaction of the soil and removal of vegetation from the riparian zone by cattle grazing, and the amount of cleared uplands due to timber harvesting has increased channel derived sediment into Benewah Creek. Percent fine sediment has contributed to a 16 percent mortality in emergent cutthroat trout fry. The little riparian vegetation that was found in Benewah Creek consisted mainly of shrubs and grasses with an average canopy cover of 3.3 percent.

Trout densities for Benewah Creek were .02 trout/m<sup>2</sup> and .01 trout/m<sup>2</sup> for 1991 and 1992, respectively. These values are far below other north Idaho cutthroat trout streams showing a similar lack of pristine cutthroat trout habitat.

Base flow below 50 percent of the annual stream flow for optimal trout habitat was partly caused by low snow pack, but also water retention time decreased due to the lack of riparian and upland vegetation.

## Evans Creek

Evans Creek supports a healthy population of cutthroat trout. Cutthroat trout densities in Evans Creek were high compared to all other streams surveyed and to other north Idaho streams. Densities of  $.43 /m^2$  were calculated for 1992. The habitat data collected for Evans Creek is optimal, resulting in these high densities of cutthroat trout. Pools comprise 24 percent of the available habitat in Evans Creek. Heavy overhanging cover (53 percent mean canopy cover) and large organic debris provided sufficient in-stream habitat for westslope cutthroat trout.

Percent emergence success was calculated at 74 percent. This mean is biased based on the low survival rate that was calculated for the lower reach. In the lower reach, one hundred percent of the riparian area had been grazed by cattle. Cattle have destroyed the integrity of the stream bank and increased instream sedimentation. The lower reach of Evans Creek served only as a migratory corridor for cutthroat trout and is not used by the resident cutthroat trout population.

Along with cattle grazing, poor forest practices also contributed to the degradation in Evans Creek. Other areas of concern include future timber sales planned for the areas and the number of instream road crossings present.

## Alder Creek

Overall habitat in Alder Creek is sufficient to sustain trout populations. Temperature ranges in Alder Creek are within the acceptable range for cutthroat trout survival, with maximum water temperatures of  $19^{\circ}C$  and  $17^{\circ}C$  for 1991 and 1992. Mean canopy cover averaged 34 percent. In Alder Creek, pools accounted for 26 percent of the total available habitat. This is the only stream in the study that contained side channels, with a total of 653 meters which provides lateral habitat for young-of-the-year fish and juvenile rearing. Ninety-one percent cutthroat survival from egg to swim-up fry existed in Alder Creek. Cutthroat trout densities were calculated at  $0.12/m^2$  for 1992. Eastern brook trout densities of  $.06/m^2$  were calculated in 1992. Cutthroat trout densities were low compared to other cutthroat trout streams in north Idaho, while eastern brook trout densities were comparable to other north Idaho streams.

The presence of eastern brook trout in Alder Creek may also be limiting the westslope cutthroat trout population.

### **1.3. Study Objectives**

The objectives of this study were to:

- **Conduct in-depth habitat evaluations of the four primary tributaries which included; estimates of amount of habitat (i.e. pools, riffle, cascades and side channels), estimate of instream and overhang cover; mass wasting (slope failure); bank cutting; vegetative type; and seral stage along stream corridor.**
- **Determine the population dynamics of trout species present in each tributary.**
- **Determine habitat utilization**
- **Determine migratory behavior patterns of trout in each stream in order to assess stocks present (adfluvial, fluvial, or resident).**
- **Assess age, growth and condition of cutthroat and bull trout.**
- **Determine extent and effectiveness of cutthroat and bull trout spawning.**
- **Identify alternatives for restoring cutthroat and bull trout**
- **Establish biological objectives based on restoration alternatives.**
- **For each drainage, list habitat improvement opportunities and cost estimates for those improvements.**

## **2.0. Methods and Materials**

### **2.1. Description of the Project Area**

The Coeur d'Alene Lake drainage system is located in the Idaho panhandle and extends approximately 9,583 square kilometers. It is divided into two subbasins, including the Coeur d'Alene River basin and the St. Joe River basin. The remainder of the drainage basin consists of streams flowing into Wolf Lodge, Corbin, Windy, Rockford, Mica, and Cougar bays of Lake Coeur d'Alene.

The project area included four tributaries located within the Coeur d'Alene drainage basin: Lake, Benewah, Evans, and Alder creeks.

The Lake Creek watershed (Figure 2.1) is located in southwest Kootenai County, Idaho and southeast Spokane County, Washington. Lake Creek discharges into Lake Coeur d'Alene at Windy Bay. Lake Creek is a third order stream and is approximately 21 kilometers long. Over half of the watershed is forested land while the remainder is agricultural land. Lake Creek is used as a domestic and limited livestock water source.

The Benewah Creek watershed (Figure 2.2) is located in Benewah County, Idaho and is a fourth order stream. Benewah Creek discharges into the southern portion of Benewah Lake, which since the raising of the water levels associated with the Post Falls Dam, is part of Lake Coeur d'Alene. Benewah Creek is approximately 24 kilometers long. Predominant land use practices within the watershed are grazing, timber, and residential uses.

The Evans Creek watershed (Figure 2.3) is located in Kootenai County, Idaho and is a second order stream. Evans Creek discharges into Medicine Lake, a lateral lake associated with the Coeur d'Alene River. Evans Creek is approximately ten kilometers long. Land uses associated with Evans Creek include forestry, grazing, and residential uses. Evans Creek is used as a domestic and livestock water source.

The Alder Creek watershed (Figure 2.4) is located in Benewah County, Idaho and is a fourth order stream. Alder Creek discharges into the St. Maries River and is approximately 20 kilometers long. The major land use practices within the watershed are private/industrial timber production and livestock grazing. Alder Creek is also used as a livestock and limited domestic water source.

### **2.2. Physical Investigations**

Physical investigations were conducted on the four tributaries and included; channel typing, habitat evaluations, riffle armor stability ratings, stream reach channel stability profiles, discharge profiles, and water quality analysis.

Figure 2.1  
Lake Creek  
Stream Reaches



**Legend**

|  |                         |
|--|-------------------------|
|  | Stream Reach Break      |
|  | Reservation Boundary    |
|  | Unimproved Roads        |
|  | Secondary Roads         |
|  | Primary Roads           |
|  | Perennial Streams takes |



It is important to remember that all maps are interpretations of what is believed to be on the ground and that the maps may not always be correct in their interpretation. The base themes for this map came from USGS 1:24000 quadrangle maps and the accuracy of this map is limited to that scale.

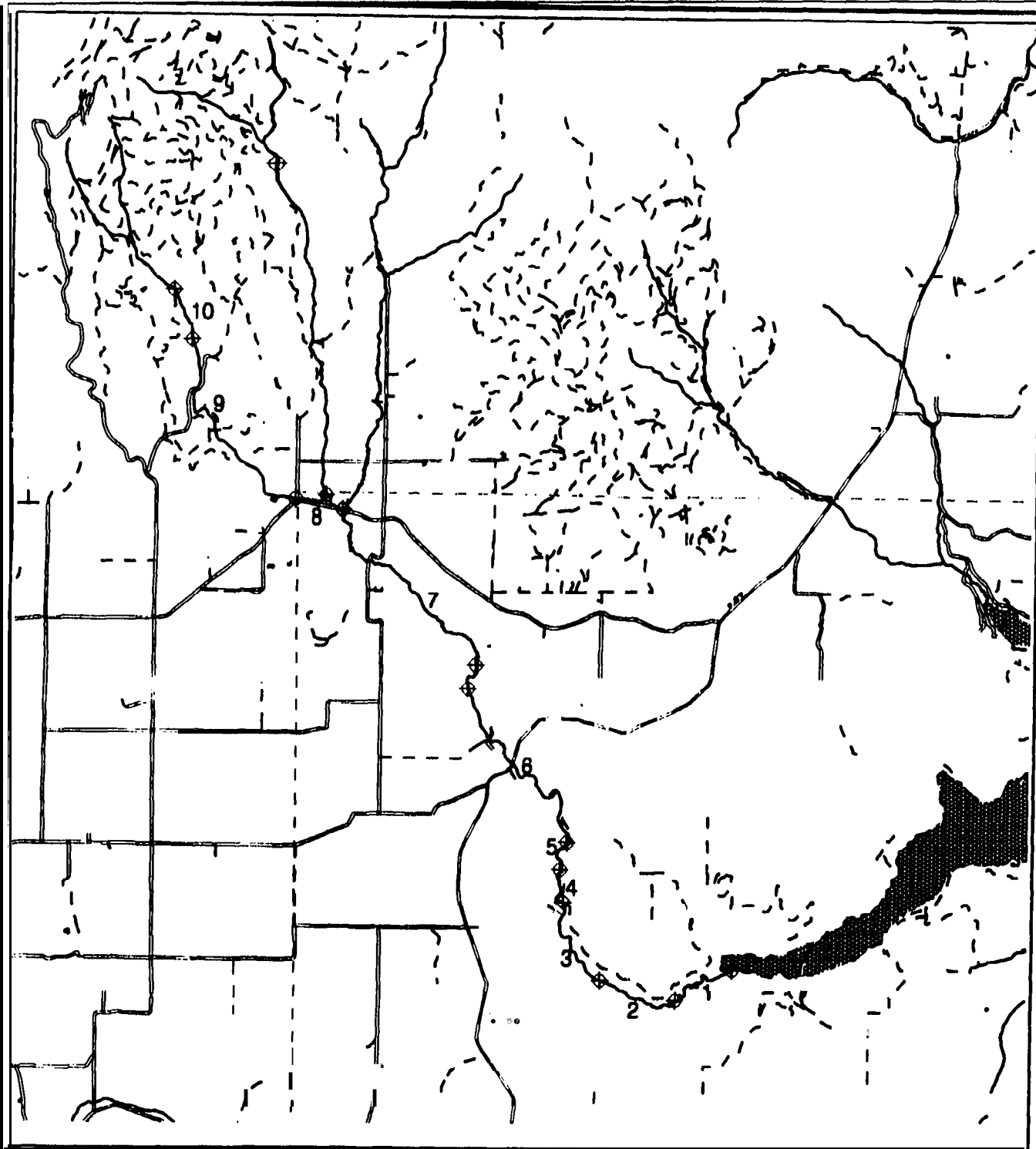




Figure 2.2  
Benewah Creek  
Stream Reaches

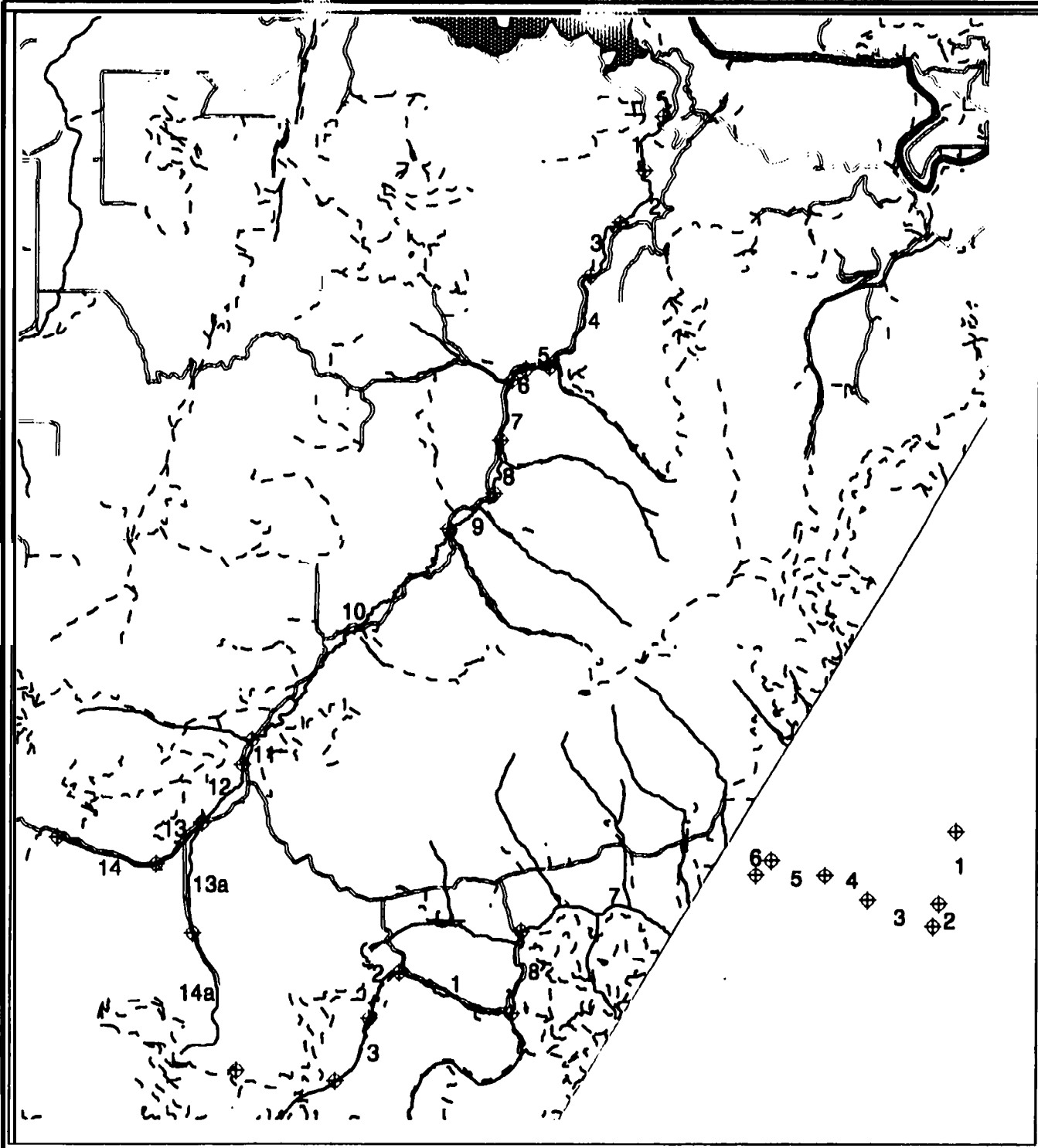


Legend

|  |                       |
|--|-----------------------|
|  | Stream Reach Break    |
|  | Reservation Boundary  |
|  | Secondary Paved Roads |
|  | Gravel Roads          |
|  | Unimproved Roads      |
|  | Perennial Streams     |
|  | Rivers                |
|  | Lakes                 |



It is important to remember that all maps are interpretations of what is believed to be on the ground and that the maps may not always be correct in their interpretation. The base themes for this map came from USGS 1:4000 quadrangle maps and the accuracy of this map is limited to that scale.



**Figure 2.3  
Evans Creek  
Stream Reaches**



**Legend**

|  |                      |
|--|----------------------|
|  | Stream Reach Break   |
|  | Reservation Boundary |
|  | Unimproved Roads     |
|  | Secondary Roads      |
|  | Primary Roads        |
|  | Perennial Streams    |
|  | Intermittent Streams |
|  | Lakes                |



It is important to remember that all maps are interpretations of what is believed to be on the ground and that the maps may not always be correct in their interpretation. The base themes for this map came from USGS 1:24000 quadrangle maps and the accuracy of this map is limited to that scale.  
Map produced by the Coeur d'Alene Tribe G.I.S.

**Figure 2.4  
Alder Creek  
Stream Reaches**

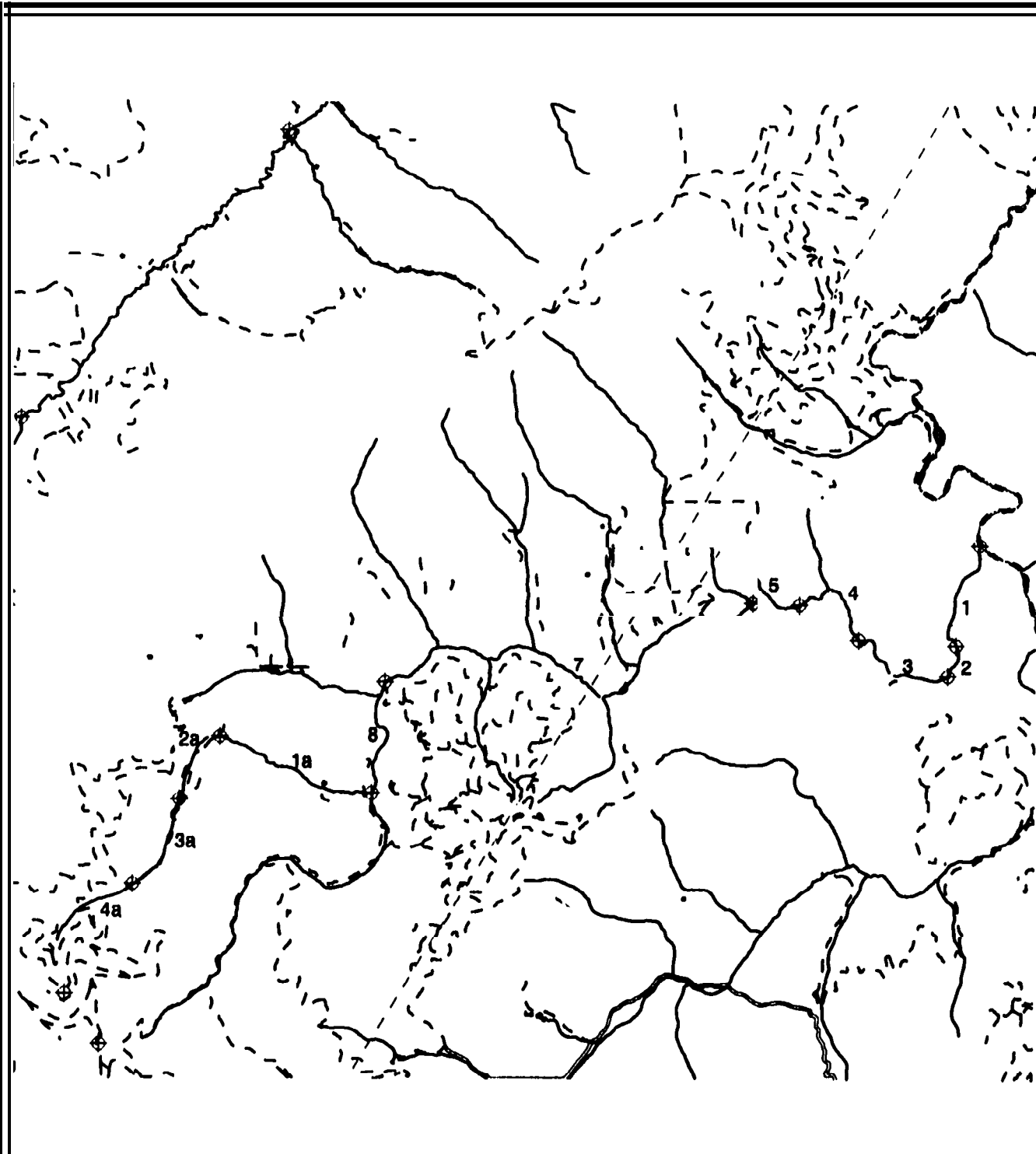


**Legend**

|  |                             |
|--|-----------------------------|
|  | <b>Stream Reach Break</b>   |
|  | <b>Reservation Boundary</b> |
|  | <b>Unimproved Roads</b>     |
|  | <b>Secondary Roads</b>      |
|  | <b>Perennial Streams</b>    |
|  | <b>Rivers</b>               |
|  | <b>Lakes</b>                |



It is important to remember that all maps are interpretations of what is believed to be on the ground and that the maps may not always be correct in their interpretation. The base themes for this map came from UBS 124000 quadrangle maps and the accuracy of this map is limited to that scale.



### **2.2.1. Channel Typing**

Channel type surveys were conducted to classify stream reaches into relatively homogeneous types according to broad geomorphological characteristics of stream morphology. Geomorphic parameters were identified and classified according to Rosgen (1991). This type of reach level classification allows for the characterization of channel processes and provides a framework to examine potential channel response, define impact zones for interpreting effects on fisheries habitat, and identify areas best suited for improvement projects.

Surveys were conducted by two person crews beginning at the mouth of streams and working in an upstream direction. Reaches were numbered in ascending order from downstream to upstream. Measured parameters included: bankfull width, bankfull depth, flood-prone width, entrenchment ratio, stream gradient, sinuosity, and dominant substrate. Measurements of bankfull width and depth were determined following visual identification of channel features related to the 1.5 year recurrence interval discharge (Williams 1978). Flood-prone width was extrapolated from a flood-prone depth estimated at two times bankfull depth (Rosgen, 1994). Entrenchment ratio was calculated as the ratio of the width of the flood-prone area to the bankfull width of the channel. Stream gradient was measured to the nearest percent using a handheld clinometer. Sinuosity was determined in the office following examination of aerial photos. Dominant substrate was determined by direct measurement of particle size as described in methods for the Riffle Armor Stability Index (Section 2.2.2). The classification scheme assigns an alphanumeric/numeric code to each stream reach based on measured parameters. Broad stream categories of A-G are assigned based on plan view morphology, entrenchment, width/depth ratio, and sinuosity. The stream types are then broken into discreet slope ranges and dominant channel-material particle sizes. The stream types are given numbers related to the median particle size diameter of channel materials (Figure 2.1). Channel types were determined following 1993 surveys. Measurements were not repeated during the 1994 field season.

### **2.2.2. Riffle Armor Stability Index**

The Riffle Armor Stability Index (Kappesser 1992) provides a quantitative determination of existing channel stability by comparing the size classes of bed material in riffle areas with the largest sizes of material capable of movement at bankfull discharge. The index draws on the principles that stream channel stability is the product of a balance between streamflow and sediment load (Leopold et. al. 1984) and that the composition of a stable riffle is representative of bedload moved at a range of floods (Lisle 1989). Marked differences in the particle size distribution of riffle areas and the geometric mean particle size on

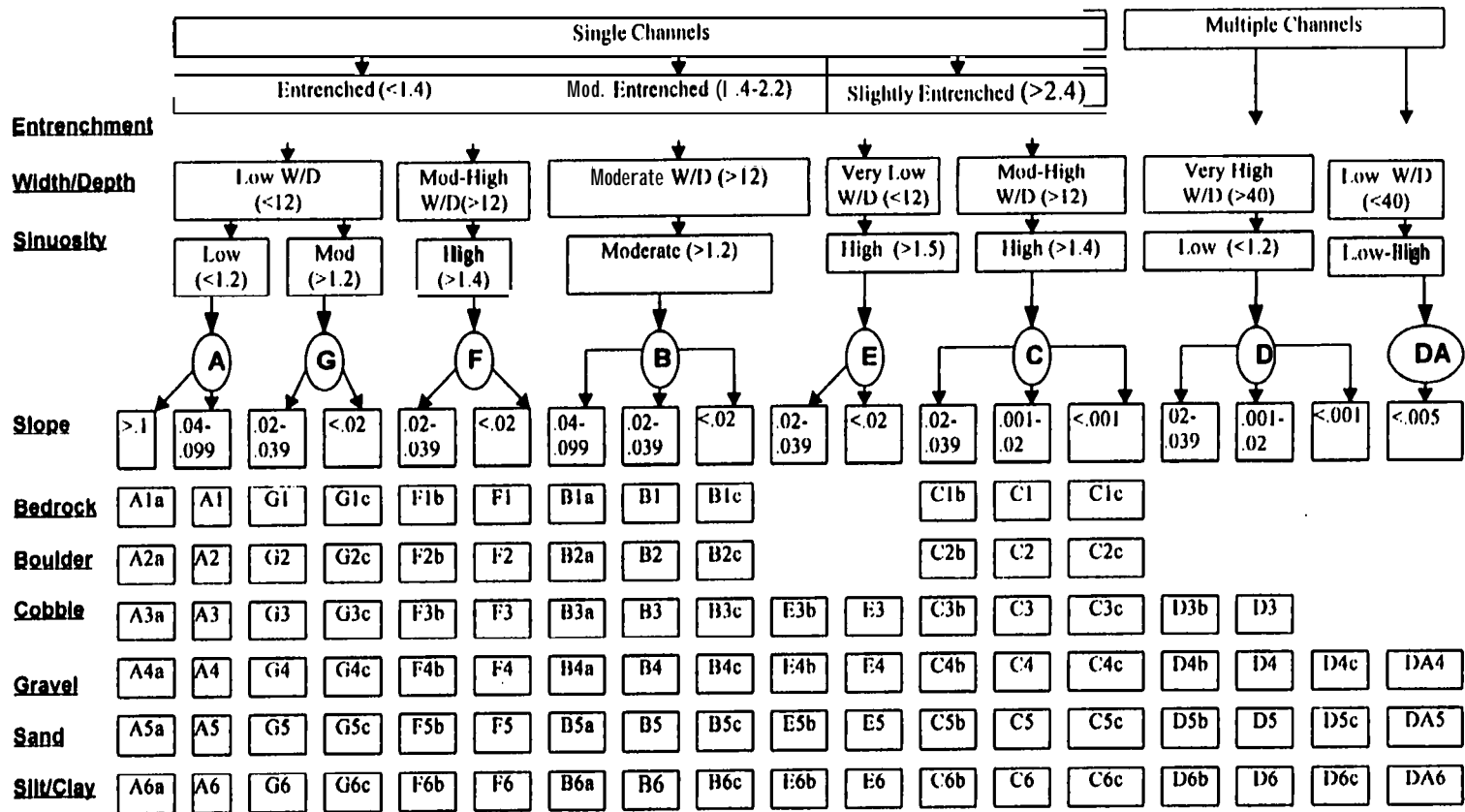


Figure 2.5. Key to channel classification (Rosgen 1994).

fresh depositional areas represents a shift in the equilibrium between streamflow and sediment supply. The index is a useful monitoring tool which provides a record of changes in channel stability over time, and an estimate of the effects of sediment transport on beneficial uses (e.g., filling of pools).

The procedure requires a set of measured field data which includes the particle size distribution of riffle material and a sample from a fresh depositional feature. Particle size distribution was determined (Wolman 1954) at three representative riffles within each of the Rosgen channel types. Sampled riffles were marked with flagging so that measurements could be repeated during 1994 surveys. In instances where channel adjustments significantly altered riffle habitats, measurements were repeated in the riffle area located immediately downstream from 1993 sample locations. Measurements of the largest particles on a fresh depositional feature were collected in close proximity to the riffles examined during each year of the survey. Channel dimensions (e.g., slope, width, depth) were taken in the same area where pebble counts were conducted so that data would be compatible for subsequent analysis. The geometric mean particle size was compared to the cumulative percent of riffle particle sizes to obtain a RASI stability number (Kappesser 1992). The RASI numbers can range from less than 50 to 100, and riffle stability is inversely related to the RASI number.

### **2.2.3. Stream Reach Inventory/Channel Stability Evaluation**

The Stream Reach Inventory/Channel Stability Evaluation (Pfancuch 1975) provides a semi-quantitative assessment of the capacity of a stream to adjust to changes in flow and/or sediment production. The procedure assigns a stability rating of excellent to poor based on the evaluation of fifteen characteristics related to the channel and adjacent floodplain (Table 2.2). Survey crews conducted Stream Reach Inventory/Channel Stability Evaluations for each Rosgen channel type. The scoring for each of the fifteen criteria considered conditions that were representative of the entire reach in question. Scores for each of the criteria were then summed to provide an overall stability rating for the reach. Surveys were repeated during the 1994 sampling effort.

**Table 2.2. Evaluating criteria for Stream Reach Inventory/Channel Stability Evaluation surveys.**

| <b>Landform</b>    | <b>Item Rated</b>  | <b>Criteria Number</b> |
|--------------------|--|------------------------|
| <b>Upper Banks</b> | <b>Landform Slope</b>  | <b>1</b>               |
|                    | <b>Mass Wasting Potential</b>                                | <b>2</b>               |
|                    | <b>Debris Jam Potential</b>                                  | <b>3</b>               |
|                    | <b>Vegetative Bank Protection</b>                            | <b>4</b>               |
| <b>Lower Banks</b> | <b>Channel Capacity</b>                                      | <b>5</b>               |
|                    | <b>Bank Rock Content</b>                                     | <b>6</b>               |
|                    | <b>Obstructions/Flow Deflectors/Sediment Traps</b>           | <b>7</b>               |
|                    | <b>Bank Cutting</b>  | <b>8</b>               |
|                    | <b>Deposition</b>  | <b>9</b>               |
| <b>Bottom</b>      | <b>Rock Angularity</b>                                       | <b>1 0</b>             |
|                    | <b>Substrate Brightness</b>                                  | <b>11</b>              |
|                    | <b>Substrate Consolidation or Particle Packing</b>           | <b>1 2</b>             |
|                    | <b>Bottom Size Distribution and Percent Stable Materials</b> | <b>1 3</b>             |
|                    | <b>Scouring and Deposition</b>                               | <b>1 4</b>             |
|                    | <b>Clinging Aquatic Vegetation</b>                           | <b>1 5</b>             |

#### **2.2.4. Habitat Evaluations.**

Habitat surveys were conducted on primary tributaries during May-October, 1993 and 1994, as described in Lillengreen et. al. (1993) with the following modifications. Stream reaches were delineated based on Rosgen channel typing (Section 2.2.1).-Data was recorded on standardized TFW forms and entered into Microsoft Excel 4.0 software. All parameters analyzed were the same as reported in Lillengreen et. al. (1993) with the exception of the incorporation of residual pool depth calculations.

Residual pool depth is the maximum depth of a pool at stage zero discharge. The measurement is recorded as the difference between the thatweg depth at the hydraulic control and the deepest point in the pool. Residual pool depth was calculated following the methods in Lisle (1989) where:

$$\text{RPD} = \text{MPD} - \text{TCD}$$

Where:

**MPD = Maximum pool depth**  
**TCD = Tail crest depth**

### **2.2.5. Stream Discharge Measurements**

Stream discharge was measured monthly from March 1993 through November 1993, and March 1994 through November 1994, as described in Lillengreen et. al (1993). In March, discharge measurements were collected more frequently to measure runoff due to rain-on-snow events. Sampling after the month of March was completed monthly but was event orientated until base flow was established in June. Sampling was then completed the third week of the month during 1993 and 1994.

### **2.2.6. Water Quality Analysis**

Water samples were collected and analyzed monthly and reported as seasonal averages. Spring samples were collected March through May. Summer samples were collected June through August and fall samples were collected September through November. Tests for conductivity, dissolved oxygen, pH, and temperature were conducted in the field using a Hydrolab Surveyor II. Water samples were also collected for laboratory analysis of nitrate, nitrite, phosphates, turbidity, and alkalinity using a LaMotte Chemical colorimetric test kit. Total dissolved solids were determined using a HANNA model 0661-10 dissolved solids tester.

## **2.3. Fisheries Surveys**

### **2.3.1. Relative Abundance**

Tributaries were sampled May through October, excluded September in 1993 and 1994, as described in Lillengreen et al (1993).

### **2.3.2. Trout Population/Biomass Estimates**

Trout populations were estimated in October 1993 using the removal-depletion method (Seber and LeCren 1967, Zippen 1958). Population sites were located within each channel type reach as outlined in Section 2.2.1. Sites within each reach were randomly distributed to include 10 percent of each habitat type present. Block nets were placed at the upstream and downstream boundaries to prevent immigration and emigration. Each section was electrofished using the standard guidelines and procedures described by Reynolds (1983). Fish were collected by spot shocking using a Smith-Root Type VII pulsed-DC backpack electrofisher. A minimum of two electrofishing passes were made for each habitat unit. Captured fish were identified, enumerated, measured (TL to nearest mm), and weighed. Cutthroat trout of 200 mm in length and larger were tagged with a Floy FD-6B numbered anchor tag.



Population estimates were calculated using the following equation (Armor et. al. 1983):

$$N = \frac{(U_1)^2}{(U_1 - U_2)},$$

Where:

**N** = estimated population size;  
**u<sub>1</sub>** = number of fish collected in the first pass;  
and,  
**U<sub>2</sub>** = number of fish collected in the second pass

The standard error of the estimate was calculated as:

$$se(N) = \sqrt{\frac{(U_1)^2(U_2)^2T}{(U_1 - U_2)^4}}$$

where:

**se(N)** = standard error of the population estimates; and  
**T** = total number of fish collected (**U<sub>1</sub>**+**U<sub>2</sub>**)

An approximate 95 percent confidence interval for N (true population size) was calculated as  $\pm 2 \times se(N)$ .

The population estimates were converted into density values (# fish/100ft) by habitat type and by reach. The confidence intervals were converted in the same manner. Total standing crop of trout in each of the four drainages was arrived at by calculating the sum of population estimates for individual reaches and habitat types, respectively.

Biomass for cutthroat and eastern brook trout within each sample site was estimated as **NW**, where **W** estimates the average weight of all fish of a species that **N** relates to. Biomass estimates were stratified into juvenile and adult age classes. Juveniles were assumed to be less than 4 years old, while the adult age class consisted of all fish 4 years old and older. Site specific back calculations of lengths at age were used to stratify fish into age classes. Total biomass of trout in each of the four study drainages was arrived at in the same manner as described above for population estimates.

### **2.3.3. Habitat Use Evaluation**

Habitat use in the four study drainages was evaluated through chi-square goodness of fit tests (Daniel, 1990) and comparisons of catch rates by habitat type. Analysis was conducted on 1993 electrofishing data. The null hypothesis tested by the chi-square analysis stated that trout populations do not differ in their relative frequency distribution among habitat types. It was assumed that each habitat type, by definition, constituted a unique set of morphologic and hydrologic characteristics (e.g. substrate size, cover type, average depth, average velocity). Catch rates served as a relative index of habitat selectivity. The assumption was that probability of capture is not biased towards species or habitat type.

### **2.3.4. Spawning Surveys**

Spawning surveys were conducted in late April and early May, 1993 to assess cutthroat trout spawning success. A two member field crew walked from the mouth of the stream to the upper limit of fish habitat in each stream. Redds were located, counted, classified, and marked on topographic maps following procedures described in Lillengreen, et. al. (1993).

### **2.3.5. Migration Trap Data**

In March, 1993 and 1994 upstream and downstream migration traps were installed in Lake, Benewah, Evans, and Alder creeks as described in Lillengreen et al (1993). The traps consisted of a weir, runway and a holding box (Figure 2.6). The design was a modification of the juvenile downstream trap found in Conlin and Tuty (1979) (Lillengreen et al. 1993). Traps were installed in pairs to allow monitoring of both upstream and downstream movements of fish (Table 2.35). Paired traps were placed approximately 200 meters apart.

Traps were check twice daily during peak spawning periods from mid-March through the middle of May and once daily afterwards until late June, when traps were removed. Fishes captured in the traps were identified, counted, measured, and weighed. A scale sample was taken to assess the age growth, condition, and stock (fluvial/adfluvial/resident) of the fish.

Table 2.3.5. Location of migration traps, 1994.

| Lake Creek                               | Benewah Creek                                 | Alder Creek        | Evans Creek        |
|--|---|--------------------|--------------------|
| <u>Lower Trap:</u>                       | <u>Lower Trae:</u>                            | <u>Lower Trap:</u> | <u>Lower Trap:</u> |
| At Hwy 95 crossing                       | At mouth of Creek                             | At mouth of creek  | At mouth of creek  |
| <u>Upper Trap:</u>                       | <u>Upper Trap:</u>                            |                    |                    |
| Upstream of confluence with Bozard Creek | Upstream of confluence with SE. Benewah Creek |                    |                    |
| <u>Bozard Creek:</u>                     | <u>S.E. Benewah:</u>                          |                    |                    |
| At mouth of creek                        | At mouth of creek                             |                    |                    |
|  | <u>Coon Creek:</u>                            |                    |                    |
|  | At mouth of creek                             |                    |                    |
|  | <u>Whitetail Creek:</u>                       |                    |                    |
|  | At mouth of creek                             |                    |                    |

### 2.3.6. Age, Growth, and Condition

Age growth and condition were determined as described in Lillengreen et al (1993).

## 2.4. Macroinvertebrate Surveys

### 2.4.1. Benthic Macroinvertebrates

Benthic macroinvertebrate samples were collected in July, August, and October 1994. Benthic macroinvertebrate densities were collected using the methods of Waters and Knapp (1961). A modified Hess sampler, with an area of 0.1 m<sup>2</sup>, and a net aperture of 390 µm, was pushed approximately 10 cm into the

substrate at three sites across the width of the stream. Stones found in the area were removed and cleaned of all attached material. The substrate was then disturbed by stirring to obtain any remaining macroinvertebrates. The sample was then preserved in 10 percent formalin and transferred to a 70 percent alcohol solution in the lab. Samples were collected in the same areas as the fish collections for feeding habits studies during all three sampling months.

#### 2.4.2. Drift Macroinvertebrates

Two drift samples were collected upstream from fish electroshocking areas in each tributary during each sampling month. Water depth was measured using a wading rod, while velocity measurements were measured directly in front of the sampler at 0.6 of the water depth, using a Price Pygmy current meter (Buchanan and Somers.1 980). These measurements allowed for the calculation of densities of organisms per volume of water passing through the drift. All samples were preserved in the field using 10 percent formalin and transferred to a 70 percent alcohol solution in the lab.

#### 2.4.3. Shannon-Weiner Diversity Index

To determine if a stream was healthy or unhealthy the Shannon-Weiner diversity index  $H'$  was used (Perkins 1982). With this method the number of species as well as the number of individuals within each species are taken into account (Krebs 1985). The lowest value would be obtained when only one species is represented in a stream. The highest value would be obtained when each species is represented by equal numbers of individuals. This equation was:

$$H' = - \sum_{i=1}^s (P_i) (\log_2 P_i),$$

where:  $H'$  = Index of species diversity;  
 $s$  = Number of species; and  
 $P_i$  = Proportion of total sample belonging to the  $i$ th species.

Values above three represent high diversity and normally indicates a healthy unpolluted community. A low diversity of less than two usually indicates an unhealthy and possibly polluted community (Her-ricks and Cairns 1974). Densities and diversities were then compared to other area streams.

## 2.5. Analysis of Migration Barriers

Survey information collected on culverts and stream channel geometry was used to assess culvert velocity barrier potential, jumping distance, and other site characteristics. Culvert diameter, bed slope, length, as well as rust level and pipe condition were noted. Stream channel geometry measurements were taken upstream to assess discharge.

Because length and velocity are the key elements in determining barrier potential, the Mannings equation was solved for velocity as follows:

$$v = 1.49/n(R)^{0.67} (sp)^{0.5}$$

Where:

$v$  = velocity in FPS (feet per second)

$n$  = Roughness in culverts assumed to be 0.035

$R$  = Hydraulic radius

$sp$  = bed slope

The wetted perimeter is computed for circular pipes by:

$$P = 1/2@d$$

and cross section area by:

$$A = 1/8 (@-\sin@)d^2$$

The maximum fish speed used in this passage analysis is

$$VF = VFB(Ffc)$$

Where:

$VFB$  = Maximum burst speed as found in Bell (1987)

$CFC$  = Coefficient of fish condition

Therefore:  $VF = (12.48)(.75)$

**VF for cutthroat trout = 9.36 (Bell, 1964)**

$$\mathbf{Lfs = (VF-VW)TF}$$

**Lfs = length fish can swim**

**VF = fish speed +9.36**

**V W = Water Velocity**

**TF = Time to fatigue**

**Passage is analyzed by computing the time it would take the fish to travel the length of the culvert. Time requirement is computed by:**

$$\mathbf{T = U(VF-Vc)}$$

**Where:**

**T = Time of passage through culvert,**

**L = length of culvert**

**vc = the velocity through culvert**

**Vf = fish speed**

### **3.0. Results**

Physical and biological surveys were conducted on each of the four drainages during 1993 and 1994. Physical surveys conducted on these drainages included, channel typing (Rosgen 1991) channel stability ratings, riffle armor stability index values, habitat evaluations, culvert analysis, stream discharge and water quality analysis. Biological surveys conducted on these drainages included relative abundance estimates, population estimates, biomass estimates. Other information was also derived from the above methods. Specific Information derived from each of the above procedures can be found in Appendices A through D.

#### **3.1. Physical Investigation Summary**

##### **3.1.1. Lake Creek**

Channel type surveys identified 11 distinct stream reaches in the Lake Creek drainage (Figure 2.1). The mainstem was divided into eight reaches, and three additional reaches were identified on the West Fork of Lake Creek. Physical summaries are given below for mainstem reaches 1-7 and for West Fork reaches 8-10. The physical conditions of reaches 9 and 10 are combined in the summary below. Additional habitat information was collected for Bozard Creek and is described below.

Reach one was classified as a C6 channel type and extends for 2,000 feet beginning at the confluence with Windy Bay (Figure 2.1). The reach is a fourth order drainage and encompasses elevations ranging from 2,100 feet to 2,180 feet. This reach is heavily influenced by Coeur d'Alene Lake level fluctuations and has slough-like characteristics. Stream gradient is less than 1 percent and channel pattern is highly sinuous (Table 3.1). Bankfull widths varied from 30-60 feet and bankfull depths vary from 3-5 feet. Channel substrate is comprised mainly of silts. The riparian area is a dense mix of mature coniferous/deciduous trees, grasses and forbs. Channel stability rated fair in 1993 and good in 1994. RASI and habitat surveys were not conducted in this reach because of the slough type nature of the reach. No barriers were observed in this reach.

Reach two extends for 5,093 meters and encompasses elevations ranging from 2,180 feet to 2,260 feet (Figure 2.1). The reach was classified as a C4 channel type (Table 3.2.). Dominant land form in the reach was floodplain with riparian vegetation consisting of a mix of conifers and deciduous trees and shrubs. Average bankfull widths were 29.5 feet and average bankfull depths were 2.7 feet. The channel was slightly entrenched (>2.4) and highly sinuous. (Appendix A).The reach had a riffle/pool ratio of 3: 1 in 1993 and 1.5:1 in 1994.

**Table 3.1. Summary table showing values for reach one, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>                     | <b>1993</b>  | <b>1994</b> |
|--------------------------------------|--|-------------|
| Stream order                         | 4th  |             |
| Elevation                            | 2,100-2,180 ft   |             |
| Soil type                            |  |             |
| Dominant land use                    | Floodplain   |             |
| Riparian ecotype                     | Coniferous/deciduous mix<br>mature overstory, grass<br>& forbs |             |
| Reach length                         | 2,000 ft   | 2,000 ft    |
| Channel type                         | C6   | C6          |
| Bankfull width                       | 30-60ft  | •           |
| Bankfull depth                       | 3-5ft  | •           |
| Width/depth ratio                    | 11.3   | *           |
| Mean stream gradient                 | <1 %   |             |
| Entrenchment                         | slight   | slight      |
| Sinuosity                            | high   | high        |
| Channel stability rating             | fair (113)   | good (73)   |
| RASI values                          | •  | •           |
| % substrate < 4mm                    | *  | •           |
| mean substrate size (mm)             | •  | •           |
| Mean % canopy cover                  | 0%   | •           |
| Barriers                             | •  |             |
| # instream woody debris              | •  |             |
| Dominant habitat complex             | •  |             |
| Riffle/pool ratio                    | •  |             |
| Residual pool depth                  | •  |             |
| Total available habitat              | •  |             |
| Westslope cutthroat<br>trout density | •  |             |



**Table 3.2. Summary table showing values for reach two, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                   | <b>1994</b>                  |
|--------------------------|-------------------------------|------------------------------|
| Stream order             | 4th                           | 4th                          |
| Elevation                | 2,180-2,260 ft                | 2,180-2,260 ft               |
| Dominant land use        | Floodplain                    | same                         |
| Riparian ecotype         | Coniferous/deciduous mix      | same                         |
| Reach length             | 5,093 m                       | 6,290 m                      |
| Channel type             | C4                            | C4                           |
| Bankfull width           | 29.5 ft                       | .                            |
| Bankfull depth           | 2.7 ft                        | .                            |
| Width/depth ratio        | 10.9                          | .                            |
| Mean stream gradient     | <1 %                          | 3.6 %                        |
| Entrenchment             | slight                        | slight                       |
| Sinuosity                | high                          | high                         |
| Channel stability rating | high fair (89)                | fair (93)                    |
| RASI values              | 62;78;48                      | 97;88                        |
| % substrate < 4mm        | 10.4                          | 4.5                          |
| mean substrate size (mm) | 88                            | 141                          |
| Mean % canopy cover      | 11.6 %                        | 46.7 %                       |
| Barriers                 | none                          | none                         |
| # instream woody debris  | 49 (0.9 /100 m <sup>2</sup> ) | 53 (0.811 00m <sup>2</sup> ) |
| Dominant habitat complex | riffle:pool                   | riffle: pool                 |
| Riffle/pool ratio        | 3.0 : 1                       | 1.5:1                        |
| Residual pool depth      | 0.6 m                         | 0.6 m                        |
| Total available habitat  | 12,797 m <sup>2</sup>         | 28,866 m <sup>2</sup>        |

Mean residual pool depth was 0.6 meters for 1993 and 1994. Channel stability rating for this reach was fair (Table 3.2) with little variation between 1993 and 1994. Channel banks were stable and had good vegetative bank protection, however, bank cutting was observed at meander bends. Some of the channel bottom was affected by scouring and deposition. RASI index values averaged 62.7 in 1993 and 92.4 in 1994, while fine sediment (<4mm) comprised up to 34 percent of the sampled substrate in 1993 and dropped to 6 percent in 1994 (Table 3.2). Dominant habitat complex for this reach was riffle-pool. Instream woody debris remained relatively equal between 1993 and 1994 at 0.9/100 m<sup>2</sup> and 0.8/100 m<sup>2</sup>, respectively. Total available habitat increased in 1994 (Table 3.2).

Reach three is 2,088 meters in length with elevations beginning at 2,260 feet and rising to 2,350 feet. (Figure 2.1). The reach was classified as a B3 type channel interspersed with short C segments (Table 3.3). Average stream gradient was 3.5 percent. Consequently, water velocities were slightly higher than in downstream reaches. Average bankfull widths were 35 feet and average bankfull depths were 2.5 feet. Riffle:pool ratio stayed relatively constant between 1993 and 1994 at 1.8:1 and residual pool depths averaged 0.5 meters. Very little large woody debris within this reach. Channel stability rating for this reach was fair (Table 3.3). This reach had upland terrace gradients ranging from 10-80 percent with some rock overhangs. Numerous areas of mass wasting were identified in this reach and contributed a relatively large volume of fine sediment. The riparian area consisted mainly of small shrubs, grasses and forbs with interspersed hardwoods. Riparian vegetation provided 18 and 21 percent canopy cover over the stream during 1993 and 1994. Bank cutting was fairly common as a result of lower vigor and density of vegetative material on the lower banks. Very little instream cover was present within this reach. RASI index values averaged 66.5 and 70.1 for 1993 and 1994, respectively. Fine sediment (<4mm) averaged 17.4 and 13.8 percent of the sampled substrate (Table 3.3). The dominant habitat type within this reach was a riffle-pool complex. Total available habitat increased in 1994 (Table 3.3).

Reach four of Lake Creek extends for 1,328 meters with elevations beginning at 2,350 feet and rising to 2,410 feet (Figure 2.1). Dominant land uses included timber harvesting and residential development. The reach was classified as a C3 channel type (Table 3.4). The channel was highly sinuous and width:depth ratio was greater than 23. The channel was only slightly incised and conveyance capacity was inadequate to contain bankfull flows and flood events. Riffle/pool sequences dominated the habitat structure in this reach and riffle/pool ratio was 3:1 in 1993 and 1.6: 1 in 1994. No large organic debris was present in this reach. Channel stability rating was fair for this reach in both 1993

**Table 3.3. Summary table showing values for reach three, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                   | <b>1994</b>              |
|--------------------------|-------------------------------|--------------------------|
| Stream order             | 4th                           | 4th                      |
| Elevation                | 2,260-2,350 ft                | 2,260-2,350 ft           |
| Dominant land use        | Floodplain                    | Floodplain               |
| Riparian ecotype         | Coniferous/ Deciduous mix     | Coniferous/Deciduous mix |
| Reach length             | 2,088 m                       | 4,809 m                  |
| Channel type             | B3                            | B3                       |
| Bankfull width           | 35ft                          | •                        |
| Bankfull depth           | 2.5 ft                        | *                        |
| Width/depth ratio        | 15.2                          | *                        |
| Mean stream gradient     | 3 - 4 %                       | 3.5 %                    |
| Entrenchment             | moderate                      | moderate                 |
| Sinuosity                | moderate                      | moderate                 |
| Channel stability rating | low fair (103)                | fair (80)                |
| RASI values              | 76; 58; 66                    | 80; 78; 52               |
| % substrate < 4mm        | 17.4 %                        | 13.8 %                   |
| mean substrate size (mm) | 145                           | 131                      |
| Mean % canopy cover      | 18 %                          | 21 %                     |
| Barriers                 | none                          | none                     |
| # instream woody debris  | 4 (CO.1 /100 m <sup>2</sup> ) | 10 (0.2 /100 ft)         |
| Dominant habitat complex | riffle / pool                 | riffle / pool            |
| Riffle/pool ratio        | 1.8 :1                        | <b>2:1</b>               |
| Residual pool depth      | 0.6 m                         | 0.4 m                    |
| Total available habitat  | 12.555 m <sup>2</sup>         | 341499 m <sup>2</sup>    |

**Table 3.4. Summary table showing values for reach four, Lake Creek, 1993 and 1994**


| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,350-2,410 ft  | 2,350-2,41 ft  |
| Dominant land use        | Forest/urban  | Forest/urban   |
| Riparian ecotype         | Coniferous/deciduous<br>with pole trees, grasses<br>and forbs | < Coniferous/deciduous mix<br>pole trees, grasses and<br>forbs |
| Reach length             | 1,328 m   | 880 m  |
| Channel type             | c 3   | c 3  |
| Bankfull width           | 50 ft   | *  |
| Bankfull depth           | 2.1 ft  | .  |
| Width/depth ratio        | 23.8  | *  |
| Mean stream gradient     | 2 %   | 3 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | high  | high   |
| Channel stability rating | medium fair (100)   | fair (86)  |
| RASI values              | 78; 67; 73  | 70; 70   |
| % substrate < 4mm        | 30.6 %  | 13.5 %   |
| mean substrate size (mm) | 135   | 141  |
| Mean % canopy cover      | 0 %   | 10 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 0 (0.1 /100 m <sup>2</sup> )                                  | 4 (0.2 /100 ft)  |
| Dominant habitat complex | riffle / pool   | riffle / pool  |
| Riffle/pool ratio        | 3:1   | 1.6 : 1  |
| Residual pool depth      | 0.5 m   | 0.5 m  |
| Total available habitat  | 22,186 m <sup>2</sup>   | 19,045 m <sup>2</sup>  |

and 1994 (Table 3.4). Vegetative root mass protection ranged from 70-90 percent. The riparian area was a coniferous/deciduous mix with pole trees, grasses and forbs. Canopy cover averaged 0-10 percent for this reach. This reach had significant bank cutting and sloughing and large amounts of fine sediment/gravel depositional areas. Areas of mass wasting were frequently observed and impact potential was judged to be high. Mean RASI values were 72.8 and 69.4 in 1993 and 1994, respectively (Table 3.4). Average fine sediment (4 mm) levels were as high as 31 percent in 1993 but much lower levels were observed in 1994 (13.5 percent). Residual pool depths averaged 0.5 meters for 1993 and 1994. Total available habitat decreased in 1994 (Table 3.4).

Reach five is a B3 channel type and extends for 525 meters with elevations ranging from 2,410 feet to 2,450 feet (Figure 2.1). The reach is forested with a riparian area consisting of mature coniferous/deciduous trees mingled with grasses and forbs. Average bankfull width was 36 feet with an average bankfull depth of 2.1 feet. Riffle to pool ratio was 3:1, and mean residual pool depth was 0.4 meters in 1993. In 1994, riffle to pool ratio was 8:1 with average residual pool depths of 0.5 meters. (Table 3.5). Pool depth was affected by sediment deposition. Mean riparian canopy cover was 19 percent for this reach in 1993 and increased to 26 percent in 1994. Channel stability ratings for this reach were low good (71) in 1993 and fair (101) in 1994 (Table 3.5). There is little mass wasting and bank cutting in this reach. Lower stream banks are armored with dense vegetation- There are moderate amounts of new gravel bars. Average riffle armor stability index values were 48 in 1993 and 73 in 1994 (Table 3.5). Quantities of fine sediment (< 4 mm) averaged 25 and 19 percent in 1993 and 1994, respectively. Total available habitat increased during 1994 (Table 3.5).

Reach six is a 873 meters section with elevations ranging from 2,450 feet to 2,480 feet (Figure 2.1). The reach was classified as a C4-C6 channel type (Table 3.6). Dominant land use is timber harvest with some livestock grazing. The riparian area consisted of a mature mix of coniferous/deciduous trees with grasses and forbs dominating the under-story community. Average bankfull width was 32 feet with an average bankfull depth of 2.1 feet. The channel was slightly entrenched and highly sinuous. Average residual pool depth was 0.7 meters in 1993 and 0.5 meters in 1994. Pool depth was affected by elevated levels of fine sediment (Table 3.6). Channel stability ratings for this reach were fair (101 and 95 in 1993 and 1994, respectively) primarily due to unstable stream banks (Table 3.6). Vegetative bank protection was limited to shallow rooted/seasonal vegetation. There was a significant amount of bank cutting, undercutting and sloughing at outcurves. Accelerated gravel bar development was noted in this reach. RASI index values averaged 73.5 and 81.4 in 1993 and 1994, respectively. Fine sediment less than 4 mm averaged 48 percent of the sampled

**Table 3.5. Summary table showing values for reach five, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,41 O-2,450 ft   | 2.41 O-2,450 ft  |
| Dominant land use        | Forested  | Forested   |
| Riparian ecotype         | Coniferous/deciduous mix<br>with grasses, forbs<br>and mature trees               | Coniferous/deciduous mix<br>grasses, forbs and mature<br>trees |
| Reach length             | 525 m   | 1,189 m  |
| Channel type             | B3  | B3   |
| Bankfull width           | 36 ft   | *  |
| Bankfull depth           | 2.1 ft  | *  |
| Width/depth ratio        | 17.1  | •  |
| Mean stream gradient     |  | 1.5 %  |
| Entrenchment             | moderate  | moderate   |
| Sinuosity                | moderate  | moderate   |
| Channel stability rating | low good (71)   | Fair (101)   |
| RASI values              | 53; 59; 34  | 70; 83; 67   |
| % substrate < 4mm        | 25 %  | 19 %   |
| mean substrate size (mm) | 113   | 173  |
| Mean % canopy cover      | 18.5 %  | 26 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 5 (, 0.1 /100 m <sup>2</sup> )  | 5 (0.1 /100 ft)  |
| Dominant habitat complex | riffle / pool   | riffle / pool  |
| Riffle/pool ratio        | <b>3:1</b>  | <b>8:1</b>   |
| Residual pool depth      | 0.4 m   | 0.5 m  |
| Total available habitat  | 8,357 m <sup>2</sup>  | 18,656 m <sup>2</sup>  |

**Table 3.6. Summary table showing values for reach six, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4 th  | 4th  |
| Elevation                | 2,450-2,480 ft  | 2,450-2,480 ft   |
| Soil type                |   |  |
| Dominant land use        | Forested/agriculture  | Forested/agriculture   |
| Riparian ecotype         | Conferous / decidous mix<br>with grasses, forbs and<br>mature trees | Conferous/decidous mix<br>grasses, forbs and<br>mature trees |
| Reach length             | 873 m   | 873 m  |
| Channel type             | C4 - C6   | C4-C6  |
| Bankfull width           | 32 ft   | •  |
| Bankfull depth           | 2.1 ft  | •  |
| Wrdth/depth ratio        | 16  | *  |
| Mean stream gradient     | 1 %   | 1.3 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | high  | high   |
| Channel stability rating | medium fair (101)   | fair (95)  |
| RASI values              | 74; 75; 72  | 81; 78; 85   |
| % substrate < 4mm        | 48 %  | 26.5 %   |
| mean substrate size (mm) | 110 mm  | 142 mm   |
| Mean % canopy cover      | 0 %   | 26.5 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 18 (<.1 /100 m <sup>2</sup> )                                       | 10 (0.12/100 m <sup>2</sup> )                                |
| Dominant habitat complex | rifle / pool  | rifle / ~001   |
| Riffle/pool ratio        | 3:1   | 2.1 : 1  |
| Residual pool depth      | 0.7 m   | 0.5 m  |
| Total available habitat  | 51,031 m <sup>2</sup>   | 64,171 m <sup>2</sup>  |

substrate in 1993, and 27 percent in 1994 (Table 3.6).

The lower boundary of reach seven is located at the highway 95 bridge crossing and continues upstream for 5,369 meters (Figure 2.1). Most of this reach is a fourth order drainage, however, the upper 500 meters is a third order drainage. Elevations range from 2,480 feet to 2,600 feet. Reach seven was classified as a E5 channel type and was highly sinuous (Table 3.7). Average bankfull width was 18 feet and average bankfull depth was 2.2 feet. The channel capacity was considered adequate to contain bankfull flows. Dominant land use was agriculture with portions of livestock grazing and forest use. Channel stability was poor to fair (Table 3.7). The riparian area was comprised mainly of shallow rooted/seasonal species and extensive bank cutting and bank sloughing were observed throughout this reach. No instream woody debris were observed in this reach and residual pools depths were declining during 1994 (Table 3.7). The channel bottom was silt covered and in a yearlong state of flux with scouring and depositic occurring frequently. Average Riffle Armor Stability Index values were 89 and in 1993 and 1994, respectively. The highest percentages of fine sediment in the watershed were found in this reach (Table 3.7). Total available habitat increased in 1994 (Table 3.7).

Reach eight of Lake Creek extends up the West Fork of Lake Creek for 3,324 meters beginning at the confluence with the mainstem (Figure 2.1). This reach is a third order drainage with elevations beginning at 2,540 feet and rising to 2,570 feet. Dominant land type in this reach is forested with some agriculture and livestock grazing use occurring. The riparian area is a coniferous/deciduous mix of young and pole size trees with grasses, forbs, shrubs intermixed. The reach was classified as an E5 channel type interspersed with short C5 segments (Table 3.8). Average bankfull width was 16.5 feet and average bankfull depth was 2.5 feet. Channel stability rated poor (129) to fair (102) in 1993 and 1994, respectively. Channel capacity was generally adequate for carrying bankfull flows, however- heavy cattle use has degraded portions of the banks creating areas of overflow. There were areas in which flow deflectors caused continuous bank failures and sloughing, resulting in channel migration. RASI surveys were not conducted in this reach due to the extremely mobile nature of the channel substrate. Vegetation provided very little bank protection and the channel bottom was in a constant state of flux as a result of scouring and depositional processes. Very little instream woody debris was present in 1993 or 1994. Riffle to pool ratios were 2.4:1 in 1993 and 1.9:1 in 1994. Continuous bank failures have reduced channel capacity. Available habitat decreased in 19 (Table 3.8).

Reach nine of Lake Creek is a second order drainages which extends for 949 meters (Figure 2.1). Elevations ranged from 2,570 feet to 2,970 feet. Dominant land type was forested. The riparian area consisted mainly of young



**Table 3.7. Summary table showing values for reach seven, Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,480-2,600 ft  | 2,480-2,600 ft   |
| Soil type                |   |  |
| Dominant land use        | Agriculture/grazing/forest                                    | Agriculture/grazing/forest                                     |
| Riparian ecotype         | Coniferous/deciduous mix<br>With shrubs, grasses and<br>forbs | Coniferous/deciduous mix<br>With shrubs, grasses, and<br>forbs |
| Reach length             | 5,369 m   | 5,739 m <sup>2</sup>   |
| Channel type             | E5-E6   | E5 - E6  |
| Bankfull width           | 18ft  | •  |
| Bankfull depth           | 2.2 ft  | *  |
| Width/depth ratio        | 8.1   | •  |
| Mean stream gradient     | ☞ ☞   | 1.5 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | very - high   | very - high  |
| Channel stability rating | high poor (118)   | fair (79)  |
| RASI values              | 80.4; 96.1; 90.6  | 68; 89; 96   |
| mean substrate size (mm) | 102   | 124  |
| Mean % canopy cover      | 0 %   | 45.8 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 49 (CO.1 /100 m <sup>2</sup> )                                | 172 (0.9 /100 ft)  |
| Dominant habitat complex | glide / pool  | riffle / pool  |
| Riffle/pool ratio        | 2.6 : 1   | 2.3 : 1  |
| Residual pool depth      | 0.6 m   | 0.4 m  |
| Total available habitat  | 49,096 m <sup>2</sup>   | 77,623 m <sup>2</sup>  |

Table 3.8. Summary table showing values for reach eight, (West fork) Lake Creek, 1993 and 1994

| PARAMETER                | 1993  | 1994   |
|--------------------------|---|--|
| Stream order             | 3rd   | 3rd  |
| Elevation                | 2.540-2.570 ft  | 2540-2.570 ft  |
| Dominant land use        | Forested/agriculture/<br>grazing  | Forested/agriculture/<br>grazing   |
| Riparian ecotype         | Coniferous/deciduous mix<br>with<br>grasses/forbs/shrubs/<br>young timber | Coniferous/deciduous mix<br>with<br>grasses/forbs/hrubs/<br>young timber |
| Reach length             | 3,324 m   | 3,265 m  |
| Channel type             | E5-C5   | E5-C5  |
| Bankfull width           | 16.5 ft   | *  |
| Bankfull depth           | 2.5 ft  | *  |
| Width/depth ratio        | 6.6   | •  |
| Mean stream gradient     | 1 %   | 1.2 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | very high   | very high  |
| Channel stability rating | medium poor (129)   | Fair (102)   |
| RASI values              | *   | •  |
| % substrate < 4mm        | •   | •  |
| mean substrate size (mm) | ☒   | •  |
| Mean % canopy cover      | 0 %   | 35.4 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 28 ( 0.8 /100 m <sup>2</sup> )  | 20 (0.8/100 m <sup>2</sup> )   |
| Dominant habitat complex | riffle/ pool  | riffle / pool  |
| Riffle/pool ratio        | 2.4 : 1   | 1.9: 1   |
| Residual pool depth      | 0.3 m   | 0.5 m  |
| Total available habitat  | 18,443 m <sup>2</sup>   | 17,003 m <sup>2</sup>  |

coniferous and deciduous trees. The reach was classified as a E6 channel type and the channel was slightly entrenched and highly sinuous (Table 3.9). Average bankfull width was 12 feet and average bankfull depth was 2.6 feet. Dominant habitat type in this reach was a step pool cascade/rapid complex. Residual pool depth was 0.1 meters in 1993 and 0.9 meters in 1994. Large accumulations of woody debris were noted in this reach (Table 3.9). Channel stability was rated as fair in both 1993 and 1994 (Table 3.9). The channel capacity was adequate to contain bankfull flows. Significant bank cutting and sloughing was evident in the lower portion of this reach, while moderate bank cutting was noted in the upper portion. In the lower portion of this reach the channel bottom is silt covered and fluctuates yearlong, however, only minor amounts of channel scour were observed in the upper portion of the reach. Results of 1993 RASI surveys calculated a mean index value of 69 and an average of 22 percent fines for the reach (Table 3.9). Total available habitat increased during 1994 to 14,528 m<sup>2</sup>.

Habitat surveys were conducted on 3,888 meters of Bozard Creek. This tributary is a third order drainage with elevation beginning at 2,540 feet and rising to 2,590 feet (Figure 2.1). Dominant land type was forested with some agriculture use occurring. The riparian area consisted of a mix of young coniferous/deciduous trees and shrubs. No channel type surveys were completed for Bozard Creek. The dominant habitat complex in this reach was riffle-glide with a riffle:pool ratio of 12:1. Mean residual pool depth was 0.3 meters (Table 3.10). Some bank cutting was observed intermittently at outcurves and constrictions. New increase in bar formation and some filling of pools contributed to the fair rating. During 1994, upper portions of this reach became subsurface and could not be utilized by fish.

### 3.1.2. Benawah Creek

Channel typing surveys identified 16 reaches in the Benawah Creek watershed. Reaches 1-14 are located on the mainstem of Benawah Creek, and reaches 13a and 14a comprise the South Fork of Benawah Creek (Figure 2.2). Physical surveys are described below for mainstem reaches 1-3 and a general description of the South Fork summarizes information related to reaches 13a and 14a.

Reach one of Benawah Creek extends from the confluence of Benawah Lake upstream for approximately one mile. Habitat surveys were conducted only on the upper 487 meters. Due to the slough type nature of the lower portion of reach one (Figure 2.2). There is limited livestock grazing within this reach. Overbank flows are common in this reach. The reach was classified as a C3 channel type. The channel was slightly entrenched with an average bankfull width-depth ratio of 39 feet (Table 3.11). Mean residual pool depth was 0.8

**Table 3.9. Summary table showing values for reach nine, (West Fork) Lake Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                | <b>1994</b>                                |
|--------------------------|--|--|
| Stream order             | 2nd  | 2nd  |
| Elevation                | 2, 570-2,970 ft                            | 2,570-2,970 ft                             |
| Dominant land use        | Forested                                   | Forested                                   |
| Riparian ecotype         | Coniferous/deciduous mix with young timber | Coniferous/deciduous mix with young timber |
| Reach length             | 949m                                       | 2,306 m <sup>2</sup>                       |
| Channel type             | E6-B4a                                     | E6-B4a                                     |
| Bankfull width           | 12ft                                       | •  |
| Bankfull depth           | 2.6 ft                                     | •  |
| Width/depth ratio        | 4.6  | •  |
| Mean stream gradient     | 3 %  | 3 %  |
| Entrenchment             | slight - moderate                          | slight-moderate                            |
| Sinuosity                | moderate - very high                       | moderate-very high                         |
| Channel stability rating | moderate fair (100)                        | fair (98)                                  |
| RASI values              | 63; 79; 66                                 | •  |
| % substrate < 4mm        | 22   | ☒  |
| mean substrate size (mm) | 66   | •  |
| Mean % canopy cover      | 77 %                                       | 41 %                                       |
| Barriers                 | none                                       | none                                       |
| # instream woody debris  | 119 (12.5 /100 m <sup>2</sup> )            | 52 (2.3 /100 m <sup>2</sup> )              |
| Dominant habitat complex | step-pool cascade/rapids                   | cascade ; pool                             |
| Riffle/pool ratio        | NA   | 1 : 1.7                                    |
| Residual pool depth      | 0.1 m                                      | 0.9 m                                      |
| Total available habitat  | 3,314 m <sup>2</sup>                       | 14,528 m <sup>2</sup>                      |

Table 3.10. Summary table showing values for Bozard, 1993 and 1994

| <b>PARAMETER</b>         | <b>1993</b>                                    | <b>1994</b>                                    |
|--------------------------|--|--|
| Stream order             | 3rd  | 3rd  |
| Elevation                | 2,540-2,590 ft                                 | 2,540-2,590 ft                                 |
| Dominant land use        | Forested / Agriculture                         | Forested/Agriculture                           |
| Riparian ecotype         | Coniferous/deciduous mix<br>shrub/young timber | Coniferous/deciduous mix<br>shrub/young timber |
| Reach length             | 3,888 m  | 3,870 m  |
| Channel type             | E5   | E5   |
| Bankfull width           | •  | •  |
| Bankfull depth           | •  | *  |
| Width/depth ratio        |  | •  |
| Mean stream gradient     | ☞ ♪  | ☞ ♪  |
| Entrenchment             | slight-moderate                                | slight-moderate                                |
| Sinuosity                | moderate-very high                             | moderate-very high                             |
| Channel stability rating | poor (119)                                     | poor (117)                                     |
| RASI values              | •  | •  |
| % substrate < 4mm        | •  | ☒  |
| mean substrate size (mm) | 1  | •  |
| Mean % canopy cover      | 0 . 0 %  | 0 . 0 %  |
| Barriers                 | none   | subsurface flow                                |
| # instream woody debris  | 0 (0/100 m <sup>2</sup> )                      | 0 (0/100 m <sup>2</sup> )                      |
| Dominant habitat complex | rifle - glide                                  | ND   |
| Riffle/pool ratio        | 12: 1  | ND   |
| Residual pool depth      | 0.3 m  | 0.0  |
| Total available habitat  | 25,757 m <sup>2</sup>                          | 0.0 m <sup>2</sup>                             |

Table 3.11. Summary table showing values for reach one, Benawah Creek, 1993 and 1994

| <b>PARAMETER</b>         | <b>1993</b>                    | <b>1994</b>                    |
|--------------------------|--------------------------------|--------------------------------|
| Stream order             | 4th                            | 4th                            |
| Elevation                | 2,250-2,250 ft.                | 2.250-2.250 ft                 |
| Dominant land use        | Wetland/floodplain             | Wetland/floodplain             |
| Riparian ecotype         | Coniferous/Deciduous<br>shrubs | Coniferous/Deciduous<br>shrubs |
| Reach length             | 471 m                          | 487 m                          |
| Channel type             | c 3                            | C3a                            |
| Bankfull width           | 158ft                          | •                              |
| Bankfull depth           | 4.0 ft                         | •                              |
| Width/depth ratio        | 39 ft                          |                                |
| Mean stream gradient     | 1.5 %                          | 1 %                            |
| Entrenchment             | slight                         | slight                         |
| Sinuosity                | high                           | high                           |
| Channel stability rating | medium fair (94)               | high fair (83)                 |
| RASI values              | 80; 69; 82                     | 90; 69; 87                     |
| % substrate < 4mm        | 12                             | 21.5                           |
| mean substrate size (mm) | 133                            | 152                            |
| Mean % canopy cover      | 0.0 %                          | 29.4 %                         |
| Barriers                 | none                           | none                           |
| # instream woody debris  | 8 (0/100 m <sup>2</sup> )      | 15 (0.4 / 100 m <sup>2</sup> ) |
| Dominant habitat complex | rifle - pool                   | rifle - pool                   |
| Rifle/pool ratio         | 2:1                            | 2 :1                           |
| Residual pool depth      | 0.8 m                          | 0.7 m                          |
| Total available habitat  | 25,223 m <sup>2</sup>          | 3,336 m <sup>2</sup>           |

meters and riffle:pool ratio was 2:1. Channel stability was rated as fair in this reach (Table 3.11). Lower banks had significant cuts (up to 2 feet high) and root mat overhangs and bank sloughing was evident. No instream woody debris was evident in this reach. Mean RASI values were 77 and 82 for 1993 and 1994, respectively (Table 3.11). Mean fine sediment values were 12 and 22 percent of sampled substrate, for 1993 and 1994, respectively. Total available habitat decreased during 1994 (Table 3.11).

Reach two of Benawah Creek extends for 1,342 meters and encompasses elevations ranging from 2,250-2,330 feet (Figure 2.2). This reach is forested and dominated by large shrubs. The reach was classified as a B2a channel type (Table 3.12). Channel capacity was adequate for existing flow conditions with the channel being moderately entrenched. Average stream gradient was 4 percent with the dominant habitat complex consisting of step-pools and riffles. A riffle to pool ratio of 4:1 was calculated for this reach and average residual pool depth was 0.4 meters in 1993, and 7: 1 and 0.4 meters in 1994 (Table 3.12). Channel stability rated good for both 1993 and 1994 (Table 3.12). Bank protection was adequate with both shallow and deep rooted species providing stability. Debris jam potential was minor. Lower banks were well armored with approximately 70 percent of the lower banks protected by large boulders. Very little cutting or deposition occurred in this reach. Seasonal blooms of aquatic vegetation were observed. The mean RASI index value was 43 and fine sediment (< 4 mm) comprised less than 15 percent of the sampled substrate (Table 3.12). Total available habitat declined in 1994 (Table 3.12).

Reach three was classified as a C2b channel type and extends for 724 meters (Figure 2.2). This reach was mostly forested with large shrubs dominating the riparian zone. Mean stream gradient was 2 percent. Average stream canopy cover was 29 percent in 1993 and 40 percent in 1994. The dominant habitat complex in this reach was riffle-pool with a riffle pool ratio of 1.3:1 in 1993 and 2.5:1 in 1994. Average residual pool depths of 0.7 meters (1993) and 0.4 meters (1994) were calculated (Table 3.13). Channel stability was considered good (Table 3.13). Upper bank slopes ranged from 3 to 60 percent and intermittent areas of mass wasting and unstable slopes were observed. Bank protection was comprised mainly of shallow rooted species. There was evidence that channel capacity was not adequate to contain bankfull flows. Lower banks were comprised of cobble and boulder while bank cutting occurred at outcurves with new coarse gravel bar development. The channel bottom was firmly embedded, with less than five percent of the bottom affected by scouring and deposition. No instream woody debris was located in this reach. Seasonal alga blooms were observed. Mean RASI values were 51 and 44 during 1993 and 1994, respectively. Fine sediment (< 4 mm) generally comprised less than 10 percent of sampled substrate (Table 3.13). Total available habitat declined in 1994 (Table 3.13).

**Table 3.12. Summary table showing values for reach two, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>               | <b>1994</b>               |
|--------------------------|---------------------------|---------------------------|
| Stream order             | 4th                       | 4th                       |
| Elevation                | 2,250-2,330 ft            | 2,250-2,330 ft            |
| Dominant land use        | forested                  | forested                  |
| Riparian ecotype         | Coniferous/ Deciduous     | Coniferous/Deciduous      |
| Reach length             | 1,342 m                   | 1,415 m                   |
| Channel type             | B2a                       | B2a                       |
| Bankfull width           | 35ft                      | •                         |
| Bankfull depth           | 2.7 ft                    | *                         |
| Width/depth ratio        | 13                        |                           |
| Mean stream gradient     | 4 %                       | 2.4 %                     |
| Entrenchment             | moderate                  | moderate                  |
| Sinuosity                | moderate                  | moderate                  |
| Channel stability rating | good (74)                 | good (80)                 |
| RASI values              | *                         | 24.1 ; 67.1 ; 38.9        |
| % substrate < 4mm        | •                         | 10.6 %                    |
| mean substrate size (mm) | •                         | 109                       |
| Mean % canopy cover      | 80 %                      | 37 %                      |
| Barriers                 | none                      | none                      |
| # instream woody debris  | 5 (0/100 m <sup>2</sup> ) | 6 (0/100 m <sup>2</sup> ) |
| Dominant habitat complex | riffle - step pool        | riffle - step pool        |
| Riffle/pool ratio        | 4:1                       | 7:1                       |
| Residual pool depth      | 0.4 m                     | 0.4 m                     |
| Total available habitat  | 51,063 m <sup>2</sup>     | 20,357 m <sup>2</sup>     |



**Table 3.13. Summary table showing values for reach three, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>               | <b>1994</b>               |
|--------------------------|---------------------------|---------------------------|
| Stream order             | 4th                       | 4th                       |
| Elevation                | 2,330-2,330 ft            | 2,330-2,330 ft            |
| Dominant land use        | forested                  | forested                  |
| Riparian ecotype         | Coniferous/ Deciduous     | Coniferous/Deciduous      |
| Reach length             | 724 m                     | 1,154 m                   |
| Channel type             | C2b                       | C2b                       |
| Bankfull width           | 41 ft                     | *                         |
| Bankfull depth           | 2.7ft                     | •                         |
| Width/depth ratio        | 15                        | 15                        |
| Mean stream gradient     | 2 %                       | 2 %                       |
| Entrenchment             | slight                    | slight                    |
| Sinuosity                | high                      | high                      |
| Channel stability rating | low good (67)             | medium good (55)          |
| RASI values              | 52; 45; 56                | 3; 43; 87                 |
| % substrate < 4mm        | 8 %                       | 6.2 %                     |
| mean substrate size (mm) | 140                       | 141                       |
| Mean % canopy cover      | 29 %                      | 40 %                      |
| Barriers                 | none                      | none                      |
| # instream woody debris  | 3 (0/100 m <sup>2</sup> ) | 1 (0/100 m <sup>2</sup> ) |
| Dominant habitat complex | riffle-pool               | riffle-pool               |
| Riffle/pool ratio        | 1.3: 1                    | 2.5 : 1                   |
| Residual pool depth      | 0.7 m                     | 0.4 m                     |
| Total available habitat  | 29,541 m <sup>2</sup>     | 22,856 m <sup>2</sup>     |

Reach four extends for 1,349 meters with elevations ranging from 2,330 to 2,380 feet (Figure 2.2). The reach is a B3c channel type with moderate entrenchment and moderate sinuosity (Table 3.14). Dominant land type in the reach was forested with large deciduous shrubs forming the dominant riparian cover. This reach has been heavily influenced by past rain-on-snow events. Average stream gradient was 1.5 percent. Riffle/pool ratio was 1 :1 in 1993 and 2:1 in 1994. Residual pool depth increased from 0.3 meters in 1993 to 0.7 meters in 1994 (Table 3.14). Channel stability rated from good to fair (Table 3.14). Overbank flows were common throughout the reach. There was no evidence of mass wasting problems, and lower bank cutting was limited to outcurve areas. There was little new bar development in this reach. The channel bottom was firmly embedded, with little scouring and deposition occurring in this reach. Seasonal alga blooms were present. Average RASI values were 46 in 1993. RASI values were considerably higher in 1994 (mean = 81). Fine sediment (< 4 mm) comprised less than 13 percent of the sampled substrate in both 1993 and 1994 (Table 3.14). Total available habitat increased in 1994 (Table 3.14).

Reach five was classified as a C1 b channel type and extends for 1,056 meters (Figure 2.2). Elevations in the reach began at 2,380 feet and rose to 2,490 feet. This reach contained a small bedrock falls which was not considered a migration barrier. This reach was forested with riparian vegetation dominated by large deciduous shrubs. Average stream gradient was 4.5 percent. Riffle to pool ratio was 1.5:1 and mean residual pool depth was 0.2 meters in 1993 and 0.6 meters in 1994. Channel stability in this reach was good (Table 3.15). Landform slope ranged from 2 to 40 percent with no evidence of mass wasting problems. Occasional overbank flows have occurred in the past. The channel area was free of debris jam materials. The lower banks were composed mainly of mostly boulder which minimized bank cutting and erosional problems. Channel substrate was predominantly bedrock with no instream woody debris. Clinging aquatic vegetation was common with both algae and moss present in pools and swifter waters. No RASI surveys were conducted in this reach because of the bedrock substrate. Total available habitat increased in 1994 (Table 3.15).

Reach six extends for 835 meters with elevations ranging from 2,490 to 2,550 feet (Figure 2.2). This reach was dominated by small shrubs, grass and forbs and was actively utilized by livestock. The reach was classified as a C3 channel type (Table 3.16). Average stream gradient was 1 percent, the channel was slightly entrenched, and sinuosity was high. Riparian vegetation provided no canopy cover and no instream woody debris was present. Dominant habitat complex was riffle/pool with a riffle pool ratio of 1.5:1. Mean residual pool depth was 0.2 meters in 1993 and 0.6 meters in 1994 (Table 3.16). Upper banks

**Table 3.14. Summary table showing values for reach four, Benawah Creek, 1993 and 1994**

| <b>P</b>                 | <b>1993</b>                                   | <b>1994</b>                               |
|--------------------------|---|---|
| Stream order             | 4th   | 4th                                       |
| Elevation                | 2,330-2,380 ft                                | 2,330-2,380 ft                            |
| Dominant land use        | forested                                      | forested                                  |
| Riparian ecotype         | Coniferous/Deciduous<br>mix with large shrubs | Coniferous/Deciduous<br>with large shrubs |
| Reach length             | 1,349 m                                       | 1,728 m.                                  |
| Channel type             | B3c   | B3c                                       |
| Bankfull width           | 55ft  | *   |
| Bankfull depth           | 2 .0ft  | *   |
| Width/depth ratio        | 27.5  | 27.5                                      |
| Mean stream gradient     | 1.5 %   | 2 %                                       |
| Entrenchment             | moderate                                      | moderate                                  |
| Sinuosity                | moderate                                      | moderate                                  |
| Channel stability rating | medium good (60)                              | high fair (78)                            |
| RASI values              | 38; 53; 46                                    | 82; 79; 83                                |
| % substrate < 4mm        | 6   | 11.7                                      |
| mean substrate size (mm) | 108   | 111                                       |
| Mean % canopy cover      |   | 17.7 %                                    |
| Barriers                 | none  | none                                      |
| # instream woody debris  | 3 (0/100 m <sup>2</sup> )                     | 3 (0/100 m <sup>2</sup> )                 |
| Dominant habitat complex | rifle - pool                                  | rifle - pool                              |
| Riffle/pool ratio        | 1 : 1   | 2 : 1                                     |
| Residual pool depth      | 0.3 m   | 0.7 m                                     |
| Total available habitat  | 32,275 m <sup>2</sup>                         | 44,093 m <sup>2</sup>                     |

**Table 3.15. Summary table showing values for reach five, Benawah Creek, 1993 and 1994**

| <b>PRAMETER</b>          | <b>1993</b>                                  | <b>1994</b>                                    |
|--------------------------|--|--|
| Stream order             | 4th  | 4th  |
| Elevation                | 2,380-2,490 ft                               | 2,380-2,490 ft                                 |
| Dominant land use        | forested                                     | forested                                       |
| Riparian ecotype         | Coniferous/Deciduous<br>mix with large srubs | Coniferous/Deciduous<br>mix with large schrubs |
| Reach length             | 1,056 m                                      | 1,244 m  |
| Channel type             | C1b  | C1b  |
| Bankfull width           | 0  | •  |
| Bankfull depth           | 0  | •  |
| Width/depth ratio        | 21.3   | 21.3   |
| Mean stream gradient     | 3 %  | 2 %  |
| Ent; enchment            | slight                                       | slight   |
| Sinuosity                | high   | high   |
| Channel stability rating | high good (48)                               | high good (48)                                 |
| RASI values              | NA   | NA   |
| % substrate < 4mm        | NA   | NA   |
| mean substrate size (mm) | NA   | NA   |
| Mean % canopy cover      | ND   | 16.8 %   |
| Barriers                 | none   | none   |
| # instream woody debris  | 0 (0/100 m <sup>2</sup> )                    | 3 (0/100 m <sup>2</sup> )                      |
| Dominant habitat complex | riffle - pool                                | riffle - step pool                             |
| Riffle/pool ratio        | 3:1  | 1:0  |
| Residual pool depth      | 0.2 m  | 0.6 m  |
| Total available habitat  | 30,956 m <sup>2</sup>                        | 29,156 m <sup>2</sup>                          |

**Table 3.16. Summary table showing values for reach six, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,490-2,550 ft  | 2,490-2,550 ft   |
| Dominant land use        | Livestock grazing   | Livestock grazing  |
| Riparian ecotype         | Coniferous/deciduous mix<br>with small shrub, grasses,<br>& forbs | Coniferous/deciduous mix<br>with small shrub/grasses/forbs |
| Reach length             | 835 m   | 797 m  |
| Channel type             | c 3   | c 3  |
| Bankfull width           | 38 ft   | *  |
| Bankfull depth           | 1.9 ft  | •  |
| Width/depth ratio        | 20  | 20   |
| Mean stream gradient     | 1 %   | 1.5 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | high  | high   |
| Channel stability rating | low good (65)   | medium good (59)   |
| RASI values              | 45; 43; 62  | 67; 29; 86   |
| % substrate < 4mm        | 11  | 13.7   |
| mean substrate size (mm) | 82  | 133  |
| Mean % canopy cover      | •   | 3 %  |
| Barriers                 | none  | none   |
| # instream woody debris  | 0 (0/100 m <sup>2</sup> )   | 0 (0/100 m <sup>2</sup> )                                  |
| Dominant habitat complex | riffle - pool   | riffle - pool  |
| Riffle/pool ratio        | 1.5: 1  | 3 : 1  |
| Residual pool depth      | 0.2 m   | 0.6 m  |
| Total available habitat  | 6.995 m <sup>2</sup>  | 12,694 m <sup>2</sup>                                      |

showed evidence of past mass wasting problems. The potential for future mass wasting problems was high, due to a lack of dense, deep rooted plant species on steeper slopes. Lower banks consisted of cobbles and boulders. Some intermittent bank cutting was observed at outcurves and new gravel bar formation was evident. The channel bottom was stable and comprised of tightly packed, well distributed particle sizes. An estimated 5 to 30 percent of the channel bottom was affected by scouring and deposition. Clinging aquatic vegetation was common throughout the reach with both seasonal algae blooms and moss in pools and swifter waters. Mean RASI values were 50 and 60 in 1993 and 1994, respectively (Table 3.16). Fine sediment (< 4 mm) comprised less than 11 percent of the sampled substrate on average for both 1993 and 1994. Total available habitat increased during 1994.

Reach seven of Benawah Creek is a C3 channel type which is 3,193 meters in length. Elevations range from 2,550 to 2,560 feet (Figure 2.2). Dominant land use was livestock grazing. Riparian areas were dominated by large shrubs which provided an average canopy cover of 45 percent. Dominant habitat complex was riffle/pool with a riffle:pool ratio of 1.2: 1 in 1993 and a riffle to pool ratio of 3.1 :1 in 1994. Mean residual pool depth was 0.7 meters. Average stream gradient was 1 percent (Table 3.17). Channel stability was rated as fair (Table 3.17). Landform slope ranged from 0 to 3 percent. There were some areas where upper bank failure had occurred. Lower banks consisted of 65 percent cobble and small boulder, with bank cutting occurring at obstructions and outcurves. The lower banks barely contained existing peak flows. Recent deposition of gravel and sand sized substrate was noted throughout the reach. The channel bottom was fairly stable with 5-30 percent of the channel bottom affected by scouring and deposition. Seasonal algal blooms were observed. Mean RASI values were 44 and 78 in 1993 and 1994, respectively (Table 3.17). Fine sediment comprised an average of 13 and 16 percent of the sampled substrate in 1993 and 1994, respectively.

Reach eight is a C6 channel type which extends for 2,753 meters and encompasses elevations ranging from 2,560 feet to 2,640 feet (Figure 4.2). The channel is slightly entrenched and highly sinuous. Mean stream gradient was 1 percent (Table 3.18). Dominant land use is livestock grazing and riparian areas are vegetated with a mix of shrubs, grasses and forbs. Average canopy cover in this reach was 21 percent with instream wood densities of less than 0.1/m<sup>2</sup>. The dominant habitat complex was riffle/pool with a mean residual pool depth of 1 .0 meters (Table 3.18). Channel stability in this reach was fair to good (Table 3.18). Overbank flows were common throughout the reach. There were continuous cuts along both lower banks, and some deposition of gravel and sand on old and some new bars. Scouring and deposition affected 5 to 30 percent of the channel bottom. Seasonal algal blooms made rocks slick, and moss was present in swifter waters. RASI values were variable with mean values increasing from 49 to 72 during the one year sample interval (Table 3.18). Measured fine sediment

**Table 3.17. Summary table showing values for reach seven, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                | <b>1994</b>                                |
|--------------------------|--|--|
| Stream order             | 4th  | 4th  |
| Elevation                | 2,550-2,560 ft                             | 2,550-2,560 ft                             |
| Dominant land use        | Livestock grazing                          | Livestock grazing                          |
| Riparian ecotype         | Coniferous/deciduous mix with large shrubs | Coniferous/deciduous mix with large shrubs |
| Reach length             | 3,193 m                                    | 3,685 m                                    |
| Channel type             | c 3  | c 3  |
| Bankfull width           | 44 ft                                      | •  |
| Bankfull depth           | 2.5 ft                                     | ☐  |
| Width/depth ratio        | 17.6                                       | 17.6                                       |
| Mean stream gradient     | 1.4 %                                      | 1.2 %                                      |
| Entrenchment             | slight                                     | slight                                     |
| Sinuosity                | high                                       | high                                       |
| Channel stability rating | high fair (77)                             | medium fair (96)                           |
| RASI values              | 60; 27; 45                                 | 72; 70; 93                                 |
| % substrate < 4mm        | 13   | 15.5                                       |
| mean substrate size (mm) | 116  | 109  |
| Mean % canopy cover      | 49 %                                       | 43 %                                       |
| Barriers                 | none                                       | none                                       |
| # instream woody debris  | 1 (0/100 m <sup>2</sup> )                  | 19 (0.01 /100 m <sup>2</sup> )             |
| Dominant habitat complex | riffle-pool                                | riffle-pool                                |
| Riffle/pool ratio        | 1.2 : 1                                    | 3.1 : 1                                    |
| Residual pool depth      | 0.7 m                                      | 0.8 m                                      |
| Total available habitat  | 53,556 m <sup>2</sup>                      | 54,177 m <sup>2</sup>                      |

**Table 3.18. Summary table showing values for reach eight, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b>   |
|--------------------------|--|---|
| Stream order             | 4th  | 4th   |
| Elevation                | 2,560-2,640 ft   | 2,560-2,640 ft  |
| Dominant land use        | Livestock grazing  | Livestock grazing                                     |
| Riparian ecotype         | Coniferous/deciduous mix<br>with shrubs, grasses,<br>& forbs | Coniferous/deciduous mix<br>with shrubs/grasses/forbs |
| Reach length             | 2,753 m  | 2,796 m   |
| Channel type             | C6   | C6  |
| Bankfull width           | 42 ft  | *   |
| Bankfull depth           | 1.6 ft   | .   |
| Width/depth ratio        | 26.2   | 26.2  |
| Mean stream gradient     | 1.8 %  | 1.8 %   |
| Entrenchment             | slight   | slight  |
| Sinuosity                | high   | high  |
| Channel stability rating | high fair (82)   | low good (73)   |
| RASI values              | 42; 33; 72   | 70.5; 76.2; 68.9                                      |
| % substrate < 4mm        | 8 %  | 14.7 %  |
| mean substrate size (mm) | 108  | 137   |
| Mean % canopy cover      | 21.4 %   | 26.1 %  |
| Barriers                 | none   | none  |
| # instream woody debris  | 24 (0/100 m <sup>2</sup> )                                   | 6 (0/100 m <sup>2</sup> )                             |
| Dominant habitat complex | riffle- pool   | riffle- pool  |
| Riffle/pool ratio        | 1.3 : 1  | 2 : 1   |
| Residual pool depth      | 1.0 m  | 0.8 m   |
| Total available habitat  | 35,793 m <sup>2</sup>  | 32,007 m <sup>2</sup>                                 |



increased from a mean of 8 percent in 1993 to 15 percent in 1994. Total available habitat decreased in 1994 (Table 3.18).

Reach nine was classified as a C5 channel type and extends for 1,032 meters (Figure 2.2). Dominant land use in the reach was livestock grazing and riparian areas were primarily vegetated with small shrubs. The channel was slightly entrenched and highly sinuous. Mean stream gradient was less than 2 percent. Mean canopy cover was 5 percent and instream woody debris densities were less than  $0.1/m^2$ . Riffle/pool ratios were 1.7:1 and 2.8:1 in 1993 and 1994, respectively. Mean residual pool depths were 0.8 meters for 1993 and 1994, respectively (Table 3.19). Channel stability ratings varied from poor to fair in 1993 and 1994, respectively and were generally less stable than downstream reaches (Table 3.19). Channel capacity was inadequate to contain bankfull flows. Lower banks were generally comprised of sand and gravel sized material. Scour and deposition were the dominant channel forming processes in the reach. The channel bottom was in a constant state of flux and deposition of fine sediment resulted in reduced pool volumes. No RASI surveys were completed for this reach. Total available habitat increased during 1994.

Reach ten has a total length of 2,846 meters and encompasses elevations ranging from 2,640 feet to 2,750 feet (Figure 2.2). The reach was classified as a C6 channel type with slight entrenchment and high sinuosity (Table 3.20). Dominant land use was livestock grazing with a riparian area comprised of large shrubs. The reach had instream woody debris densities of  $8.5/100 m^2$ . Average residual pool depth was 0.9 meters in 1993 and 0.5 meters in 1994 (Table 3.20). Channel stability was rated as fair (Table 3.20). Landform slope ranged from zero to twenty five percent with no evidence of mass wasting. Stream bank vegetation was comprised of shallow rooted species. Debris jam potential was high for the entire reach. Channel capacity was insufficient to contain peak flows. There was evidence of frequent channel migration, and a significant amount of bank cutting and sloughing had occurred. A moderate amount of gravel and sand deposition was identified on old and new bars. The channel bottom consisted of a loose assortment of well-rounded material. Stable bottom materials occurred on 20-50 percent of the reach. Seasonal algal blooms were observed. Dominant habitat complex in this reach was riffle-pool with a riie to pool ratio of 3.5:1 for 1993 and 1994. Total available habitat decreased during 1994.

Reach eleven was classified as a E6 channel type and extended for 589 meters (Figure 2.2). Elevations ranged from 2,750 feet to 2,760 feet. Dominant land use was livestock grazing and riparian areas were primarily vegetated with large shrubs. Riparian vegetation provided 21 percent stream canopy cover and instream large woody debris densities were less than  $0.1/100m^2$ . The channel was slightly entrenched (width:depth ratio = 3.9) and highly sinuous (Table 3.21).

**Table 3.19. Summary table showing values for reach nine, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                              | <b>1994</b>                             |
|--------------------------|--|---|
| Stream order             | 4th                                      | 4th                                     |
| Elevation                | 2,640-2,640 ft                           | 2,640-2,640 ft                          |
| Dominant land use        | Livestock grazing                        | Livestock grazing                       |
| Riparian ecotype         | Coniferous/ Deciduous<br>mix with shrubs | Coniferous/deciduous<br>mix with shrubs |
| Reach length             | 1,032 m                                  | 868 m                                   |
| Channel type             | C5                                       | C5                                      |
| Bankfull width           | 80 ft                                    | .                                       |
| Bankfull depth           | 6 ft                                     | .                                       |
| Width/depth ratio        | 13.3                                     |   |
| Mean stream gradient     | 1.75 %                                   | 1.3 %                                   |
| Entrenchment             | slight                                   | slight                                  |
| Sinuosity                | high                                     | high                                    |
| Channel stability rating | poor (122)                               | fair (90)                               |
| RASI values              | NA                                       | NA                                      |
| % substrate < 4mm        | NA                                       | NA                                      |
| mean substrate size (mm) | NA                                       | NA                                      |
| Mean % canopy cover      | 45 %                                     | 10.5 %                                  |
| Barriers                 | none                                     | none                                    |
| # instream woody debris  | 29 (0/100 m <sup>2</sup> )               | 74 (0/100 m <sup>2</sup> )              |
| Dominant habitat complex | riffle - pool                            | riffle - pool                           |
| Riffle/pool ratio        | 1.7: 1                                   | 2.8 : 1                                 |
| Residual pool depth      | 0.8 m                                    | 0.8 m                                   |
| Total available habitat  | 27,231 m <sup>2</sup>                    | 60,016 m <sup>2</sup>                   |

**Table 3.20. Summary table showing values for reach ten, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                              | <b>1994</b>                             |
|--------------------------|--|---|
| Stream order             | 4th                                      | 4th                                     |
| Elevation                | 2,640-2,750 ft                           | 2,640-2,750 ft                          |
| Dominant land use        | Livestock grazing                        | Livestock grazing                       |
| Riparian ecotype         | Coniferous/ Deciduous<br>mix with shrubs | Coniferous/deciduous<br>mix with shrubs |
| Reach length             | 2,846 m                                  | 693 m                                   |
| Channel type             | C6                                       | C6                                      |
| Bankfull width           | 30ft                                     | •                                       |
| Bankfull depth           | 2ft                                      | •                                       |
| Width/depth ratio        | 13.3                                     | *                                       |
| Mean stream gradient     | 1 %                                      | 1.2 %                                   |
| Entrenchment             | slight                                   | slight                                  |
| Sinuosity                | high                                     | high                                    |
| Channel stability rating | medium fair (100)                        | •                                       |
| RASI values              | •  | •                                       |
| % substrate < 4mm        | •  | •                                       |
| mean substrate size (mm) | •  | •                                       |
| Mean % canopy cover      | 0%                                       | 0%                                      |
| Barriers                 | none                                     | none                                    |
| # instream woody debris  | 242 (8.5/100 m <sup>2</sup> )            | 58 (8.2/100 m <sup>2</sup> )            |
| Dominant habitat complex | riffle - pool                            | riffle - pool                           |
| Riffle/pool ratio        | 3.5 : 1                                  | 3.5 : 1                                 |
| Residual pool depth      | 0.9 m                                    | 0.5 m                                   |
| Total available habitat  | 39,637 m <sup>2</sup>                    | 5,689 m <sup>2</sup>                    |

**Table 3.21. Summary table showing values for reach eleven, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                             | <b>1994</b>                             |
|--------------------------|---|---|
| Stream order             | 4th                                     | 4th                                     |
| Elevation                | 2,750-2,760 ft                          | 2,750-2,760 ft                          |
| Dominant land use        | Livestock grazing                       | Livestock grazing                       |
| Riparian ecotype         | Coniferous/Deciduous<br>mix with shrubs | Coniferous/deciduous<br>mix with shrubs |
| Reach length             | 589 m                                   | 1,145 m                                 |
| Channel type             | E6                                      | E6                                      |
| Bankfull width           | 27 ft                                   | •                                       |
| Bankfull depth           | 6.9 ft                                  | •                                       |
| Width/depth ratio        | 3.9                                     | 11.0                                    |
| Mean stream gradient     | 1 %                                     | 1.2 %                                   |
| Entrenchment             | slight                                  | slight                                  |
| Sinuosity                | very high                               | very high                               |
| Channel stability rating | high poor (125)                         | high poor (115)                         |
| RASI values              | NA                                      | NA                                      |
| % substrate < 4mm        | NA                                      | NA                                      |
| mean substrate size (mm) | NA                                      | NA                                      |
| Mean % canopy cover      | 21.4 %                                  | 35 %                                    |
| Barriers                 | none                                    | none                                    |
| # instream woody debris  | 25 (0.1/100 m <sup>2</sup> )            | 32 (0.1/100 m <sup>2</sup> )            |
| Dominant habitat complex | riffle - pool                           | riffle - pool                           |
| Riffle/pool ratio        | 2.4 : 1                                 | 2.3 : 1                                 |
| Residual pool depth      | 0.7 m                                   | 0.4 m                                   |
| Total available habitat  | 6,237 m <sup>2</sup>                    | 4,439 m <sup>2</sup>                    |

Mean stream gradient was 1 percent and mean residual pool depth was 0.7 and 0.4 meters for 1993, and 1994, respectively (Table 3.21). Channel stability was rated as poor (Table 3.21). Landform slope ranged from 0-5 percent and no mass wasting was evident on the upper banks. Shallow rooted plant species provided discontinuous bank cover. Lower banks lacked a cohesive structure, resulting in continuous bank cutting and extensive fine sediment deposition. Debris jam potential was estimated to be high. Channel capacity was adequate to contain peak flows. Channel substrate was comprised of a loose assemblage of fine gravels and sand and appeared to be in a constant state of flux. Dominant habitat complex was riffle pool with total available habitat declining in 1994.

Reach twelve is 1,257 meters in length and encompasses elevations ranging from 2,760 feet to 2940 feet (Figure 2.2). The reach was classified as a C5-C6 channel type (Table 3.22). This is a fourth order drainage with an average stream gradient of less than 1 percent. Dominant land use was forest production and livestock grazing. Riparian areas were dominated by large shrub species. Average canopy closure was 52 percent and instream woody debris densities were 28.1/100 m<sup>2</sup> in 1993. In 1994 average canopy closure was 44 percent and instream woody debris densities were 13.2/100 m<sup>2</sup>. Riffle to pool ratios was 1.3:1 and mean residual pool depths were 0.4 meters in 1993. IN 1994 riffle-pool ratio was 2: 1 and residual pool depth was 0.4 meters. (Table 3.22). Channel stability was rated as poor (Table 3.22). Landform slope was less than one percent and no sediment sources were associated with the floodplain. Riparian vegetation showed signs of overgrazing and shallow-rooted plant species provided only discontinuous bank cover. Overbank flows were common in stream segments degraded by cattle use. Lower banks consisted primarily of gravel and smaller sized material and bank failure resulting from undercutting was common. Extensive deposition of fine sediment accelerated bar development. The channel bottom was composed of loose, well rounded sediments with evidence of scouring and deposition occurring year long. Aquatic vegetation was absent in this reach.

Reach thirteen extends 1,508 meters up the mainstem of Benawah Creek beginning at the confluence with the South Fork Benawah Creek (Figure 2.2). The reach is a second order drainage and encompasses elevations ranging from 2,940 to 3,060 feet. The dominant land type was forest with a riparian area comprised mainly of young coniferous and deciduous trees. The reach was classified as a E4 channel type characterized by very high sinuosity and slight entrenchment. Mean stream gradient was 2 percent (Table 3.23). Average stream canopy cover was 75 percent with woody debris densities of 32.4/100 m<sup>2</sup> in 1993. In 1994, average stream canopy cover was 84 percent with woody debris densities of 17.2/100 m<sup>2</sup>. Mean residual pool depth was 0.4 meters in

**Table 3.22. Summary table showing values for reach twelve, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                             | <b>1994</b>                             |
|--------------------------|---|---|
| Stream order             | 4th                                     | 4th                                     |
| Elevation                | 2,760-2,940 ft                          | 2,760-2,940 ft                          |
| Dominant land use        | Forest/Livestock grazing                | Forest/Livestock grazing                |
| Riparian ewtype          | Coniferous/deciduous<br>mix with shrubs | Coniferous/deciduous<br>mix with shrubs |
| Reach length             | 1,257 m                                 | 1,412 m                                 |
| Channel type             | C5 - C6                                 | C5 - C6                                 |
| Bankfull width           | NA                                      | *                                       |
| Bankfull depth           | 3.0 ft                                  | •                                       |
| Width/depth ratio        | NA                                      | *                                       |
| Mean stream gradient     | <1 %                                    | 1.2 %                                   |
| Entrenchment             | slight                                  | slight                                  |
| Sinuosity                | high                                    | high                                    |
| Channel stability rating | mediuim poor (129)                      | poor (115)                              |
| RASI values              | NA                                      | NA                                      |
| % substrate < 4mm        | NA                                      | NA                                      |
| mean substrate size (mm) | NA                                      | NA                                      |
| Mean % canopy cover      | 51.6 %                                  | 44 %                                    |
| Barriers                 | none                                    | none                                    |
| # instream woody debris  | 355 (28.1/1 00 m <sup>2</sup> )         | 186 (13.2/100 m <sup>2</sup> )          |
| Dominant habitat complex | riffle - pool                           | riffle - pool                           |
| Riffle/pool ratio        | 1.3 : 1                                 | 2 : 1                                   |
| Residual pool depth      | 0.4 m                                   | 0.4 m                                   |
| Total available habitat  | 10,197 m <sup>2</sup>                   | 12,394 m <sup>2</sup>                   |

**Table 3.23. Summary table showing values for reach thirteen, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                               | <b>1994</b>                               |
|--------------------------|---|---|
| Stream order             | 2nd                                       | 2nd                                       |
| Elevation                | 2,940-3,060 ft                            | 2,940-3,060 ft                            |
| Dominant land use        | Forest                                    | Forest                                    |
| Riparian ecotype         | Coniferous/Deciduous mix with young trees | Coniferous/Deciduous mix with young trees |
| Reach length             | 1,508 m                                   | 761 m                                     |
| Channel type             | E4  | E4  |
| Bankfull width           | 15ft                                      | •   |
| Bankfull depth           | 2.5 ft                                    | •   |
| Width/depth ratio        | 6   |   |
| Mean stream gradient     | 2 %                                       | 3.8 %                                     |
| Entrenchment             | slight                                    | slight                                    |
| Sinuosity                | very high                                 | very high                                 |
| Channel stability rating | high fair (83)                            | fair (107)                                |
| RASI values              | NA  | NA  |
| % substrate < 4mm        | NA  | NA  |
| mean substrate size (mm) | NA  | NA  |
| Mean % canopy cover      | 75.3 %                                    | 84 %                                      |
| Barriers                 | none                                      | none                                      |
| # instream woody debris  | 488 (32.41/100 m <sup>2</sup> )           | 131 (17.2/100 m <sup>2</sup> )            |
| Dominant habitat complex | riffle-pool /step pool rapid              | riffle-step pool                          |
| Riffle/pool ratio        | 1 : 1.2                                   | 2 : 1                                     |
| Residual pool depth      | 0.4 m                                     | 0.7 m                                     |
| Total available habitat  | 13,783 m <sup>2</sup>                     | 6,873 m <sup>2</sup>                      |

1993 and 0.7 meters in 1994. This reach had a channel stability rating of fair (Table 3.23). Upper bank slope ranged from 0-20 percent and included examples of historic bank failures that had stabilized over time. Debris jam potential was minimal with only small woody debris present in the channel. Vegetation covered 70-90 percent of the upper banks, but plant communities lacked the species diversity found in other reaches. The channel capacity was adequate for containing existing peak flows. Lower banks contained small cobble-sized materials with only minor cutting occurring at obstructions and outcurves. New bar development was minimal. The channel bottom was moderately stable and very little scouring and deposition was observed in the reach. The dominant habitat complex was a riffle-step pool complex. Total available habitat declined in 1994.

Reach fourteen consisted of 3,293 meters of the South Fork Benawah Creek (denoted on Figure 2.2 as reaches 13a and 14a) beginning at the confluence with the mainstem. Dominant land use is livestock grazing and riparian areas are dominated by young conifers and deciduous trees, and shrubs. The reach was classified as E3b and E4 channel types and encompassed elevations ranging from 2,647 to 3,061 feet (Table 3.24). Average stream canopy cover was 52 percent and densities of instream woody debris were 12.111 00m<sup>2</sup> in 1993. In 1994, mean stream canopy cover was 76 percent with densities of instream woody debris of 2.9/100m<sup>2</sup>. The dominant habitat complex was riffle-pool/step pool rapid. Riffle to pool ratio was 1.1: 1 and 2.7: 1 in 1993 and 1994, respectively. Mean residual pool depth remained constant in 1993 and 1994 at 0.2 meters (Table 3.24). Channel stability was rated fair to good (Table 3.24). Upper bank slopes ranged from 0-15 percent with no evidence of mass wasting. Debris jam potential was minimal with mostly small twigs and limbs present. Upper bank vegetation consisted mainly of shrubs and covered 50 to 70 percent of the surface area. The channel capacity was adequate for existing peak flows. Minor bank cutting was observed at obstructions and outcurves. Some fresh deposition of gravel sized material was noted on point bars. Bottom materials were moderately packed and stable. RASI surveys were conducted at one location on the South Fork Benawah Creek in 1993. The index value of 50 suggested an equilibrium had been reached between scour and depositional processes (Table 3.24). Levels of fine sediment were higher than other downstream reaches (20 percent).

### 3.1.3. Evans Creek

Channel typing surveys identified 10 reaches in the Evans Creek watershed. Reaches 1-7 are located on the mainstem, reach 5a encompasses the North Fork drainage, and reaches 5b and 6b are located on the South Fork Evans Creek (Figure 2.3). Physical conditions are described below for reaches 1-7 on the mainstem, reach 5a on the North Fork, and 5b and 6b on the South



**Table 3.24. Summary table showing values for reach fourteen, Benawah Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b>  |
|--------------------------|--|--|
| Stream order             | 3rd  | 3rd  |
| Elevation                | 2,847-3,061 ft                                     | 2,847-3,061 ft                                     |
| Dominant land use        | Forest /Livestock grazing                          | Forest/Livestock grazing                           |
| Riparian ecotype         | Coniferous/Deciduous mix<br>with shrub/young trees | Coniferous/Deciduous mix<br>with shrub/young trees |
| Reach length             | 3,293 m  | 3,493 m  |
| Channel type             | E3b - E4   | E3b - E4   |
| Bankfull width           | 22 ft  | •  |
| Bankfull depth           | 2.2 ft   | •  |
| Width/depth ratio        | □□   | *  |
| Mean stream gradient     | 3 %  | <1 %   |
| Entrenchment             | slight   | slight   |
| Sinuosity                | very high  | very high  |
| Channel stability rating | low good (76)                                      | medium good (92)                                   |
| RASI values              | 50   | •  |
| % substrate < 4mm        | 20.1   | •  |
| mean substrate size (mm) | 76   | •  |
| Mean % canopy cover      | 51.7 %   | 76 %   |
| Barriers                 | none   | none   |
| # instream woody debris  | 398 (0/100 m <sup>2</sup> )                        | 103 (0/1 00 m <sup>2</sup> )                       |
| Dominant habitat complex | riffle - pool/step pool rapid                      | riffle / step pool                                 |
| Riffle/pool ratio        | 1.1 : 1  | 2.7 : 1  |
| Residual pool depth      | 0.2 m  | 0.2 m  |
| Total available habitat  | 54,795 m <sup>2</sup>                              | 24,604 m <sup>2</sup>                              |

Fork. For general descriptions to summarize conditions, mainstem reaches 5-7 were combined and reaches 5b and 6b of the South Fork were combined.

Reach one of Evans Creek was classified as a C6 channel type with a total length of 880 meters (Figure 2.3). Evans Creek is a third order drainage at the mouth. Dominant land use within this reach was livestock grazing. The riparian area was dominated by grasses and small shrubs. The channel is highly sinuous and mean stream gradient is less than 1 percent (Table 3.25). Average canopy cover in this reach was 28 percent in 1993 and 16 percent in 1994. No instream woody debris was present. Riffle to pool ratio was 1.7:1 and mean residual pool depth was 0.4 meters (Table 3.25). This reach received a poor channel stability rating as a result of a combination of unstable conditions on the upper and lower banks and channel. Landform slope ranged from 0-3 percent and no evidence of mass wasting problems were present, however, upper bank vegetation was shallow rooted and discontinuous, covering less than 50 percent of the surface area. Debris jam potential was limited to small limbs and twigs. The channel capacity had been reduced as a result of frequent bank failures and was inadequate to contain bankfull flows. The lower banks consisted mainly of highly erodible fine sediment. The channel bottom was unstable and accelerated bar development was observed throughout this reach.

Reach two extended 1,581 meters and was classified as a C3 channel type (Figure 2.3). Dominant land use within this reach was livestock grazing. The riparian area was dominated by mature trees and shrubs. Average canopy cover in this reach was 52 percent with instream woody debris loads of 1.5/100 m<sup>2</sup> in 1993. In 1994, average canopy cover was 34 percent with instream woody debris densities 9.4/100m<sup>2</sup>. Residual pool depths averaged 0.2 meters in 1993 and 0.6 meters in 1994 (Table 3.26). Channel stability rated from poor to fair (Table 3.26). Upper bank slopes were generally less than 3 percent and no mass wasting was evident. Bank vegetation consisted of discontinuous, shallow rooted species with a density of less than 50 percent. Channel capacity was reduced due to channel alterations and allowed for occasional overbank flooding. There was significant amounts of bank cutting and sloughing, as well as accelerated bar development. Riffle armor stability index values were highly variable, ranging from 39.8 to 75.2 (mean=59) in 1993 and 89.4 to 97.5 (mean=93) in 1994 (Table 3.26). Fine sediment (< 4 mm) generally comprised less than 10 percent of the sampled substrate with the exception of a single measurement (36 percent) taken in 1993 (Table 3.26). Dominant habitat complex was riffle-pool with total available habitat decreasing in 1994.

Reach three is 1,722 meters in length and encompasses elevations ranging from 2,130 to 2,270 feet (Figure 2.3). Dominant land use within this reach was forested. The riparian area was dominated by mature coniferous

**Table 3.25. Summary table showing values for reach one, Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                           | <b>1994</b>                           |
|--------------------------|---------------------------------------|---------------------------------------|
| Stream order             | 3rd                                   | 3rd                                   |
| Elevation                | 2,130-2,130 ft                        | 2,130-2,130 ft                        |
| Dominant land use        | Livestock grazing                     | Livestock grazing                     |
| Riparian ecotype         | Coniferous/deciduous<br>grasses/forbs | Coniferous/deciduous<br>grasses/forbs |
| Reach length             | 880 m                                 | 1,583 m                               |
| Channel type             | C6                                    | C6                                    |
| Bankfull width           | 21 ft                                 | •                                     |
| Bankfull depth           | 2 ft                                  | •                                     |
| Width/depth ratio        | 10                                    | •                                     |
| Mean stream gradient     | <1 %                                  | 1.3 %                                 |
| Entrenchment             | slight                                | slight                                |
| Sinuosity                | high                                  | high                                  |
| Channel stability rating | high poor (122)                       | high poor (125)                       |
| RASI values              | NA                                    | NA                                    |
| % substrate < 4mm        | NA                                    | NA                                    |
| mean substrate size (mm) | NA                                    | NA                                    |
| Mean % canopy cover      | 28.3 %                                | 16.0 %                                |
| Barriers                 | none                                  | none                                  |
| # instream woody debris  | 0 (0/100 m <sup>2</sup> )             | 5 (0/100 m <sup>2</sup> )             |
| Dominant habitat complex | pool-riffle                           | riffle-pool                           |
| Riffle/pool ratio        | 1.7 : 1                               | 1 : 1                                 |
| Residual pool depth      | 0.4 m                                 | 0.5 m                                 |
| Total available habitat  | 35,957 m <sup>2</sup>                 | 9,556 m <sup>2</sup>                  |

**Table 3.26. Summary table showing values for reach two, Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b>  |
|--------------------------|--|--|
| Stream order             | 3rd  | 3rd  |
| Elevation                | 2,130-2,130 ft                                       | 2,130-2,130 ft                                       |
| Dominant land use        | Forest /Livestock grazing                            | Forest/Livestock grazing                             |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees/shrubs | Coniferous/deciduous mix<br>with mature trees/shrubs |
| Reach length             | 1,581 m  | 965 m  |
| Channel type             | c3   | c3   |
| Bankfull width           | 25ft   | *  |
| Bankfull depth           | 3ft  | *  |
| Width/depth ratio        | 7.8  | 6.0  |
| Mean stream gradient     | 1 %  | 1.3 %  |
| Entrenchment             | slight   | slight   |
| Sinuosity                | high   | high   |
| Channel stability rating | high poor (119)                                      | medium fair (101)                                    |
| RASI values              | 75; 40; 63   | 91; 98; 89   |
| % substrate < 4mm        | 14 %   | 6.6 %  |
| mean substrate size (mm) | 66   | 145  |
| Mean % canopy cover      | 51.8 %   | 34.1 %   |
| Barriers                 | none   | none   |
| # instream woody debris  | 76 (4.0/100 m <sup>2</sup> )                         | 26 (0/100 m <sup>2</sup> )                           |
| Dominant habitat complex | riffle - pool  | riffle - pool  |
| Riffle/pool ratio        | 1.2: 1   | 1.2: 1   |
| Residual pool depth      | 0.2 m  | 0.6 m  |
| Total available habitat  | 75,090 m <sup>2</sup>                                | 25,006 m <sup>2</sup>                                |

**Table 3.27. Summary table showing values for reach three, Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                     | <b>1994</b>                                     |
|--------------------------|---|---|
| Stream order             | 3rd   | 3rd   |
| Elevation                | 2,130-2,270 ft                                  | 2,130-2,270 ft                                  |
| Dominant land use        | forested  | forested  |
| Riparian ecotype         | Coniferous/deciduous mix<br>mature trees/shrubs | Coniferous/deciduous mix<br>mature trees/shrubs |
| Reach length             | 1,722 m   | 927 m   |
| Channel type             | E3b   | E3b   |
| Bankfull width           | ft  | *   |
| Bankfull depth           | ft  | •   |
| Width/depth ratio        |   | •   |
| Mean stream gradient     | 2 %   | 1.9 %   |
| Entrenchment             | slight  | slight  |
| Sinuosity                | very high                                       | very high                                       |
| Channel stability rating | fair (95)                                       | fair (94)                                       |
| RASI values              | 67; 82; 60                                      | 84; --; 92                                      |
| % substrate < 4mm        | 19  | 18.9  |
| mean substrate size (mm) | 98  | 136   |
| Mean % canopy cover      | 65.2 %  | 44.1 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 56 (3.2000 m <sup>2</sup> )                     | 90 (9.7/100 m <sup>2</sup> )                    |
| Dominant habitat complex | riffle - pool                                   | riffle - pool                                   |
| Riffle/pool ratio        | 1.6: 1  | 1.2: 1  |
| Residual pool depth      | 1.5 m   | 0.5 m   |
| Total available habitat  | 21.471 m <sup>2</sup>                           | 14.471 m <sup>2</sup>                           |

and deciduous trees and large shrubs. The reach was classified as a E3b channel type (Table 3.27). Mean stream gradient was 2 percent, channel entrenchment was slight ( $>2.4$ ) and sinuosity was very high ( $>1.5$ ). Average stream canopy cover was 65 percent and instream woody debris loads were 3.3/100m<sup>2</sup> in 1993. In 1994, average stream canopy cover was 44% with instream woody debris densities of 9.7/100m<sup>2</sup>. Riffle to pool ratio stayed relatively unchanged for 1993 and 1994 at 1.6: 1 and 1.2:1, respectively. Mean residual pool depths decreased in 1994 from 1.5 meters in 1993 to 0.5 meters in 1994 (Table 3.27). This decrease corresponded to a decrease in total available habitat. Channel stability rated fair (Table 3.27). Density of bank vegetation was 70-90 percent. Debris jam potential was abundant but mostly in the form of small twigs and limbs. The channel capacity was adequate to contain peak flows. Slight bank cutting occurred at flow deflectors and outcurves, and some new point bar development was noted. Mean RASI values were higher in 1994 than in 1993 (88 and 70, respectively). Mean percent fines ( $< 4$  mm) were 17 and 19 for 1993 and 1994, respectively (Table 3.27).

Reach four is 1,800 meters in length and encompasses elevations ranging from 2,270 to 2,580 feet (Figure 2.3). Dominant land use was forested with the riparian area consisting of a mix of mature coniferous and deciduous trees. The reach was classified as a B3a channel type with moderate entrenchment and sinuosity (Table 3.28). Mean stream gradient was 3 percent. Average canopy cover was 69 percent with instream woody debris at densities of 1 X100 m<sup>2</sup> in 1993. In 1994, average stream canopy cover decreased to 44 percent while instream woody debris increased to 10.2/100 m<sup>2</sup>. Riffle to pool ratio was 1:32 and mean residual pool depth was 0.3 meters in 1993. In 1994, the riffle to pool ratio was 1.2:1 with a mean residual pool depth of 0.3 meters. (Table 3.28). Channel stability rating fair for this reach (Table 3.28). Upper bank slopes ranged from 2040 percent with intermittent, older bank failures noted. Bank vegetation density was greater than 90 percent with high diversity of both shallow and deep rooted species. The channel capacity was adequate for existing flow conditions. There was a moderate amount of lower bank cutting at obstructions and outcurves, and some new gravel bar development in this reach. The channel bottom throughout the reach was fairly stable with approximately 5-12 percent of the bottom substrate affected by scouring and deposition. Algal blooms and moss were observed in this reach in slack and swift water. Mean RASI values were considerably higher in 1994 than in 1993 (89 compared with 41) as were the geometric mean diameters of point bar particles (164 mm compared with 84 mm). Fine sediment ( $< 4$  mm) comprised an average of 13 and 22 percent in 1993 and 1994, respectively (Table 3.28).

Channel surveys were conducted on 990 meters of the mainstem upstream of the three forks area (Figure 2.3). This is a second order drainage

**Table 3.28. Summary table showing values for reach four, Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                   | <b>1994</b>                                   |
|--------------------------|---|---|
| Stream order             | 3rd   | 3rd   |
| Elevation                | 2,270 -2,580 ft                               | 2,270-2,580 ft                                |
| Dominant land use        | Forested                                      | Forested                                      |
| Riparian ecotype         | Coniferous/deciduous<br>mix with mature trees | Coniferous/deciduous<br>mix with mature trees |
| Reach length             | 1,800 m                                       | 1,818 m                                       |
| Channel type             | B3a   | B3a   |
| Bankfull width           | •   | •   |
| Bankfull depth           | •   | •   |
| Width/depth ratio        | •   | ∞   |
| Mean stream gradient     | 3 %   | 2.8 %   |
| Entrenchment             | moderate                                      | moderate                                      |
| Sinuosity                | moderate                                      | moderate                                      |
| Channel stability rating | low good (77)                                 | fair (86)                                     |
| RASI values              | 59; 61; 57                                    | 90; 93; 85                                    |
| % substrate < 4mm        | 13  | 21.8  |
| mean substrate size (mm) | 84  | 164   |
| Mean % canopy cover      | 69.4 %  | 44 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 21 (1.2/l 00 ft)                              | 186 (10.2/l 00 linear ft)                     |
| Dominant habitat complex | rifle - pool                                  | rifle - pool                                  |
| Riffle/pool ratio        | 1 : 32  | 1.2: 1  |
| Residual pool depth      | 0.3 m   | 0.3 m   |
| Total available habitat  | 25,541 m <sup>2</sup>                         | 30,505 m <sup>2</sup>                         |

**Table 3.29. Summary table showing values for reach five, Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                   | <b>1994</b>                                   |
|--------------------------|---|---|
| Stream order             | 2nd   | 2nd   |
| Elevation                | 2,580-3,340 ft                                | 2,580-3340 ft                                 |
| Dominant land use        | Forested                                      | Forested                                      |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees | Coniferous/deciduous mix<br>with mature trees |
| Reach length             | 990 m   | 735 m   |
| Channel typt             | A2a - B3a                                     | A2a - B3a                                     |
| Bankfull widt:           | •   | ☒   |
| Bankfull depth           | •   | *   |
| Width/depth ratio        | 17.3  | •   |
| Mean stream gradient     | 1.6 %   | 3.1 %   |
| Entrenchment             | moderate                                      | moderate                                      |
| Sinuosity                | moderate                                      | moderate                                      |
| Channel stability rating | good (67)                                     | good (72)                                     |
| RASI values              | 43; 52; 31                                    | 60; 62; 64                                    |
| % substrate < 4mm        | 22 %  | 13.6 %  |
| mean substrate size (mm) | 75  | 123   |
| Mean % canopy cover      | 37.5 %  | 59.7 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 90 (9.1 /100 m <sup>2</sup> )                 | 191 (25.91100 m <sup>2</sup> )                |
| Dominant habitat complex | step - pool : rapid                           | step - pool : rapid                           |
| Riffle/pool ratio        | 1 : 1.3                                       | 1 :6  |
| Residual pool depth      | 0.5 m   | 0.5 m   |
| Total available habitat  | 10,696 m <sup>2</sup>                         | 17,779 m <sup>2</sup>                         |



with elevations beginning at 2,580 feet and rising to 3,340 feet. The reach was comprised of A and B channel types with moderate to high entrenchment ratios (< 2.2) and low to moderate sinuosity (Table 3.29). Stream gradient varied from 7 to 10 percent. Average stream canopy cover was 37.5 percent with instream woody debris densities of 9.11100 m<sup>2</sup> in 1993. In 1994, average stream canopy cover was 59.7 percent with instream woody debris densities of 25.9/100 m<sup>2</sup>. Riffle to pool ratio was 1: 1.3 and mean residual pool depth was 0.5 meters. Channel stability was rated good (Table 3.29). Upper bank slopes ranged from 1 O-90 percent with no evidence of mass wasting problems. Upper bank vegetation density was greater than 90 percent with both shallow and deep rooted species represented. The channel capacity was adequate for existing peak flows. There were few obstructions or flow deflectors in this reach. Lower bank cutting was minimal with only a few areas of new gravel bar development. The channel bottom was stable with abundant moss and algal growth noted throughout the reach. The Old Rainbow Mine is located in the upper segment of reach 7 (Figure 4.3). In this area some small bank failures were observed on steeper, less stable slopes. Lower bank cutting was observed at obstructions and outcurves, with moderate amounts of new bar development from coarse gravels. Mean RASI values were 42 and 62 for 1993 and 1994, respectively. Fine sediment comprised up to 23 percent of the sampled substrate in 1993 and up to 17 percent in 1994 (Table 3.29).

Physical surveys were conducted on 977 meters of the North Fork Evans Creek encompassing elevations ranging from 2,580 to 3,170 feet (Figure 2.3). This reach is a second order stream. The dominant land use was forested with some areas of livestock grazing. The riparian area consisted of mature coniferous and deciduous trees and shrubs. This drainage included B3a and E3b channel types (Table 3.30). Average stream gradient was 10 percent. Average canopy cover in this reach was 34 percent with instream woody debris densities of 1.61100 m<sup>2</sup>. Riffle to pool ratio was 3:1 and average residual pool depth was 0.3 meters (Table 3.30). No significant changes were reported in 1994. Channel stability ranged from fair to good (Table 3.30). Upper bank slopes ranged from 0-45 percent with no evidence of mass wasting problems. Bank vegetation density varied from 70-90 percent. Plant density was less vigorous in some areas as a result of encroachment of the north fork road. There were fair amount of debris jam potential in this reach. The channel capacity was adequate for existing peak flows. Small amounts of bank cutting occurred at obstructions and outcurves with some new bar development resulting from deposition of gravels. Clinging aquatic vegetation was abundant throughout this reach.

Surveys were conducted along 694 meters of the South Fork Evans Creek (Figure 2.3). This is a second order drainage with elevations beginning at 2,580 feet and rising to 3,040 feet. Dominant land use is forested with the

**Table 3.30. Summary table showing values for reach six, (north fork) Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                             | <b>1994</b>                             |
|--------------------------|---|---|
| Stream order             | 2nd                                     | 2nd                                     |
| Elevation                | 2,580-3,170 ft                          | 2,580-3,170 ft                          |
| Dominant land use        | Forested/Livestock grazing              | Forested/Livestock grazing              |
| Riparian ecotype         | Coniferous/Deciduous with mature timber | Coniferous/Deciduous with mature timber |
| Reach length             | 977 m                                   | 1,015 m                                 |
| Channel type             | B3a - E3b                               | B3a - E3b                               |
| Bankfull width           | .                                       | *                                       |
| Bankfull depth           | .                                       | .                                       |
| Width/depth ratio        | 6.3                                     | .                                       |
| Mean stream gradient     | 10%                                     | 10.8 %                                  |
| Entrenchment             | slight                                  | slight                                  |
| Sinuosity                | high                                    | high                                    |
| Channel stability rating | good (70)                               | good (72)                               |
| RASI values              | NA                                      | NA                                      |
| % substrate < 4mm        | NA                                      | NA                                      |
| mean substrate size (mm) | NA                                      | NA                                      |
| Mean % canopy cover      | 34%                                     | 39 %                                    |
| Barriers                 | none                                    | none                                    |
| # instream woody debris  | 50 (1.6 /100 ft)                        | 35 (1 /100 ft)                          |
| Dominant habitat complex | step-pool cascade                       | step-pool cascade                       |
| Riffle/pool ratio        | 3 : 1                                   | 5 : 1                                   |
| Residual pool depth      | 0.3 m                                   | 0.2 m                                   |
| Total available habitat  | 5,996 m <sup>2</sup>                    | 6,526 m <sup>2</sup>                    |

**Table 3.31. Summary table showing values for reach seven, (south fork) Evans Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>                                   | <b>1994</b>                                   |
|--------------------------|---|---|
| Stream order             | 2nd   | 2nd   |
| Elevation                | 2,580-3,040 ft                                | 2,580-3,040 ft                                |
| Dominant land use        | Forested                                      | Forested                                      |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees | Coniferous/deciduous mix<br>with mature trees |
| Reach length             | 694 m   | 665 m   |
| Channel type             | A2 - A4                                       | A2 - A4                                       |
| Bankfull width           | .   | .   |
| Bankfull depth           | *   | .   |
| Width/depth ratio        | 11.1  | .   |
| Mean stream gradient     | 8.1 %   | 8.1 %   |
| Entrenchment             | high  | high  |
| Sinuosity                | low   | low   |
| Channel stability rating | fair (76)                                     | fair (82)                                     |
| RASI values              | 63 ; 85 ; 72                                  | 65 ; 78 ; —                                   |
| % substrate < 4mm        | 23 %  | 38.5%   |
| mean substrate size (mm) | 80  | 67  |
| Mean % canopy cover      | 70.0 %  | 66 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 86 (3.8 /100 ft )                             | 34 (2 /100 ft )                               |
| Dominant habitat complex | step - pool : rapid                           | step - pool : rapid                           |
| Riffle/pool ratio        | NA  | 5 : 1   |
| Residual pool depth      | NA  | 0.2   |
| Total available habitat  | 2,827 m <sup>2</sup>                          | 7,960 m <sup>2</sup>                          |

riparian area consisting of mature coniferous and deciduous trees. This reach was classified as an A channel type. Average stream gradient was 8 percent with some areas in excess of 15 percent. Average canopy cover in this reach was 70 percent with instream woody debris densities of 3.81/100 m<sup>2</sup>. The dominant habitat complex in this reach was step-pool/rapid (Table 3.31). Channel stability was rated fair to good (Table 3.31). Upper bank slopes ranged from 30-75 percent with some infrequent, small upper bank failures, and some older healed bank failures. Large debris dams were common throughout the reach with abundant recruitable sized materials located within the floodplain. Upper bank vegetation varied from a vigorous assemblage of deep and shallow rooted species with a density greater than 90 percent to less vigorous communities. The channel capacity was barely adequate to contain peak flows. Most bank cutting occurred as a result of obstructions created by debris dams. Some new gravel bar development was noted in the more sinuous segments of the reach. Increased filling of pools with silts and sands was observed in the upper reaches of this segment. There was abundant clinging aquatic vegetation throughout this reach. RASI values ranged from 63.4 to 84.6 and percent fine sediment was greater than in all downstream reaches (mean=39 percent).

#### 3.1.4. Alder Creek

Channel surveys identified 13 reaches in the Alder Creek watershed (Figure 2.4). Nine of these reaches were located in the mainstem of Alder Creek and four reaches were located in the North Fork drainage. Of the nine reaches identified in the mainstem of Alder Creek, reaches one and two were combined for general descriptions and were labeled reach one. Reaches three and four were also combined and labeled reach two. Reaches five and six were also combined and labeled reach three. In the North Fork drainage reach 2a and 3a were combined into reach two, North Fork Alder Creek.

Reach one is 881 meters in length and encompasses elevations ranging from 2,260 to 2,350 feet (Figure 2.4). The dominant land type in this reach is forested/floodplain with a riparian area consisting of young coniferous/deciduous trees, shrubs, forbs and grasses. The reach was classified as a B3a-B3c channel type (Table 3.32). Average stream canopy cover was 32 percent with instream woody debris density of 1.61/100 m<sup>2</sup> for 1993. In 1994 average stream canopy was 22 percent with instream woody debris densities of 32/100 m<sup>2</sup>. Riffle to pool ratio was 1-4:1 and mean residual pool depth was 0.6 meters (Table 3.32). Channel stability was rated as fair to good (Table 3.32). Upper bank slopes ranged from 5-60 percent with no evidence of mass wasting problems. Vegetation density on the upper slopes was quite vigorous (90 percent density). Debris jam potential was moderate, consisting mostly of small twigs and limbs. The channel capacity is barely adequate to contain existing

**Table 3.32. Summary table showing values for reach one, Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,260-2,350 ft  | 2,260-2,350 ft   |
| Dominant land use        | Forested  | Forested   |
| Riparian ecotype         | Coniferous/deciduous mix<br>with young trees, shrubs<br>& grasses | Coniferous/deciduous mix<br>young timber/shrubs &<br>8 grasses |
| Reach length             | 881 m   | 876 m  |
| Channel type             | B3a - B3c   | B3a - B3c  |
| Bankfull width           | ft  | •  |
| Bankfull depth           | ft  | •  |
| Width/depth ratio        |   | •  |
| Mean stream gradient     | 2 %   | 5 %  |
| Entrenchment             | moderate  | moderate   |
| Sinuosity                | moderate  | moderate   |
| Channel stability rating | fair (88)   | good (76)  |
| RASI values              | 32; 38; 29  | 58; 69; 39   |
| % substrate < 4mm        | 7.7   | 9.5  |
| mean substrate size (mm) | 57  | 136  |
| Mean % canopy cover      | 32%   | 22%  |
| Barriers                 | none  | none   |
| # instream woody debris  | 14 (1.61100 m <sup>2</sup> )                                      | 28(3.2/100 m <sup>2</sup> )                                    |
| Dominant habitat complex | rifle - pool  | rifle - -001   |
| Rifle/pool ratio         | 1.4: 1  | 1 :1   |
| Residual pool depth      | 0.6 m   | 0.7 m  |
| Total available habitat  | 28,242 m <sup>2</sup>   | 18,411 m <sup>2</sup>  |

peak flows. Bank cutting is limited to outcurves and constrictions with moderate amounts of sand and gravel deposition. Mean RASI values were 33 and 55 in 1993 and 1994, respectively. Fine sediment (< 4 mm) comprised less than 10 percent of the sampled substrate in both 1993 and 1994 (Table 3.32). Dominant habitat complex consisted of riffle-pool with total available habitat declining in 1994.

Reach 2 extends for 4,188 meters and encompass elevations ranging from 2,350 to 2,630 feet (Figure 2.4). The upstream end of this reach is delineated by a twenty-five foot bedrock waterfall. This waterfall acts as a fish passage barrier except for in extreme high water years. Dominant land type was forested with livestock grazing occurring on the adjacent uplands. The riparian area was dominated by mature coniferous and deciduous trees with shrubs, forbs, and grasses present. The reaches included A2 and A3 channel types. The channel was highly entrenched, had low sinuosity, and a mean stream gradient of 3 percent (Table 3.33). Average stream canopy cover was 30 percent and instream woody debris densities were 0.5/100 feet. Riffle to pool ratio was 1.2: 1 and mean residual pool depth was 0.6 meters (Table 3.33). Channel stability rated fair to good (Table 3.33). Upper bank slopes ranged from 65-86 percent and were marked by frequent bank failures and small areas of mass wasting. Vegetation density was greater than 90 percent. Debris jam potential was moderate and consisted mainly of small woody debris. The channel capacity was adequate for existing peak flows. Bank cutting occurred at obstructions and outcurves with some newly developed gravel bars noted. Clinging aquatic vegetation was common in most areas of the reach. Mean RASI values were 25 and 59 in 1993 and 1994, respectively. Fines (c 4 mm) comprised an average of 7 percent of the sampled substrate in 1993 and 11 percent in 1994 (Table 3.33).

Reach 3 of Alder Creek were classified as C3 type channel with some higher gradient areas resembling E3b type channels (Table 3.36). The reach encompasses 2,439 meters and elevations vary from 2,163 to 2,700 feet (Figure 2.4). Dominant land use in this reach was livestock grazing. The riparian area consisted of shrubs, grasses and forbes. Average canopy cover was 32 percent with very little instream woody debris present (c 0.1/100 feet). Dominant habitat complex was riffle pool with one area of this reach consisting of a bedrock/boulder cascade area. Riffle to pool ratio was 1.3:1 and mean residual pool depth was 1.2 meters during 1993. In 1994, riffle to pool ratio was 4.4:1 with a mean residual pool depth of 0.2 meters. This reach of Alder Creek had a channel stability rating of good (Table 3.34). Upper bank slopes were predominantly less than 30 percent, though intermittent portions had slopes up to 80 percent. Debris jam materials consisted of small limbs and twigs. Vegetation density was 70-90 percent, but consisted mostly of shallow rooted species. The channel capacity was adequate for existing peak flows. Bank

**Table 3.33. Summary table showing values for reach two, Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,350-2,630 ft  | 2,350-2,630 ft   |
| Dominant land use        | Livestock grazing/forested  | Livestock grazing/forested   |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees, shrubs,<br>forbs and grasses | Coniferous/deciduous mix<br>mature trees, shrubs,<br>forbs and grasses |
| Reach length             | 4,188 m   | 4,207 m  |
| Channel type             | A2-A3   | A2-A3  |
| Bankfull width           | ft  | •  |
| Bankfull depth           | •   | •  |
| Width/depth ratio        | •   | •  |
| Mean stream gradient     | 3 %   | 4 %  |
| Entrenchment             | high  | high   |
| Sinuosity                | low   | low  |
| Channel stability rating | slog (73)   | CI& (73)   |
| RASI values              | 33; 25; 16  | 62; 66; 50   |
| % substrate < 4mm        | 7.3   | 10.5   |
| Barriers                 | none  | none   |
| # instream woody debris  | 70 (0.5 /100 ft)  | 50 (6.4 /100 ft)   |
| Dominant habitat complex | riffle - pool   | riffle - pool  |
| Riffle/pool ratio        | 1.2 : 1   | 4: 1   |
| Residual pool depth      | 0.6 m   | 0.8 m  |
| Total available habitat  | 30.361 m <sup>2</sup>   | 96.260 m <sup>2</sup>  |

Table 3.34. Summary table showing values for reach three, Alder Creek, 1993 and 1994

| PARAMETER                | 1993  | 1994  |
|--------------------------|---|---|
| Stream order             | 4th   | 4th   |
| Elevation                | 2,163-2,700 ft  | 2,163-2,700 ft  |
| Dominant land use        | Livestock grazing                                       | Livestock grazing                                       |
| Riparian ecotype         | Coniferous/deciduous mix<br>with shrubs, forbs, grasses | Coniferous/deciduous mix<br>with shrubs, forbs, grasses |
| Reach length             | 2,490 m   | 2,439 m   |
| Channel type             | C3 - E3b  | C3 - E3b  |
| Bankfull height          | ft  | *   |
| Bankfull                 | ft  | •   |
| Width/depth ratio        | 11.3  | *   |
| Mean stream gradient     | 2 %   | 2.5 %   |
| Entrenchment             | slight  | slight  |
| Sinuosity                | high  | high  |
| Channel stability rating | <b>god</b> (68)   | <b>good</b> (72)  |
| RASI values              | 28; 40; 31  | 60; 46; 71  |
| % substrate < 4mm        | 9.6   | 9.4   |
| mean substrate size (mm) | 48  | 162   |
| Mean % canopy cover      | 32 %  | 32 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 14 (0.1 /100 ft)  | 15 (0.2 /100 ft)  |
| Dominant habitat complex | riffle - pool   | riffle - pool   |
| Riffle/pool ratio        | 1.3: 1  | 4.4 : 1   |
| Residual pool depth      | 1.2 m   | 0.2 m   |
| Total available habitat  | 42.423 m <sup>2</sup>                                   | 55.005 m <sup>2</sup>                                   |



cutting occurred at flow deflectors and outcurve areas. Some new bar development had occurred as evidenced by coarse gravel deposition. The channel bottom was fairly stable. Clinging aquatic vegetation was common throughout the reach with both algae and moss present. Mean RASI values were 33 and 59 in 1993 and 1994, respectively. Fine sediment (< 4 mm) comprised less than 10 percent of the sampled substrate in both 1993 and 1994 (Table 4.4.3).

Reach four extended for 5,053 meters and encompassed elevations ranging from 2,700 to 2,900 feet (Figure 2.4). Dominant land use in this reach was livestock grazing. The riparian area consisted of mature deciduous/coniferous trees, shrubs, grasses and forbs. The channel was classified as a C3 channel type with a mean stream gradient of 1.5 percent (Table 3.35). Average canopy cover was much lower than other reaches in the drainage (6 percent) with instream woody debris densities of 0.51/100 feet. Riffle to pool ratio was 1.2:1 in 1993 and 2.6:1 in 1994. Mean residual pool depth was 0.5 meters for both 1993 and 1994. Channel stability was rated as fair (Table 3.35). Upper bank slopes ranged from 2-35 percent with no evidence of mass wasting problems. Vegetative bank protection was provided by both deep and shallow rooted species. Vegetative cover, however, was largely absent from outside meander bends, leaving these areas more susceptible to erosive action. Debris jam material consisted of small woody debris. The channel capacity was ample for existing peak flows. Significant amounts of bank cutting in this reach provided material for new bar development. There was a fair amount of algae in this reach with some areas of moss growth noted. Mean RASI values were 80 and 71 in 1993 and 1994, respectively. Fines (< 4 mm) comprised as much as 25.6 percent (mean = 20%) of the sampled substrate in 1993 and 22.3 percent (mean= 18%) in 1994 (Table 3.35).

Reach five of Alder Creek was classified as a C6 channel type with a channel bottom comprised of silt/clay and a mean stream gradient of less than 1 percent (Table 3.36). The reach extended for 1,763 meters and encompassed elevations ranging from 2,900 to 2,940 (Figure 2.4). Dominant land uses were livestock grazing and timber harvest. The riparian area consisted of mature deciduous and coniferous trees interspersed with shrubs, forbes and grasses. Average canopy cover was 18 percent for this reach in 1993 and in 1994 averaged 44 percent. Instream woody debris densities were higher than other reaches in the drainage (1.1/100 ft (1993) & 1.6/100 ft (1994)). Riffle to pool ratio was 1 :1 and mean residual pool depth was 0.5 meters in 1993. In 1994, the riffle to pool ratio increased to 6.8:1 with a corresponding increase in residual pool depth 0.7 meters (Table 3.36). This reach had a channel stability rating of poor (Table 3.36). Upper bank slopes ranged from 1-3 percent with no evidence of mass wasting associated with the floodplain. Vegetative density ranged from

**Table 3.35. Summary table showing values for reach four, Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 4th   | 4th  |
| Elevation                | 2,700-2,900 ft  | 2,700-2,900 ft   |
| Dominant land use        | Livestock grazing   | Livestock grazing  |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees, shrubs,<br>forbs and grasses | Coniferous/deciduous<br>mature trees, shrub<br>forbs and grasses |
| Reach length             | 5,053 m   | 5,806 m  |
| Channel type             | c 3   | c 3  |
| Bankfull width           | ft  | •  |
| Bankfull depth           | ft  | •  |
| Width/depth ratio        | > 12  | •  |
| Mean stream gradient     | 1.5 %   | 2.6 %  |
| Entrenchment             | slight  | slight   |
| Sinuosity                | high  | high   |
| Channel stability rating | fair (79)   | fair (90)  |
| RASI values              | 83; 67; 91  | 69; 88; 56   |
| % substrate < 4mm        | 20  | 18.0   |
| mean substrate size (mm) | 91  | 1 2 4  |
| Mean % canopy cover      | 6.2 %   | 46.6 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 77 (0.5 /100 ft)  | 116 (0.6 /100 ft)  |
| Dominant habitat complex | riffle : pool   | riffle : pool  |
| Riffle/pool ratio        | 1.2: 1  | 2.6 : 1  |
| Residual pool depth      | 0.5 m   | 0.5 m  |
| Total available habitat  | 96,846 m <sup>2</sup>   | 80,418 m <sup>2</sup>  |

**Table 3.36. Summary table showing values for reach five, Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>   |
|--------------------------|---|---|
| Stream order             | 4th   | 4th   |
| Elevation                | 2, 900-2,940 ft   | 2,900-2,940 ft  |
| Dominant land use        | Forested/Livestock grazing  | Forested/Livestock grazing  |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees, shrubs,<br>forbs and grasses | Coniferous/deciduous mix<br>with mature trees, shrubs,<br>forbs and grasses |
| Reach length             | 1,763 m   | 1,465 m   |
| Channel type             | C6  | C6  |
| Bankfull width           | ft  | *   |
| Bankfull depth           | ft  | •   |
| Width/depth ratio        | 12.5  | •   |
| Mean stream gradient     | 1.4 %   | 3.6 %   |
| Entrenchment             | slight  | slight  |
| Sinuosity                | high  | high  |
| Channel stability rating | poor (128)  | poor (118)  |
| RASI values              | NA  | NA  |
| % substrate < 4mm        | NA  | NA  |
| mean substrate size (mm) | NA  | NA  |
| Mean % canopy cover      | 17.7 %  | 43.7 %  |
| Barriers                 | none  | none  |
| # instream woody debris  | 65 (1.1 /100 ft)  | 76 (1.6 /100 ft)  |
| Dominant habitat complex | rifle - pool  | rifle - pool  |
| Riffle/pool ratio        | 1 : 1   | 6.8 : 1   |
| Residual pool depth      | 0.5 m   | 0.7 m   |
| Total available habitat  | 37,282 m <sup>2</sup>   | 45,789 m <sup>2</sup>   |

50-70 percent with a shallow and discontinuous root mass. Due to increased beaver activity in this reach there was increased amounts of debris jam material. There were occasional overbank flows attributable to beaver activity and cattle use. Continuous bank cutting and erosion was observed and accompanied by accelerated bar development. Most of the channel bottom is in a yearly state of flux with minimal aquatic vegetation present.

Reach six of Alder Creek was classified as a E3b type channel (Table 3.37) with elevation beginning at 2,940 feet and rising to 3,330 feet. The reach extended for 6,630 meters (Figure 2.4). The dominant land use in this reach was livestock grazing and timber harvest. The riparian area consisted of mature coniferous and deciduous trees mixed with grasses and forbs. Average canopy cover in this reach was 25 percent with instream woody debris densities of 0.6/100 feet during 1993. In 1994, average stream canopy cover was 68 percent with instream woody debris densities of 1.5/100 ft. Beaver ponds and associated downstream riffles comprised the dominant habitat complex. Riffle to pool ratio was 1:1.7 and mean residual pool depths were 0.6 meters in 1993. In 1994, riffle to pool ratio was 2.4:1 with a residual pool depth of 0.4 meters (Table 3.37). The reach had a channel stability rating of fair (Table 3.37). Upper bank slopes averaged 2 percent with no mass wasting problems evident. Debris jam potential was limited to small woody debris. Vegetative cover was provided by shallow rooted species with a density of 50-70 percent. The channel capacity was adequate for peak flows. Bank cutting occurred at obstructions and outcurves. There was a moderate amount of new bar development, consisting of sand and pea sized gravel particles. RASI values could not be calculated in this reach due to the sand bottom.

Reach 1 of the North Fork Alder Creek was classified as an E3b type channel with elevations beginning at 2,940 feet and rising to 3,070 feet (Figure 2.4). Reach length was 1,260 meters. The channel was very sinuous ( $>1.5$ ) and slightly entrenched ( $>2.4$ ) and had a mean stream gradient of 3 percent (Table 3.38). The dominant land use in this reach was livestock grazing with the riparian area consisting of forbs and grasses. Average canopy cover was 36 percent with instream woody debris densities of 0.6/100 feet. Dominant habitat complex was riffle/pool with a riffle to pool ratio of 1.3:1 during 1993. In 1994, the riffle to pool ratio shifted to 16:1. Mean residual pool depth was 0.4 meters in 1993 and decreased to 0.2 meters in 1994 (Table 3.38). Channel stability in this reach was good to fair (Table 3.38). The upper slopes averaged 3 percent with no mass wasting problems evident. Bank protection consisted of discontinuous and shallow root masses with a density of 70-90 percent. The channel capacity was adequate for existing peak flows. Some bank cutting and sloughing was observed at obstructions and outcurves. Little new bar development was observed in this reach. Rocks were slick throughout the reach from seasonal blooms of algae. Mean RASI values were 65 and 79 in 1993 and 1994,

**Table 3.37. Summary table showing values for reach six, Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>   | <b>1994</b>  |
|--------------------------|---|--|
| Stream order             | 2nd   | 2nd  |
| Elevation                | 2,940-3,330 ft  | 2,940-3,330 ft   |
| Dominant land use        | Forested/Livestock grazing  | Forested/Livestock grazing                                     |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees, forbs<br>and grasses | Coniferous/deciduous mix<br>mature trees, forbs<br>and grasses |
| Reach length             | 6,630 m   | 6,278 m  |
| Channel type             | E3b   | E3b  |
| Bankfull width           | ft  | *  |
| Bankfull depth           | ft  | *  |
| Width/depth ratio        | 4.4   | •  |
| Mean stream gradient     | 1 %   | 1.8%   |
| Entrenchment             | slight  | slight   |
| Sinuosity                | very high   | very high  |
| Channel stability rating | fair (103)  | fair (80)  |
| RASI values              | NA  | NA   |
| % substrate < 4mm        | NA  | NA   |
| mean substrate size (mm) | NA  | NA   |
| Mean % canopy cover      | 25 %  | 68 %   |
| Barriers                 | none  | none   |
| # instream woody debris  | 127 (0.6 /100 ft)   | 307 (1.5 /100 ft)  |
| Dominant habitat complex | beaver ponds; riffle  | riffle; pool   |
| Riffle/pool ratio        | 1 : 1.7   | 2.4 : 1  |
| Residual pool depth      | 0.6 m   | 0.4 m  |
| Total available habitat  | 161,234 m <sup>2</sup>  | 73,119 m <sup>2</sup>  |

Table 3.38. Summary table showing values for reach one, North Fork Alder Creek, 1993 and 1994

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b>  |
|--------------------------|--|--|
| Stream order             | 2nd  | 2nd  |
| Elevation                | 2,940-3,070 ft                                       | 2,940-3,070 ft                                     |
| Dominant land use        | Livestock grazing                                    | Livestock grazing                                  |
| Riparian ecotype         | Coniferous / deciduous mix<br>with forbs and grasses | Coniferous/Deciduous mix<br>with forbs and grasses |
| Reach length             | 1,260 m  | 1,677 m  |
| Channel type             | E3b  | E3b  |
| Bankfull width           | ft   | •  |
| Bankfull depth           | ft   | •  |
| Width/depth ratio        | 4..6   | •  |
| Mean stream gradient     | 3 %  | 3.1 %  |
| Entrenchment             | slight   | slight   |
| Sinuosity                | very high  | very high  |
| Channel stability rating | good (73)  | fair (91)  |
| RASI values              | 55; 79; 51   | 64; 43   |
| % substrate < 4mm        | 22 %   | 29.9 %   |
| mean substrate size (mm) | 53   | 103  |
| Mean % canopy cover      | 36.3 %   | 8%   |
| Barriers                 | none   | none   |
| # instream woody debris  | 24 (0.6 /100 ft)                                     | 22 (0.4 /100 ft)                                   |
| Dominant habitat complex | riffle - pool  | riffle - pool                                      |
| Riffle/pool ratio        | 1.3 :1   | 16: 1  |
| Residual pool depth      | 0.4 m  | 0.2 m  |
| Total available habitat  | 11,206 m <sup>2</sup>                                | 5,651 m <sup>2</sup>                               |

respectively. Fine sediment (< 4 mm) comprised 22 percent of the sampled substrate in 1993, and 30 percent in 1994 (Table 3.38).

Reach two (comprising stream reaches 2a and 3a) of North Fork Alder Creek were classified as B3a type channels with isolated areas having characteristics of D3b type channels (Table 3.39). Reach length was 3,461 meters with elevations ranging from 3,070 to 3,280 feet (Figure 2.4). The dominant land type in this reach was forested wetland with the riparian area composed of mature coniferous and deciduous trees interspersed with large shrubs. Average canopy cover in this reach was 30 percent with instream woody debris densities of 0.4/100 feet during 1993. In 1994, average canopy cover was 87 percent with instream woody debris loads of 4.8/100 feet. Dominant habitat complex was beaver pond/riffle with a riffle to pool ratio of 1:1.6. Mean residual pool depth was 0.4 meters (Table 3.39). The channel stability in these reaches was fair (Table 3.39). The upper bank slopes averaged 5 percent with no mass wasting problems evident. Plant density was approximately 70 percent with some deep rooted species interspersed throughout the reach. Channel capacity was adequate for existing peak flows, but due to beaver activity there was evidence of some overbank flows. Frequent obstructions created sediment traps throughout the reach and bank cutting was significant. Some new gravel bar development had occurred with slight amounts of fines being deposited. Clinging aquatic vegetation was common even in swift waters. One RASI survey site was sampled in reach 2a in 1994. The RASI value was 69.3 and fine sediment (< 4 mm) comprised 36 percent of the sampled substrate (Table 3.39). Mean RASI values for reach 3a were 71 and 52 in 1993 and 1994, respectively. Fine sediment in reach 3a comprised between 26 and 30 percent of the sampled substrate on average.

Reach three of North Fork Alder Creek was classified as a B3a type channel with elevations beginning at 3,280 feet and rising to 3,620 feet. Reach length was 3,130 meters (Figure 2.4). The dominant land type within this reach was forested with the riparian area consisting of young and mature coniferous and deciduous trees interspersed with shrubs. Average canopy cover was 39 percent with instream woody debris densities of 0.4/100 feet. Dominant habitat complex was riffle/pool with a riffle to pool ratio of 2.3:1. Mean residual pool depth was 0.2 meters. Channel stability was rated as fair to good (Table 3.40). Upper bank slopes ranged from 2-30 percent with no evidence of mass wasting problems. Plant density in this reach was at least 90 percent with a vigorous community of both shallow and deep rooted species. There were large amounts of debris jam material located throughout the reach. Channel capacity was adequate for existing peak flows. Bank cutting occurred at outcurves with little new bar development observed. Clinging aquatic vegetation was abundant throughout the reach with both algae and moss present. The mean RASI value

**Table 3.39. Summary table showing values for reach two, North Fork Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b>  |
|--------------------------|--|--|
| Stream order             | 2nd  | 2nd  |
| Elevation                | 3,070-3,280 ft   | 3,070-3,280 ft   |
| Dominant land use        | Forested   | Forested   |
| Riparian ecotype         | Coniferous/deciduous mix<br>with mature trees & shrubs | Coniferous/deciduous mix<br>with mature trees & shrubs |
| Reach length             | 3,461 m  | 4,250 m  |
| Channel type             | B3a - D3b  | B3a - D3b  |
| Bankfull width           | ft   | *  |
| Bankfull depth           | ft   | •  |
| Width/depth ratio        | > 40 , 5.6   | *  |
| Mean stream gradient     | 2.5 %  | 8.7 %  |
| Entrenchment             | moderate   | moderate   |
| Sinuosity                | moderate   | moderate   |
| Channel stability rating | fair (78)  | fair (86)  |
| RASI values              | 66; 77   | 46; 66; 46   |
| % substrate < 4mm        | 29.5 %   | 25.7 %   |
| mean substrate size (mm) | 46   | 47   |
| Mean % canopy cover      | 35.2 %   | 87.2 %   |
| Barriers                 | none   | none (gradient)  |
| # instream woody debris  | 46 (0.4 /100 ft)                                       | 663 (4.8 /100 ft)                                      |
| Dominant habitat complex | beaver pond - riffle                                   | riffle - beaver pond                                   |
| Riffle/pool ratio        | 1 : 1.6  | 1 : 1  |
| Residual pool depth      | 0.4 m  | 0.5 m  |
| Total available habitat  | 99,826 m <sup>2</sup>                                  | 69.288 ft <sup>2</sup>                                 |



**Table 3.40. Summary table showing values for reach three, North Fork Alder Creek, 1993 and 1994**

| <b>PARAMETER</b>         | <b>1993</b>  | <b>1994</b> |
|--------------------------|--|-------------|
| Stream order             | 2nd  |             |
| Elevation                | 3,280-3,620 ft   |             |
| Dominant land use        | Forested   |             |
| Riparian ecotype         | Coniferous/deciduous mix with<br>mature, young trees, & shrubs |             |
| Reach length             | 3,130 m  |             |
| Channel type             | B3a  |             |
| Bankfull width           | ft   | •           |
| Bankfull depth           | ft   | •           |
| Width/depth ratio        | 6  | *           |
| Mean stream gradient     | 5 %  |             |
| Entrenchment             | moderate   |             |
| Sinuosity                | moderate   |             |
| Channel stability rating | 0  |             |
| RASI values              | 45; 40; 38   | *           |
| % substrate < 4mm        | 19.6   | *           |
| mean substrate size (mm) | 39   | •           |
| Mean % canopy cover      | 39 %   | •           |
| Barriers                 | none   |             |
| # instream woody debris  | 476 (0.4 /100 ft)  |             |
| Dominant habitat complex | riffle - pool  |             |
| Riffle/pool ratio        | 2.3 : 1  |             |
| Residual pool depth      | 0.2 m  |             |
| Total available habitat  | 4,682 m <sup>2</sup>   |             |

was 41 in 1993 and fine sediment (< 4 mm) comprised between 15 and 22 percent (mean=20 percent) of the sampled substrate (Table 3.40).

## 3.2. BIOLOGICAL EVALUATION

### 3.2.1. Relative Abundance

#### 3.2.1.1. 1993 Lake, Benewah, Evans, and Alder Creeks

In May, June, July, August, and October of 1993, a total of 15.8 electroshocking hours were spent collecting relative abundance information. A total of 1,625 fish were collected from Lake, Benewah, Evans, and Alder creeks (Table 3.41). In Lake Creek, a total of 257 fish were captured with sculpin *spp.* being the most abundant at 40.1 percent. In Benewah Creek, 700 fish were captured with reidside shiners comprising 45.9 percent of the total catch. In Evans Creek, a total of 386 fish were captured with cutthroat trout the most abundant at 53.4 percent. A total of 293 fish were captured in Alder Creek, with sculpin *spp.* being the most abundant at 34.5 percent. Cutthroat trout densities were highest in Evans Creek at 53.4 percent followed by, Alder (22.5 percent), Lake (16.0 percent) and Benewah creeks (11 .0 percent) (Table 3.42). Relative abundance data for each month, reach, and stream can be found in Appendix E.

#### Lake Creek

In spring, summer, and fall a total of three, one hundred seventy-two, and seventy fish were collected in Lake Creek, respectively (Table 3.43). No trout species were collected in the spring sample (Table 3.43). Of the 172 fish captured in the summer, thirty (17.4 percent) were cutthroat trout, with the remaining 82 percent being comprised of sculpin *spp.*, longnose dace, longnose sucker, and reidside shiner (Table 3.44). Of the thirty cutthroat trout captured in the summer, two were 0+ of age, three were 1+ and 2+ years of age, respectively. Ten were 3+ years of age, ten were 4+ years of age and two were 5+ years of age (Table 3.45). During Fall, eleven (15.7 percent) were cutthroat trout and one (1.4 percent) was a bull trout. The remaining 82.8 percent were comprised of sculpin *spp.*, and longnose suckers. Of the eleven cutthroat captured in the fall, two were 1+ years of age, four were 2+ years of age, two cutthroat trout were 3+ years of age, respectively. Two cutthroat trout 4+ years of age and one was 5+ years of age (Table 3.45).

#### Benewah Creek

In spring, summer and fall a total of two hundred eighty-six, three hundred thirty-one, and eighty-five fish were collected from Benewah Creek, respectively (Table 3.46). Of the 286 fish collected in spring, thirty-six (12.6 percent) were

**Table 3.41. Number of each species of fish caught by electrofishing at each Coeur d'Alene tributary during 1993.**

| <b>Tributary</b>          | <b>Lake</b> | <b>Benewah</b> | <b>Evans</b> | <b>Alder</b> |
|---------------------------|-------------|----------------|--------------|--------------|
| <b>Shock time (hours)</b> | <b>2.7</b>  | <b>5.3</b>     | <b>3.9</b>   | <b>3.9</b>   |
| <b>Species</b>            |             |                |              |              |
| Cutthroat trout           | 41          | 77             | 206          | 66           |
| Eastern brook trout       |             | 1              |              | 87           |
| Bull trout                | 1           |                |              |              |
| Rainbow trout             |             | 2              |              |              |
| Sculpin spp.              | 103         | 49             | 176          | 101          |
| Largemouth bass           |             | 2              |              |              |
| Longnosesucker            | 48          | 128            | 3            | 32           |
| Longnosedace              | 51          | 73             |              | 4            |
| Western speckled dace     | 1           | 36             |              |              |
| Brown bullhead            |             | 2              |              | 3            |
| Redside shiner            | 12          | 321            |              |              |
| Northern squawfish        |             | 8              |              |              |
| Fumpkinseed               |             | 1              | 1            |              |
| <b>TOTAL</b>              | <b>257</b>  | <b>700</b>     | <b>386</b>   | <b>293</b>   |

**Table 3.42. Percent (%) of each species of fish caught by electrofishing at each Coeur d'Alene tributary during 1993.**

| <b>Species</b>        | <b>Lake</b> | <b>Benewah</b> | <b>Evans</b> | <b>Alder</b> |
|-----------------------|-------------|----------------|--------------|--------------|
| Cutthroat trout       | 16.0        | 11.0           | 53.4         | 22.5         |
| Eastern brook trout   |             | 0.1            |              | 29.7         |
| Bull trout            | 0.4         |                |              |              |
| Rainbow trout         |             | 0.3            |              |              |
| Sculpin spp.          | 40.1        | 7.0            | 45.6         | 34.5         |
| Largemouth bass       |             | 0.3            |              |              |
| Longnose sucker       | 18.7        | 18.3           | 0.8          | 10.9         |
| Lonanose dace         | 19.8        | 10.4           |              | 1.4          |
| Western speckled dace | 0.4         | 5.1            |              |              |
| Brown bullhead        |             | 0.3            |              | 1.0          |
| Redside shiner        | 4.7         | 45.9           |              |              |
| Northern squawfish    |             | 1.1            |              |              |
| Pumpkinseed           | 0.1         |                | 0.3          |              |

**Table 3.43. Number of each species of fish caught by electrofishing in Lake Creek during May - October, 1993.**

| <b>Species</b>          | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>September<br/>(October)</b> |
|-------------------------|------------------------------|---------------------------------|--------------------------------|
| <b>Shock time (min)</b> | <b>12.4</b>                  | <b>88.4</b>                     | <b>58.6</b>                    |
| Cutthroat trout         |                              | 30                              | 11                             |
| Bull trout              |                              |                                 | 1                              |
| Sculpin spp.            | 2                            | 58                              | 43                             |
| Longnose sucker         |                              | 33                              | 15                             |
| Longnosedace            |                              | 39                              |                                |
| Western speckled dace   | 1                            |                                 |                                |
| Redside shiner          |                              | 12                              |                                |
| <b>TOTAL</b>            | <b>3</b>                     | <b>172</b>                      | <b>70</b>                      |

**Table 3.44. Percent (%) of each species of fish caught by electrofishing in Lake Creek during May - October, 1993.**

| <b>Species</b>        | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|-----------------------|------------------------------|---------------------------------|---------------------------|
| Cutthroat trout       |                              | 17.4                            | 15.7                      |
| Bull trout            |                              |                                 | 1.4                       |
| Largemouth bass       |                              |                                 |                           |
| Sculpin spp.          | 66.7                         | 33.7                            | 61.4                      |
| Longnose sucker       |                              | 19.2                            | 21.4                      |
| Longnose dace         |                              | 22.7                            |                           |
| Western speckled dace | 33.3                         |                                 |                           |
| Redside shiner        |                              | 7.0                             |                           |

**Table 3.45. Electrofishing relative abundance for salmonid species by age class in Lake Creek, 1993.**

| <b>Cutthroat trout</b> |               |               |             |
|------------------------|---------------|---------------|-------------|
| <b>Age</b>             | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 0+                     | 0             | 2             | 0           |
| 1+                     | 0             | 3             | 2           |
| 2+                     | 0             | 3             | 4           |
| 3+                     | 0             | 10            | 2           |
| 4+                     | 0             | 10            | 2           |
| 5+                     | 2             | 2             | 1           |
| 6+                     | 2             | 30            | 11          |

**Table 3.46. Number of each species of fish caught by electrofishing in Benewah Creek during May - October, 1993.**

| <b>Species</b>          | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|-------------------------|------------------------------|---------------------------------|---------------------------|
| <b>Shock time (min)</b> | <b>129.7</b>                 | <b>124.3</b>                    | <b>NA</b>                 |
| Cutthroat trout         | 36                           | 22                              | 19                        |
| Rainbow trout           |                              |                                 | 2                         |
| Eastern brook trout     |                              |                                 | 1                         |
| Sculpin <i>spp.</i>     | 9                            | 18                              | 22                        |
| Long nose sucker        | 48                           | 60                              | 20                        |
| Longnosedace            | 12                           | 61                              |                           |
| Western speckled dace   | 16                           | 20                              |                           |
| Northern sauawfish      | 6                            | 2                               |                           |
| Largemouth bass         |                              |                                 | 2                         |
| Redside shiner          | 158                          | 144                             | 19                        |
| Pumpkinseed             |                              | 3                               |                           |
| Brown bullhead          | 1                            | 1                               |                           |
| <b>TOTAL</b>            | <b>286</b>                   | <b>331</b>                      | <b>85</b>                 |

cutthroat trout, one hundred fifty-eight (55.2 percent) were redbase shiners, and forty-eight (16.8 percent) were longnose sucker. The other 15.3 percent consisted of western speckled dace, longnose dace, sculpin *spp*, northern squawfish and brown bullhead (Table 3.47). Of the thirty-six cutthroat captured, four were 0+ years of age, eleven were 1+ years of age, nine were 2+ years of age, six were 3+ years of age, two were 4+ years of age, one was 5+ years of age, and four were 6+ years of age. (Table 3.48). Of the 331 fish captured in summer 22 (6.6 percent) were cutthroat trout, 144 (43.5 percent) were redbase shiner, sixty-one (18.4 percent) were longnose dace, sixty (18.1 percent) were longnose sucker. The other 7.2 percent consisted of sculpin *spp*, brown bullhead, northern squawfish and pumpkinseed. Of the 22 cutthroat captured; one was 0+ and 1+ years of age, respectively, nine were 2+ years of age, five were 3+ years of age, four were 4+ years of age, one was 5+ years of age and one was 6+ years of age. Of the 85 fish collected from Benewah Creek in the fall, nineteen (22.4 percent) were cutthroat trout, twenty-two (25.9 percent) were sculpin *spp*, twenty (23.5 percent) were longnose sucker, nineteen (22.4 percent) were redbase shiner. The other six percent consisted of largemouth bass, rainbow trout, and eastern brook trout. Of the nineteen cutthroat trout captured in the fall one was 2+ years of age, eight were 3+ years of age, seven were 4+ years of age, two were 5+ years of age, and one was 6+ years of age (Table 3.48).

**Table 3.47. Percent (%) of each species of fish caught by electrofishing in Benawah Creek during May - October, 1993.**

| <b>Species</b>        | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|-----------------------|------------------------------|---------------------------------|---------------------------|
| Cutthroat trout       | 12.6                         | 6.6                             | 22.4                      |
| Rainbow trout         |                              |                                 | 2.4                       |
| Eastern brook trout   |                              |                                 | 1.2                       |
| Sculpin <i>spp.</i>   | 3.1                          | 5.4                             | 25.9                      |
| Longnose sucker       | 16.8                         | 18.1                            | 23.5                      |
| Longnosedace          | 4.2                          | 18.4                            |                           |
| Western speckled dace | 5.6                          | 6.0                             |                           |
| Northern squawfish    | 2.1                          | 0.6                             |                           |
| Largemouth bass       |                              |                                 | 2.4                       |
| Redside shiner        | 55.2                         | 43.5                            | 22.4                      |
| Pumpkinseed           |                              | 0.3                             |                           |
| Brown bullhead        | 0.3                          | 0.9                             |                           |

**Table 3.48. Electrofishing relative abundance for salmonid species by age class in Benawah Creek, 1993.**

| <b>Cutthroat trout</b> |               |               |             |
|------------------------|---------------|---------------|-------------|
| <b>Age</b>             | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 0+                     | 4             | 1             | 0           |
| 1+                     | 11            | 1             | 0           |
| 2+                     | 9             | 9             | 1           |
| 3+                     | 6             | 5             | 8           |
| 4+                     | 2             | 4             | 7           |
| 5+                     | 1             | 1             | 2           |
| 6+                     | 4             | 1             | 1           |

### **Evans Creek**

A total of seven, two hundred sixty-two, and one hundred sixteen fish were captured in Evans Creek during spring, summer, and fall, respectively (Table 3.49). Twenty-nine percent of the fish caught in the spring were cutthroat trout and the remaining 71 percent were sculpin *spp.* (Table 3.50). Of the 2 cutthroat trout captured in spring, one was 1+ years of age and one was 2+ years of age (Table 3.51) In summer, 135 (51.5 percent) of the fish collected were cutthroat trout, with the remainder being sculpin species, longnose sucker and pumpkinseed. Of the one hundred thirty-five cutthroat, nine were 0+ years of age, thirty were 1+ years of age, 30 were 2+ years of age, twenty-six were 3+

years of age, twenty-seven were 4+ years of age, seven were 5+ years of age and six were 6+ years of age (Table 3.51). Of the one hundred sixteen fish collected from Evans Creek in the fall sample, 69 (59.5 percent) were cutthroat trout, and 47 (40.5 percent) were sculpin species (Table 3.50). Of the sixty-nine fish captured in Evans during the fall, six were 1+ years of age, twenty-three were 2+ years of age, eighteen were 3+ years of age, twelve were 4+ years of age, five were 5+ years of age, and five were 6+ years of age (Table 3.51).

**Table 3.49. Number of each species of fish caught by electrofishing in Evans Creek during May - October, 1993.**

| Species             | Spring<br>(May-June) | Summer<br>(July-August) | September<br>(October) |
|---------------------|----------------------|-------------------------|------------------------|
| Shock time (min)    | 49.4                 | 1 1 7 . 2               | 74.0                   |
| Cutthroat trout     | 2                    | 135                     | 69                     |
| Sculpin <i>spp.</i> | 5                    | 124                     | 47                     |
| Longnose sucker     |                      | 3                       |                        |
| Pumpkinseed         |                      | 1                       |                        |
| <b>TOTAL</b>        | <b>7</b>             | <b>262</b>              | <b>176</b>             |

**Table 3.50. Percent (%) of each species of fish caught by electrofishing in Evans Creek during May - October, 1993.**

| Species             | Spring<br>(May-June) | Summer<br>(July-August) | Fall<br>(October) |
|---------------------|----------------------|-------------------------|-------------------|
| Cutthroat trout     | 28.6                 | 51.5                    | 59.5              |
| Sculpin <i>spp.</i> | 71.4                 | 47.3                    | 40.5              |
| Longnose sucker     |                      | 1.1                     |                   |
| Pumpkinseed         |                      | 0.4                     |                   |

### **Alder Creek**

A total of forty-three, one hundred eighty-nine, and sixty-one fish were captured in Alder Creek in spring, summer and fall, respectively (Table 3.52). Of the forty-three fish collected in the spring sample, 25.6 percent were eastern brook trout, 18.6 percent were cutthroat trout and the remaining 55.8 percent were sculpin species, longnose sucker, brown bullhead, and longnose dace (Table 3.53). Of the eight cutthroat trout captured in Alder Creek during spring, five were 2+ years of age, one was 3+ years of age, one was 4+ years of age and one was 5+ years of age (Table 3.54). Of the one hundred eighty-nine fish

**Table 3.51. Electrofishing relative abundance for salmonid species by age class in Evans Creek, 1993.**

| Age | Cutthroat trout |        |      |
|-----|-----------------|--------|------|
|     | Spring          | Summer | Fall |
| 0+  | 0               | 9      | 0    |
| 1+  | 1               | 30     | 6    |
| 2+  | 1               | 30     | 23   |
| 3+  | 0               | 26     | 18   |
| 4   | 0               | 27     | 12   |
| 5+  | 0               | 7      | 5    |
| 6+  | 0               | 6      | 5    |

captured from Alder Creek in July, 22.2 percent were cutthroat trout, 28.0 percent were eastern brook trout and the remaining 49.8 percent were sculpin species, longnose sucker, and longnose dace. Of the forty-two cutthroat trout captured in the summer, two were 1+ years of age, seven were 2+ years of age, thirteen were 3+ years of age, ten were 4+ years of age, eight were 5+ years of age, and two were 6+ years of age (Table 3.54). In the fall sample, sixty-one fish were collected in Alder Creek. Of those sixty-one fish, twenty-three (37.7 percent) were eastern brook trout, sixteen (26.2 percent) were cutthroat trout and sculpin species, respectively, and six (9.8 percent) were longnose sucker (Table 3.53). Of the sixteen cutthroat trout captured, one was 1+ years of age, one was 2+ years of age, five were 3+ years of age, four were 4+ years of age and six were 5+ years of age (Table 3.54).

**Table 3.52. Number of each species of fish caught by electrofishing in Alder Creek during May - October, 1993.**

| Species                 | Spring<br>(May-June) | Summer<br>(July-August) | September<br>(October) |
|-------------------------|----------------------|-------------------------|------------------------|
| <b>Shock time (min)</b> | <b>54.6</b>          | <b>120.6</b>            | <b>58.0</b>            |
| Cutthroat trout         | 8                    | 42                      | 16                     |
| Eastern brook trout     | 11                   | 53                      | 23                     |
| Sculpin spp.            | 14                   | 71                      | 16                     |
| Longnose sucker         | 5                    | 21                      | 6                      |
| Redside shiner          |                      |                         |                        |
| Longnosedace            | 2                    | 2                       |                        |
| Brown bullhead          | 3                    |                         |                        |
| <b>TOTAL</b>            | <b>43</b>            | <b>189</b>              | <b>61</b>              |



**Table 3.53. Percent (%) of each species of fish caught by electrofishing in Alder Creek during May - October, 1993.**

| <b>Species</b>      | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|---------------------|------------------------------|---------------------------------|---------------------------|
| Cutthroat trout     | 18.6                         | 22.2                            | 26.2                      |
| Eastern brook trout | 25.6                         | 28.0                            | 37.7                      |
| Sculpin <i>spp.</i> | 32.6                         | 37.6                            | 26.2                      |
| Longnose sucker     | 11.6                         | 11.1                            | 9.8                       |
| Longnosedace        | 4.7                          | 1.1                             |                           |
| Brown bullhead      | 7.0                          |                                 |                           |

**Table 3.54. Electrofishing relative abundance for salmonid species by age class in Alder Creek, 1993.**

| <b>Cutthroat trout</b> |               |               |             |
|------------------------|---------------|---------------|-------------|
| <b>Age</b>             | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 0+                     | 0             | 0             | 0           |
| 1+                     | 0             | 2             | 1           |
| 2+                     | 5             | 7             | 1           |
| 3+                     | 1             | 13            | 5           |
| 4+                     | 1             | 10            | 4           |
| 5+                     | 1             | 8             | 6           |
| 6+                     | 0             | 2             | 0           |

### 3.2.1.2. 1994 Lake, Benewah, Evans, and Alder Creeks

In May, June, July, August, and October of 1994, a total of 6.7 electroshocking hours were spent collecting relative abundance information. A total of 1,370 fish were collected from the four tributaries (Table 3.53). In Lake Creek, a total of two hundred twenty-four fish were captured with sculpin *spp* being the most abundant at 37.1 percent. In Benewah Creek, six hundred fifty-nine fish were captured with reidside shiners comprising 72.2 percent of the total catch. In Evans Creek, a total of one hundred seventy-eight fish were captured with sculpin *spp.* the most abundant at 47.8 percent. A total of three hundred nine fish were captured in Alder Creek, with eastern brook trout the most abundant at 42.1 percent. Cutthroat trout densities were highest in Evans Creek at 43.3 percent followed by Alder (23.3 percent), Lake (15.2 percent) and Benewah (2.8 percent). Relative abundance data for each month, reach and stream can be found in Appendix E.

**Table 3.55. Number of each species of fish caught by electrofishing at each Coeur d'Alene tributary during 1994.**

| <b>Species</b>            | <b>Lake</b> | <b>Benewah</b> | <b>Evans</b> | <b>Alder</b> |
|---------------------------|-------------|----------------|--------------|--------------|
| <b>Shock time (hours)</b> | <b>1.54</b> | <b>1.04</b>    | <b>1.60</b>  | <b>2.54</b>  |
| Cutthroat trout           | 34          | 18             | 77           | 72           |
| Eastern brook trout       |             |                |              | 130          |
| Sculpin spp.              | 83          | 71             | 85           | 87           |
| Largemouth bass           |             |                | 5            |              |
| Longnose sucker           | 54          | 87             | 4            | 20           |
| Western speckled dace     |             | 4              |              |              |
| Brown bullhead            |             |                | 1            |              |
| Redside shiner            | 49          | 459            | 6            |              |
| Northern squawfish        |             |                |              |              |
| <b>TOTAL</b>              | <b>224</b>  | <b>659</b>     | <b>178</b>   | <b>309</b>   |

**Table 3.56. Percent of each species of fish caught by electrofishing at each Coeur d'Alene tributary during 1994**

| <b>Species</b>        | <b>Lake</b> | <b>Benewah</b> | <b>Evans</b> | <b>Alder</b> |
|-----------------------|-------------|----------------|--------------|--------------|
| Cutthroat trout       | 15.2        | 2.8            | 43.3         | 23.3         |
| Eastern brook trout   |             |                |              | 42.1         |
| Sculpin spp.          | 37.1        | 11.2           | 47.8         | 28.2         |
| Largemouth bass       |             |                | 2.8          |              |
| Longnose sucker       | 24.1        | 13.7           | 2.2          | 6.5          |
| Western speckled dace |             | 1.8            |              |              |
| Brown bullhead        |             |                | 0.6          |              |
| Redside shiner        | 21.9        | 72.2           | 3.4          |              |
| Northern squawfish    |             | 0.2            |              |              |

### Lake Creek

In spring, summer, and fall a total of fifty-three, fifty-five, and one hundred sixteen fish were collected in Lake Creek, respectively (Table 3.55). Cutthroat trout comprised 15 percent of the spring sample. The remaining 85 percent of the catch was comprised of sculpin *spp*, longnose sucker and redbside shiners. Of the eight cutthroat captured, seven were 2+ fish and one was a 1+ fish (Table 3.57). Of the fifty-five fish collected in the summer sample, 32.7 percent of the

catch were cutthroat trout. The remaining 67.3 percent were comprised of sculpin, longnose sucker, redbase shiner and western speckled dace (Table 3.56). All eighteen trout captured were 0+ of age (Table 3.57). Of the one hundred sixteen fish captured during the fall sample, 6.9 percent were cutthroat trout. Sculpin *spp.*, longnose sucker, and redbase shiner comprised the remaining 93 percent. Of the eight cutthroat captured, five were 0+ of age and three were 1+ of age.

**Table 3.57. Number of each species of fish caught by electrofishing in Lake Creek during May - October, 1994.**

| <b>Species</b>          | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>September<br/>(October)</b> |
|-------------------------|------------------------------|---------------------------------|--------------------------------|
| <b>Shock time (min)</b> | 25.7                         | 29.6                            | 37.2                           |
| Cutthroat trout         | 8                            | 18                              | 8                              |
| Sculpin <i>spp.</i>     | 32                           | 10                              | 41                             |
| Longnose sucker         | 11                           | 14                              | 29                             |
| Redside shiner          | 2                            | 9                               | 38                             |
| Western speckled dace   |                              | 4                               |                                |
| <b>TOTAL</b>            | <b>53</b>                    | <b>55</b>                       | <b>116</b>                     |

**Table 3.58. Percent (%) of each species of fish caught by electrofishing in Lake Creek during May - October, 1994.**

| <b>Species</b>        | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|-----------------------|------------------------------|---------------------------------|---------------------------|
| Cutthroat trout       | 15.1                         | 32.7                            | 6.9                       |
| Sculpin <i>spp.</i>   | 60.4                         | 18.2                            | 35.3                      |
| Longnose sucker       | 20.8                         | 25.4                            | 25.0                      |
| Redside shiner        | 3.8                          | 16.4                            | 32.8                      |
| Western speckled dace |                              | 7.3                             |                           |

**Table 3.59. Electrofishing relative abundance for salmonid species by age class in Lake Creek, 1994.**

| <b>Cutthroat trout</b> |               |               |             |
|------------------------|---------------|---------------|-------------|
| <b>Age</b>             | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 0+                     | 0             | 18            | 5           |
| 1+                     | 1             | 0             | 3           |
| 2+                     | 7             | 0             | 0           |

## Benewah Creek

In Benewah Creek, ninety-five, one hundred fifty-nine, and three hundred eighty-two fish were collected during the spring, summer and fall sample, respectively (Table 3.60). Cutthroat trout comprised 13.7 percent of the spring catch, while the remaining 86 percent was made up of sculpin *spp.*, longnose sucker, northern squawfish and redbside shiner. Of the twelve cutthroat captured, five were 1+ of age, three were 2+ of age, one was 3+ and 5+ of age and two were 6+ of age (Table 3.62). In the summer sample, cutthroat trout comprised 0.6 percent of the catch. Redside shiner, longnose sucker and sculpin *spp.* made up the remainder of the catch. The one cutthroat trout captured in the summer sample was 2+ years of age. In the fall sample, cutthroat trout comprised 1 percent of the catch with the remainder being sculpin *spp.*, longnose sucker and redbside shiner. All four cutthroat trout captured were 2+ years of age (Table 3.62).

**Table 3.60. Number of each species of fish caught by electrofishing in Benewah Creek during May - October, 1994**

| Species                 | Spring<br>(May-June) | Summer<br>(July-August) | Fall<br>(October) |
|-------------------------|----------------------|-------------------------|-------------------|
| <b>Shock time (min)</b> | <b>40.1</b>          | <b>26.3</b>             | <b>62.4</b>       |
| Cutthroat trout         | 13                   | 1                       | 4                 |
| Sculpin <i>spp.</i>     | 17                   | 21                      | 33                |
| Longnose sucker         | 46                   | 6                       | 35                |
| Northern squawfish      | 1                    |                         |                   |
| Redside shiner          | 18                   | 131                     | 310               |
| <b>TOTAL</b>            | <b>95</b>            | <b>159</b>              | <b>382</b>        |

**Table 3.61. Percent of each species of fish caught by electrofishing in Benewah Creek during May - October, 1994.**

| Species             | Spring<br>(May-June) | Summer<br>(July-August) | Fall<br>(October) |
|---------------------|----------------------|-------------------------|-------------------|
| Cutthroat trout     | 13.7                 | 0.6                     | 1.0               |
| Sculpin <i>spp.</i> | 17.9                 | 13.2                    | 8.6               |
| Longnose sucker     | 48.4                 | 3.8                     | 9.2               |
| Northern squawfish  | 1.1                  |                         |                   |
| Redside shiner      | 18.9                 | 82.4                    | 81.2              |

**Table 3.62. Electrofishing relative abundance for salmonid species by age class in Benawah Creek, 1994.**

| Age | Cutthroat trout |        |      |
|-----|-----------------|--------|------|
|     | Spring          | Summer | Fall |
| 1+  | 5               | 0      | 0    |
| 2+  | 3               | 1      | 4    |
| 3+  | 1               | 0      | 0    |
| 5+  | 1               | 0      | 0    |
| 6+  | 2               | 0      | 0    |

### Evans Creek

In Evans Creek, sixty-three, fifty-eight, and fifty-seven fish were collected during the spring, summer and fall sample. Cutthroat trout comprised 50.8 percent of the catch during the spring sample while sculpins comprised the other 49.2 percent (Table 3.63). Of the thirty-two cutthroat trout captured, one was 0+ of age, nine were 1+ of age, fifteen were 2+ of age, six were 3+ of age and one was 4+ of age (Table 3.65). In the summer sample, cutthroat trout comprised 31 percent of the catch with sculpin *spp.*, longnose sucker, redbside shiner and brown bullhead comprising the remainder of the catch. Of the twenty-one cutthroat captured, one was 0+ of age, eight were 1+ of age, seven were 2+ of age and five were 3+ of age. Cutthroat trout comprised 47 percent of the fall catch, while sculpin and largemouth bass comprised the remainder. Of the thirty cutthroat trout collected, three were 0+ years of age, sixteen were 1+ years of age, four were 2+ years of age, six were 3+ years of age, and one was 4+ years of age (Table 3.65).

**Table 3.63. Number of each species of fish caught by electrofishing in Evans Creek during May - October, 1994.**

| Species             | Spring<br>(May-June) | Summer<br>(July-August) | September<br>(October) |
|---------------------|----------------------|-------------------------|------------------------|
| Shock time (min)    | 34.9                 | 22.7                    | 38.6                   |
| Cutthroat trout     | 32                   | 18                      | 27                     |
| Sculpin <i>spp.</i> | 31                   | 29                      | 25                     |
| Longnose sucker     |                      | 4                       |                        |
| Redside shiner      |                      | 6                       |                        |
| Brown bullhead      |                      | 1                       |                        |
| Largemouth bass     |                      |                         | 5                      |
| <b>TOTAL</b>        | <b>63</b>            | <b>58</b>               | <b>57</b>              |

**Table 3.64. Percent (%) of each species of fish caught by electrofishing in Evans Creek during May - October, 1994.**

| Species             | Spring<br>(May-June) | Summer<br>(July-August) | Fall<br>(October) |
|---------------------|----------------------|-------------------------|-------------------|
| Cutthroat trout     | 50.8                 | 31.0                    | 47.4              |
| Sculpin <i>spp.</i> | 49.2                 | 50.0                    | 43.8              |
| Longnose sucker     |                      | 6.9                     |                   |
| Redside shiner      |                      | 10.3                    |                   |
| Brown bullhead      |                      | 1.7                     |                   |
| Largemouth bass     |                      |                         | 8.8               |

**Table 3.65. Electrofishing relative abundance for salmonid species by age class in Evans Creek, 1994.**

| Age | Cutthroat trout |        |      |
|-----|-----------------|--------|------|
|     | Spring          | Summer | Fall |
| 0+  | 1               | 1      | 3    |
| 1+  | 9               | 8      | 16   |
| 2+  | 15              | 7      | 4    |
| 3+  | 6               | 5      | 6    |
| 4+  | 1               | 0      | 1    |

### Alder Creek

During the spring, summer, and fall sampling in Alder Creek, sixty-six, one hundred twenty-five, and one hundred eighteen fish were captured (Table 3.66). Cutthroat trout and eastern brook trout comprised 24.2 percent and 42.4 percent of the catch during the spring sample. Other species captured included sculpin and longnose sucker (Table 3.66). Of the sixteen cutthroat trout collected, one was 1+ years of age, twelve were 2+ years of age, and three were 3+ years of age. Of the twenty-eight eastern brook trout collected, eight were 2+ years of age, twelve were 3+ years of age and eight were 4+ years of age (Table 3.68). In the summer sample cutthroat trout comprised 20.8 percent of the catch and eastern brook trout comprised 48.8 percent of the catch. Sculpin *spp.* and longnose sucker comprised the remainder of the catch (Table 3.66). Of the twenty-six cutthroat trout collected, five were 1+ years of age, twelve were 2+ years of age, seven were 3+ years of age and two were 4+ years of age. Of the sixty-one eastern brook trout collected, eighteen were 1+ years of age, seventeen were 2+ years of age, fourteen were 3+ years of age and twelve were 4+ years of age (Table 3.68). In the fall sample cutthroat and eastern brook trout comprised 60 percent of the catch, while sculpin *spp.* comprised the remainder

(Table 3.67). Seven of the thirty cutthroat trout collected were 1+ years of age, fifteen were 2+ years of age, four were 3+ and 4+ years of age (Table 3.68). Of the forty-one eastern brook trout collected, eight were 1+ years of age, seventeen were 2+ years of age, fifteen were 3+ years of age and one was 4+ years of age (Table 3.68).

**Table 3.66. Number of each species of fish caught by electrofishing in Alder Creek during May - October, 1994.**

| <b>Species</b>          | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>September<br/>(October)</b> |
|-------------------------|------------------------------|---------------------------------|--------------------------------|
| <b>Shock time (min)</b> | 38.3                         | 43.6                            | 70.6                           |
| Cutthroat trout         | 16                           | 26                              | 30                             |
| Eastern brook trout     | 28                           | 61                              | 41                             |
| Sculpin spp.            | 14                           | 26                              | 47                             |
| Longnose sucker         | 8                            | 12                              |                                |
| <i>TOTAL</i>            | <b>66</b>                    | <b>125</b>                      | <b>118</b>                     |

**Table 3.67. Percent (%) of each species of fish caught by electrofishing in Alder Creek during May - October, 1994.**

| <b>Species</b>      | <b>Spring<br/>(May-June)</b> | <b>Summer<br/>(July-August)</b> | <b>Fall<br/>(October)</b> |
|---------------------|------------------------------|---------------------------------|---------------------------|
| Cutthroat trout     | 24.2                         | 20.8                            | 25.4                      |
| Eastern brook trout | 42.4                         | 48.8                            | 34.7                      |
| Sculpin sm.         | 21.2                         | 20.8                            | 39.8                      |
| Longnose sucker     | 12.1                         | 9.6                             |                           |

**Table 3.68. Electrofishing relative abundance for salmonid species by age class in Alder Creek, 1994.**

| <b>Cutthroat trout</b> |               |               |             |
|------------------------|---------------|---------------|-------------|
| <b>Age</b>             | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 1+                     | 1             | 5             | 7           |
| 2+                     | 12            | 12            | 15          |
| 3+                     | 3             | 7             | 4           |
| 4+                     | 0             | 2             | 4           |

| <b>Eastern brook trout</b> |               |               |             |
|----------------------------|---------------|---------------|-------------|
| <b>Age</b>                 | <b>Spring</b> | <b>Summer</b> | <b>Fall</b> |
| 1+                         | 0             | 18            | 8           |
| 2+                         | 8             | 17            | 17          |
| 3+                         | 12            | 14            | 15          |
| 4+                         | 8             | 12            | 1           |

### 3.2.2. Trout Population and Biomass Estimates by Habitat Type

A total of 3,490 fish were caught during electrofishing surveys conducted in October, 1993. Approximately 29 percent of the catch was comprised of trout (n cutthroat= 600; n brook= 416). Population and biomass estimates for each study drainage are summarized below by reach and habitat type.

#### 3.2.2.1. Lake Creek

Cutthroat trout populations in the Lake Creek drainage were estimated at 1,457 fish with a total biomass of 57 pounds and a mean density of 2.6 trout/100 m<sup>2</sup> (Table 3.69). Adult fish (4+ and older) were not represented in the population estimate. Estimated cutthroat trout densities by reach ranged from 0.3-18.2/100 m<sup>2</sup>, with the highest densities found in reach 1 of West Fork Lake Creek. The lowest trout densities were found in mainstem reaches 6 and 7. Density estimates by habitat type ranged from 0.2-128.2/100 m<sup>2</sup> (Table 3.70). Cutthroat trout densities were greatest in plunge pool habitat and lowest in dammed pool and glide habitats. Total estimated biomass was greatest in low gradient riffles, which comprised 37 percent of the total habitat.

#### 3.2.2.2. Benewah Creek

Total trout population in the Benewah Creek drainage was estimated at 1,637 fish, with cutthroat (53 percent) and brook trout (47 percent) present in almost equal proportions (Table 3.71 & 3.73). Total cutthroat trout biomass was estimated at 77.2 pounds, with adults accounting for 29 percent of the total. Total estimated brook trout biomass was considerably greater (110.4 pounds), due to a relatively higher proportion of adults in the population. Cutthroat trout



**Table 3.69 Estimates of cutthroat trout abundance and biomass by reach. 1993.**

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST.(N)  | SE(N)        | #/100 ft.  | BIOMASS (lbs) |            |
|-------------------|-------|--------------|---------------|--------------|------------|---------------|------------|
|                   |       |              |               |              |            | JUVENILES     | ADULTS     |
| LAKE CREEK/CTT    | 2     | 1550         | 50.0          | 0.0          | 0.9        | 7.5           | 0.0        |
|                   | 3     | 3060         | 30.0          | 0.0          | 1.0        | 5.2           | 0.0        |
|                   | 4     | 5890         | 90.0          | 0.0          | 1.5        | 5.1           | 0.0        |
|                   | 5     | 3090         | 50.0          | 0.0          | 1.6        | 4.8           | 0.0        |
|                   | 6     | 19920        | 50.0          | 0.0          | 0.3        | 5.5           | 0.0        |
|                   | 7     | 1890         | 10.0          | 0.0          | 0.5        | 0.8           | 0.0        |
|                   | WFI   | 2860         | 520.5         | 70.9         | 18.2       | 17.0          | 0.0        |
|                   | VW2   | 11570        | 474.0         | 41.8         | 4.1        | 7.6           | 0.0        |
|                   | BOZI  | 1840         | 182.9         | 34.8         | 9.9        | 3.6           | 0.0        |
| <b>TOTALS</b>     |       | <b>55670</b> | <b>1457.4</b> | <b>147.5</b> | <b>2.6</b> | <b>57.0</b>   | <b>0.0</b> |

97

**Table 3.70 Estimates of cutthroat trout abundance and biomass by habitat type, 1993.**

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N)  | SE(N)        | #/100 ft.  | BIOMASS (lbs) |            |
|-------------------|--------------|--------------|---------------|--------------|------------|---------------|------------|
|                   |              |              |               |              |            | JUVENILES     | ADULTS     |
| LAKE CREEK/CTT    | DMP/BVP      | 2820         | 10.0          | 0.0          | 0.4        | 1.1           | 0.0        |
|                   | GLD          | 12920        | 70.0          | 0.0          | 0.5        | 5.4           | 0.0        |
|                   | LGR          | 20610        | 504.3         | 77.6         | 2.4        | 18.7          | 0.0        |
|                   | PKW          | 2400         | 163.3         | 7.8          | 6.8        | 3.2           | 0.0        |
|                   | PLP          | 410          | 277.0         | 22.8         | 67.6       | 7.6           | 0.0        |
|                   | RPD          | 2000         | 135.0         | 26.0         | 6.8        | 2.7           | 0.0        |
|                   | SCH          | 450          | 20.0          | 0.0          | 4.4        | 2.6           | 0.0        |
|                   | SCP          | 2620         | 80.0          | 0.0          | 3.1        | 11.1          | 0.0        |
|                   | SPC          | 11440        | 197.8         | 26.5         | 1.7        | 4.5           | 0.0        |
| <b>TOTALS</b>     |              | <b>55670</b> | <b>1457.4</b> | <b>160.7</b> | <b>2.6</b> | <b>57.0</b>   | <b>0.0</b> |

Table 3.71 Estimates of cutthroat trout abundance and biomass by reach, 1993.

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST.(N) | SE(N) | #/100 ft. | BIOMASS (lbs) |        |
|-------------------|-------|--------------|--------------|-------|-----------|---------------|--------|
|                   |       |              |              |       |           | JUVENILES     | ADULTS |
| BENEWAH CREEK/CTT | 1     | 13980        | 192          | 22.1  | 1.4       | 14.31         | 7.1    |
|                   | 2     | 11110        | 100          | 0.01  | 0.91      | 6.6           | 3.61   |
|                   | 3     | 3920         | 30           | 0.0   | 0.81      | 3.51          | 0.0    |
|                   | 4     | 4100         | 70           | 0.0   | ---       | 7.0           | 0.0    |
|                   | 5     | 5190         | 30           | 0.0   | 1.7       | 1.8           | 4.4    |
|                   | 6     | 7520         | 40           | 0.01  | 0.5       | 2.4           | 0.0    |
|                   | 7     | 14600        | 80           | 0.0   | 0.5       | 3.6           | 3.5    |
|                   | 8     | 10120        | 20           | 0.0   | 0.2       | 1.3           | 0.0    |
|                   | 9     | 3660         | 10           | 0.0   | 0.3       | 0.0           | 0.0    |
|                   | 10    | 6240         | 70           | 0.0   | 1.1       | 2.5           | 4.0    |
|                   | 11    | 6830         | 30           | 0.0   | 0.4       | 2.4           | 0.0    |
|                   | 12    | 3630         | 30           | 0.0   | 0.3       | 2.4           | 0.0    |
|                   | 13    | 3070         | 173          | 9.9   | 5.6       | 6.5           | 0.0    |
| <b>TOTALS</b>     |       | 93970        | 875          | 32.1  | 0.9       | 64.5          | 22.7   |

Table 3.72 Estimates of cutthroat trout abundance and biomass by habitat type, 1993.

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N) | SE(N) | #/100 ft. | BIOMASS (lbs) |        |
|-------------------|--------------|--------------|--------------|-------|-----------|---------------|--------|
|                   |              |              |              |       |           | JUVENILES     | ADULTS |
| BENEWAH CREEK/CTT | DMP/BVP      | 4320         | 0            | 0.0   | 0.0       | 0.0           | 0.0    |
|                   | GLD          | 15860        | 90           | 0.0   | 0.61      | 4.8           | 3.7    |
|                   | LGR          | 45800        | 273          | 9.9   | 0.6       | 15.31         | 3.5    |
|                   | PKW          | 11750        | 110          | 21.2  | 0.3       | 9.3           | 0.0    |
|                   | PLP          | 1200         | 60           | 0.0   | 5.0       | 3.0           | 7.4    |
|                   | SCP          | 5410         | 140          | 0.0   | 2.6       | 8.9           | 4.4    |
|                   | SPC          | 9630         | 202          | 6.3   | 2.1       | 13.2          | 3.6    |
| <b>TOTALS</b>     |              | 93970        | 875          | 37.5  | 0.9       | 64.5          | 22.7   |

Table 3.73 Estimates of brook trout abundance and biomass by reach, 1993.

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST.(N) | SE(N)       | #/100 ft.  | BIOMASS (lbs) |             |
|-------------------|-------|--------------|--------------|-------------|------------|---------------|-------------|
|                   |       |              |              |             |            | JUVENILES     | ADULTS      |
| BENEWAH CREEK/EBT | 1     | 13980        | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 2     | 11110        | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 3     | 3920         | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 4     | 4100         | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 5     | 5190         | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 6     | 7520         | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | 7     | 14600        | 30           | 0.0         | 0.2        | 2.2           | 0.0         |
|                   | 8     | 10120        | 20           | 0.0         | 0.2        | 0.0           | 9.9         |
|                   | 9     | 3660         | 20           | 0.0         | 0.5        | 0.0           | 9.3         |
|                   | 10    | 6240         | 30           | 0.0         | 0.5        | 2.0           | 5.5         |
|                   | 11    | 6830         | 362          | 18.4        | 5.3        | 27.0          | 17.3        |
|                   | 12    | 3630         | 190          | 0.0         | 5.2        | 14.4          | 13.6        |
|                   | 13    | 3070         | 110          | 0.0         | 3.6        | 5.9           | 3.1         |
| <b>TOTALS</b>     |       | <b>93970</b> | <b>762</b>   | <b>18.4</b> | <b>0.8</b> | <b>51.6</b>   | <b>58.8</b> |

Table 3.74 Estimates of brook trout abundance and biomass by habitat type, 1993.

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N) | SE(N)       | #/100 ft.  | BIOMASS (lbs) |             |
|-------------------|--------------|--------------|--------------|-------------|------------|---------------|-------------|
|                   |              |              |              |             |            | JUVENILES     | ADULTS      |
| BENEWAH CREEK/EBT | DMP/BVP      | 43:20        |              |             |            | 2.6           | 4.2         |
|                   | GID          | 15860        | 2950         | 10.4        | 1.2        | 26.7          | 3.7         |
|                   | LGR          | 45800        | 170          | 0           | 0.41       | 6.71          | 24.91       |
|                   | PKW          | 11750        | 0            | 0.0         | 0.0        | 0.0           | 0.0         |
|                   | PLP          | 1200         | 20           | 0           | 1.7        | 0.1           | 3.1         |
|                   | SCP          | 5410         | 220          | 0           | 4.1        | 14.8          | 22.9        |
|                   | SPC          | 9630         | 10           | 0           | 0.1        | 0.7           | 0.0         |
| <b>TOTALS</b>     |              | <b>93970</b> | <b>762</b>   | <b>18.4</b> | <b>0.8</b> | <b>51.6</b>   | <b>58.8</b> |

were distributed in fairly even numbers throughout the drainage, with the highest densities occurring in reach 13 (8.0 trout/100 m<sup>2</sup>). Cutthroat trout biomass, however, was disproportionately higher in the lower reaches (55 percent of total in reaches 1-4). The highest estimates of cutthroat trout density and biomass occurred in plunge pool habitat (4.5 fish/100 m<sup>2</sup> and 0.78 lbs./100 m<sup>2</sup>). Brook trout were only found in reaches 7-13, with the highest densities occurring in the uppermost reaches (Table 3.73 & 3.74). Brook trout were found in the highest densities in scour pool habitat (2.2 fish/100 m<sup>2</sup>).

### **3.2.2.3. Evans Creek**

Cutthroat trout populations in the Evans Creek drainage were estimated at 2,480 fish with a total biomass of 151 pounds and a mean density of 4.7 trout/100 m<sup>2</sup> (Table 3.75). Biomass of juvenile fish was 3.7 times greater than adult biomass. Estimated cutthroat trout densities by reach ranged from 0.0-7.3/100 m<sup>2</sup>. Density estimates were generally higher in the lower reaches of the drainage. Density estimates by habitat type ranged from 0.0-24.8/100 m<sup>2</sup> (Table 3.76). Cutthroat trout densities were greatest in plunge pool and scour pool habitats (24.8/100 m<sup>2</sup> and 20.7/100 m<sup>2</sup>, respectively). The highest estimate of biomass per unit area, however, was calculated for scour pool habitat (1.8 lbs./100 m<sup>2</sup>).

### **3.2.2.4. Alder Creek**

Total trout population in the Alder Creek drainage was estimated at 5,380 fish, with cutthroat trout comprising 30 percent of the total and brook trout the remaining 70 percent (Table 3.77 & 3.79). Total cutthroat trout biomass was estimated at 102.7 pounds. Juveniles accounting for 79 percent of the total biomass. Total estimated brook trout biomass was 298 pounds, with juveniles and adults comprising almost equal proportions. Cutthroat trout were found in increasingly greater numbers from reach 1 to 4, however, none were found upstream of reach 5. Total cutthroat trout biomass was greatest in reach 3 (45 percent of total). The highest estimates of cutthroat trout density and biomass occurred in glide habitat (2.9 fish/100 m<sup>2</sup> and 0.15 lbs./100 m<sup>2</sup>). Brook trout were found in all sample reaches except for reach 1, and densities generally increased in an upstream direction (Table 3.79). The highest brook trout density occurred in reach 3 of North Fork Alder Creek (61.8 fish/100 m<sup>2</sup>). The highest biomass estimates per unit area were found in mainstem reach 5 (0.48 lbs./100 m<sup>2</sup>). Brook trout were found in the highest densities and biomass in scour hole habitat (12.2 fish/100 m<sup>2</sup>; 1.1 lbs./100 m<sup>2</sup>) (Table 3.80).

Table 3.75 Estimates of cutthroat trout abundance and biomass by reach, 1993.

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST.(N) | SE(N) | #/100 ft. | BIOMASS (lbs) |             |
|-------------------|-------|--------------|--------------|-------|-----------|---------------|-------------|
|                   |       |              |              |       |           | JUVENILES     | ADULTS      |
| EVANS CREEK/CTT   | 1     | 3840         | 0            | 0.0   | 0.0       | 0.0           | 0.0         |
|                   | 2     | 115530       | 455          | 34.6  | 2.9       | 20.2          | 11.5        |
|                   | 3     | 21220        | 771          | 21.5  | 3.6       | 42.4          | 11.9        |
|                   | 4     | 22050        | 878          | 42.7  | 4.0       | 41.6          | 8.6         |
|                   | 5     | 5470         | 203          | 36.0  | 3.7       | 9.9           | 0.0         |
|                   | EF1   | 69601        | 123          | 9.9   | 1.8       | 3.5           | 0.0         |
|                   | WF1   | 3090         | 50           | 0.0   | 1.6       | 1.2           | 0.0         |
| <b>TOTALS</b>     |       | <b>83160</b> | 2480         | 144.8 | 3.0       | 118.91        | <b>32.1</b> |

Table 3.76 Estimates of cutthroat trout abundance and biomass by habitat type, 1993.

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N) | SE(N) | #/100 ft. | BIOMASS (lbs) |             |
|-------------------|--------------|--------------|--------------|-------|-----------|---------------|-------------|
|                   |              |              |              |       |           | JUVENILES     | ADULTS      |
| EVANS CREEK/CTT   | DMP/BVP      | 1720         | 166          | 12.2  | 9.6       | 8.3           | 4.4         |
|                   | EDP          | 850          | 60           | 0.0   | 7.1       | 1.2           | 0.0         |
|                   | GLD          | 3810         | 0            | 0.0   | 0.0       | 0.0           | 0.0         |
|                   | LGR          | 16810        | 400          | 37.3  | 2.4       | 18.9          | 11.5        |
|                   | PLP          | 760          | 209          | 15.4  | 27.5      | 12.4          | 2.9         |
|                   | RPD          | 17630        | 439          | 13.1  | 2.5       | 15.2          | 0.0         |
|                   | SCH          | 660          | 203          | 22.1  | 30.7      | 8.7           | 0.0         |
|                   | SCP          | 2480         | 629          | 44.8  | 25.4      | 40.7          | 13.3        |
|                   | SFC          | 100          | 0            | 0.0   | 0.0       | 0.0           | 0.0         |
|                   | SPC          | 13120        | 374          | 21.0  | 2.8       | 13.7          | 0.0         |
| <b>TOTALS</b>     |              | <b>57940</b> | 2480         | 165.9 | 4.3       | 118.9         | <b>32.1</b> |

**Table 3.77 Estimates of cutthroat trout abundance and biomass by reach, 1993.**

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST.(N) | SE(N)        | #/100 ft.  | BIOMASS (lbs) |             |
|-------------------|-------|--------------|--------------|--------------|------------|---------------|-------------|
|                   |       |              |              |              |            | JUVENILES     | ADULTS      |
| ALDER CREEK/CTT   | 1     | 8840         | 130          | 53.7         | 1.5        | 4.9           | 0.0         |
|                   | 2     | 15530        | 220          | 49.2         | 1.4        | 11.0          | 2.9         |
|                   | 3     | 21220        | 640          | 116.9        | 3.0        | 35.9          | 10.5        |
|                   | 4     | 22050        | 630          | 59.9         | 2.9        | 28.7          | 7.7         |
|                   | 5     | 5470         | 20           | 0.0          | 0.4        | 1.1           | 0.0         |
|                   | 6     | 6960         | 0            | 0.0          | 0.0        | 0.0           | 0.0         |
|                   | NF1   | 3090         | 0            | 0.0          | 0.0        | 0.0           | 0.0         |
|                   | NF2   | 3090         | 0            | 0.0          | 0.0        | 0.0           | 0.0         |
|                   | NF3   | 2090         | 0            | 0.0          | 0.0        | 0.0           | 0.0         |
| <b>TOTALS</b>     |       | <b>88340</b> | <b>1640</b>  | <b>279.8</b> | <b>1.9</b> | <b>81.6</b>   | <b>21.1</b> |

**Table 3.78 Estimates of cutthroat trout abundance and biomass by habitat type, 1993.**

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N) | SE(N)        | #/100 ft.  | BIOMASS (lbs) |             |
|-------------------|--------------|--------------|--------------|--------------|------------|---------------|-------------|
|                   |              |              |              |              |            | JUVENILES     | ADULTS      |
| ALDER CREEK/CTT   | DMP/BVP      | 7390         | 60           | 34.6         | 0.8        | 3.2           | 0.0         |
|                   | GLD          | 3530         | 210          | 116.8        | 5.9        | 11.0          | 0.0         |
|                   | LGR          | 52240        | 860          | 37.6         | 1.6        | 42.0          | 10.5        |
|                   | PKW          | 4500         | 100          | 41.1         | 2.2        | 4.2           | 0.0         |
|                   | PLP          | 2720         | 40           | 0.0          | 1.5        | 2.6           | 0.0         |
|                   | SCH          | 2450         | 30           | 0.0          | 1.2        | 1.2           | 0.0         |
|                   | SCP          | 7070         | 190          | 41.6         | 2.7        | 11.7          | 7.7         |
|                   | SPC          | 8440         | 150          | 53.7         | 1.8        | 5.7           | 2.9         |
| <b>TOTALS</b>     |              | <b>88340</b> | <b>1640</b>  | <b>325.5</b> | <b>1.9</b> | <b>81.6</b>   | <b>21.1</b> |

Table 3.79 Estimates of brook trout abundance and biomass by reach, 1993.

| TRIBUTARY/SPECIES | REACH | LENGTH (ft.) | POP. EST. (N) | SE(N) | #/100 ft. | BIOMASS (lbs) |        |
|-------------------|-------|--------------|---------------|-------|-----------|---------------|--------|
|                   |       |              |               |       |           | JUVENILES     | ADULTS |
| ALDER CREEK/EBT   | 1     | 8640         | 0             | 0.0   | 0.0       | 0.0           | 0.0    |
|                   | 2     | 15530        | 100           | 6.3   | 0.61      | 2.71          | 16.4   |
|                   | 3     | 21220        | 310           | 115.2 | 1.5       | 19.1          | 36.5   |
|                   | 4     | 22050        | 960           | 81.3  | 4.4       | 31.7          | 43.1   |
|                   | 5     | 5470         | 640           | 30.3  | 11.7      | 39.3          | 20.2   |
|                   | 6     | 6960         | 590           | 44.2  | 8.5       | 30.1          | 9.4    |
|                   | NF1   | 3090         | 390           | 49.1  | 12.6      | 22.8          | 9.3    |
|                   | NF2   | 3090         | 290           | 0.0   | 9.4       | 11.3          | 0.0    |
|                   | NF3   | 2090         | 460           | 41.3  | 22.0      | 6.1           | 0.0    |
| TOTALS            |       | 883401       | 3740          | 387.8 | 4.2       | 163.0         | 135.0  |

Table 3.80 Estimates of brook trout abundance and biomass by habitat type, 1993.

| TRIBUTARY/SPECIES | HABITAT TYPE | LENGTH (ft.) | POP. EST.(N) | SE(N) | #/100 ft. | BIOMASS (lbs) |        |
|-------------------|--------------|--------------|--------------|-------|-----------|---------------|--------|
|                   |              |              |              |       |           | JUVENILES     | ADULTS |
| ALDER CREEK/EBT   | DMP/BVP      | 7390         | 160          | 0.0   | 2.2       | 6.5           | 5.4    |
|                   | GLD          | 3530         | 230          | 22.8  | 6.5       | 12.9          | 13.8   |
|                   | LGR          | 52240        | 1340         | 116.9 | 2.6       | 60.8          | 33.7   |
|                   | PKW          | 4500         | 0            | 0.0   | 0.0       | 0.0           | 0.0    |
|                   | PLP          | 2720         | 270          | 0.0   | 9.9       | 12.2          | 13.6   |
|                   | SCH          | 2450         | 370          | 34.7  | 15.1      | 27.7          | 6.8    |
|                   | SCP          | 7070         | 900          | 103.1 | 12.7      | 39.1          | 58.9   |
|                   | SPC          | 8440         | 470          | 30.0  | 5.6       | 3.1           | 2.8    |
| TOTALS            |              | 88340        | 3740         | 307.5 | 4.2       | 182.4         | 135.0  |

### 3.2.3. Habitat Use Evaluation

Chi-square goodness of fit tests and analysis of catch rates indicate that cutthroat and brook trout were not distributed uniformly ( $p < 0.001$ ) among the available habitat types in the four study drainages. (Table 3.81) Drainage specific catch rates showed considerable variation in selection of habitat types, whereas combined catch rates for both cutthroat and brook trout showed a strong selective tendency for deep water habitat with slower velocities.

**BVP/DMP** - Combined catch rates were lower in beaver created pools (BVP) and dammed pools (DMP) than all other habitat types utilized by cutthroat and brook trout (0.14 trout/min and 0.13 trout/min, respectively). Chi-square analysis also indicated that trout generally underutilized this habitat type. Observed catch rates for cutthroat and brook trout were 2.3 and 1.7 times lower than expected if populations had been equally distributed among habitat types (Table 3.81).

**EDP** - Cutthroat trout were observed utilizing eddy pools (EDP) only in the Evans Creek drainage. Combined catch rates, however, were 1.3 times high than expected if populations had been equally distributed among habitat types. Chi-square analysis of the combined observed versus expected values indicated that cutthroat trout selected for eddy pool habitat (Table 3.81). No brook trout were observed in this habitat type.

**GLD** - Combined catch rates for cutthroat trout (0.15 trout/min) were the second lowest of all habitat types. Catch rates for brook trout fell in the median range for all habitat types (0.24 trout/min). Chi-square analysis indicated that cutthroat trout selected against glide habitats, while brook trout were found 1.1 times more frequently than expected (Table 3.81).

**LGR** - Combined catch rates for cutthroat trout and brook trout were similar (0.28 trout/min and 0.21 trout/min, respectively) in low gradient riffles (LGR). Chi-square analysis indicated that cutthroat and brook trout were not observed as frequently as expected if populations had been equally distributed among all habitat types (Table 3.81).

**PKW** - Combined catch rates for cutthroat trout were 0.28 trout/min in pocket water habitat (PKW). No brook trout were observed in this habitat type. Chi-square analysis indicated that cutthroat trout generally underutilized this habitat type. Catch rates were 1.2 times less than expected if populations had been equally distributed among all habitat types (Table 3.81).

**PLP** - Combined catch rates for cutthroat trout were higher in plunge pools (PLP) than in any other habitat type (0.71 trout/min). Combined catch rates for brook trout were 0.34 trout/min. Chi-square analysis indicated that trout selected for this habitat type. Observed catch rates for cutthroat and brook trout were 2.2



and 1.5 times higher than expected if populations had been equally distributed among habitat types (Table 3.81).

**RPD** - Combined catch rates for cutthroat trout (0.55 trout/min) were second highest among all habitat types. Chi-square analysis indicated that trout selected for this habitat type. Observed catch rates were 1.7 times higher than expected if populations had been equally distributed among habitat types (Table 3.81). No brook trout were observed in rapids.

**SCH** - Combined catch rates for cutthroat and brook trout were 0.43 trout/min and 0.63 trout/min, respectively. Catch rates for brook trout in scour holes (SCH) were higher than in any other habitat type. Chi-square analysis indicated that trout selected for this habitat type. Observed catch rates for cutthroat and brook trout were 1.3 and 2.8 times greater than expected (Table 3.81).

**SCP** - Combined catch rates for cutthroat and brook trout in scour pools (SCP) were similar (0.45 trout/min and 0.46 trout/min). Catch rates for brook trout were second highest among all habitat types. Observed catch rates were 1.4 and 2.1 times higher than expected for cutthroat and brook trout, respectively (Table 3.81). Chi-square analysis indicated that trout selected for this habitat type.

**SPC** - Combined catch rates for step-pool cascades (SPC) were 0.37 trout/min for cutthroat trout and 0.20 trout/min for brook trout. Chi-square analysis indicated that cutthroat trout selected for this habitat type, while brook trout selected against it. Observed catch rates were 1.2 times higher than expected for cutthroat and 1.1 times lower than expected for brook trout (Table 3.81).

#### **3.2.4. Spawning Surveys**

Spawning surveys were conducted on Lake, Benawah, Evans, and Alder creeks during late May. During 1993 and 1994 the entire stream length was surveyed to locate and identify redds. Because spawning surveys were conducted during spring runoff, it was difficult to see the bottom of the stream channel, especially in mainstem areas. No redds were located in any drainages.

#### **3.2.5. Trout Migration Analysis**

Traps were monitored in Lake, Benawah, Evans, and Alder creeks from mid-March through mid-May, 1994. Migration traps were not successfully installed in 1993 due to severe runoff conditions. A total of 1,348 fish were caught in all traps in 1994. Cutthroat trout comprised the greatest proportion of the catch (95 percent) with largescale sucker, reddsideshiner, brook trout, longnose sucker, and pumpkinseed comprising the remainder of the catch.

**Table 3.81 Chi-Square analysis of combined observed and expected values for electrofishing catches by habitat type, 1993.**

| Habitat Type | Effort (min) | C R  |       |              | EBT  |       |              |
|--------------|--------------|------|-------|--------------|------|-------|--------------|
|              |              | Obs. | Exp.  | Mag. p-Value | Obs. | Exp.  | Mag. p-Value |
| BVP/DMP      | 160          | 22   | 51.5  | -2.3         | 21   | 35.7  | -1.7         |
| EDP          | 15           | 6    | 4.8   | 1.3          | 0    | 3.3   |              |
| GLD          | 203          | 30   | 65.2  | -2.2         | 49   | 45.2  | 1.1          |
| LGR          | 668          | 188  | 214.9 | -1.1         | 139  | 149.0 | -1.1         |
| PKW          | 127          | 35   | 40.8  | -1.2         | 0    | 28.3  |              |
| PLP          | 79           | 56   | 25.4  | 2.2          | 27   | 17.6  | 1.5          |
| RPD          | 100          | 55   | 32.2  | 1.7          |      |       |              |
| SCH          | 55           | 24   | 17.8  | 1.3          | 35   | 12.3  | 2.8          |
| SCP          | 216          | 97   | 69.5  | 1.4          | 99   | 48.2  | 2.1          |
| SFC          | 8            | 0    | 2.6   |              |      |       |              |
| SPC          | 234          | 87   | 75.3  | 1.2          | 46   | 52.2  | -1.1         |
| TOTAL        | 1866         | 600  | 600   | ~0.001       | 416  | 416.0 | <0.001       |

A total of six hundred ninety-eight fish were caught in Lake Creek traps, of which 99 percent were cutthroat trout. Analysis of cutthroat trout age class structure showed that age 2+ fish were the most abundant age class followed by 6+, 5+ and 7+ (Figure 3.1). Immature trout (age 0+ through 3+) comprised 64 percent of the population and adult trout (4+ and older) made up 36 percent of the total catch.

Upstream migration of cutthroat trout in the Lake Creek drainage was concentrated within the period March 20-April 23. Approximately, 71 percent of all upstream migration occurred during this period. Upstream movement past the lower Lake Creek trap consisted primarily of mature fish (age 5+ - 7+) greater than 277 mm total length (Figure 3.2). Most of these fish were thought to belong to adfluvial stocks. In contrast, upstream movement at the upper Lake Creek trap was dominated by age 2+ fish (Figure 3.3). Trap data indicated that spawning activity of adfluvial stocks occurred primarily below the confluence with Bozard Creek, in reaches 1-7.

A total of five hundred eighty-seven fish were trapped in Benawah Creek, of which 90 percent were cutthroat trout and 1.7 percent were brook trout. Analysis of cutthroat trout age class structure showed that age 2+ and 3+ fish were the most abundant age classes, accounting for 56 percent of the population. Age 7+ and 5+ trout comprised 18 and 17 percent of the population, respectively (Figure 3.4).

Upstream migration of cutthroat trout in Benawah Creek conformed to a bimodal distribution, with mature fish (age 5+ - 7+) migrating past the lower trap during the period April 3 to April 30, while the number of juvenile migrants peaked during the week May 15-21 (Figure 3.5). The age distribution of upstream migrants at the lower Benawah Creek trap consisted primarily of age classes 5+ and 7+ (28 and 30 percent of the total catch, respectively). Age 2+ fish comprised 21 percent of the catch in lower Benawah Creek. The timing of adult migration into Benawah Creek generally coincided with runs into Lake Creek. A total of 48 fish were caught in the trap set at the mouth of SE. Benawah Creek. These fish were primarily age class 3+ and 2+, 50 and 19 percent of the total catch, respectively. Twenty-seven percent were mature fish (age classes 5-7). An additional nine trout were caught at the mouth of Whitetail Creek, however, only one of these fish was a mature adult.

Fish trapped in the Alder and Evans creek drainages accounted for less than 10 percent of the total catch. Thirty-nine fish were trapped in Alder Creek, of which three were mature cutthroat trout greater than 320 mm total length. A total of eighty-four fish were caught in Evans Creek, none of which were trout species.

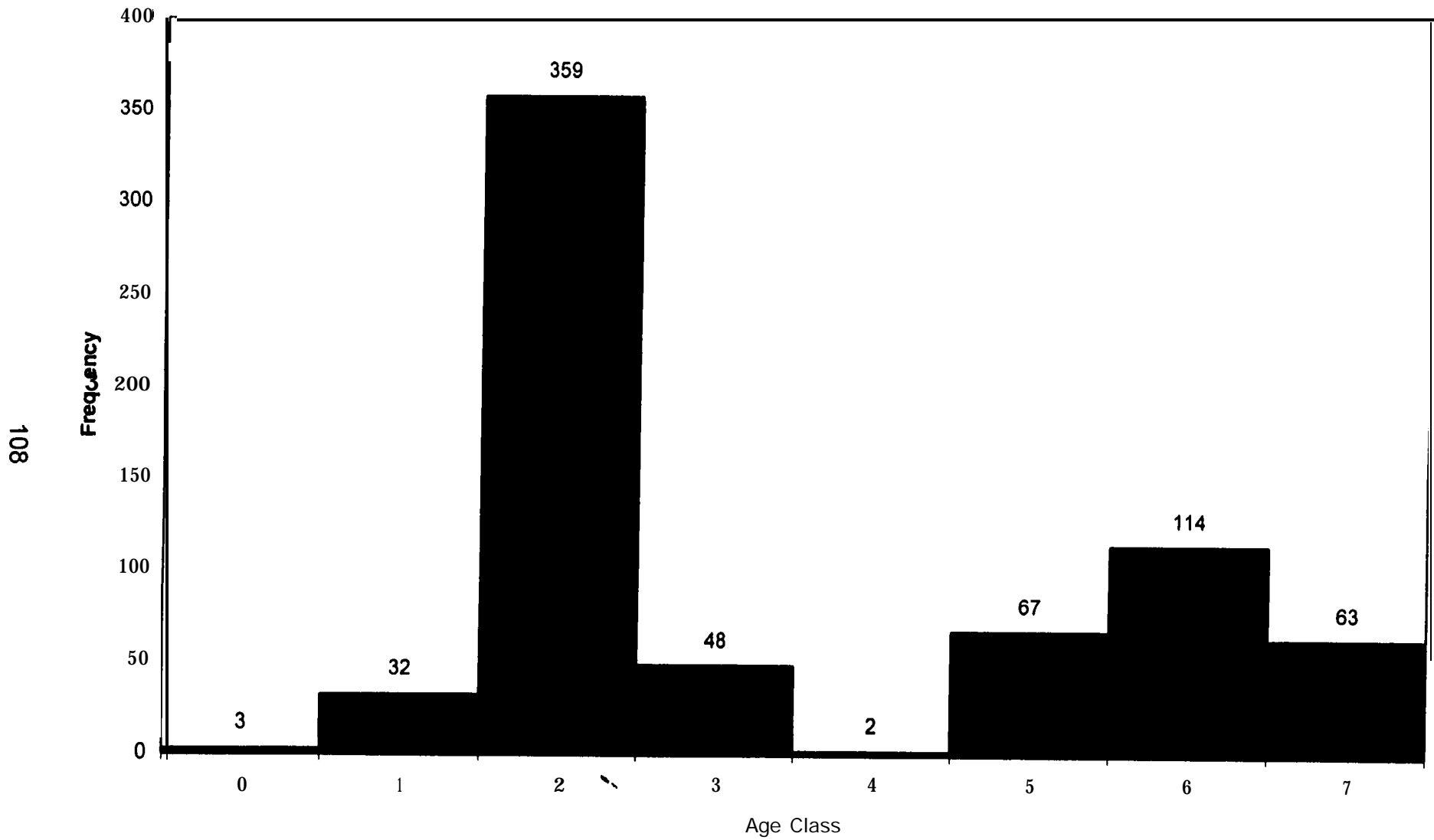


Figure 3.1 Age frequency of cutthroat trout in Lake Creek, March-May, 1994.

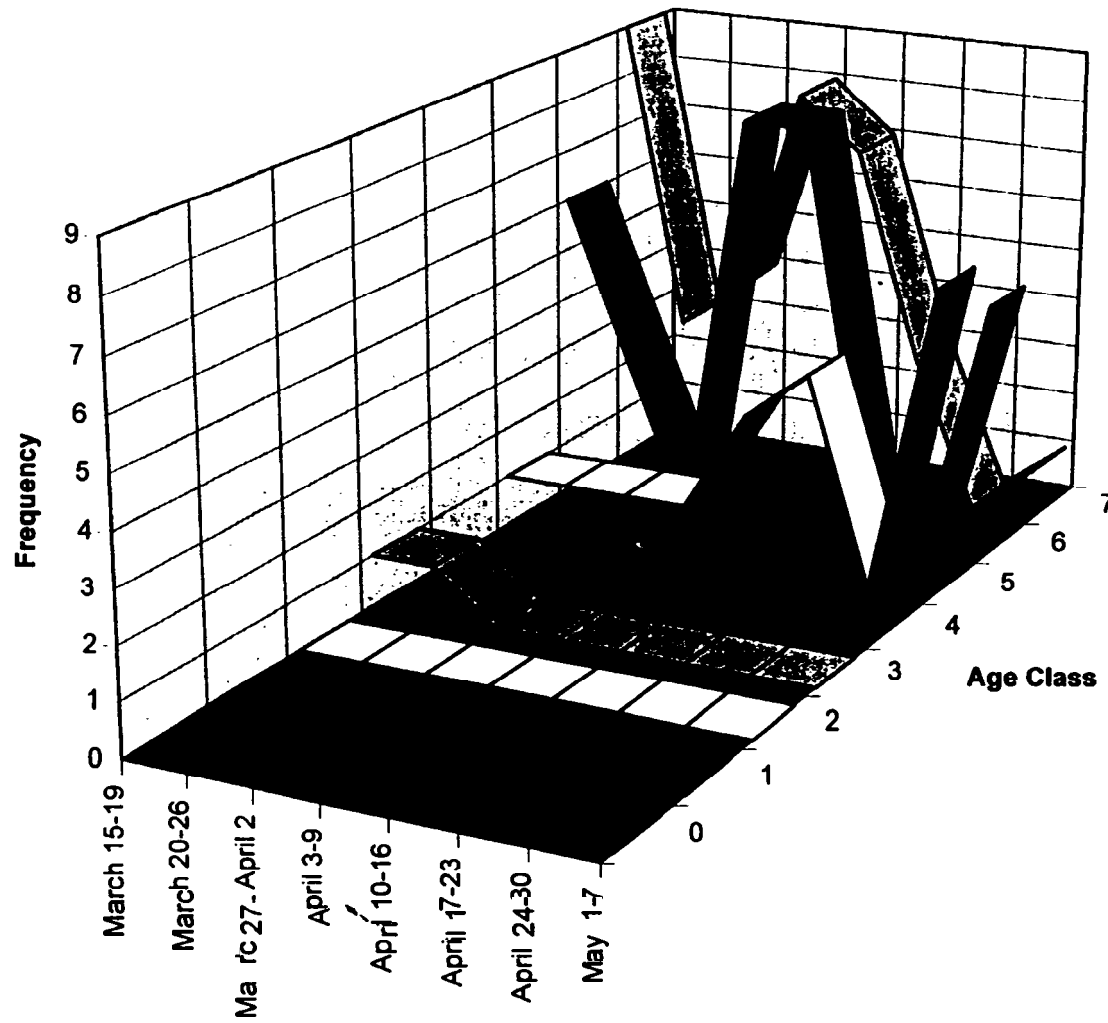


Figure 3.2 Timing of upstream migration for cutthroat trout in lower Lake Creek.

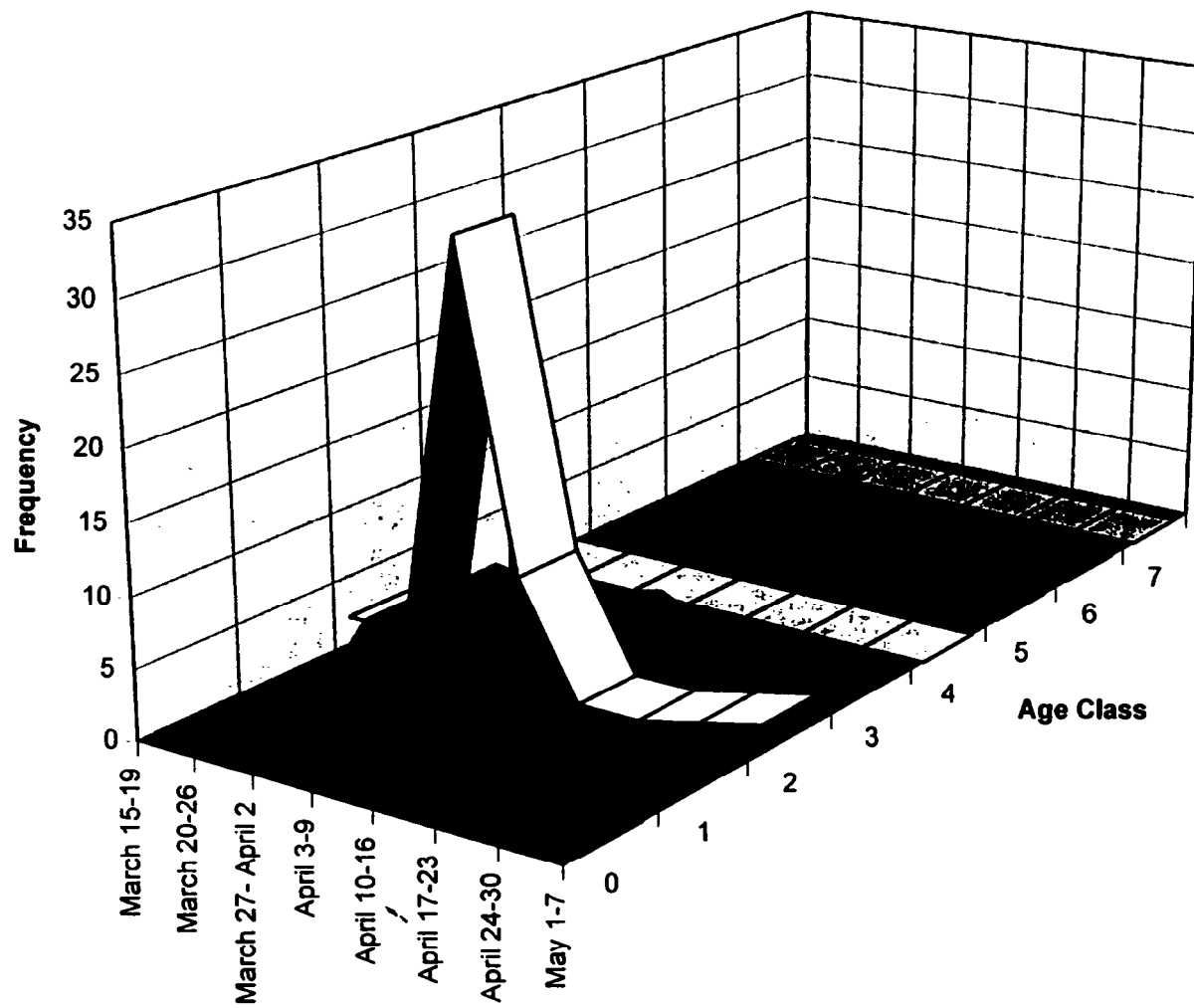
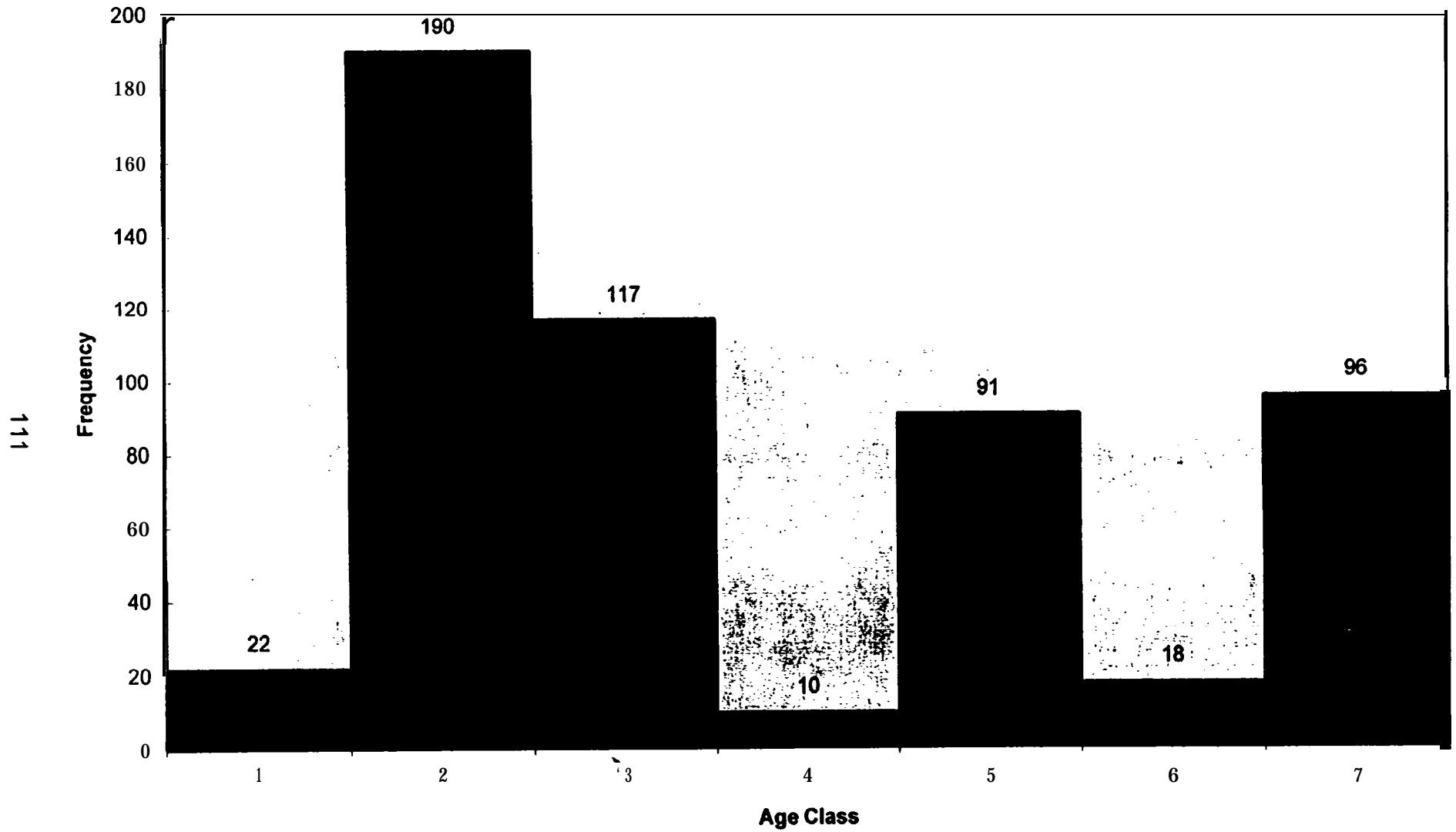


Figure 3.3 Timing of upstream migration for cutthroat trout in upper Lake Creek.



**Figure 3.4 Age frequency of cutthroat trout in Benewah Creek, March-May, 1994.**

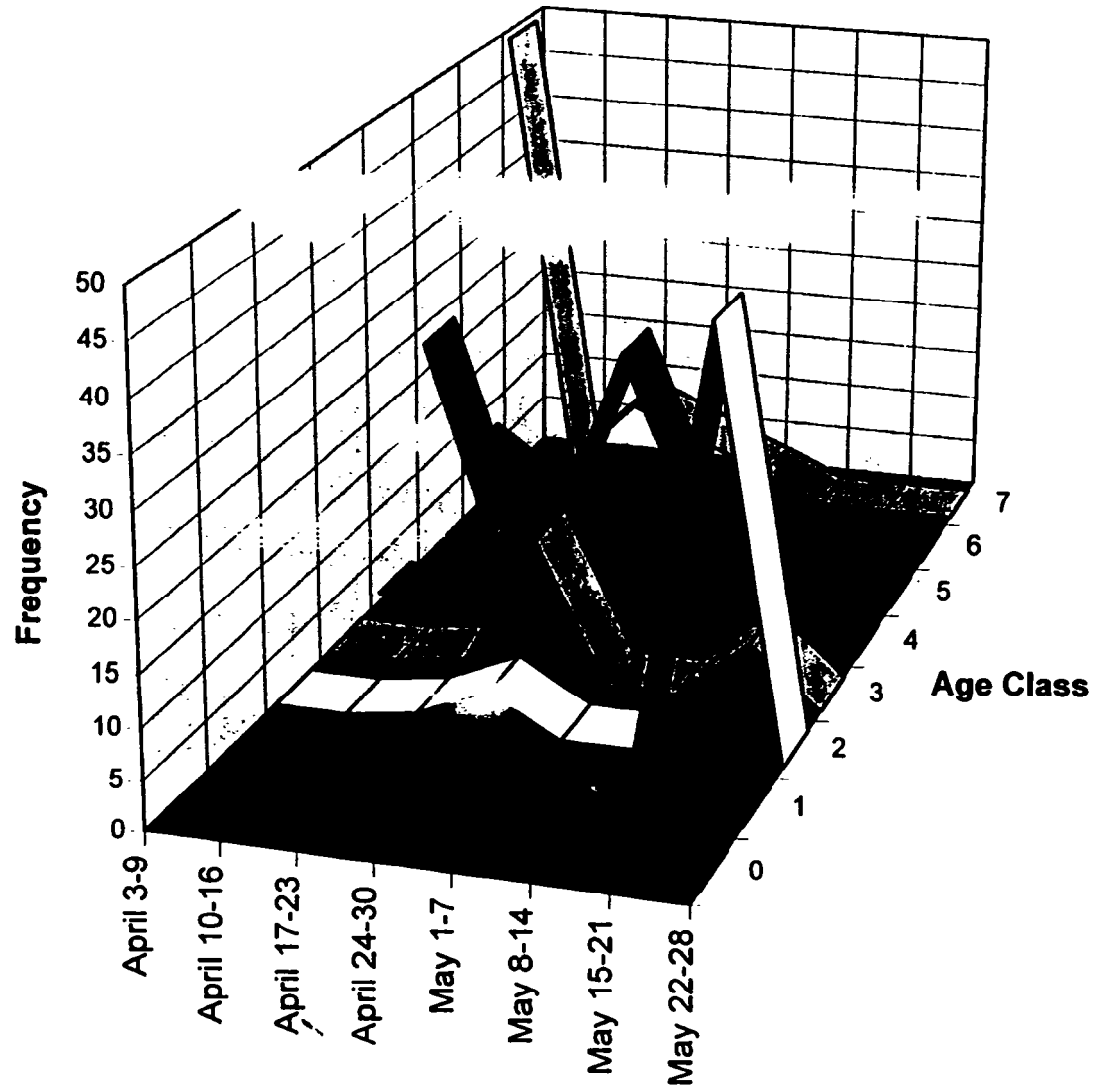


Figure 3.5 Timing of upstream migration for cutthroat trout in lower Benawah Creek.



### **3.2.6. Age, Growth and Condition**

#### **3.2.6.1. 1993**

##### **Lake Creek**

A total of 145 **scales** were collected from cutthroat trout in Lake Creek for age determination. Back-calculated lengths for cutthroat trout at the first annulus ranged from 60 to 74 mm with a grand mean of 66.2 mm (Table 3.82). At the end of the second years growth, lengths ranged from 79 to 105 mm with a grand mean of 92 mm. At the end of the third years growth, lengths ranged from 109 to 144 mm with a grand mean of 122 mm. At the end of the fourth years growth lengths ranged from 144 to 172 mm with a grand mean of 158 mm. At the end of the fifth year of growth, lengths ranged from 163 to 198 with a grand mean of 192 mm. At the end of the sixth year of growth lengths averaged 183 mm.

Mean lengths, weights and condition factors for each age class of cutthroat trout in Lake Creek are listed in Table 3.83. Mean condition factors ranged from 0.9 for 2+ and 3+ age classes to 1.2 for 1+ age classes with an overall mean of 1.0.

##### **Benewah Creek**

A total of 125 scales were collected from cutthroat trout in Benewah Creek for age determination. Back-calculated lengths for cutthroat trout at the first annulus ranged from 72 to 80 mm with a grand mean of 74 mm. At the second years of growth, lengths ranged from 95 to 106 mm with a grand mean of 100 mm. At the end of the third years of growth, lengths ranged from 118 to 149 mm with a grand mean of 126 mm. At the end of the fourth years of growth, lengths ranged from 139 to 190 mm with a grand mean of 151 mm. At the end of the fifth year of growth, lengths ranged from 171 to 223 mm with a grand mean of 190 mm. Lengths at the end of six years of growth were 247 mm (Table 3.84).

Mean lengths, weights and condition factors for each age class of cutthroat trout in Benewah Creek are listed in Table 3.85. Mean condition factors ranged from 0.8 for 2+ age class to 1 .1 for the 6+ **age** class with an overall mean of 1.0.

##### **Evans Creek**

A total of 365 scales were collected from cutthroat trout in Evans Creek for age determination. Back-calculated lengths for cutthroat trout at the first annulus ranged from 68 to 78 mm with a grand mean of 71 mm. At the end of the second years growth, lengths ranged from 93 to 97 mm with a grand mean of 94 mm. At the end of the third year of growth, lengths ranged from 114 to 123 mm with a grand mean of 115 mm. At the end of the fourth year of growth,

Table 3.82 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Lake Creek, 1993.

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |          |             |              |              |              |              |          |
|--|----------|-------------|--------------|--------------|--------------|--------------|----------|
| <b>Cohort</b>  | <b>N</b> | <b>1</b>    | <b>2</b>     | <b>3</b>     | <b>4</b>     | <b>5</b>     | <b>6</b> |
| 1992   | 24       | 66.4 ± 8.4  |              |              |              |              |          |
| 1991   | 40       | 66.9 ± 7.3  | 91.8 ± 12.1  |              |              |              |          |
| 1990   | 37       | 62.1 ± 8.2  | 86.4 ± 13.1  | 109.5 ± 16.4 |              |              |          |
| 1989   | 31       | 68.0 ± 24.6 | 95.2 ± 17.9  | 129.9 ± 23.7 | 154.9 ± 23.1 |              |          |
| 1988   | 10       | 74.2 ± 11.8 | 104.9 ± 18.3 | 144.2 ± 27.9 | 171.7 ± 43.5 | 198.3 ± 53.6 |          |
| 1987   | 1        | 59.8        | 79.2         | 124.4        | 143.8        | 163.2        | 182.5    |
| GRAND MEAN   | 145      | 66.2 ± 8.9  | 91.9 ± 14.4  | 122.1 ± 22.4 | 158.3 ± 27.8 | 192.4 ± 50.4 | 182.5    |
| MEAN ANNUAL GROWTH INCREMENT                         |          | 66          | 26           | 30           | 36           | 34           | -        |

114

Table 3.83 Mean lengths, weights and condition factors for each age class of cutthroat trout in Lake Creek, 1993.

| <b>Age</b> | <b>N</b> | <b>Mean weight (g) ±SD</b> | <b>Mean length (mm) ±SD</b> | <b>Mean K<sub>H</sub> ±SD</b> |
|------------|----------|----------------------------|-----------------------------|-------------------------------|
| 0+         | 14       | 58.7 ± 10.2                | 2.2 ± 1.2                   | 1.0 ± 0.2                     |
| 1+         | 23       | 77.3 ± 12.3                | 5.6 ± 2.7                   | 1.2 ± 0.2                     |
| 2+         | 40       | 108.6 ± 15.9               | 11.4 ± 5.0                  | 0.9 ± 0.2                     |
| 3+         | 37       | 122.5 ± 19.6               | 18.1 ± 11.1                 | 0.9 ± 0.2                     |
| 4+         | 30       | 171.2 ± 19.5               | 52.8 ± 18.5                 | 1.0 ± 0.1                     |
| 5+         | 10       | 212.3 ± 51.7               | 130.8 ± 18.4                | 1.0 ± 0.2                     |
| 6+         | 1        | 189                        | 67                          | 1.0                           |
| Total      | 158      |                            |                             | 1.0 ± 0.2                     |

Table 3.84 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Benawah Creek (including tributaries), 1994.

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |          |             |              |              |              |              |              |
|--|----------|-------------|--------------|--------------|--------------|--------------|--------------|
| <b>Cohort</b>  | <b>N</b> | <b>1</b>    | <b>2</b>     | <b>3</b>     | <b>4</b>     | <b>5</b>     | <b>6</b>     |
| 1992   | 7        | 71.6 ± 3.4  |              |              |              |              |              |
| 1991   | 23       | 76.1 ± 9.4  | 103.9 ± 14.8 |              |              |              |              |
| 1990   | 18       | 79.6 ± 10.9 | 104.9 ± 15.2 | 128.3 ± 17.1 |              |              |              |
| 1989   | 42       | 72.7 ± 8.2  | 95.1 ± 11.3  | 117.9 ± 12.7 | 139.1 ± 14.9 |              |              |
| 1988   | 22       | 72.1 ± 6.5  | 97.2 ± 13.9  | 126.9 ± 26.6 | 152.0 ± 35.6 | 170.7 ± 38.9 |              |
| 1987   | 13       | 74.3 ± 9.0  | 105.6 ± 15.1 | 149.2 ± 42.6 | 189.7 ± 55.9 | 222.9 ± 68.9 | 247.7 ± 69.0 |
| GRAND MEAN<br>MEAN ANNUAL<br>GROWTH<br>INCREMENT     | 125      | 74.3 ± 8.7  | 99.9 ± 14.1  | 126.2 ± 24.9 | 151.3 ± 36.0 | 190.1 ± 57.2 | 246.7 ± 69.0 |
|  |          | 74          | 26           | 26           | 25           | 39           | 57           |

Table 3.85 Mean lengths, weights and condition factors for each age class of cutthroat trout in Benawah Creek (including tributaries), 1994.

| <b>Age</b> | <b>N</b> | <b>Mean weight (g) ±SD</b> | <b>Mean length (mm) ±SD</b> | <b>Mean K<sub>tl</sub> ±SD</b> |
|------------|----------|----------------------------|-----------------------------|--------------------------------|
| 0+         | 4        | 56.3 ± 3.0                 | 1.8 ± 0.5                   | 1.0 ± 0.4                      |
| 1+         | 36       | 75.3 ± 16.8                | 4.6 ± 3.8                   | 1.0 ± 0.4                      |
| 2+         | 32       | 118.8 ± 17.4               | 14.9 ± 6.5                  | 0.8 ± 0.3                      |
| 3+         | 27       | 137.7 ± 21.0               | 26.6 ± 1.8                  | 0.9 ± 0.1                      |
| 4+         | 51       | 147.1 ± 19.2               | 29.6 ± 18.3                 | 0.9 ± 0.2                      |
| 5+         | 31       | 175.4 ± 38.0               | 59.0 ± 49.8                 | 1.0 ± 0.2                      |
| 6+         | 12       | 253.9 ± 75.4               | 230.2 ± 199.7               | 1.1 ± 0.3                      |
| Total      | 193      |                            |                             | 1.0 ± 0.3                      |

lengths ranged from 137 to 148 mm with a grand mean of 138 mm. At the end of the fifth year of growth, lengths ranged from 164 to 168 mm with a grand mean of 165 mm. At the end of the sixth year of growth, lengths were 192 mm (Table 3.86).

Mean lengths, weights and condition factors for each age class of cutthroat trout in Evans Creek are listed in Table 3.87. Mean condition factors range from 0.9 for age classes 1+,2+ and 3+, to 1.1 for 6+ fish with an overall mean 0.9.

### **Alder Creek**

A total of 227 scales were collected from cutthroat trout in Alder Creek for age determination. Back-calculated lengths for cutthroat trout at the first annulus ranged from 58 to 77 mm with a grand mean of 65 mm. At the end of the second years growth, lengths ranged from 80 to 119 mm with a grand mean of 94 mm. At the end of the third year of growth, lengths ranged from 109 to 150 mm with a grand mean of 120 mm. At the end of the fourth year of growth, lengths ranged from 134 to 192 mm with a grand mean of 141 mm. At the end of the fifth year of growth, lengths ranged from 156 to 223 with a grand mean of 189. At the end of the sixth year of growth, lengths averaged 255 mm (Table 3.88).

Mean lengths, weights and condition factors for each age class of cutthroat trout in Alder Creek are listed in Table 3.89. Mean condition factors ranged from 0.9 for 2+,3+,4+ and 6+ fish to 1.1 for 0+ fish with an overall mean of 0.9.

A total of 357 scales were collected from eastern brook trout in Alder Creek for age determination. Back-calculated lengths for eastern brook trout at the first annulus ranged from 54 to 63 mm with a grand mean of 60 mm. At the end of the second year of growth, lengths ranged from 82 to 99 mm with a grand mean of 92 mm. At the end of the third year of growth, lengths ranged from 126 to 135 mm with a grand mean of 129 mm. At the end of the fourth year of growth, lengths ranged from 158 to 171 mm with a grand mean of 159 mm. At the end of the fifth years growth, lengths ranged from 189 to 203 mm with a grand mean of 191 mm. At the end of the sixth year of growth lengths were calculated at 239 mm (Table 3.90).

Mean lengths, weights, and condition factors for each age class of eastern brook trout in Alder Creek are listed in Table 3.91. Mean condition factors ranged from 0.9 for 1+ and 2+ fish to 1.1 for 0+ fish, with an overall mean of 0.9 mm.

**Table 3.86 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Evans Creek, 1994.**

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |            |                    |                    |                     |                     |                     |                     |
|--|------------|--------------------|--------------------|---------------------|---------------------|---------------------|---------------------|
| <b>Cohort</b>  | <b>N</b>   | <b>1</b>           | <b>2</b>           | <b>3</b>            | <b>4</b>            | <b>5</b>            | <b>6</b>            |
| 1992   | 74         | 78.2 ± 12.6        |                    |                     |                     |                     |                     |
| 1991   | 93         | 67.9 ± 8.5         | 94.4 ± 13.9        |                     |                     |                     |                     |
| 1990   | 101        | 69.7 ± 17.3        | 93.6 ± 22.9        | 114.4 ± 22.0        |                     |                     |                     |
| 1989   | 78         | 69.9 ± 12.6        | 92.8 ± 16.0        | 114.6 ± 23.7        | 137.4 ± 23.1        |                     |                     |
| 1988   | 17         | 71.5 ± 12.6        | 93.9 ± 14.2        | 116.3 ± 17.7        | 139.2 ± 22.5        | 164.1 ± 27.9        |                     |
| 1987   | 2          | 72.3 ± 17.4        | 97.4 ± 26.1        | 122.6 ± 34.8        | 148.0 ± 43.4        | 167.6 ± 44.8        | 192.3 ± 39.4        |
| <b>GRAND MEAN</b>                                    | <b>365</b> | <b>71.1 ± 13.7</b> | <b>93.7 ± 18.0</b> | <b>114.7 ± 22.3</b> | <b>137.9 ± 23.1</b> | <b>164.5 ± 28.4</b> | <b>192.3 ± 39.4</b> |
| <b>MEAN ANNUAL GROWTH INCREMENT</b>                  |            | 71                 | 23                 | 21                  | 23                  | 27                  | 27                  |

**Table 3.87 Mean lengths, weights and condition factors for each age class of cutthroat trout in Evans Creek, 1994.**

| <b>Age</b> | <b>N</b>  | <b>Mean weight (g) ±SD</b> | <b>Mean length (mm) ±SD</b> | <b>Mean K<sub>II</sub> ±SD</b> |
|------------|-----------|----------------------------|-----------------------------|--------------------------------|
| <b>0+</b>  | <b>19</b> | <b>65.1 ± 12.6</b>         | <b>2.9 ± 2.0</b>            | <b>1.0 ± 0.4</b>               |
| <b>1+</b>  | 74        | 94.3 ± 12.7                | 7.8 ± 5.7                   | 0.9 ± 0.1                      |
| <b>2+</b>  | 84        | 113.2 ± 15.4               | 13.4 ± 7.1                  | 0.9 ± 0.2                      |
| <b>3+</b>  | 98        | 129.1 ± 21.9               | 21.5 ± 14.1                 | 0.9 ± 0.1                      |
| <b>4+</b>  | <b>76</b> | <b>150.3 ± 24.8</b>        | <b>36.2 ± 21.4</b>          | 1.0 ± 0.2                      |
| <b>5+</b>  | 17        | 175.8 ± 28.8               | 61.5 ± 41.4                 | <b>1.0 ± 0.2</b>               |
| <b>6+</b>  | 2         | 207.5 ± 47.4               | 112.5 ± 95.5                | 1.1 ± 0.3                      |
| Total      | 368       |                            |                             | <b>0.9 ± 0.5</b>               |

**Table 3.88** Mean back-calculated lengths at the end of each years growth (**annulus** formation) for each age class of cutthroat trout in Alder Creek, 1994.

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |            |                    |                    |                     |                     |                     |              |
|--|------------|--------------------|--------------------|---------------------|---------------------|---------------------|--------------|
| <b>Cohort</b>  | <b>N</b>   | <b>1</b>           | <b>2</b>           | <b>3</b>            | <b>4</b>            | <b>5</b>            | <b>6</b>     |
| 1992   | 14         | 62.2 ± 8.8         |                    |                     |                     |                     |              |
| 1991   | 42         | 64.0 ± 7.8         | 97.9 ± 9.7         |                     |                     |                     |              |
| 1990   | 82         | 68.1 ± 20.9        | 99.7 ± 26.1        | 123.9 ± 22.3        |                     |                     |              |
| 1989   | 72         | 63.9 ± 10.9        | 90.3 ± 14.8        | 118.7 ± 18.3        | 141.5 ± 24.8        |                     |              |
| 1988   | 16         | 58.3 ± 10.5        | 80.4 ± 17.4        | 109.1 ± 26.5        | 134.4 ± 30.5        | 155.5 ± 34.2        |              |
| 1987   | 1          | 77.1               | 118.9              | 150.2               | 191.9               | 223.2               | 254.6        |
| <b>GRAND MEAN</b>                                    | <b>227</b> | <b>65.0 ± 15.0</b> | <b>94.3 ± 21.0</b> | <b>120.4 ± 21.7</b> | <b>140.6 ± 26.4</b> | <b>189.1 ± 36.7</b> | <b>254.6</b> |
| <b>MEAN ANNUAL GROWTH INCREMENT</b>                  |            | <b>65</b>          | <b>29</b>          | <b>26</b>           | <b>21</b>           | <b>48</b>           | <b>66</b>    |

**Table 3.89** Mean lengths, weights and condition factors for each age class of cutthroat trout in Alder Creek, 1994.

| <b>Age</b> | <b>N</b> | <b>Mean weight(g) ±SD</b> | <b>Mean length (mm) ±SD</b> | <b>Mean K<sub>tt</sub> ±SD</b> |
|------------|----------|---------------------------|-----------------------------|--------------------------------|
| 0+         | 3        | 64.3 ± 3.8                | 3.0 ± 1.0                   | 1.1 io.3                       |
| 1+         | 8        | 71.8 ± 8.2                | 3.5 ± 1.1                   | 1.0 ± 0.2                      |
| 2+         | 36       | 118.6 ± 12.7              | 15.3 ± 7.0                  | 0.9 ± 0.3                      |
| 3+         | 80       | 142.3 ± 18.9              | 26.4 ± 12.6                 | 0.9 ± 0.2                      |
| 4+         | 72       | 150.9 ± 27.8              | 34.5 ± 27.5                 | 0.9 ± 0.1                      |
| 5+         | 16       | 163.0 ± 33.6              | 51.4 ± 41.7                 | 1.0 ± 0.2                      |
| 6+         | 1        | 265                       | 170                         | 0.9                            |
| Total      | 225      |                           |                             | 0.9 ± 0.2                      |

Table 3.90 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of eastern brook trout in Alder Creek, 1994.

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |            |                    |                    |                     |                     |                     |                     |
|--|------------|--------------------|--------------------|---------------------|---------------------|---------------------|---------------------|
| <b>Cohort</b>  | <b>N</b>   | <b>1</b>           | <b>2</b>           | <b>3</b>            | <b>4</b>            | <b>5</b>            | <b>6</b>            |
| 1992   | 52         | 54.8 ± 8.6         |                    |                     |                     |                     |                     |
| 1991   | 77         | 56.8 ± 13.6        | 81.7 ± 21.9        |                     |                     |                     |                     |
| 1990   | 106        | 62.6 ± 16.3        | 99.3 ± 24.1        | 131.1 ± 33.3        |                     |                     |                     |
| 1989   | 76         | 61.4 ± 13.2        | 92.9 ± 18.0        | 127.2 ± 31.4        | 157.7 ± 28.8        |                     |                     |
| 1988   | 38         | 60.0 ± 11.4        | 91.8 ± 14.7        | 126.1 ± 21.9        | 157.7 ± 26.0        | 188.6 ± 29.1        |                     |
| 1987   | 8          | 60.7 ± 11.9        | 93.4 ± 15.6        | 134.7 ± 19.2        | 170.6 ± 21.9        | 203.1 ± 23.0        | 239.1 ± 26.3        |
| <b>GRAND MEAN</b>                                    | <b>357</b> | <b>59.6 ± 13.8</b> | <b>92.2 ± 21.8</b> | <b>129.1 ± 30.4</b> | <b>158.6 ± 27.6</b> | <b>191.4 ± 28.4</b> | <b>239.1 ± 26.3</b> |
| <b>MEAN ANNUAL GROWTH INCREMENT</b>                  |            | <b>60</b>          | <b>32</b>          | <b>37</b>           | <b>30</b>           | <b>32</b>           | <b>47</b>           |

Table 3.91 Mean lengths, weights and condition factors for each age class of eastern brook trout in Alder Creek, 1994.

| <b>Age</b>   | <b>N</b>   | <b>Mean length (mm) ±SD</b> | <b>Mean weight (g) ±SD</b> | <b>Mean K<sub>f</sub> ±SD</b> |
|--------------|------------|-----------------------------|----------------------------|-------------------------------|
| 0+           | 3          | 30.7 ± 8.4                  | 2.7 ± 1.5                  | 1.1 ± 0.3                     |
| 1+           | 52         | 71.2 ± 9.4                  | 3.4 ± 1.7                  | 0.9 ± 0.3                     |
| 2+           | 77         | 98.8 ± 30.6                 | 10.8 ± 14.3                | 0.9 ± 0.2                     |
| 3+           | 104        | 147.5 ± 31.0                | 35.3 ± 26.9                | 1.0 ± 0.2                     |
| 4+           | 74         | 173.3 ± 25.4                | 53.9 ± 25.5                | 1.0 ± 0.2                     |
| 5+           | 38         | 199.6 ± 31.4                | 83.8 ± 35.9                | 1.0 ± 0.3                     |
| 6+           | 38         | 244.4 ± 19.3                | 150.9 ± 28.4               | 1.0 ± 0.2                     |
| <b>Total</b> | <b>354</b> |                             |                            | <b>0.9 ± 0.2</b>              |

### 3.2.6.2. 1994

#### Lake Creek

A total of 730 scales were collected from cutthroat trout in Lake Creek for age determination. Back-calculated lengths for cutthroat trout at the first annulus ranged from 51 to 89 mm with a grand mean of 71.4 mm (Table 3.92). At the end of the second years growth, lengths ranged from 88 to 143 mm with a grand mean of 114 mm. At the end of the third years growth, lengths ranged from 121 to 211 mm with a grand mean of 152 mm. At the end of the fourth years growth lengths ranged from 175 to 258 with a grand mean of 231 mm. At the end of the fifth year of growth, length: ranged from 261 to 321 with a grand mean of 277 mm. At the end of the sixth year of growth lengths ranged from 309 mm to 328 mm with a grand mean of 327 mm. At the end of the seventh year of growth lengths averaged 351 mm.

Mean lengths, weights and condition factors for each age class of cutthroat trout in Lake Creek are listed in Table 3.93. Mean condition factors ranged from 0.8 for 2+ age classes to 1.3 for 0+ age classes with an overall mean of 0.9.

#### Benewah Creek

A total of 681 scales were collected from cutthroat trout in Benewah Creek for age determination during 1994. Back-calculated lengths for cutthroat trout at the first annulus ranged from 62.5 mm to 68 mm with a grand mean of 65 mm. At the end of the second years growth, lengths ranged from 108 mm to 115 mm with a grand mean of 112 mm. At the end of the third years of growth lengths ranged from 149 mm to 163 mm with a grand mean of 155 mm. At the end of the fourth year of growth, lengths ranged from 184 mm to 206 mm with a grand mean of 199 mm. At the end of the fifth year of growth, lengths ranged from 264 mm to 280 mm with a grand mean of 279 mm. At the end of the sixth year of growth, lengths ranged from 321 mm to 337 mm with a grand mean of 337 mm. At the end of the seventh year of growth lengths averaged 342 mm (Table 3.94)

Mean lengths, weights and condition factors for each age class of cutthroat trout in Benewah Creek are listed in Table 3.95. Mean condition factors ranged from 0.8 for 2+ age classes to 1.0 for 5+ age classes with an overall mean of 1.0.

A total of 100 scales were collected from eastern brook trout in Benewah Creek for age determination during 1994. Back-calculated lengths for eastern brook trout at the first annulus ranged from 57 mm to 66 mm with a grand mean of 63 mm. At the end of the second year of growth, lengths ranged from 112 mm to 117 mm with an grand mean of 115 mm. At the end of the third year of



Table 3.92 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Lake Creek, 1994.

| MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS |     |             |              |              |               |              |              |              |
|---|-----|-------------|--------------|--------------|---------------|--------------|--------------|--------------|
| Cohort  | N   | 1           | 2            | 3            | 4             | 5            | 6            | 7            |
| 1993  | 56  | 63.8±9.9    |              |              |               |              |              |              |
| 1992  | 168 | 71.1 ±5.3   | 113.5 ± 9.3  |              |               |              |              |              |
| 1991  | 273 | 50.9±18.2   | 88.4 ± 20.8  | 120.8 ± 18.5 |               |              |              |              |
| 1990  | 20  | 75.3 ± 6.5  | 118.5 ± 9.9  | 154.2 ± 13.9 | 174.8 ± 37.6  |              |              |              |
| 1989  | 4   | 89.4 ± 24.1 | 143.1 ± 37.3 | 211.2 ± 46.2 | 258.4 ± 46.0  | 320.8 ± 41.5 |              |              |
| 1988  | 200 | 76.5 ± 8.2  | 121.8 ± 13.2 | 175.0 ± 19.9 | 237.4 ± 131.4 | 277.3 ± 26.3 | 327.6 ± 29.7 |              |
| 1987  | 9   | 71.8 ± 4.6  | 118.6 ± 15.8 | 166.4 ± 14.4 | 212.0 ± 13.0  | 260.7 ± 20.5 | 308.7 ± 22.7 | 350.8 ± 24.6 |
| GRAND MEAN                                    | 730 | 71.4 ± 7.9  | 114.2 ± 12.3 | 151.9 ± 27.6 | 231.4 ± 123.7 | 277.4 ± 27.2 | 326.8 ± 29.7 | 350.8±24.6   |
| MEAN ANNUAL GROWTH INCREMENT                  |     | 71          | 43           | 38           | 79            | 46           | 50           | 24           |

Table 3.93 Mean lengths, weights and condition factors for each age class of cutthroat trout in Lake Creek, 1994.

| Age   | N   | Mean weight (g) ±SD | Mean length (mm) ±SD | Mean K <sub>II</sub> ±SD |
|-------|-----|---------------------|----------------------|--------------------------|
| 0+    | 3   | 1.7 ± 0.6           | 50.0 ± 6.2           | 1.3 ± 0.1                |
| 1+    | 58  | 3.8 ± 1.7           | 69.8 ± 13.1          | 1.1 ± 0.4                |
| 2+    | 168 | 14.3 ± 3.8          | 118.7 ± 8.4          | 0.8 ± 0.2                |
| 3+    | 273 | 22.9 ± 6.1          | 136.6 ± 10.9         | 0.9 ± 0.2                |
| 4+    | 20  | 52.9 ± 19.9         | 182.4 ± 20.4         | 0.9 ± 0.2                |
| 5+    | 4   | 284.0 ± 93.1        | 325.5 ± 42.7         | 0.8 ± 0.2                |
| 6+    | 200 | 299.1 ± 79.2        | 333.4 ± 21.5         | 0.8 ± 0.2                |
| 7+    | 9   | 430.6 ± 142.9       | 359.0 ± 22.7         | 0.9 ± 0.2                |
| Total | 733 |                     |                      | 0.9                      |

**Table 3.94 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Benawah Creek, 1994.**

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |          |            |              |              |              |              |               |               |
|--|----------|------------|--------------|--------------|--------------|--------------|---------------|---------------|
| <b>Cohort</b>  | <b>N</b> | <b>1</b>   | <b>2</b>     | <b>3</b>     | <b>4</b>     | <b>5</b>     | <b>6</b>      | <b>7</b>      |
| 1993   | 17       | 63.8 ± 9.4 |              |              |              |              |               |               |
| 1992   | 214      | 64.6 ± 5.9 | 112.0 ± 10.1 |              |              |              |               |               |
| 1991   | 218      | 63.9 ± 5.1 | 111.4 ± 8.1  | 148.9 ± 12.5 |              |              |               |               |
| 1990   | 49       | 63.9 ± 7.8 | 110.6 ± 10.5 | 150.8 ± 13.1 | 183.6 ± 14.1 |              |               |               |
| 1989   | 9        | 66.4 ± 6.2 | 114.5 ± 13.4 | 159.5 ± 17.7 | 199.9 ± 21.3 | 261.3 ± 23.2 |               |               |
| 1988   | 161      | 67.9 ± 4.9 | 114.8 ± 7.7  | 163.4 ± 12.8 | 205.6 ± 16.4 | 279.9 ± 19.6 | 337.1 ± 20.54 |               |
| 1987   | 65       | 62.5 ± 1.8 | 108.3 ± 6.9  | 148.9 ± 6.0  | 189.9 ± 7.1  | 264.1 ± 12.6 | 321.2 ± 18.2  | 342.26 ± 51.8 |
| GRAND MEAN   | 681      | 65.1 ± 5.9 | 112.3 ± 9.1  | 154.6 ± 14.4 | 198.7 ± 17.9 | 278.5 ± 20.2 | 336.6 ± 20.6  | 342.3 ± 51.8  |
| MEAN ANNUAL GROWTH INCREMENT                         |          | 65         | 47           | 43           | 44           | 80           | 58            | 5             |

122

**Table 3.95 Mean lengths, weights and condition factors for each age class of cutthroat trout in Benawah Creek, 1994.**

| <b>Age</b> | <b>N</b> | <b>Mean length (mm) ±SD</b> | <b>Mean weight (g) ±SD</b> | <b>Mean K<sub>ij</sub> ±SD</b> |
|------------|----------|-----------------------------|----------------------------|--------------------------------|
| 0+         | 2        | 55.5 ± 3.5                  | 3.0 ± 0.0                  | 1.78 i 0.34                    |
| 1+         | 16       | 76.2 i 13.3                 | 4.5 ± 2.0                  | 1.00 ± 0.25                    |
| 2+         | 80       | 120.4 ± 11.5                | 14.1 ± 5.2                 | 0.78 i 0.17                    |
| 3+         | 211      | 154.9 i 16.4                | 31.4 i 12.7                | 0.83 ± 0.21                    |
| 4+         | 49       | 186.9 i 15.2                | 58.2 ± 17.4                | 0.87 k 0.16                    |
| 5+         | 10       | 271.3 i 25.7                | 191.0 i 56.6               | 1.04 ± 0.69                    |
| 6+         | 169      | 339.0 ± 22.9                | 345.5 k 65.3               | 0.89 k 0.22                    |
| 7+         | 5        | 379.4 ± 20.7                | 450.4 ± 137.2              | 0.81 f 0.12                    |
| Total      | 542      |                             |                            | 1                              |

growth, lengths averaged 158 mm. At the end of the fourth year of growth, lengths averaged 202 mm (Table 3.96).

Mean lengths, weights and condition factors for each age class of eastern brook trout in Benewah Creek are listed in Table 3.97. Mean overall condition factors were 0.9.

### **Evans Creek**

A total of 221 scales were collected from cutthroat trout in Evans Creek for age determination during 1994. Back-calculated lengths for cutthroat trout at the first **annulus** ranged from 55 mm to 61 mm with a grand mean of 56 mm. At the end of the second year of growth, lengths ranged from 109 mm to 121 mm with a grand mean of 111 mm. At the end of the third year of growth, lengths ranged from 167 mm to 171 mm with a grand mean of 163 mm. At the end of the fourth year of growth lengths averaged 221 mm (Table 3.98).

Mean lengths, weights, and condition factors for each age class of cutthroat trout in Evans Creek are listed in Table 3.99. Mean condition factors for cutthroat trout ranged from 0.7 for 3+ age classes to 1.1 for 4+ age classes with an overall mean of 0.9.

### **Alder Creek**

A total of 92 scales were collected from cutthroat trout in Alder Creek for age determination during 1994. Back-calculated lengths for cutthroat trout at the first **annulus** ranged from 60 mm to 65 mm with a grand mean of 62 mm. At the end of the second year of growth lengths ranged from 108 mm to 119 mm with a grand mean of 111 mm. At the end of the third year of growth, lengths ranged from 151 mm to 161 mm with a grand mean of 158 mm. At the end of the fourth years of growth, lengths averaged 198 mm (Table 3.100).

Mean lengths, weights, and condition factors for each age class of cutthroat trout in Alder Creek are listed in Table 3.101. Mean condition factors for cutthroat ranged from 0.8 for 2+ and 3+ age classes to 1.6 for 0+ age class with an overall mean of 0.9.

A total of 452 scales were collected from eastern brook trout in Alder Creek for age determination during 1994. Back-calculated lengths for eastern brook trout at the first **annulus** ranged from 54 mm to 62 mm with a grand mean of 59 mm. At the end of the second years of growth, average lengths ranged from 95 mm to 104 mm with a grand mean of 101 mm. At the end of the third years of growth lengths ranged. from 150 mm to 159 mm with a grand mean of 154 mm. At the end of the fourth years of growth, lengths averaged 199 mm (Table 3.102).

**Table 3.96 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of eastern Brook trout in Benewah Creek, 1994.**

| <b>MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS</b> |            |                   |                    |                     |                      |
|--|------------|-------------------|--------------------|---------------------|----------------------|
| <b>Cohort</b>  | <b>N</b>   | <b>1</b>          | <b>2</b>           | <b>3</b>            | <b>4</b>             |
| 1993   | 38         | 66.3 ± 8.0        |                    |                     |                      |
| 1992   | 24         | 60.2 ± 6.9        | 117.2 ± 7.9        |                     |                      |
| 1991   | 31         | 60.8 ± 5.1        | 114.1 ± 8.4        | 158.9 ± 14.0        |                      |
| 1990   | 10         |                   |                    |                     | 202.0 ± 14.9         |
| <b>GRAND MEAN</b>                                    | <b>100</b> | <b>57.1 ± 5.7</b> | <b>111.9 ± 9.4</b> | <b>158.2 ± 8.9</b>  | <b>202.0 ± 14.9</b>  |
| <b>MEAN ANNUAL</b>                                   |            | <b>62.5 ± 7.4</b> | <b>115.1 ± 8.4</b> | <b>158.7 ± 13.1</b> | <b>202.0 ± 14.99</b> |
| <b>GROWTH INCREMENT</b>                              |            | <b>71</b>         | <b>52</b>          | <b>43</b>           | <b>A3</b>            |

**Table 3.97 Mean lengths, weights and condition factors for each age class of eastern Brook trout in Benewah Creek, 1994.**

| <b>Age</b>   | <b>N</b>   | <b>Mean length (mm) ±SD</b> | <b>Mean weight (g) ±SD</b> | <b>Mean K<sub>f</sub> ±SD</b> |
|--------------|------------|-----------------------------|----------------------------|-------------------------------|
| <b>0+</b>    | 2          | 62 ± 16.9                   | 3.0 ± 1.4                  | 1.3 ± 0.4                     |
| <b>1+</b>    | 39         | 76.6 ± 7.6                  | 33 ± 1.3                   | 0.9 ± 0.2                     |
| <b>2+</b>    | 23         | 127.7 ± 9.7                 |                            | 0.9 ± 0.2                     |
| <b>3+</b>    | 32         | 165.3 ± 17.9                | 43.6 ± 17.0                | 0.9 ± 0.2                     |
| <b>4+</b>    | 7          | 210 ± 19.2                  | 94.6 ± 26.2                | 1.0 ± 0.2                     |
| <b>Total</b> | <b>103</b> |                             |                            | <b>0.91 ± 0.17</b>            |

Table 3.98 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Evans Creek, 1994.

| Cohort                       | MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS |            |              |              |              |
|------------------------------|---|------------|--------------|--------------|--------------|
|                              | N   | 1          | 2            | 3            | 4            |
| 1993                         | 46  | 60.75±9.56 |              |              |              |
| 1992                         | 124   | 54.5 ±7.34 | 108.7 ± 15.9 |              |              |
| 1991                         | 36  | 57.9±4.7   | 117.3 ± 9.5  | 166.8 ± 96   |              |
| 1990                         | 15  | 59.4 ± 4.4 | 120.6 ± 11.7 | 171.1 ± 12.5 | 221.4 ± 22.3 |
| GRAND MEAN                   | 221   | 56.4± 11.2 | 111.0 ± 20.1 | 162.9 ± 29.5 | 221.4 ± 22.3 |
| MEAN ANNUAL GROWTH INCREMENT |   | 56         | 55           | 52           | 58           |

Table 3.99 Mean lengths, weights and condition factors for each age class of cutthroat trout in EvansCreek, 1994.

| Age   | N   | Mean length (mm) ±SD | Mean weight (g) ±SD | Mean Ktl ±SD       |
|-------|-----|----------------------|---------------------|--------------------|
| 0+    | 14  | 59.1 5.8             | 2.1 0.7             | 1.03 0.31          |
| 1+    | 45  | <b>81.9 ± 10.0</b>   | 5.1 ± 1.9           | 0.90 ± 0.19        |
| 2+    | 122 | <b>120.8 ± 17.9</b>  | 16.0i8.8            | <b>0.85 ± 0.14</b> |
| 3+    | 33  | 169.0i14.1           | 45.65i15.26         | 0.72 it.11         |
| 4+    | 13  | 232.4 k23.8          | 136.1 f46.8         | 1.07 k0.25         |
| Total | 272 |                      |                     | <b>0.89 ± 0.18</b> |

**Table 3.100 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of cutthroat trout in Alder Creek, 1994.**

| Cohort                       | MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS |            |             |              |              |
|------------------------------|---|------------|-------------|--------------|--------------|
|                              | N   | 1          | 2           | 3            | 4            |
| 1993                         | 23  | 62.4 ± 5.8 |             |              |              |
| 1992                         | 47  | 60.0 ± 4.7 | 108.1 ± 9.1 |              |              |
| 1991                         | 16  | 64.8 ± 3.4 | 118.8 ± 9.4 | 161.0 ± 15.2 |              |
| 1990                         | 6   | 61.6 ± 1.6 | 110.4 ± 3.6 | 150.7 ± 6.2  | 198.0 ± 7.14 |
| GRAND MEAN                   | 92  | 61.5 ± 4.9 | 110.7 ± 9.8 | 158.2 ± 14.0 | 198.0 ± 7.14 |
| MEAN ANNUAL GROWTH INCREMENT |   | 62         | 49          | 47           | 40           |

**Table 3.101 Mean lengths, weights and condition factors for each age class of cutthroat trout in Alder Creek, 1994.**

| Age   | N   | Mean length (mm) ±SD | Mean weight (g) ±SD | Mean K <sub>tl</sub> ±SD |
|-------|-----|----------------------|---------------------|--------------------------|
| 0+    | 5   | 58.0 ± 4.2           | 3.0 ± 0.0           | 1.58 ± 0.34              |
| 1+    | 28  | 76.1 ± 11.3          | 4.5 ± 2.4           | 0.96 ± 0.24              |
| 2+    | 108 | 121.1 ± 13.2         | 15.3 ± 5.6          | 0.83 ± 0.16              |
| 3+    | 41  | 165.7 ± 17.3         | 40.2 ± 16.1         | 0.84 ± 0.13              |
| 4+    | 14  | 204.8 ± 15.2         | 77.2 ± 18.7         | 0.89 ± 0.13              |
| 6+    | 1   | 325                  | 326                 | 0.95                     |
| Total | 197 |                      |                     | 0.88 ± 0.21              |

**Table 3.102 Mean back-calculated lengths at the end of each years growth (annulus formation) for each age class of eastern Brook trout in Alder Creek, 1994.**

**MEAN ± S.D. BACK CALCULATED LENGTH AT ANNULUS**

| <b>Cohort</b>                       | <b>N</b>   | <b>1</b>           | <b>2</b>            | <b>3</b>            | <b>4</b>            |
|-------------------------------------|------------|--------------------|---------------------|---------------------|---------------------|
| 1993                                | 110        | 53.9 ± 6.7         |                     |                     |                     |
| 1992                                | 158        | 58.8 ± 6.9         | 95.5 ± 12.6         |                     |                     |
| 1991                                | 114        | 59.4 ± 6.7         | 94.8 ± 14.9         | 150.1 ± 13.6        |                     |
| 1990                                | 70         | 61.9 ± 7.9         | 103.6 ± 16.6        | 158.9 ± 18.5        | 198.4 ± 20.9        |
| <b>GRAND MEAN</b>                   | <b>452</b> | <b>59.0 ± 17.3</b> | <b>100.9 ± 52.1</b> | <b>153.6 ± 16.1</b> | <b>198.6 ± 20.7</b> |
| <b>MEAN ANNUAL GROWTH INCREMENT</b> |            | 59                 | 42                  | 53                  | 45                  |

**Table 3.103 Mean lengths, weights and condition factors for each age class of eastern Brook trout in Alder Creek, 1994.**

| <b>Age</b>   | <b>N</b>   | <b>Mean length (mm) ±SD</b> | <b>Mean weight (g) ±SD</b> | <b>Mean K<sub>f</sub> ±SD</b> |
|--------------|------------|-----------------------------|----------------------------|-------------------------------|
| 0+           | 4          | 53.3 ± 5.4                  | 1.75 ± 0.5                 | 1.15 ± 0.24                   |
| 1+           | 108        | 73.02 ± 8.6                 | 4.16 ± 1.61                | 1.05 ± 0.23                   |
| 2+           | 158        | 127.9 ± 16.7                | 19.4 ± 8.3                 | 0.9 ± 0.21                    |
| 3+           | 114        | 166.0 ± 14.1                | 43.0 ± 13.4                | 0.9 ± 0.12                    |
| 4+           | 70         | 213.3 ± 20.4                | 101.9 ± 40.7               | 1.02 ± 0.23                   |
| <b>Total</b> | <b>454</b> |                             |                            | <b>0.96 ± 0.22</b>            |

Mean lengths, weights, and condition factors for each age class of eastern brook trout in Alder Creek are listed in Table 3.103. Mean condition factors for eastern brook trout ranged from 0.9 for 2+ and 3+ ages classes to 1.2 for 0+ age class with an overall mean of 1.0.

### **3.3. Water Quality and Quantity Analysis**

Stream discharge measurements, monthly temperature profiles, and seasonal water quality analysis were completed between March through November 1993 and 1994. Stream discharge measurements were collected monthly except for March. During March, discharge measurements were collected three times to account for rain-or snow events. Sampling after March was completed monthly but was event oriented until base flow was established in June. Sampling was then conducted the third week of the month. Temperature profiles were determined on same schedule as discharge measurements. Water quality samples were collected and analyzed monthly and reported as seasonal averages. Spring samples were those samples collected March through May. Summer samples were those samples collected June through August and fall samples were those samples collected September through November, 1993. Precipitation were also obtained for the months March through November and compared to recorded discharge measurements to show the relationship between precipitation and run-off. Precipitation and average maximum air temperatures were obtained from the Plummer field station and the St. Maries field station. Plummer data was used to determine Lake and Benewah creek conditions, and the St. Maries field station was used to approximate the conditions for Alder and Evans creeks, respectively.

Water quality data was analyzed monthly and reported seasonally. Spring samples were those samples collected March through May. Summer samples were those samples collected June through August, and Fall samples were those samples collected September through November, 1993. For monthly water quality values reference Appendix E.

#### **3.3.1. 1993**

Tables 3.104, 3.105, 3.106, and 3.107 shows the calculated discharge measurements, observed temperatures data, recorded precipitation data and average maximum ambient air temperature for Lake, Benewah, Evans, and Alder creeks, respectively during 1993.

The 1993 field season was more of a usual precipitation year for the panhandle of Idaho. Snow packs in higher elevation were average while snow packs in the lower elevation drainages were greater than normal percent. In March three-foot snow packs in lower elevation drainages resulted in higher than



**Table 3.104 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Lake Creek, 1993.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-09-93</b> | <b>41</b>                  | <b>1.8</b>                        | <b>33</b>                              | <b>36</b>                           |
| <b>03-16-93</b> | <b>57</b>                  | <b>0.2</b>                        | <b>38</b>                              | <b>46</b>                           |
| <b>03-26-93</b> | <b>700</b>                 | <b>1.0</b>                        | <b>33</b>                              | <b>48</b>                           |
| <b>04-15-93</b> | <b>51</b>                  | <b>0.8</b>                        | <b>43</b>                              | <b>53</b>                           |
| <b>05-05-93</b> | <b>840</b>                 | <b>2.8</b>                        |  | <b>55</b>                           |
| <b>06-22-93</b> | <b>4</b>                   | <b>3.7</b>                        | <b>56</b>                              | <b>73</b>                           |
| <b>07-23-93</b> | <b>6</b>                   | <b>4.1</b>                        | <b>60</b>                              | <b>64</b>                           |
| <b>08-23-93</b> | <b>1</b>                   | <b>0.7</b>                        | <b>57</b>                              | <b>79</b>                           |
| <b>09-23-93</b> | <b>2</b>                   | <b>0.8</b>                        | <b>44</b>                              | <b>71</b>                           |
| <b>10-23-93</b> | <b>1</b>                   | <b>1.1</b>                        | <b>36</b>                              | <b>63</b>                           |
| <b>11-19-93</b> |                            | <b>1.0</b>                        |  | <b>48</b>                           |

**Table 3. 105 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Benawah Creek, 1993.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-09-93</b> | <b>32</b>                  | <b>1.8</b>                        | <b>38</b>                              | <b>38</b>                           |
| <b>03-16-93</b> | <b>69</b>                  | <b>0.2</b>                        | <b>40</b>                              | <b>46</b>                           |
| <b>03-26-93</b> | <b>350</b>                 | <b>1.0</b>                        | <b>33</b>                              | <b>48</b>                           |
| <b>04-15-93</b> | <b>86</b>                  | <b>0.8</b>                        | <b>34</b>                              | <b>53</b>                           |
| <b>05-05-93</b> | <b>676</b>                 | <b>2.8</b>                        | <b>34</b>                              | <b>55</b>                           |
| <b>06-22-93</b> | <b>3</b>                   | <b>3.7</b>                        | <b>60</b>                              | <b>73</b>                           |
| <b>07-23-93</b> | <b>12</b>                  | <b>4.1</b>                        | <b>66</b>                              | <b>64</b>                           |
| <b>08-23-93</b> | <b>1</b>                   | <b>0.7</b>                        | <b>53</b>                              | <b>79</b>                           |
| <b>09-23-93</b> | <b>&lt;1</b>               | <b>0.8</b>                        | <b>51</b>                              | <b>71</b>                           |
| <b>10-23-93</b> | <b>2</b>                   | <b>1.1</b>                        | <b>45</b>                              | <b>63</b>                           |
| <b>11-19-93</b> | <b>4</b>                   | <b>1.0</b>                        | <b>33</b>                              | <b>48</b>                           |

**Table 3.106 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Evans Creek, 1993.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-09-93</b> | <b>20</b>                  | <b>2</b>                          | <b>37</b>                              | <b>38</b>                           |
| <b>03-16-93</b> | <b>27</b>                  | <b>1</b>                          | <b>39</b>                              | <b>45</b>                           |
| <b>03-26-93</b> | <b>182</b>                 | <b>1</b>                          | <b>33</b>                              | <b>50</b>                           |
| <b>04-15-93</b> | <b>48</b>                  | <b>2.2</b>                        |  | <b>52</b>                           |
| <b>05-05-93</b> | <b>456</b>                 | <b>3.5</b>                        | <b>34</b>                              | <b>58</b>                           |
| <b>06-22-93</b> | <b>5</b>                   | <b>2.2</b>                        | <b>51</b>                              | <b>75</b>                           |
| <b>07-23-93</b> | <b>7</b>                   | <b>5.5</b>                        | <b>53</b>                              | <b>70</b>                           |
| <b>08-23-93</b> | <b>4</b>                   | <b>0.3</b>                        | <b>53</b>                              | <b>78</b>                           |
| <b>09-23-93</b> | <b>3</b>                   | <b>0.5</b>                        | <b>48</b>                              | <b>71.</b>                          |
| <b>10-23-93</b> | <b>2</b>                   | <b>0.9</b>                        | <b>43</b>                              | <b>66</b>                           |
| <b>11-19-93</b> | <b>7</b>                   | <b>0.9</b>                        | <b>35</b>                              | <b>47</b>                           |

**Table 3.107 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Alder Creek, 1993.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-09-93</b> | 12                         | 2                                 | 36                                     | 38                                  |
| 03-16-93        | 38                         | 1                                 | 37                                     | 45.                                 |
| 03-26-93        | 168                        | 1                                 | 33                                     | 50                                  |
| 04-15-93        | 45                         | 2.2                               | 34                                     | 52                                  |
| 05-05-93        | 289                        | 3.5                               | 34                                     | 58                                  |
| 06-22-93        | 3                          | 2.2                               | 55                                     | 75                                  |
| 07-23-93        | 5                          | 5.5                               | 55                                     | 70                                  |
| 08-23-93        | 2                          | 0.3                               | 55                                     | 78                                  |
| 09-23-93        | 1                          | 0.5                               | 48                                     | 71                                  |
| 10-23-93        | 1                          | 0.9                               | 37                                     | 66                                  |
| 11-19-93        | 2                          | 0.9                               | 33                                     | 47                                  |

**Table 3.108 Seasonal (spring) water quality parameters for selected Coeur d'Alene tributaries during 1993.**

| Location | pH      | D.O.      | Cond.     | Redox   | Alk.  | NO2      | NO3       | PO4     | TSS     | Turbidity |
|----------|---------|-----------|-----------|---------|-------|----------|-----------|---------|---------|-----------|
| Lake     | 6.8-7.2 | 11.8-13.0 | .069-.104 | 332     | 30-40 | .004-.03 | 1.03-3.58 | .36-2.3 | 42-265  | 35-195    |
| Benewah  | 7.0-7.2 | 13.0-13.5 | .042-.060 | 308     | 30-50 | .01-.03  | 0-1.28    | .07-.40 | 1-640   | 40-164    |
| Evans    | 7.0-7.2 | 14.5-22.0 | .023-.056 | 291-302 | 20-30 | 0-.04    | 0         | .18-.91 | 3.0-6.7 | 35->500   |
| Alder    | 6.8-7.2 | .041-.073 | 10.9-16.4 | 289     | 30-40 | .02-.07  | .02-1.3   | .01-.7  | 18-5940 | 27-320    |

\*Total dissolved solids and turbidity not determined.

**Table 3.109 Seasonal (summer) water quality parameters for selected Coeur d'Alene tributaries during 1993.**

| Location | pH      | D.O.     | Cond.     | Redox     | Alk.  | NO2      | NO3      | PO4       | TSS    | Turbidity |
|----------|---------|----------|-----------|-----------|-------|----------|----------|-----------|--------|-----------|
| Lake     | 7.0-7.2 | 1.2-15.0 | .045-.056 | 240-299   | 30    | .01-.06  | 0-0.0    | 0.23-0.44 | 5-280  | 9-75      |
| Benewah  | 7.9-7.4 | 2.2-11.0 | .042-.071 | 260-286   | 30-50 | 0.0-.06  | 0.0      | 0.12-1.43 | 2-640  | 12-87     |
| Evans    | 266-313 | 6.7-7.0  | 10.6-13.8 | .009-.035 | 30    | 0.0-0.08 | 0.0      | 0.0-0.18  | 2-460  | 0-30      |
| Alder    | 270-288 | 6.3-7.2  | 9.6-15.7  | .091-.092 | 30-60 | .01-.03  | 0.0-0.06 | .01-3.0   | 6-5940 | 18-320    |

\*Total dissolved solids and turbidity not determined.

**Table 3.110 Seasonal (fall) water quality parameters for selected Coeur d'Alene tributaries during 1993.**

| Location | pH      | D.O.      | Cond.     | Redox   | Alk.  | NO2      | NO3      | PO4       | TSS   | Turbidity |
|----------|---------|-----------|-----------|---------|-------|----------|----------|-----------|-------|-----------|
| Lake     | 7.0-7.2 | 10.4-15.1 | .060-.061 | 469-942 | 40-50 | 0.0-0.03 | 0.0      | 0.07-1.92 | 10-87 | 18-28     |
| Benewah  | 6.4-7.5 | 10.6-14.8 | .069-.090 | 415-939 | 50    | 0.0-0.03 | 0.0      | 0.0-0.52  | 30    | 12-24     |
| Evans    | 6.7-7.0 | 10.3-13.8 | .000-.035 | 133-910 | 40-50 | 0.0-0.03 | 0.0      | 0.07-1.68 | 2     | 0-12      |
| Alder    | 6.3-7.2 | 9.6-15.7  | .091-.092 | 313-842 | 50-70 | 0.0-0.03 | 0.0-0.04 | 0.01-1.18 | 8     | 18-172    |

\*Total dissolved solids and turbidity not determined.

normal spring discharge profiles. Snow melt began in March when air temperatures began to rise above freezing during daylight hours. This was exacerbated by precipitation in the form of rain during the day, snow at night, which resulted in the first rain-on-snow event. The first rain-on-snow event was recorded on March 23 when discharge in all our drainages spiked to 700, 350, 182, and 168 cfs for Lake, Benewah, Evans, and Alder creeks, respectively. Corresponding drops in discharge was reported as precipitation values dropped. Air and water temperatures continued to increase, while above normal precipitation was received. This resulted in a bankfull flow event for most of the drainages, except for Lake Creek, which reached flood stage. Discharge values peaked at 840, 676, 456, and 289 cfs for Lake, Benewah, Evans, and Alder creeks, respectively. Within a month, discharge values in all drainages had dropped significantly with corresponding increases in ambient air and water temperatures. Base flow was achieved in all drainages in October.

### **Water Quality Analysis**

Table 3.108 shows water quality values for the spring sampling period. All parameters were within acceptable limits for fish except for nitrate, phosphates, total suspended sediments, and turbidity. Spring Nitrate values ranged from 0.0 mg/l in Evans Creek to 3.58 mg/l in Lake Creek. Standards for nitrate are 0.4 mg/l. Phosphate values ranged from 0.01 mg/l in Alder Creek to 2.3 mg/l in Lake Creek, EPA accepted standards for phosphate is 0.15 mg/l. Fish appear relatively indifferent to nitrate and phosphate, however both compounds are associated with eutrophication. Excess amounts of these compounds can result in algal blooms which ultimately may result in fish kills. Turbidity and TSS values ranged from 27 NTU's in Alder Creek to >500 NTU's in Evans Creek. Large quantities of suspended sediments can be carried for short periods of time without detriment to fish. However, long exposures to increased suspended sediments results in lowered primary food production. Streams with less than 25 ppm may be expected to support good fresh water fisheries, while streams with values between 80-400 should not be considered good areas for supporting fresh water fisheries.

Table 3.109 shows water quality values for the summer sampling period. All parameters were within acceptable limits for fish except for phosphate. Phosphate values were above acceptable limits in all four drainages with values ranging from 0.18 mg/l in Evans to 3.0 mg/l in Alder.

Table 3.110 shows water quality values for the fall sampling period. All parameters except for phosphate and TSS were within acceptable limits: Phosphate standard were exceeded in all drainages in the fall with values ranging from 0.52 in Benewah to 1.92 mg/l in Lake Creek. Total suspended sediment values were above accepted limits of 25 NTU's in Lake Creek at 87 NTU's.

### 3.3.2. 1994

Table 3.11 I-3.1 14 shows the calculated discharge measurements, observed temperatures data, recorded precipitation data and average maximum ambient air temperature for Lake, Benewah, Evans, and Alder creeks, respectively during 1994.

Discharge and precipitation values for 1994 were lower than normal. This resulted in lower discharge profiles than those reported in 1993. Base flow was achieved earlier than in 1993. All streams reached base flow levels in August except Evans Creek.

Table 3.108 shows water quality values for the spring through fall sampling period. All parameters were within acceptable ranges for cutthroat trout except for phosphate in Lake Creek during the fall sampling period and turbidity values in all drainages, except Evans Creek during the spring sampling period.

## 3.4. MACROINVERTEBRATE SURVEYS

### 3.4.1. Hess and Drift Samples

A total of 102 Hess samples were collected in Alder, Benewah, Evans, and Lake Creeks in 1994. Mean annual densities of benthic macroinvertebrates ranged from a low of 181 0/m<sup>2</sup> in Lake Creek to a high of 3180/m<sup>2</sup> in Evans Creek. Mean annual densities in Alder Creek and Benewah Creek were 2270/m<sup>2</sup> and 2280/m<sup>2</sup>, respectively (Table 3.118). Macroinvertebrate densities for the study streams are generally comparable or greater than other third order streams in north Idaho. Chironomidae larvae were the most abundant macroinvertebrate in Alder Creek (30.58 percent), Benewah Creek (21.51 percent), and Evans Creek (33.28 percent). Elmidae larvae were the most abundant taxon in Lake Creek (25.22 percent), while aquatic Oligochaetes were the second most abundant taxon (15.24 percent). Heptageniidae were the second most abundant taxon in Alder Creek (11.29 percent), and Evans Creek (15.25 percent), while Baetidae were the second most abundant in Benewah Creek (13.52 percent). These results are similar to those reported in 1991 (Lillengreen, et. al. 1991).

A total of 63 drift samples were collected from study streams in 1994. Mean densities of benthic invertebrates ranged from a low of 53.7/100 m<sup>3</sup> in Benewah Creek, to a high of 249.5/100 m<sup>3</sup> in Alder Creek (Table 3.118). Average densities in Benewah (53.7000 m<sup>3</sup>), Evans (223.41100 m<sup>3</sup>), and Lake Creek (72.0/100 m<sup>3</sup>) were lower than in 1991 (Lillengreen, et al. 1991). while average density in Alder Creek (201.3/100 m<sup>3</sup>) was higher. Chironomidae were the most abundant taxon collected in Alder Creek (43.46 percent) and Evans

**Table 3.111 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Laka Creek, 1994.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-04-94</b> | 42.5                       | 2.7                               | 2.0                                    | 44                                  |
| 03-21-94        | 38.4                       | 2.6                               | 2.4                                    | 49                                  |
| 04-18-94        | 18.9                       | 2.0                               | 9.9                                    | 56                                  |
| 05-16-94        | 6.0                        | 1.1                               | 9.6                                    | 66                                  |
| 06-22-94        | 1.8                        | 5.1                               | 15.9                                   | 67                                  |
| 07-18-94        | 0.3                        | 0.8                               | 20.1                                   | 79                                  |
| 08-29-94        | <0.1                       | 0.1                               | 11.3                                   | 89                                  |
| 09-19-94        | 0.5                        | 0.6                               | 14.9                                   | 78                                  |
| 10-17-94        | 0.6                        | 1.2                               | 5.2                                    | 63.                                 |
| 11-15-94        | 2.9                        | 6.9                               | 1.0                                    | 44                                  |



**Table 3.112 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Benawah Creek, 1994.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-04-94</b> | 71.6                       | 2.7                               | 5.0                                    | 44                                  |
| 03-21-94        | 32.1                       | 2.6                               | 2.0                                    | 49                                  |
| 04-18-94        | 2.9                        | 2.0                               | 16.8                                   | 56                                  |
| 05-16-94        | 7.2                        | 1.1                               | 17.4                                   | 66                                  |
| 06-22-94        | 2.4                        | 5.1                               | 28.1                                   | 67                                  |
| 07-18-94        | 2.2                        | 0.8                               | 23.3                                   | 79                                  |
| 08-29-94        | <0.1                       | 0.1                               | 17.1                                   | 89                                  |
| 09-19-94        | 0.6                        | 0.6                               | 10.2                                   | 78                                  |
| 10-17-94        | 1.3                        | 1.2                               | 7.8                                    | 63.                                 |
| 11-15-94        | 7.5                        | 6.9                               | 1.7                                    | 44                                  |

**Table 3.113 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Evans Creek, 1994.**

| <b>Date</b> | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-------------|----------------------------|-----------------------------------|--|-------------------------------------|
| 03-04-94    | 31.8                       | 1.8                               | 3.5                                    | 44                                  |
| 03-21-94    | 15.7                       | 1.6                               | 3.7                                    | 49                                  |
| 04-18-94    | 32.8                       | 1.2                               | 7.0                                    | 59                                  |
| 05-16-94    | 15.7                       | 1.1                               | 8.2                                    | 68                                  |
| 06-22-94    | 2.6                        | 2.8                               | 12.0                                   | 71                                  |
| 07-18-94    | 3.5                        | 0.4                               | 14.0                                   | 82                                  |
| 08-29-94    | 1.8                        | •                                 | 12.7                                   | 89                                  |
| 09-19-94    | 1.4                        | 0.7                               | 11.6                                   | 78                                  |
| 10-17-94    | 1.3                        | 0.5                               | 6.4                                    | *                                   |
| 11-15-94    | 3.6                        | *                                 | 2.7                                    | •                                   |

**Table 3.114 Stream Discharge, temperature, month to date precipitation values and average maximum air temperature for Alder Creek, 1994.**

| <b>Date</b>     | <b>Discharge<br/>(cfs)</b> | <b>Precipitation<br/>(inches)</b> | <b>Average<br/>Stream Temp.<br/>°F</b> | <b>Average<br/>Air Temp.<br/>°F</b> |
|-----------------|----------------------------|-----------------------------------|--|-------------------------------------|
| <b>03-04-94</b> | 36.3                       | 1.8                               | 32                                     | 44                                  |
| 03-21-94        | 19.0                       | 1.6                               | 1.1                                    | 49                                  |
| 04-18-94        | 14.5                       | 1.2                               | 41                                     | 59                                  |
| 05-16-94        | 9.0                        | 1.1                               | 9.8                                    | 68                                  |
| 06-22-94        | 1.8                        | 2.8                               | •                                      | 71                                  |
| 07-18-94        | •                          | 0.4                               | •                                      | a2                                  |
| 08-29-94        | <0.5                       | •                                 | 16.5                                   | 89                                  |
| 09-19-94        | <0.5                       | 0.7                               | 13.3                                   | 78                                  |
| 10-17-94        | 1.5                        | 0.5                               | 5.7                                    | •                                   |
| 11-15-94        | 2.8                        | •                                 | 0.9                                    | •                                   |

**Table 3.115 Seasonal (spring) water quality parameters for selected Coeur d'Alene tributaries during 1994.**

| Location | pH             | D.O.      | Cond.            | Redox          | Alk.        | NO2       | NO3      | PO4       | TSS       | Turbidity |
|----------|----------------|-----------|------------------|----------------|-------------|-----------|----------|-----------|-----------|-----------|
| Lake     | 6.5-7.4        | 9.4-13.8  | <b>.038-.055</b> | 357-385        | <b>3-40</b> | 0.03      | 0.0-0.06 | 0.01-0.07 | 7         | 9-40      |
| Benewah  | <b>7.0-7.8</b> | 8.8-14.1  | <b>.040-.057</b> | 336-416        | 5-60        | 0.0-0.03  | 0.0-0.0  | 0.01-0.07 | <b>10</b> | 9-30      |
| Evans    | <b>6.8-7.2</b> | 10.9-13.6 | 0.011-0.022      | <b>328-407</b> | 4-50        | 0.02-0.05 | 0.0      | 0.01-0.07 | 6         | 0.5       |
| Alder    | 7.5-7.8        | 11.3-14.0 | <b>.039-.067</b> | 278-379        | <b>5-60</b> | 0.1-0.3   | 0.0-0.0  | 0.0-0.07  | .         | 5-18      |

\*Parameters not determined.

**Table 3.116 Seasonal (summer) water quality parameters for selected Coeur d'Alene tributaries during 1994.**

| Location | pH              | D.O.            | Cond.            | Redox   | Alk.         | NO2              | NO3              | PO4              | TSS | Turbidity |
|----------|-----------------|-----------------|------------------|---------|--------------|------------------|------------------|------------------|-----|-----------|
| Lake     | 7.3-8.9         | 7.7-9.7         | <b>.020-.040</b> | 176-337 | <b>40 to</b> | <b>0.05-0.08</b> | 0.09             | 0.12-0.40        | .   | 9-15      |
| Benewah  | <b>8.3-10.7</b> | 8.2             | <b>.059-.094</b> | 160-366 | 35-50        | .                | 0.01-6.0         | <b>0.07-0.18</b> | *   | 12-50     |
| Evans    | <b>7.0-8.5</b>  | <b>9.2-10.5</b> | <b>0.24-0.28</b> | 279-360 | 35-70        | 0.03-0.07        | 0.0              | 0.01-0.07        | .   | 0-5       |
| Alder    | 6.7-8.2         | 7.4-9.2         | <b>.076-.096</b> | 277-493 | 50-75        | 0.04             | <b>0.05-0.08</b> | 0.07-0.12        | *   | 21-24     |

\*Parameters not determined.

**Table 3.117 Seasonal (fall) water quality parameters for selected Coeur d'Alene tributaries during 1994.**

| Location | pH      | D.O.            | Cond.            | Redox   | Alk.         | NO2              | NO3              | PO4              | TSS | Turbidity |
|----------|---------|-----------------|------------------|---------|--------------|------------------|------------------|------------------|-----|-----------|
| Lake     | 7.0-7.2 | 8.8-11.2        | <b>.050-.080</b> | 366-455 | <b>50-60</b> | 0.03-0.060       | .                | <b>0.18-0.27</b> | .   | 9-18      |
| Benewah  | 7.3-7.9 | <b>9.3-10.8</b> | <b>.062-.090</b> | 338-394 | 70           | .                | <b>0.04-0.06</b> | 0.07-0.12        | .   | 9-12      |
| Evans    | 6.9-7.1 | 9.4-10.6        | <b>.030-.031</b> | 356-466 | 50           | <b>0.03-0.06</b> | 0.0              | 0.01             | .   | 5         |
| Alder    | 7.0-7.2 | 8.3-9.9         | <b>.093-.096</b> | 277-354 | 70           | 0.05             | 0.06             | <b>0.07-0.18</b> | .   | 15-24     |

\*Parameters not determined.

Table 3.118 Mean density of macroinvertebrates from Hess (#/m<sup>2</sup>) and Drift (#/100m<sup>3</sup>) samples, 1994. Sample size in parentheses.

| Hess Samples  |                   |                  |                   |                  |
|---------------|-------------------|------------------|-------------------|------------------|
| Date          | Stream            |                  |                   |                  |
|               | Alder             | Benewah          | Evans             | Lake             |
| June          | 2720 (9)          | 2530 (9)         | 1990 (9)          | 1890 (9)         |
| August        | 1850 (9)          | 890 (9)          | 3100 (9)          | 1100 (6)         |
| October       | 2230 (9)          | 3420 (9)         | 4460 (9)          | 1240 (9)         |
| <b>Annual</b> | <b>2270 (27)</b>  | <b>2280 (27)</b> | <b>3180 (27)</b>  | <b>1810 (21)</b> |
| Drift Samples |                   |                  |                   |                  |
| Date          | Stream            |                  |                   |                  |
|               | Alder             | Benewah          | Evans             | Lake             |
| June          | 52.7 (6)          | 38.7 (6)         | 49.9 (6)          | 89.4 (6)         |
| August        | 624.5 (6)         | 73.4 (4)         | 480.1 (6)         | 24.5 (1)         |
| October       | 71.3 (6)          | 30.9 (6)         | 140.1 (6)         | 18.1 (4)         |
| <b>Annual</b> | <b>249.5 (18)</b> | <b>53.7 (16)</b> | <b>223.4 (18)</b> | <b>72.0 (11)</b> |

Table 3.119 Shannon-Weiner diversity index and Family-level Biotic Index(FBI) for macroinvertebrate communities, 1994.

| Shannon-Weiner   |             |             |             |             |
|------------------|-------------|-------------|-------------|-------------|
|                  | Stream      |             |             |             |
|                  | Alder       | Benewah     | Evans       | Lake        |
| # of Taxa        | 44          | 44          | 43          | 38          |
| # of Individuals | 6122        | 6155        | 8599        | 3806        |
| <b>Index</b>     | <b>2.42</b> | <b>2.53</b> | <b>2.34</b> | <b>2.31</b> |
| FBI              |             |             |             |             |
|                  | Stream      |             |             |             |
|                  | Alder       | Benewah     | Evans       | Lake        |
| June             | 3.45        | 4.32        | 4.54        | 4.33        |
| August           | 5.06        | 4.76        | 4.10        | 4.72        |
| October          | 4.06        |             | 3.00        | 4.75        |
| <b>Annual</b>    | <b>4.19</b> | <b>3.98</b> | <b>3.91</b> | <b>4.17</b> |

Creek (49.30 percent). Elmidae were the second most abundant taxon in Alder Creek (11.00 percent), while Baetidae were second most abundant in Evans Creek (8.64 percent). Baetidae were most abundant in Benewah Creek (27.37 percent), followed by Chironomidae (26.67 percent). The most abundant taxa in Lake Creek were Heptageniidae and Chironomidae (24.25 percent and 23.75 percent, respectively).

### **3.4.2. Biological Assessment**

Shannon-Weiner diversity values were highest in Benewah Creek (2.53) and lowest in Lake Creek (2.31). Values for Alder and Evans Creeks were 2.42 and 2.34, respectively (Table 3.119). Shannon-Weiner diversity values for drift samples ranged from 2.1 in Evans Creek to 2.25 in Benewah Creek. These values were lower than those reported in 1991 (Lillengreen et al. 1991). Terrestrial invertebrates were not included in the 1994 index calculations and may have resulted in lower diversity values.

Average annual Family-level Biotic Index (FBI) values ranged from 3.91 in Evans Creek to 4.19 in Alder Creek. Values were 3.98 for Benewah Creek and 4.17 for Lake Creek, respectively (Table 3.119). These results suggest that water quality ranges from very good (possible slight organic pollution) to good (some organic pollution probable) in the study drainages. Values calculated from drift samples were slightly higher than those for Hess samples. These values fall within the range reported for similar streams in the region (Lester 1994).

### **3.5. Analysis of Migration Barriers**

A total of 19 culverts were analyzed during 1993 to determine if passage barriers or velocity barriers existed. Only one passage barrier existed, while several potential velocity/distance barriers existed. In the Benewah Creek drainage, a culvert connecting a side tributary, Windfall Creek was determined to be a jumping pool distance barrier. Replacement of this culvert would access 2-3 miles of spawning/rearing habitat for westslope cutthroat trout.

## **4.0 Discussion**

Fisheries resources are an integral part of the Coeur d'Alene Tribe's cultural heritage. Anadromous and resident salmonids were a critical component of the tribes annual subsistence requirements. The Coeur d'Alene Tribe, however, lost their salmon fishery early in the twentieth century with the construction of Chief Joseph and Grand Coulee Dams on the Columbia River. These actions taken by non-tribal entities forced the Coeur d'Alene Tribe to rely solely on the resident fish resources of Lake Coeur d'Alene.

Historical evidence suggests that cutthroat trout were an abundant and extremely important part of the Coeur d'Alene Tribe's resident fishery. Cutthroat trout were the most abundant trout species in the Coeur d'Alene system. Large numbers of cutthroat trout were harvested in Lake Coeur d'Alene by boat fishing, as well as trapping in tributaries during spring spawning runs (Walker 1977; Peltier 1975; Scott 1968). Since 1932, the cutthroat population has declined significantly. The present ecosystem bears little resemblance to habitat composition, diversity and structure of the historic ecosystem.

A literature review determined optimal habitat conditions for cutthroat and bull trout (Graves et al. 1990; Lillengreen et al, 1993). These were then compared to existing conditions reported in the study drainages. Data compiled in 1993 and 1994 included: stream discharge rates, temperature, water quality, available habitat, large woody debris densities, substrate and percentage of fine sediment. Biological data included trout population estimates, biomass estimates, stock assessment, and benthic macroinvertebrate densities. All data were combined to determine factors affecting, and ways to increase, westslope cutthroat and bull trout populations on the Reservation.

### **4.1 Physical Characteristics**

#### **4.1.1 Water Quantity and Quality**

In the four study drainages, base flow, temperature, and water quality are limiting the quality of trout habitat. Data collected in 1993 and 1994 indicated that average annual flow values reported were considered poor for maintaining quality trout. Average annual flows of >50 percent are considered excellent for maintaining quality trout habitat, 25-50 percent is considered fair, and base flows less <25 percent is considered poor habitat (Lillengreen et al, 1993; Hickman and Raleigh 1982; Wesche 1980). Lillengreen et al, (1993) reported that average annual flows in 1991 and 1992 were also poor (e.g. Lake Creek values for 1991 and 1992 were 13.4 and 25.2 percent, respectively). Drought conditions have existed during the past four years which may have impacted water yield to streams which may have contributed to these low flow conditions. Data collected indicated that extreme peaks take place during spring run-off

followed by periods of extreme low flows. This can be associated with land use practices in rich upland and riparian land clearing results in loss of water retention, and contributes to the “flashy” nature of these drainages.

Based on water temperature data collected during 1993 and 1994 (Tables 3.103, 3.110) it appears that temperatures may be limiting cutthroat trout survival. (Lillengreen et al, 1993) reported that these conditions also existed in 1991 and 1992. Evans and Alder creeks were within acceptable temperature ranges for westslope cutthroat trout in 1993. However, in 1994, both streams exceeded optimal temperature ranges.

Based on results, water quality data is inconclusive. Phosphate, turbidity, total suspended solids, and dissolved oxygen exceeded EPA approved limits reported for optimal trout habitat conditions for short durations. These values may indicate a problem exists with water quality, however, additional data needs to be collected to determine the extent of the potential problem.

#### **4.1.2 Habitat**

Current habitat conditions for westslope cutthroat trout in Lake, Benawah, Evans and Alder creeks were determined. Parameters used for evaluation included residual pool depth, average canopy cover, number of large woody debris within wetted perimeter of the channel (instream cover); riffle:pool ratio and average percent fine sediment (< 4 mm). All habitat parameters were measured for each stream reach in each of the four drainages.

Pools are important components of fish habitat because they provide cover, protection and play a role in maintaining optimal temperature conditions for trout. Residual pool depth is defined as the depth of the pool during low flow or no flow conditions. Optimal pool depth for trout habitat can be described as pools greater than 1.5 meters in streams less than 5 meters wide or pools greater than 2 meters deep in streams greater than 5 meters wide (Hickman and Raleigh, 1981). All of the study drainages failed to meet the pool depth requirements for optimum trout habitat (Table 4.1).

Canopy cover is an important component of fish habitat because it provides temperature regulation, controls and maintains watershed erosion and streambank integrity. Average canopy cover between 50-100 percent shade is acceptable for trout habitat in streams less than 50 feet wide (Idyll 1942; Martin et. al. 1981). Stream side buffers of approximately 33 meters in which 80 percent of the buffer is well rooted will maintain adequate erosion control (Raleigh and Duff 1981). All mainstem sections of the drainages lacked sufficient riparian buffers for thermal regulation and bank stability, while forested headwater drainages contained sufficient riparian buffers to maintained temperature and bank stability (Table 4.1).



The need for large woody debris is widely accepted in the scientific literature, however optimal levels of woody debris reported are variable. Woody debris plays a significant role in channel geomorphology, (i.e. stream stability/habitat diversity) as well as contributes to overwinter fish habitat and plays an important role in summer juvenile cover (Val. Crispen and Roberts, 1993) Large woody debris is also important in the production of benthic invertebrates. Large woody debris is conspicuously absent in most of the four drainages. Removal of woody debris from stream channels, forest management and agricultural practices, and road construction have all contributed to the current conditions. This results in homogeneous stream reaches dominated by riffles and degraded to boulder-bedrock substrate. (Val Crispen and Roberts, 1993). These conditions are evident throughout the four study drainages. All lower mainstem reaches are devoid of large woody debris, while smaller headwater, drainages in forested areas still retain some level of woody debris (Table 4.1).

Riffle pool ratio can be used as an indicator of habitat diversity. An optimal riffle pool ratio also provides for adequate food production. Habitat diversity (complexity) is important for westslope cutthroat trout, especially juveniles (Griffith, 1970; Rieman et al, 1989). Rieman et al (1989) reported that manipulation of stream complexity by artificially increasing the amount of lateral habitat resulted in a proportional increase in cutthroat numbers. Lere (1982) found westslope cutthroat trout densities were correlated to pool riffle periodicity. Optimal pool-riffle ratios that provide an optimal proportion of rearing and food producing areas are 1:1 (Hynes, 1970; Raleigh et. al. 1982; Rieman and Apperson 1989). Average riffle:pool ratio for the four study drainages did not meet this target value. Most all drainages, with the exception of Alder Creek had riffle:pool ratios of greater than 1 :1. (Table 4.1).

Substrate size and the amount of fine sediment are important to cutthroat trout for spawning success, food production, overwintering and rearing habitat. For successful spawning and reproduction, cutthroat trout require an adequate amount of gravels between 2.0 and 6.0 cm in diameter with less than 10 percent fine sediments. Substrate is also important in overwintering habitat. For optimal winter and escape cover of fry and juveniles, 10 percent of the substrate ranges between 1040 centimeters in diameter (Hickman and Raleigh, 1981). Small fish utilize substrate for hiding and in extreme environmental conditions, such as high velocities and ice formation, fry and subadults burrow into substrate (Everest, 1969, Bjorn et al 1982). Fine sediment also contributes to the reduction of carrying capacity of essential pool habitat and can eventually eliminate pool habitat (Bjom et al 1977; Rhodes and Jone 1991). Average percent fines for all drainages were above the recommended 10 percent fines value (Table 4.1).

**Table 4.1. Current (1994) and optimal habitat conditions for westslope cutthroat trout.**

| Habitat Characteristic               | <i>Lake Creek</i>            | <i>Benewah Creek</i> | <i>Evans Creek</i> | <i>Alder Creek</i> | Optimal Condition |
|--------------------------------------|------------------------------|----------------------|--------------------|--------------------|-------------------|
| Average Residual Pool Depth          | (1993) 0.5 m<br>(1994) 0.5 m | 0.6 m<br>0.6 m       | 0.5 m<br>0.4 m.    | 0.6 m<br>0.4 m     | 1.5 m             |
| Average Canopy Cover                 | (1993) 6.9%<br>(1994) 30.2%  | 30.3%<br>27.0%       | 50.9%<br>43.3%     | 24.8%<br>34.2%     | 75%               |
| # Large Woody Debris/Lineal Distance | (1993) .3/m2<br>(1994) .5/m2 | 4.9/m2<br>7.6/m2     | 3.3/m2<br>6.9/m2   | 0.1/m2<br>.8/m2    |                   |
| Riffle: Pool Ratio                   | (1993) 2.7:1<br>(1994) 2.8:1 | 1.4:1<br>3.2:1       | 1.4:1<br>1.5:1     | 1.3:1<br>4.2:1     | 1:1               |
| Average Percent Fine Sediment        | (1993)18.8%<br>(1994)11.0%   | 5.6%<br>15.6%        | 13%<br>14.2%       | 12.8%<br>11.4%     | <10%              |

## 4.2 Biological Assessment

Cutthroat trout populations within their native range are considered to be declining region wide. The range of native trout have been severely reduced in the drainages of the Coeur d' Alene Indian Reservation. Multiple drainages that historically contained westslope cutthroat trout are currently devoid of viable populations. However, viable populations of cutthroat trout currently exist within selected study drainages. Bull trout are currently classified as a Category 1 species and are being considered by the U.S. Fish and Wildlife Service for listing as threatened or endangered. Baseline surveys on the Coeur d' Alene Indian Reservation revealed a conspicuous absence of bull trout (Lillengreen 1993).

Habitat use evaluation information indicated that cutthroat trout populations were selecting for certain habitat types (i.e. plunge pool) thus, indicating that all available habitat was not being utilized. Chi-square goodness of fit tests and analysis of catch rates indicated that cutthroat and brook trout were not distributed uniformly among available habitat types in the four study drainages. Drainage specific catch rate showed considerable variation in selection of habitat types, whereas combined catch rates for both cutthroat and brook trout showed a strong selective tendency for deep water habitat with slower velocities. Cutthroat trout densities were highest in plunge pool habitat

for all drainages except Alder Creek in which densities were highest in glide habitat. Not enough information was collected to determine why cutthroat trout selected for specific habitat types. Literature reports that habitat quality is related to depth and cover components (Hickman and Ralieggh, 1981). Assumptions could be made that habitat types selected by cutthroat trout in this study contained adequate depth and cover requirements. However, more data needs to be collected and will be addressed in future monitoring efforts.

Population and density estimates indicated that cutthroat trout populations and densities were below those reported in literature. Irving (1987) summarized several studies and concluded that "good" rearing habitat may support up to 200 fry/100 m<sup>2</sup>. Mean densities may approach 20 fish /100 m<sup>2</sup> for age one and two fish. Mean densities in all drainages ranged from 0.4 fish/100 m<sup>2</sup> to 4.7 ctt/100 m<sup>2</sup>. Some headwater reaches contained cutthroat trout densities that fell within the "good" rearing habitat range as reported by Irving (1987). However, cutthroat trout densities reported in mainstem drainages were well below those values reported by Irving (1987).

In order to assess habitat restoration alternatives, an accurate stock assessment needs to be completed. Part of this stock assessment is identifying the adfluvial, fluvial, and resident components of the population. Migratory analysis in combination with age and growth analysis indicated that an adfluvial stock of westslope cutthroat trout potentially existed in Lake Creek (Lillengreen *et. al.* 1993). Benewah Creek also contained a large number of adult westslope cutthroat trout migrants. However, determination of adfluvial or fluvial origin based on back calculated lengths was inconclusive. Fish trapped in the Alder and Evans creek drainages accounted for less than 10 percent of the total catch. No trout species were caught in the Evans Creek trap, while three mature adults were captured in Alder Creek. Stock determination was inconclusive with data collected. Further investigations need to be conducted.

### **4.3 Conclusion**

Habitat degradation of riparian/stream ecosystems are a result of the cumulative effects of trapping, livestock grazing, dam construction, logging, mining, the introduction of exotic species, channelization, urbanization, road construction, irrigation withdrawals, etc. In many instances, these cumulative effects, over time, have harmed the native fisheries because of their potential to alter ecosystem processes (Platts, 1991; Swanston, 1991). For example, severe prolonged reduction of riparian vigor may result in loss of bank stability, creating a long term source of sediment. A shift in the equilibrium between streamflow and sediment will produce a change in bar profile and bedload transport. When transport capacity is exceeded aggradation occurs and there is a decrease in bed material size (Leopold *et al.* 1964; Kappesser, 1992). The relationship between aquatic communities and specific habitat parameters (e.g. stream flow,

water quality, substrate size) is such that anthropogenic disturbance, which varies from the natural disturbance regime, can significantly alter the productivity of ecosystems by affecting species composition and diversity (Bjorn et al., 1974; Brusven and Prather, 1974; Hausle and Coble, 1976). Conditions currently existing in study drainages are characterized by declines in biological diversity and ecosystem productivity, and are the focus of interest for restoration efforts on the reservation.

Given this scenario and the limitations placed upon it by current management activities and conflicting economic demands, the following goals and objectives have been identified for the Coeur d'Alene Tribe fisheries enhancement program:

- Reestablish and protect self-sustaining populations of native cutthroat and bull trout which were historically prominent in the Lake Coeur d'Alene system.
- Develop and maintain continuous, healthy riparian corridors which support the full range of ecological and hydrological processes.
- Manage the riparian/aquatic interface for both wildlife and limited domestic use, while protecting water quality, public health, and the fisheries resource.
- Create long-term fishing opportunities for the local community.
- Develop and coordinate community and agency coalitions which would address issues related to habitat restoration.
- Develop agreements with private landholders to implement site specific restoration projects and encourage commitments to cost-sharing opportunities.

In order to achieve the above goals the Coeur d'Alene Tribe has identified the following objectives;

1. Conduct an extensive habitat restoration program in four major drainages on the Reservation which includes; Lake Benewah, Evans and Alder Creeks.
2. Develop harvest opportunities for the local community through a sustainable cutthroat trout fishery and an interim put and take rainbow trout pond program.
3. Purchase "critical watershed areas" for protection of fisheries habitat.

4. Educate and inform the public about issues related to cutthroat and bull trout habitat restoration.
5. Design, construct, operate and maintain a trout production facility on the Coeur d'Alene Reservation and
6. Implement a five-year monitoring program to evaluate the effectiveness of the above recommendations.

In order to protect stocks of westslope cutthroat trout restrictive fishing regulations have been adopted by the Coeur d'Alene Tribe. Complete stream fishing closures exist in Benewah and Lake creeks, while limited harvest opportunities exist in Alder and Evans creeks. These restrictive regulations will protect declining stocks from mortality due to angler harvest during spawning migrations. The Idaho Department of Fish and Game also has a moratorium on bull trout harvest in the Lake Coeur d'Alene Drainage as well as restrictive regulations pertaining to the harvest of westslope cutthroat trout.

## **LITERATURE CITED**

- Armour, C.L., K.P. Burnham, and W.S. Platts. 1983. Field methods and statistical analyses for monitoring small salmonid streams. U.S.D.I., Fish and Wildlife Service FWS/OBS-83/33.**
- Bjornn, T.C., M.A. Brusven, M.P. Nolnau, J.H. Milligan, R.A. Klamt, E. Chacho, and C. Schaye. 1977. Transport of granitic sediment in streams and its effects on insects and fish. College of Forestry, Wildlife and Range Sciences, Bulletin 17. University of Idaho, Moscow, ID.**
- Brusven, M.A., and K. Prather. 1974. Influence of stream sediments on distribution of macrobenthos. J. Ent. Soc. B.C. 71:25-32.**
- Buchanan, T. J. and W. P Somers, 1980. Discharge measurements at gaging stations: Techniques of Water Research Investigations of the U. S. Geology Survey, Book 3, Chapter A8, 69 pp.**
- Conlin, K. and B.D. Tutty. 1979. Juvenile salmon field trapping manual. Dept. of Fisheries and Oceans. Fisheries and Marine Service Resource Services Branch. Habitat Protection Division, Vancouver, British Columbia, 136pp.**
- Daniel, W.W. 1990. Applied nonparametric statistics, second edition. PWS-Kent Publishing Company, Boston, Massachusetts.**
- Everest, F.H. 1969. Habitat selection and spatial interaction of juvenile chinook salmon and steelhead trout in two Idaho streams. Ph.D. Dissertation, University of Idaho, Moscow, ID. 77 pp.**
- Fajen, O. F. and R. E. Wehnes. 1981. Missouri's method of evaluating stream habitat: in Armantrout (ed) Acquisition and Utilization of Aquatic Habitat Inventory Information, American Fisheries Society, Portland, p. 178-185.**
- Graves, S., K. L. Lillengreen, D. C. Johnson and A. T. Scholz. 1990. fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation, Draft Annual Report; Upper Columbia United Tribes Fisheries Center, Cheney, WA. 80 pp.**
- Griffith, J.S. 1970. interaction of brook trout and cutthroat trout in small streams. Ph.D. Dissertation, University of Idaho, Moscow, Id.**
- Hausle, D.A., and D.W. Coble. 1976. Influence of sand in redds on survival and emergence of brook trout (*Salvelinus fontinalis*). Trans. Am. Fish. Soc. 105(1):57-63.**

- Herricks, E. E. and J. Carina Jr. 1974. Rehabilitation of streams receiving acid mine drainage: Virginia River Resources Research Center, Virginia Polytechnic and State University, Blacksburg, Virginia, Bulletin 66. 284 pp.
- Hickman, T. and R.F. Raleigh. 1982. Habitat Suitability index models: Cutthroat trout. U.S.D.I. Fish and Wildlife Service. FWS/OBS-82/10.5 38 pp.
- Hynes, H.B.N. 1970. The ecology of running waters. Univ. Toronto Press, Canada. 555 pp.
- Idyll, C. 1942. Food of rainbow, cutthroat, and brown trout in the Cowichan River System, British Columbia. Journal of the Fisheries Research Board of Canada. 5: 448-458.
- Irving, O.B. 1987. Cutthroat trout abundance, potential yield, and interaction with brook trout in Priest Lake tributaries. M.S. thesis, Univ. of Idaho, Moscow, Idaho. 213 pp.
- Kappesser, Gary. 1992. Riffle Stability Index, Version 4.1. U.S.D.A., U.S. Forest Service, 7 pp.
- Krebs, C. J. 1985. Ecology: The Experimental Analysis of Distribution and Abundance, Third Edition: Harper Collins Publishers, New York, 800 pp.
- Leopold, L.B., M.G. Wolman and J.P. Miller. 1964. Fluvial processes in geomorphology. W H Freeman. San Francisco, CA. 522 pp.
- Lillengreen, K.L., Tami Skillingstad, and Allan T. Scholz. 1993. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation. Bonneville Power Administration. Division of Fish and Wildlife. Portland, Or. Project #90-44. 218 pp.
- Lisle, T.E. 1989. Using "residual depths" to monitor pool depths independently of discharge. U.S.D.A. Forest Service Res. note psw-394 Pacific Southwest Experimental Station. Berkeley, CA. 4 pp.
- Martin, D. J., E.O. Sala, S.T. White, J.A. June, W.J. Foris, and G.L. Lucchetti. 1981. The impact of managed streamside timber removal on cutthroat trout and the stream ecosystem. Fisheries Research Institute, Final report FRI-UI-8107. University of Washington, Seattle, WA.
- Peltier, J. 1975. Manners and customs of the Coeur d'Alene Indians. News Review Publishing Co., Moscow, ID. 84 pp.
- Perkins, J. L. 1982. Shannon-Wiener Journal of the Water Pollution Control Federation, V. 54, p 100-1050.

- Pfancuch, Dale J., 1975. Stream Reach Inventory and Channel Stability Evaluation, U.S.D.A., U.S. Forest Service, 15 pp.**
- Platts, W.S. 1991. Livestock grazing. American Fisheries Society Special Publication 19:389-424.**
- Raleigh, R.F., and D.A. Duff. 1981. Trout stream habitat improvement: ecology and management. Pages 67-77 in W. King (ed.). Proceedings of the Flathead River basin bull trout biology and population dynamics modeling information exchange. 24-25 July; British Columbia Ministry of Environment, Cranbrook, British Columbia, Canada.**
- Raleigh, R, F., T.J. Hickman, K.L. Nelson, and O.E. Maughan. 1982. Riverine habitat evaluation procedures for rainbow trout. Proceedings of the trout stream improvement workshop. November, 1980. Asheville, N.C.**
- Reiman, B.E., and K.A. Apperson. 1989. Status and analysis of salmonid fisheries: Westslope cutthroat trout synopsis and analysis of fishery information. Idaho Dept. of Fish and Game, Job Performance Report, Project F-73-R-1 1. Boise, ID.**
- Reynolds, J. B., 1983, Electrofishing: In Nielson, L. A., D. L. Johnson, and S. S Lampton (eds.) Fisheries Techniques, American Fisheries Society, Bethesda, MD, p 147-163.**
- Rhodes, J. and R. Jones. 1991. Riparian area management; current management situation and the need for improved riparian management. U.S.D.A. U.S. Forest Service, Clearwater National Forest. 20 pp.**
- Rosgen, Dave, L., Classification of Natural Rivers, U.S.D.A., U.S. Forest Service, 1991.**
- Scott, 0.1968. Pioneer Days on the Shadowy St. Joe. Caxton Printers, Caldwell, ID 372 pp.**
- Swanston, D.N. 1991. Natural Processes. American Fisheries Society Special Publication 19: 139-1 79.**
- Val Crispin, R.H. and D. Roberts. 1993. changes in instream habitat, large woody debris, and salmon habitat after the restructuring of a coastal Oregon stream. North American Journal of Fisheries Management Vol. 13 pp. 96-102.**
- Walker, W.D., Jr. 1977. Indians of Idaho. University of Idaho Press. Moscow, ID. 207 pp.**



- Waters, T.F. and R. J. Knapp. 1961. An improved stream bottom fauna sampler. Transactions of the American Fisheries Society, V. 90. pp. 225-229.**
- Wesche, T.A. 1980. The WRll trout cover rating method development and application. Water Resources Research Institute. University of Wyoming Project # B-032-Wyo. Completion report present to Wyoming Game and Fish Dept. Laramie, WY. 46 pp.**
- Williams, G.P. 1978. Bank-ful discharge of rivers. Water Resources Research. Vol. 14(6) pp. 1141-1154.**
- Wolman. M.G., 1954. A method of sampling coarse river bed material, Trans. , 1. Geophys. Union, 35(6), 951-956,**

## **Appendix A**

**Appendix A.1. Summary of Lake Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 1         | Mainstem       | C6           | Slight       | Mod - High        | High      | 1               | Silt               |
| 2         | Mainstem       | C 4          | Slight       | Mod - High        | High      | 1               | Gravel             |
| 3         | Mainstem       | B3           | Moderate     | Moderate          | Moderate  | 3-4             | Cobble             |
| 4         | Mainstem       | C 3          | Slight       | High              | High      | 2               | Cobble             |
| 5         | Mainstem       | B3           | Moderate     | Moderate          | Moderate  | 3               | Cobble             |
| 6         | Mainstem       | C4-C6        | Slight       | Mod - High        | High      | 1               | Gravel-Silt        |
| 7         | Mainstem       | E6           | Slight       | Verv Low          | Very High | 1               | Silt               |
| 8         | Mainstem       | E5           | Slight       | Verv Low          | Very High | 1               | Sand               |
| 8b        | West           | E5-C5        | Slight       | Very Low          | Very High | 1               | Sand               |
| 9b        | west           | E6           | Slight       | Very Low          | Very High | 1               | Silt               |
| 10b       | West           | B4a          | Moderate     | Moderate          | Moderate  | 5               | Gravel             |

**Appendix A.2. Summary of Benawah Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 1         | Mainstem       | C3           | Slight       | 39.0              | High      | 2               | Cobble             |
| 2         | Mainstem       | B2a          | Moderate     | 13.0              | Moderate  | 4               | Boulder            |
| 3         | Mainstem       | C2b          | Slight       | 15.0              | High      | 2               | Boulder            |
| 4         | Mainstem       | B3c          | Moderate     | 27.5              | Moderate  | 2               | Boulder            |
| 5         | Mainstem       | C1b          | Slight       | 21.3              | High      | 4-5             | Bedrock            |
| 6         | Mainstem       | C3           | Slight       | 20.0              | High      | 1               | Cobble             |
| 7         | Mainstem       | C3           | Slight       | 17.6              | High      | 1               | Cobble             |
| 8         | Mainstem       | C6           | Slight       | 26.2              | High      | 1               | Cobble             |

**Appendix A.2. (continued): Summary of Benawah Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 9         | Mainstem       | C5           | Slight       | 13.3              | High      | <1              | Sand               |
| 10        | Mainstem       | C6           | Slight       | 13.3              | High      | 1               | Silt               |
| 11        | Mainstem       | E6           | Slight       | 3.9               | Very High | <1              | Silt               |
| 12        | Mainstem       | C5-C6        | Slight       | —                 | High      | <1              | Sand-Silt          |
| 13        | Mainstem       | E 4          | Slight       | 6.0               | Very High | 2               | Gravel             |
| 13a       | West Fork      | E4           | Slight       | 10.0              | Very High | 2-3             | Gravel             |
| 14a       | West Fork      | E3b          | Slight       | 12.0              | Very High | 3 4             | Cobble             |

**Appendix A.3. Summary of Evans Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 1         | Mainstem       | C6           | Slight       | —                 | High      | <1              | Silt               |
| 2         | Mainstem       | C3           | Slight       | 7.8               | High      | 1               | Cobble             |
| 3         | Mainstem       | E3b          | Slight       | —                 | Very High | 3               | Cobble             |
| 4         | Mainstem       | B3b          | Moderate     | —                 | Moderate  | —               | Cobble             |
| 5         | Mainstem       | A4           | Moderate     | 17.3              | Moderate  | 7               | Cobble-Boulder     |
| 6         | Mainstem       | A2a          | High         | 8.4               | Low       | 10              | Boulder-Cobble     |
| 7         | Mainstem       | B3a          | Moderate     | 14.4              | Moderate  | 9               | Cobble             |
| 5a        | North Fork     | B3a          | Slight       | 6.3               | Moderate  | 8               | Cobble-Boulder     |

**Appendix A.3. (continued): Summary of Evans Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 5b        | South Fork     | A 2          | High         | 11.1              | Low       | 7-10            | Gravel             |
| 6b        | south Fork     | B3a          | High         | 6.3               | Low       | 16              | Boulder-Cobble     |

**Appendix A.4. Summary of Alder Creek channel type survey, 1993.**

| Reach Num | Reach Location | Channel Type | Entrenchment | Width/Depth Ratio | Sinuosity | Stream Gradient | Dominant Substrate |
|-----------|----------------|--------------|--------------|-------------------|-----------|-----------------|--------------------|
| 1         | Mainstem       | B3           | Moderate     |                   | Moderate  | 2               | Cobble             |
| 2,3,4     | Mainstem       | A2-A3        | High         | 11.5              | Low       | 3               | Boulder-Cobble     |
| 5,6       | Mainstem       | E3b          | Slight       | 11.3              | High      | 2.5             | Cobble             |
| 7         | Mainstem       | C3           | Slight       | >12               | High      | 1.5             | Cobble             |
| 8         | Mainstem       | C6           | Slight       | 12.5              | High      | <1              | Silt               |
| 9         | Mainstem       | E3b          | Slight       | 4.4               | Very High | 1               | Cobble             |
| 1a        | North Fork     | E3b          | Slight       | 4.6               | Very High | 3               | Cobble             |
| 2a,3a     | North Fork     | B3a-D3b      | Moderate     | 5.6 -- >40        | Moderate  | 2.5             | Cobble             |
| 4a        | North Fork     | B3a          | Moderate     | 6                 | Moderate  | 5               | Cobble             |

## Appendix B

**Table B.1 Lake Creek Stream Reach Inventory and Channel Stability Evaluation Summary.**

| Reach | 1993      |        | 1994      |        |
|-------|-----------|--------|-----------|--------|
|       | Stability | Rating | Stability | Rating |
| 1     | 113       | Fair   | 73        | Good   |
| 2     | 89        | Fair   | 93        | Fair   |
| 3     | 103       | Fair   | 80        | Fair   |
| 4     | 100       | Fair   | 86        | Fair   |
| 5     | 71        | Good   | 101       | Fair   |
| 6     | 101       | Fair   | 95        | Good   |
| 7     | 118       | Poor   | 79        | Fair   |
| 8     | 119       | Poor   | 117       | Poor   |
| 8b    | 129       | Poor   | 102       | Fair   |
| 9b    | 112       | Fair   | 113       | Fair   |
| 10b   | 87        | Fair   | 83        | Fair   |

**Table B.2 Benewah Creek Stream Reach Inventory and Channel Stability Evaluation Summary.**

| Reach | 1993      |        | 1994      |        |
|-------|-----------|--------|-----------|--------|
|       | Stability | Rating | Stability | Rating |
| 1     | 94        | Fair   | 83        | Fair   |
| 2     | 74        | Good   | 80        | Fair   |
| 3     | 67        | Good   | 67        | Good   |
| 4     | 60        | Good   | 78        | Fair   |
| 5     |           |        | 48        | Good   |
| 6     | 65        | Good   | 59        | Good   |
| 7     | 77        | Fair   | 96        | Fair   |
| 8     | 82        | Fair   | 74        | Good   |
| 9     | 122       | Poor   | 90        | Fair   |
| 10    | 100       | Fair   |           |        |
| 11    | 125       | Poor   | 117       | Poor   |
| 12    | 129       | Poor   | 115       | Poor   |
| 13    | 83        | Fair   | 107       | Poor   |
| 13a   | 77        | Fair   | 92        | Fair   |
| 14a   | 75        | Good   |           |        |

**Table B.3 Evans Creek Stream Reach Inventory and Channel Stability Evaluation Summary.**

| Reach | 1993      |        | 1994      |        |
|-------|-----------|--------|-----------|--------|
|       | Stability | Rating | Stability | Rating |
| 1     | 124       | Poor   | 125       | Poor   |
| 2     | 119       | Poor   | 101       | Fair   |
| 3     | 95        | Fair   | 94        | Fair   |
| 4     | 77        | Fair   | 86        | Fair   |
| 5     | 67        | Good   | 73        | Good   |
| 6     | 65        | Good   | 72        | Good   |
| 7     | 70        | Good   |           |        |
| 5a    | 73        | Good   | 83        | Fair   |
| 5b    | 86        | Fair   | 84        | Fair   |
| 6b    | 66        | Good   | 80        | Fair   |

**Table B-4 Alder Creek Stream Reach Inventory and Channel Stability Evaluation Summary.**

| Reach | 1993      |        | 1994      |        |
|-------|-----------|--------|-----------|--------|
|       | Stability | Rating | Stability | Rating |
| 1     | 88        | Fair   | 76        | Good   |
| 2     | 82        | Fair   | 78        | Fair   |
| 3     | 73        | Good   | 80        | Fair   |
| 4     | 73        | Good   | 65        | Good   |
| 5     | 82        | Fair   | 81        | Fair   |
| 6     | 53        | Good   | 72        | Good   |
| 7     | 79        | Fair   | 90        | Fair   |
| 8     | 128       | Poor   | 118       | Poor   |
| 9     | 103       | Fair   | 80        | Fair   |
| 1a    | 58        | Good   | 79        | Fair   |
| 2a    | 89        | Fair   | 102       | Fair   |
| 3a    | 105       | Fair   | 93        | Fair   |
| 4a    | 51        | Good   | 78        | Fair   |



## **Appendix C**

Appendix C.1 Results of Riffle Armour Stability Surveys for Lake Creek, 1993/1994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
| 2            |               | 62.0        |      | 69             |      | 28.4           |      |
|              | 2             | 78.1        | 96.8 | 96             | 142  | 24.0           | 3.1  |
|              | 3             | 48.0        | 88.0 | 98             | 139  | 10.4           | 5.9  |
| 3            | 1             | 75.8        | 80.3 | 119            | 107  | 21.2           | 10.8 |
|              | 2             | 58.0        | 77.9 | 150            | 134  | 19.6           | 11.3 |
|              | 3             | 65.8        | 52.0 | 165            | 152  | 11.2           | 19.4 |
| 4            | 1             | 78.0        | 69.1 | 153            | 148  | 24.6           | 17.5 |
|              | 2             | 67.4        |      | 120            |      | 26.6           |      |
|              | 3             | 72.9        | 69.6 | 133            | 134  | 40.7           | 9.5  |
| 5            | 1             | 52.6        | 69.5 | 94             | 171  | 37.3           | 15.7 |
|              | 2             | 59.1        | 82.5 | 144            | 176  | 16.8           | 14.4 |
|              | 3             | 33.6        | 66.9 | 101            | 173  | 19.9           | 26.9 |
| 6            | 1             | 74.2        | 81.4 | 130            | 151  | 34.5           | 34.0 |
|              | 2             | 74.8        | 78.3 | 85             | 134  | 54.5           | 22.5 |
|              | 3             | 71.6        | 84.5 | 115            | 142  | 55.5           | 23.1 |
| 7            | 1             | 80.4        | 68.0 | 123            | 113  | 46.7           | 43.1 |
|              | 2             | 96.1        | 88.6 | 98             | 139  | 61.2           | 61.7 |
|              | 3             | 90.6        | 95.8 | 86             | 121  | 83.7           | 31.7 |
| 10b          | 1             | 62.6        |      | 61             |      | 17.5           |      |
|              | 2             | 62.6        |      | 61             |      | 17.5           |      |
|              | 3             | 66.0        |      | 53             |      | 30.0           |      |

Appendix C.2 Results of Riffle Armour Stability Surveys for Benawah 1993/1994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
| 1            | 1             | 79.5        | 89.8 | 130            | 130  | 26.4           | 24.8 |
|              | 2             | 69.3        | 68.9 | 133            | 175  | 8.7            | 12.3 |
|              | 3             | 82.4        | 87.1 | 136            | 152  | 1.5            | 27.3 |
| 2            | 1             |             | 24.1 |                | 129  |                | 5.3  |
|              | 2             |             | 67.1 |                | 126  |                | 11.7 |
|              | 3             |             | 38.9 |                | 132  |                | 14.7 |
| 3            | 2             | 52.3        | 2.7  | 126            | 135  | 11.3           | 2.4  |
|              |               | 44.7        | 43.0 | 131            | 152  | 6.5            | 6.7  |
|              | 3             | 55.5        | 86.7 | 162            | 138  | 6.3            | 9.4  |
| 4            | 1             | 38.2        | 82.0 | 97             | 124  | 2.5            | 11.0 |
|              | 2             | 53.3        | 78.8 | 111            | 113  | 6.2            | 12.7 |
|              | 3             | 45.5        | 83.4 | 112            | 131  | 6.0            | 11.1 |
| 6            | 1             | 44.8        | 66.8 | 127            | 148  | 10.0           | 15.3 |
|              | 2             | 43.0        | 28.9 | 107            | 129  | 13.5           | 9.8  |
|              | 3             | 62.3        | 85.7 | 134            | 122  | 9.0            | 10.6 |
| 7            | 1             | 60.2        | 77.0 | 142            | 94   | 14.5           | 17.2 |
|              | 2             | 27.0        | 69.6 | 79             |      | 13.3           | 9.3  |

Appendix C.2 (continued): Results of Riffle Armour Stability Surveys for Benawah 1993/1994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
|              | 3             | 45.1        | 92.7 | 127            | 117  | 11.2           | 20.0 |
| 8            | 1             | 41.6        | 70.5 | 108            | 164  | 9.8            | 27.5 |
|              | 2             | 33.0        | 76.2 | 63             | 123  | 5.0            | 8.2  |
|              | 3             | 72.4        | 68.9 | 154            | 124  | 9.2            | 8.5  |
| 13a          | 1             | 49.5        |      | 78             |      | 20.1           |      |

Appendix C.3 Results of Riffle Armour Stability Surveys for Evans Creek, 1993/1994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
|              |               | 75.2        | 91.4 | 44             |      | 25.7           | 3.5  |
|              | 2             | 39.8        | 97.5 | 71             | 139  | 2.0            | 10.3 |
|              | 3             | 62.5        | 89.4 | 82             | 154  | 4.5            | 6.1  |
| 3            | 1             | 67.3        | 84.0 | 87             | 130  | 11.8           | 24.2 |
|              | 2             | 82.2        |      | 113            |      | 17.3           |      |
|              | 3             | 59.6        | 91.5 | 94             | 141  | 23.2           | 13.5 |
| 4            | 1             | 59.1        | 90.3 | 95             | 172  | 11.5           | 25.8 |
|              | 2             | 61.0        | 92.6 | 76             | 183  | 10.0           | 22.6 |
|              | 3             | 56.7        | 84.6 | 81             | 137  | 16.0           | 17.0 |
| 5            | 1             | 42.5        | 60.0 | 73             | 121  | 20.9           | 16.5 |
|              | 2             | 51.8        | 61.9 | 77             | 113  | 23.3           | 13.1 |
|              | 3             | 31.4        | 64.3 | 76             | 135  | 20.8           | 11.1 |
| 5b           | 1             | 63.4        | 64.8 | 71             | 54   |                | 37.9 |
|              | 2             | 84.6        | 78.3 | 93             | 79   |                | 39.5 |
|              | 3             | 71.8        |      | 75             |      |                |      |

Appendix C.4 Results of Riffle Armour Stability Surveys for Alder Creek, 1993/1994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
| 1            | 1             | 31.9        | 58.0 | 57             | 135  | 7.7            | 11.3 |
|              | 2             | 37.9        | 68.6 | 58             | 141  | 7.1            | 11.1 |
|              | 3             | 28.5        | 39.2 | 62             | 132  | 11.0           | 6.2  |
| 3            | 1             | 33.4        | 62.7 | 66             | 161  | 13.9           | 10.1 |
|              | 2             | 24.8        | 65.0 | 62             | 163  | 5.1            | 10.3 |
|              | 3             | 15.6        | 50.0 | 54             | 156  | 2.8            | 11.1 |
| 5            | 1             | 27.8        | 59.7 | 46             | 173  | 6.5            | 9.2  |
|              | 2             | 40.3        | 46.2 | 52             | 146  | 11.0           | 11.4 |
|              | 3             | 30.5        | 71.1 | 47             | 167  | 11.2           | 7.6  |
| 7            | 1             | 83.2        | 68.8 | 123            | 96   | 19.8           | 22.3 |
|              | 2             | 66.6        | 87.8 | 83             | 154  | 14.6           | 14.9 |

Appendix C.4 (continued): Riffle Armour Stability Surveys for Alder Creek, 1993/ 994.

| Reach Number | Sample Number | Index Value |      | Geometric Mean |      | Percent < 4 mm |      |
|--------------|---------------|-------------|------|----------------|------|----------------|------|
|              |               | 1993        | 1994 | 1993           | 1994 | 1993           | 1994 |
|              | 3             | 90.8        | 55.5 | 66             | 123  | 25.6           | 16.9 |
| 1a           | 1             | 55          | 80.1 |                | 107  |                | 32.3 |
|              | 2             | 79.4        | 85.8 | 64             | 107  | 23.5           | 26.9 |
|              | 3             | 50.7        | 72.1 | 43             | 98   | 20.2           | 30.5 |
| 2a           | 1             |             | 69.3 |                | 58   |                | 36.1 |
| 3A           | 1             | 65.9        | 45.6 | 43             | 47   | 29.4           | 22.9 |
|              | 2             | 76.7        | 65.6 | 48             | 43   | 30.1           | 34.6 |
|              | 3             |             | 46.2 |                | 50   |                | 19.7 |
| 4A           | 1             | 45.0        |      | 42             |      | 22.0           |      |
|              | 2             | 39.5        |      | 45             |      | 15.1           |      |
|              | 3             | 37.5        |      | 30             |      | 21.6           |      |

## Appendix D

**Appendix D.I. Summary report for Reach # 2 of the Lake Creek Watershed data collected during 1993 and 1994.**

| Parameter   | 1993             | 1994         |
|---|------------------|--------------|
| Elevation   | 2,180 - 2,260 ft | same         |
| Total length  | 5,093 ft         | 6,260 ft     |
| Stream order  | 4                | 4            |
| Mean stream gradient                                    | 0.6              | 3.6          |
| Riffle/pool ratio                                       | 3.0 : 1          | 1.6: 1       |
| Land use  |                  |              |
| <b>Forest</b>   |                  |              |
| <b>Agriculture</b>                                      |                  |              |
| <b>Livestock grazing</b>                                |                  |              |
| <b>Mining</b>   |                  |              |
| <b>Wetland</b>  |                  |              |
| <b>Floodplain</b>                                       | <b>100%</b>      | <b>100 %</b> |
| <b>Other (includes residential, right of way, etc.)</b> |                  |              |
| Vegetative type   |                  |              |
| <b>Decidious</b>  |                  |              |
| <b>Coniferous</b>                                       |                  |              |
| <b>Mixed</b>  | <b>100%</b>      | <b>100 %</b> |
| Seal stage  |                  |              |
| <b>Grass/forb</b>                                       | 45.6 %           | 45.6 %       |
| <b>Shrub</b>  |                  |              |
| <b>Pole</b>   |                  |              |
| <b>Young</b>  |                  |              |
| <b>Mature</b>   | 54.4 %           | 54.4 %       |
| <b>Old growth</b>                                       |                  |              |
| <b>Other</b>  |                  |              |
| x Canopy cover  | 11.6             | 46.7         |
| #Woody debris   |                  |              |
| Logs  | 49               | 48           |
| <b>Root wads</b>  | 0                | 5            |

**Appendix D.2 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for Lake Creek during 1993.**

| Habitat                    | Frequency | %         | Total Area   | % Area | Residual       |
|----------------------------|-----------|-----------|--------------|--------|----------------|
| Type                       |           | Frequency | (sq. meters) |        | pool depth (m) |
| <b>Rapid</b>               |           |           |              |        |                |
| <b>Step pool cascade</b>   |           |           |              |        |                |
| <b>Slip face cascade</b>   |           |           |              |        |                |
| Total Cascades             |           |           |              |        |                |
| <b>Pocketwater</b>         | 1         | 2.9       | 275.1        | 2.1    |                |
| <b>Glide</b>               | 6         | 17.0      | 2,090        | 16.3   |                |
| <b>Run</b>                 |           |           |              |        |                |
| <b>Low gradient riffle</b> | 14        | 40.0      | 7,276        | 56.9   |                |
| Total Riffles              | 21        | 59.9      | 9,542.0      | 75.3   |                |
| <b>Dammed pool</b>         | 1         | 2.9       | 541.2        | 4.2    |                |
| <b>Eddy pool</b>           |           |           |              |        |                |
| <b>Plunge pool</b>         |           |           |              |        |                |
| <b>Scour pool</b>          | 12        | 34.3      | 924          | 7.2    | 0.6            |
| <b>Scour hole</b>          |           |           |              |        |                |
| <b>Beaver pond</b>         | 1         | 2.9       | 1,689        | 13.2   | 0.6            |
| Total Pools                | 14        | 40.1      | 3,155        | 24.6   |                |
| <b>Secondary channel</b>   |           |           |              |        |                |
| GrandTotals                | 35        | 100.0     | 12,792.0     | 99.9   |                |

**Appendix D.3 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for Lake Creek during 1994.**

| Habit                      | Frequency | %         | Total Area   | % Area | Residual       |
|----------------------------|-----------|-----------|--------------|--------|----------------|
| Type                       |           | Frequency | (sq. meters) |        | pool depth (m) |
| <b>Rapid</b>               |           |           |              |        |                |
| <b>Step pool cascade</b>   | 2         | 4.2       | 2,292        | 7.9    |                |
| <b>Slip face cascade</b>   | 1         | 2.1       | 112          | 0.4    |                |
| Total Cascades             | 3         | 6.3       | 2,404        | 8.3    |                |
| <b>Pocketwater</b>         | 1         | 2.1       | 1,372        | 4.8    |                |
| <b>Glide</b>               | 5         | 16.6      | 3,734        | 12.9   |                |
| <b>Run</b>                 |           |           |              |        |                |
| <b>Low gradient riffle</b> | 21        | 44.7      | 17,639       | 61.1   |                |
| Total Riffles              | 27        | 63.4      | 22,745       | 78.8   |                |
| <b>Dammed pool</b>         |           |           |              |        |                |
| <b>Eddy pool</b>           |           |           |              |        |                |
| <b>Plunge pool</b>         |           |           |              |        |                |
| <b>Scour pool</b>          | 11        | 23.4      | 2,475        | 8.6    | 0.7            |
| <b>Scour hole</b>          | 6         | 12.8      | 1,242        | 4.3    | 0.5            |
| <b>Beaver pond</b>         |           |           |              |        |                |
| Total Pools                | 17        | 36.2      |              |        |                |
| <b>Secondary channel</b>   |           |           |              |        |                |
| Grand Totals               | 47        |           | 28,866       |        |                |

**Appendix D.4. . Summary report for Reach # 3 of the Lake Creek Watershed  
data collected during 1993 and 1994.**

| Parameter  | 1993          | 1994        |
|--|---------------|-------------|
| Elevation  | 2,060 - 2,350 |             |
| Total length   | 6,346 ft      | 4,809 ft    |
| Stream order   | 4             | 4           |
| Mean stream gradient                                   | 1.5           | 1.8         |
| Riffelpoolratio  | 1.8:1         | 1.6 : 1     |
| Land use   |               |             |
| <b>Forest</b>  |               |             |
| <b>Agriculture</b>                                     |               |             |
| <b>Livestock grazing</b>                               |               |             |
| <b>Mining</b>  |               |             |
| <b>Wetland</b>   |               |             |
| <b>Floodplain</b>                                      | <b>100%</b>   | <b>100%</b> |
| <b>Other (includes residential,right of way, etc.)</b> |               |             |
| Vegetative type  |               |             |
| <b>Decidious</b>                                       |               |             |
| <b>Coniferous</b>                                      |               |             |
| <b>Mixed</b>   | <b>100 %</b>  | <b>100%</b> |
| Seral stage  |               |             |
| Grass/forb   |               |             |
| <b>Shrub</b>   | <b>100 %</b>  | <b>100%</b> |
| <b>Pole</b>  |               |             |
| <b>Young</b>   |               |             |
| <b>Mature</b>  |               |             |
| <b>Old growth</b>                                      |               |             |
| <b>Other</b>   |               |             |
| x Canopy cover   | 18            | 21          |
| #Woody debris  |               |             |
| Logs   | 3             | <b>10</b>   |
| <b>Root wads</b>                                       | 1             | <b>0</b>    |



**Appendix D.5 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Lake Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 1                | 9.1                | 290.2                          | 2.3           |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 6                | 54.6               | 11,762.2                       | 93.7          |                                |
| <b>Total Riffle</b>   | <b>7</b>         | <b>63.7</b>        | <b>12,052.4</b>                | <b>98.0</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 4                | 36.4               | 502.8                          | 4.0           | 0.6                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>4</b>         | <b>35.4</b>        | <b>502.8</b>                   | <b>4.0</b>    |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>11</b>        | <b>100.0</b>       | <b>12,555.0</b>                | <b>100.0</b>  |                                |

**Appendix D.6 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Lake Creek during 1994.**

| <b>Habitat Type</b>  | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                |                  |                    |                                |               |                                |
| Step pool cascade    | 3                | 12                 | 1,179                          | 3.4           |                                |
| Slip face cascade    |                  |                    |                                |               |                                |
| <b>Total Cascade</b> | <b>3</b>         | <b>12</b>          | <b>1,179</b>                   | <b>3.4</b>    |                                |
| Pocketwater          | 2                | 8                  | 2,721                          | 7.9           |                                |
| Glide                |                  |                    |                                |               |                                |
| Run                  |                  |                    |                                |               |                                |
| Low gradient riffle  | 12               | 48                 | 28,623                         | 83.0          |                                |
| <b>Total Riffles</b> | <b>14</b>        | <b>56</b>          | <b>31,344</b>                  | <b>90.9</b>   |                                |
| Dammed pool          |                  |                    |                                |               |                                |
| Eddy pool            |                  |                    |                                |               |                                |
| Plunge pool          |                  |                    |                                |               |                                |
| Scour pool           | 4                | 16                 | 998                            | 2.9           | 0.4                            |
| Scour hole           | 3                | 12                 | 873                            | 2.5           | 0.3                            |
| Beaver pond          |                  |                    |                                |               |                                |
| <b>Total Pools</b>   | <b>7</b>         | <b>28</b>          | <b>1,871</b>                   |               |                                |
| Secondary channel    |                  |                    |                                |               |                                |
| <b>Grand Totals</b>  | <b>25</b>        |                    | <b>34,499</b>                  |               |                                |

**Appendix D.7. . Summary report for Reach # 4 of the Lake Creek Watershed  
data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>     |
|--|-------------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,350 - 2,410 ft</b> |                 |
| <b>Total length</b>                              | <b>4,048 ft</b>         | <b>2,583 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                | <b>4</b>        |
| <b>Mean stream gradient</b>                      | <b>2.0</b>              | <b>3.0</b>      |
| <b>Riffle/poolratio</b>                          | <b>2.9 : 1</b>          | <b>1.6 : 1</b>  |
| <b>Land use</b>                                  |                         |                 |
| Forest   | 92.2 %                  | 92.2 %          |
| Agriculture                                      |                         |                 |
| Livestock grazing                                |                         |                 |
| Mining   |                         |                 |
| Wetland  |                         |                 |
| Floodplain                                       |                         |                 |
| Other (includes residential, right of way, etc.) | 7.8 %                   | 7.8 %           |
| <b>Vegetative type</b>                           |                         |                 |
| Decidious  |                         |                 |
| Coniferous                                       |                         |                 |
| Mixed  | 100%                    | 100%            |
| <b>Seral stage</b>                               |                         |                 |
| Grass/forb                                       | 27.5 %                  | 27.5 %          |
| Shrub  |                         |                 |
| Pole ( shrub ? )                                 | 72.5 %                  | 72.5 %          |
| Young  |                         |                 |
| Mature   |                         |                 |
| Old growth                                       |                         |                 |
| Other  |                         |                 |
| <b>x Canopy cover</b>                            | <b>0.0</b>              | <b>10 %</b>     |
| <b># Woody debris</b>                            |                         |                 |
| Logs   |                         | 4               |
| Root wads  |                         | 0               |

**Appendix D.8 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Lake Creek during 1993.**

| Habitat Type               | Frequency | % Frequency | Total Area (sq. meters) | % Area      | Residual pool depth (m) |
|----------------------------|-----------|-------------|-------------------------|-------------|-------------------------|
| <b>Step pool cascade</b>   | <b>4</b>  | <b>12.9</b> | <b>2,737.1</b>          | <b>12.3</b> |                         |
| <b>Slip face cascade</b>   |           |             |                         |             |                         |
| Total Cascades             | 4         | 12.9        | 2,737.1                 | 12.3        |                         |
| <b>Pocketwater</b>         |           |             |                         |             |                         |
| <b>Glide</b>               | <b>5</b>  | <b>16.1</b> | <b>4,737.1</b>          | <b>21.3</b> |                         |
| <b>Run</b>                 |           |             |                         |             |                         |
| <b>Low gradient riffle</b> | <b>15</b> | <b>48.4</b> | <b>13,776.1</b>         | <b>62.1</b> |                         |
| Total Riffles              | 20        | 64.5        | 18,502.3                | 83.4        |                         |
| <b>Dammed pool</b>         |           |             |                         |             |                         |
| <b>Eddy pool</b>           |           |             |                         |             |                         |
| <b>Plunge pool</b>         |           |             |                         |             |                         |
| <b>Scour pool</b>          | <b>6</b>  | <b>19.4</b> | <b>814.7</b>            | <b>3.7</b>  |                         |
| <b>Scour hole</b>          | <b>1</b>  | <b>3.2</b>  | <b>131.7</b>            | <b>0.6</b>  |                         |
| <b>Beaver pond</b>         |           |             |                         |             |                         |
| Total Pools                | 7         | 22.6        | 946.4                   | 4.3         |                         |
| <b>Secondary channel</b>   |           |             |                         |             |                         |
| Grand Totals               | 31        |             | 22,185.6                |             |                         |

**Appendix D.9 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Lake Creek during 1994.**

| Habitat Type               | Frequency | % Frequency | Total Area (sq. meters) | % Area      | Residual pool depth (m) |
|----------------------------|-----------|-------------|-------------------------|-------------|-------------------------|
| <b>Step pool cascade</b>   | <b>2</b>  | <b>6.7</b>  | <b>1,652</b>            | <b>6.7</b>  |                         |
| <b>Slip face cascade</b>   |           |             |                         |             |                         |
| Total Cascades             | 2         | 6.7         | 1,652                   | 8.7         |                         |
| <b>Pocketwater</b>         |           |             |                         |             |                         |
| <b>Glide</b>               | <b>1</b>  | <b>6.7</b>  | <b>643</b>              | <b>3.4</b>  |                         |
| <b>Run</b>                 |           |             |                         |             |                         |
| <b>Low gradient riffle</b> | <b>7</b>  | <b>46.0</b> | <b>16,003</b>           | <b>84.0</b> |                         |
| Total Riffles              | 8         | 52.7        | 16,646                  | 87.4        |                         |
| <b>Dammed pool</b>         |           |             |                         |             |                         |
| <b>Eddy pool</b>           |           |             |                         |             |                         |
| <b>Plunge pool</b>         |           |             |                         |             |                         |
| <b>Scour pool</b>          | <b>5</b>  | <b>33</b>   | <b>748</b>              | <b>3.9</b>  | <b>0.5</b>              |
| <b>Scour hole</b>          |           |             |                         |             |                         |
| <b>Beaver pond</b>         |           |             |                         |             |                         |
| Total Pools                | 5         | 33          | 748                     | 3.9         |                         |
| <b>Secondary channel</b>   |           |             |                         |             |                         |
| Grand Totals               | 15        |             | 19,945                  |             |                         |

**Appendix D.10. Summary report for Reach # 5 of the Lake Creek Watershed data collected during 1993 and 1994.**

| Parameter  | 1993             | 1994         |
|--|------------------|--------------|
| Elevation  | 2,410 - 2,450 ft | same         |
| Total length   | <b>1,603 ft</b>  | 3,902 ft     |
| Stream order   | 4                | 4            |
| Mean stream gradient                                   | 1.0 %            | 1.5%         |
| Riie/pool ratio  | <b>2.9 : 1</b>   | 8.1 : 1      |
| Land use   |                  |              |
| <b>Forest</b>  | <b>100 %</b>     | <b>100 %</b> |
| <b>Agriculture</b>                                     |                  |              |
| <b>Livestock grazing</b>                               |                  |              |
| <b>Mining</b>  |                  |              |
| <b>Wetland</b>   |                  |              |
| <b>Floodplain</b>                                      |                  |              |
| <b>Other (includes residential.right of way, etc.)</b> |                  |              |
| Vegetative type  |                  |              |
| <b>Decidious</b>                                       |                  |              |
| <b>Coniferous</b>                                      |                  |              |
| <b>Mixed</b>   | <b>100 %</b>     | <b>100%</b>  |
| Seal stage   |                  |              |
| <b>Grass/forb</b>                                      | <b>61.6</b>      | <b>61.6</b>  |
| <b>Shrub</b>   |                  |              |
| <b>Pole</b>  |                  |              |
| <b>Young</b>   |                  |              |
| <b>Mature</b>  | <b>38.4</b>      | <b>38.4</b>  |
| <b>Old growth</b>                                      |                  |              |
| <b>Other</b>   |                  |              |
| x Canopy cover   | 18.5 %           | <b>26 %</b>  |
| #Woody debris  |                  |              |
| <b>Logs</b>  | <b>5</b>         | <b>4</b>     |
| <b>Root wads</b>                                       |                  |              |

**Appendix D.11 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Lake Creek during 1993.**

| Habitat Type        | Frequency | % Frequency | Total Area (sq. meters) | % Area | Residual pool depth (m) |
|---------------------|-----------|-------------|-------------------------|--------|-------------------------|
| <b>Rapid</b>        |           |             |                         |        |                         |
| Step pool cascade   |           |             |                         |        |                         |
| Slip face cascade   |           |             |                         |        |                         |
| Total Cascades      |           |             |                         |        |                         |
| <b>Pocketwater</b>  |           |             |                         |        |                         |
| Glide               | 4         | 33.3        | 2,372.3                 | 28.4   |                         |
| Run                 |           |             |                         |        |                         |
| Low gradient riffle | 5         | 41.7        | 4,812.2                 | 57.6   |                         |
| Total Riffles       | 9         | 75.0        | 7,184.5                 | 86.0   |                         |
| <b>Dammed pool</b>  |           |             |                         |        |                         |
| Eddy pool           |           |             |                         |        |                         |
| Plunge pool         |           |             |                         |        |                         |
| Scour pool          | 3         | 25.0        | 1,172.9                 | 14.0   | 0.31                    |
| Scour hole          |           |             |                         |        |                         |
| Beaver pond         |           |             |                         |        |                         |
| Total Pools         | 3         | 25.0        | 1,172.9                 | 14.0   |                         |
| Secondary channel   |           |             |                         |        |                         |
| <b>Grand Totals</b> | <b>12</b> |             | <b>8,357.4</b>          |        |                         |

**Appendix D.12 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Lake Creek during 1994.**

| Habitat Type        | Frequency | % Frequency | Total Area (sq. meters) | % Area | Residual pool depth (m) |
|---------------------|-----------|-------------|-------------------------|--------|-------------------------|
| <b>Rapid</b>        |           |             |                         |        |                         |
| Step pool cascade   |           |             |                         |        |                         |
| Slip face cascade   |           |             |                         |        |                         |
| Total Cascades      |           |             |                         |        |                         |
| <b>Pocketwater</b>  |           |             |                         |        |                         |
| Glide               | 7         | 39          | 5,574                   | 29.9   |                         |
| Run                 |           |             |                         |        |                         |
| Low gradient riffle | 9         | 50          | 12,456                  | 66.8   |                         |
| Total Riffles       | 16        | 89          | 18,030                  | 96.7   |                         |
| <b>Dammed pool</b>  |           |             |                         |        |                         |
| Eddy pool           |           |             |                         |        |                         |
| Plunge pool         |           |             |                         |        |                         |
| Scour pool          |           |             |                         |        |                         |
| Scour hole          | 2         | 11.1        | 626                     | 3.4    | 0.5                     |
| Beaver pond         |           |             |                         |        |                         |
| Total Pools         | 2         | 11.1        | 626                     | 3.4    |                         |
| Secondary channel   |           |             |                         |        |                         |
| Grand Totals        | 18        |             | 18,656                  |        |                         |

**Appendix 0.13. . Summary report for Reach # 6 of the Lake Creek  
Watershed data collected during 1993 and 1994.**

| Parameter  | 1993                    | 1994           |
|--|-------------------------|----------------|
| Elevation  | <b>2,450 - 2,480 ft</b> | same           |
| Total length                                     | <b>8,705 ft</b>         | <b>8,248ft</b> |
| Stream order                                     | 4                       | 4              |
| Mean stream gradient                             | 1.7                     | 1.3            |
| Riffle:pool ratio                                | 2.9 : 1                 | 2.1 : 1        |
| Land use   |                         |                |
| <b>Forest</b>                                    | <b>80.4</b>             | <b>80.4</b>    |
| <b>Agriculture</b>                               | <b>19.6</b>             | <b>19.6</b>    |
| Livestock grazing                                |                         |                |
| Mining   |                         |                |
| Wetland  |                         |                |
| Floodplain                                       |                         |                |
| Other (includes residential, right of way, etc.) |                         |                |
| Vegetative type                                  |                         |                |
| Deciduous  |                         |                |
| Coniferous                                       |                         |                |
| Mixed  | <b>100 %</b>            | <b>100 %</b>   |
| Seral stage                                      |                         |                |
| <b>Grass/forb</b>                                | <b>28.8</b>             | <b>28.8</b>    |
| <b>Shrub</b>                                     | <b>2.9</b>              | <b>2.9</b>     |
| Pole   |                         |                |
| Young  |                         |                |
| <b>Mature</b>                                    | 68.3                    | 68.3           |
| Old growth                                       |                         |                |
| Other  |                         |                |
| x Canopy cover                                   | 0                       | 26.5           |
| # Woody debris                                   |                         |                |
| Logs   | <b>16</b>               | <b>10</b>      |
| <b>Root wads</b>                                 | 2                       | 0              |

**Appendix 0.14 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Lake Creek during 1993.**

| Habitat                    | Frequency  | %           | Total Area    | % Area      | Residual       |
|----------------------------|------------|-------------|---------------|-------------|----------------|
| Type                       | Frequency  |             | (sq. meters)  |             | pool depth (m) |
| <b>Rapid</b>               |            |             |               |             |                |
| <b>Step pool cascade</b>   |            |             |               |             |                |
| <b>Slip face cascade</b>   |            |             |               |             |                |
| Total Cascades             |            |             |               |             |                |
| <b>Pocketwater</b>         |            |             |               |             |                |
| <b>Glide</b>               | 42         | 30.9        | 20,069        | 39.3        |                |
| <b>Run</b>                 |            |             |               |             |                |
| <b>Low gradient riffle</b> | 57         | 41.9        | 21,954        | 43.0        |                |
| Total Riffles              | 99         | 72.8        | 42,023        | 82.3        |                |
| <b>Dammed pool</b>         | 1          | 0.7         | 692           | 1.4         | 0.8            |
| <b>Eddy pool</b>           |            |             |               |             |                |
| <b>Plunge pool</b>         |            |             |               |             |                |
| <b>Scour pool</b>          | 32         | 23.5        | 7,737         | 15.2        | 0.6            |
| <b>Scour hole</b>          | 1          | 0.7         | 434           | 0.9         | 0.8            |
| <b>Beaver pond</b>         |            |             |               |             |                |
| <b>Total Pools</b>         | <b>34</b>  | <b>24.9</b> | <b>8,863</b>  | <b>17.5</b> | <b>0.7</b>     |
| Secondary channel          |            |             |               |             |                |
| <b>Grand Totals</b>        | <b>136</b> |             | <b>51,031</b> |             |                |

**Appendix D.15 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Lake Creek during 1994.**

| Habit                      | Frequency | %           | Total Area    | % Area      | Residual       |
|----------------------------|-----------|-------------|---------------|-------------|----------------|
| Type                       | Frequency |             | (sq. meters)  |             | pool depth (m) |
| <b>Rapid</b>               |           |             |               |             |                |
| <b>Step pool cascade</b>   |           |             |               |             |                |
| <b>Slip face cascade</b>   |           |             |               |             |                |
| Total Cascades             |           |             |               |             |                |
| <b>Pocketwater</b>         |           |             |               |             |                |
| <b>Glide</b>               | 11        | 20.8        | 9,566         | 14.9        |                |
| <b>Run</b>                 |           |             |               |             |                |
| <b>Low gradient riffle</b> | 25        | 47.1        | 26,175        | 40.8        |                |
| Total Riffles              | <b>36</b> | <b>67.9</b> | <b>35,741</b> | <b>55.7</b> |                |
| <b>Dammed pool</b>         |           |             |               |             |                |
| <b>Eddy pool</b>           |           |             |               |             |                |
| <b>Plunge pool</b>         |           |             |               |             |                |
| <b>Scour pool</b>          | 11        | 20.8        | 4,785         | 7.5         | 0.4            |
| <b>Scour hole</b>          | 4         | 7.5         | 1,878         | 2.9         | 0.6            |
| <b>Beaver pond</b>         | 2         | 3.8         | 21,766        | 33.9        | 0.6            |
| Total Pools                | 17        | 32.1        | 28,429        | 44.3        |                |
| Secondary channel          |           |             |               |             |                |
| <b>Grand Totals</b>        | <b>53</b> |             | <b>64,171</b> |             |                |

**Appendix D.16 Summary report for Reach # 7 of the Lake Creek Watershed  
data collected during 1993 and 1994.**

| Parameter   | 1993                    | 1994           |
|---|-------------------------|----------------|
| Elevation   | <b>2,480 - 2,600 ft</b> | same           |
| Total length  | 16,365 ft               | 18,828 ft      |
| Stream order  | 4                       | 4              |
| Mean stream gradient                                    | 1.5                     | same           |
| Riffle/pool ratio                                       | 2.6 : 1                 | <b>2.3 : 1</b> |
| Land use  |                         |                |
| <b>Forest</b>   | <b>9.1</b>              | <b>9.1</b>     |
| <b>Agriculture</b>                                      | <b>82.6</b>             | <b>82.6</b>    |
| <b>Livestock grazing</b>                                | <b>8.3</b>              | <b>8.3</b>     |
| <b>Mining</b>   |                         |                |
| <b>Wetland</b>  |                         |                |
| <b>Floodplain</b>                                       |                         |                |
| <b>Other (includes residential, right of way, etc.)</b> |                         |                |
| Vegetative type   |                         |                |
| <b>Deciduous</b>  |                         |                |
| <b>Coniferous</b>                                       |                         |                |
| <b>Mixed</b>  | <b>100 %</b>            | <b>100 %</b>   |
| Seal stage  |                         |                |
| <b>Grass forb</b>                                       | <b>58.3</b>             | <b>58.3</b>    |
| <b>Shrub</b>  | <b>41.7</b>             | <b>41.7</b>    |
| <b>Pole</b>   |                         |                |
| <b>Young</b>  |                         |                |
| <b>Mature</b>   |                         |                |
| <b>Old growth</b>                                       |                         |                |
| <b>Other</b>  |                         |                |
| x Canopy cover  | 0                       | 45.8           |
| # Woody debris  |                         |                |
| <b>Logs</b>   | <b>48</b>               | <b>168</b>     |
| <b>Root wads</b>  | <b>1</b>                | <b>4</b>       |



**Appendix 0.17 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 7 for Lake Creek during 1993.**

| Habitat                    | Frequency | %            | Total Area      | % Area       | Residual       |
|----------------------------|-----------|--------------|-----------------|--------------|----------------|
| Type                       | Frequency |              | (sq. meters)    |              | pool depth (m) |
| <b>Rapid</b>               |           |              |                 |              |                |
| <b>Step pool cascade</b>   |           |              |                 |              |                |
| <b>Slip face cascade</b>   |           |              |                 |              |                |
| Total Cascades             |           |              |                 |              |                |
| <b>Pocketwater</b>         |           |              |                 |              |                |
| <b>Glide</b>               | <b>19</b> | <b>44.2</b>  | <b>40,547.5</b> | <b>82.6</b>  |                |
| <b>Run</b>                 |           |              |                 |              |                |
| <b>Low gradient riffle</b> | 12        | 27.9         | 5,674.3         | 11.6         |                |
| Total Riffles              | 31        | 72.1         | 46,221.8        | 94.2         |                |
| <b>Dammed pool</b>         | <b>4</b>  | <b>9.3</b>   | <b>1,475.8</b>  | <b>3.0</b>   | <b>0.6</b>     |
| <b>Eddy pool</b>           |           |              |                 |              |                |
| <b>Plunge pool</b>         |           |              |                 |              |                |
| <b>Scour pool</b>          | <b>6</b>  | <b>14.0</b>  | <b>1,146.4</b>  | <b>2.3</b>   | <b>0.7</b>     |
| <b>Scour hole</b>          | <b>2</b>  | <b>4.7</b>   | <b>252.0</b>    | <b>0.5</b>   | <b>0.5</b>     |
| <b>Beaver pond</b>         |           |              |                 |              |                |
| Total Pools                | 12        | 28.0         | 2,874.2         | 5.8          |                |
| <b>Secondary channel</b>   |           |              |                 |              |                |
| <b>Grand Totals</b>        | <b>43</b> | <b>100.1</b> | <b>49,096.0</b> | <b>100.0</b> |                |

**Appendix D-18 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 7 for Lake Creek during 1994.**

| Habitat                    | Frequency | %         | Total Area    | % Area      | Residual       |
|----------------------------|-----------|-----------|---------------|-------------|----------------|
| Type                       | Frequency |           | (sq. meters)  |             | pool depth (m) |
| <b>Rapid</b>               |           |           |               |             |                |
| <b>Step pool cascade</b>   |           |           |               |             |                |
| <b>Slip face cascade</b>   |           |           |               |             |                |
| Total Cascades             |           |           |               |             |                |
| <b>Pocketwater</b>         | 1         | 2.5       | 181           | 0.2         |                |
| <b>Glide</b>               | <b>16</b> | <b>40</b> | <b>44,509</b> | <b>57.3</b> |                |
| <b>Run</b>                 |           |           |               |             |                |
| <b>Low gradient riffle</b> | 11        | 27.5      | 14,911        | 19.3        |                |
| Total Riffles              | <b>28</b> | <b>70</b> | <b>59,601</b> | <b>76.6</b> |                |
| <b>Dammed pool</b>         |           |           |               |             |                |
| <b>Eddy pool</b>           |           |           |               |             |                |
| <b>Plunge pool</b>         |           |           |               |             |                |
| <b>Scour pool</b>          | 5         | 12.5      | 2,282         | 2.9         | 0.4            |
| <b>Scour hole</b>          |           |           |               |             |                |
| <b>Beaver pond</b>         | 7         | 17.5      | 15,740        | 20.3        | 0.4            |
| Total Pools                | <b>12</b> | <b>30</b> | <b>18,022</b> | <b>23.2</b> |                |
| <b>Secondary channel</b>   |           |           |               |             |                |
| <b>Grand Totals</b>        | <b>40</b> |           | <b>77,625</b> |             |                |

**Appendix D.19. Summary report for Reach # 1 of the West Fork Lake Creek Watershed data collected during 1993 and 1994.**

| Parameter  | 1993                 | 1994         |
|--|----------------------|--------------|
| Elevation  | <b>2,540 - 2,570</b> | same         |
| Total length   | 10,132 ft            | 10,712 m     |
| Stream order   | <b>3</b>             | <b>3</b>     |
| Mean stream gradient                                   | 1.0                  | 1.2          |
| Riffle/pool ratio                                      | <b>2.4: 1</b>        | 1.9: 1       |
| Land use   |                      |              |
| <b>Forest</b>  | <b>68.7</b>          | <b>68.7</b>  |
| <b>Agriculture</b>                                     | <b>19.7</b>          | <b>19.7</b>  |
| <b>Livestock grazing</b>                               | <b>11.6</b>          | <b>11.6</b>  |
| <b>Mining</b>  |                      |              |
| <b>Wetland</b>   |                      |              |
| <b>Floodplain</b>                                      |                      |              |
| <b>Other (includes residential.right of way, etc.)</b> |                      |              |
| Vegetative type  |                      |              |
| <b>Decidious</b>                                       |                      |              |
| <b>Coniferous</b>                                      |                      |              |
| <b>Mixed</b>   | <b>100 %</b>         | <b>100 %</b> |
| Seral stage  |                      |              |
| <b>Grass/forb</b>                                      | <b>16.2</b>          | <b>16.2</b>  |
| <b>Shrub</b>   | <b>58.9</b>          | <b>58.9</b>  |
| <b>Pole</b>  | <b>15.1</b>          | <b>15.5</b>  |
| <b>Young</b>   | <b>9.9</b>           | <b>9.9</b>   |
| <b>Mature</b>  |                      |              |
| <b>Old growth</b>                                      |                      |              |
| <b>Other</b>   |                      |              |
| x Canopy cover   | 0                    | <b>35.4</b>  |
| #Woody debris  |                      |              |
| Logs   | <b>28</b>            | <b>18</b>    |
| <b>Root wads</b>                                       | <b>0</b>             | <b>0</b>     |

**Appendix D.20 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for West Fork Lake Creek during 1993.**

| Habitat                  | Frequency | %           | Total Area      | % Area       | Residual       |
|--------------------------|-----------|-------------|-----------------|--------------|----------------|
| Type                     |           | Frequency   | (sq. meters)    |              | pool depth (m) |
| <b>Rapid</b>             |           |             |                 |              |                |
| <b>Step pool cascade</b> |           |             |                 |              |                |
| <b>Slip face cascade</b> |           |             |                 |              |                |
| Total Cascades           |           |             |                 |              |                |
| <b>Pocketwater</b>       | 1         | <b>2.9</b>  | <b>674</b>      | <b>3.7</b>   |                |
| <b>Glide</b>             | 9         | <b>26.5</b> | <b>3,521.4</b>  | <b>19.1</b>  |                |
| <b>Run</b>               |           |             |                 |              |                |
| Low gradient rime        | 14        | <b>41.2</b> | <b>13,904.7</b> | <b>75.4</b>  |                |
| Total Riffles            | 24        | 70.6        | 18,100.1        | 98.2         |                |
| <b>Dammed pool</b>       |           |             |                 |              |                |
| <b>Eddy pool</b>         |           |             |                 |              |                |
| <b>Plunge pool</b>       | 5         | <b>14.7</b> | <b>182.6</b>    | <b>1.0</b>   |                |
| <b>Scour pool</b>        | 5         | 14.7        | <b>160.3</b>    | <b>0.9</b>   |                |
| <b>Scour hole</b>        |           |             |                 |              |                |
| <b>Beaver pond</b>       |           |             |                 |              |                |
| Total Pools              | 10        | 29.4        | 342.9           | 1.9          |                |
| <b>Secondary channel</b> |           |             |                 |              |                |
| <b>Grand Totals</b>      | <b>34</b> | <b>100</b>  | <b>18,443.0</b> | <b>100.1</b> |                |

**Appendix D.21 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #1 for West Fork Lake Creek during 1994.**

| Habitat                  | Frequency | %           | Total Area    | % Area      | Residual    |
|--------------------------|-----------|-------------|---------------|-------------|-------------|
| Type                     |           | Frequency   | (sq. meters)  |             | + depth (m) |
| <b>Rapid</b>             |           |             |               |             |             |
| <b>Step pool cascade</b> | 1         | <b>1.9</b>  | 366           | 2.1         |             |
| <b>Slip face cascade</b> |           |             |               |             |             |
| Total Cascades           | 1         | 1.9         | 366           | 2.1         |             |
| <b>Pocketwater</b>       |           |             |               |             |             |
| <b>Glide</b>             | 125       | 22.6        | 5,384         | <b>31.7</b> |             |
| <b>Run</b>               |           |             |               |             |             |
| Low gradient riffle      | 22        | 41.5        | <b>10,220</b> | <b>60.1</b> |             |
| Total Riffles            | 147       | 64.1        | 15,604        | <b>91.8</b> |             |
| <b>Dammed pool</b>       |           |             |               |             |             |
| <b>Eddy pool</b>         |           |             |               |             |             |
| <b>Plunge pool</b>       | 1         | <b>1.9</b>  | 10            | 0.6         | 0.6         |
| <b>Scour pool</b>        | 13        | <b>24.5</b> | <b>607</b>    | 36          | <b>0.4</b>  |
| <b>Scour hole</b>        | 4         | <b>4.5</b>  | <b>317</b>    | 1.9         | <b>0.5</b>  |
| <b>Beaver pond</b>       |           |             |               |             |             |
| Total Pools              | 18        | <b>30.9</b> | <b>934</b>    | 6.1         |             |
| <b>Secondary channel</b> |           |             |               |             |             |
| <b>Grand Totals</b>      | <b>53</b> |             | <b>17,003</b> |             |             |

**Appendix 0.22 Summary report for Reach # 2 of the West Fork Lake Creek  
Watershed data collected during 1993 and 1994.**

| Parameter  | 1993             | 1994        |
|--|------------------|-------------|
| Elevation  | 2,570 - 2,970 ft | same        |
| Total length   | 2,893 ft         | 7,556 ft    |
| Stream order   | 3                | 4.1         |
| Mean stream gradient                                   | 4.0              | 4.1         |
| Rifflepoolratio  | NA               | 1 :1.7      |
| Land use   |                  |             |
| <b>Forest</b>  | <b>100%</b>      | <b>100%</b> |
| <b>Agriculture</b>                                     |                  |             |
| <b>Livestock grazing</b>                               |                  |             |
| <b>Mining</b>  |                  |             |
| <b>Wetland</b>   |                  |             |
| <b>Floodplain</b>                                      |                  |             |
| <b>Other (includes residential right of way, etc.)</b> |                  |             |
| Vegetative type  |                  |             |
| <b>Decidious</b>                                       |                  |             |
| <b>Coniferous</b>                                      | 6                | 6           |
| <b>Mixed</b>   | 93               | 93          |
| Seal stage   |                  |             |
| <b>Grass/forb</b>                                      |                  |             |
| <b>Shrub</b>   |                  |             |
| <b>Pole</b>  |                  |             |
| <b>Young</b>   | <b>100</b>       | <b>100</b>  |
| <b>Mature</b>  |                  |             |
| <b>Old growth</b>                                      |                  |             |
| <b>Other</b>   |                  |             |
| x Canopy cover   | 77               | 41          |
| # Woody debris   |                  |             |
| <b>Logs</b>  | <b>106</b>       | <b>48</b>   |
| <b>Root wads</b>                                       | <b>13</b>        | <b>4</b>    |

**Appendix 0.23 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for West Fork Lake Creek during 1993.**

| Habitat                    | Frequency | %            | Total Area     | % Area       | Residual       |
|----------------------------|-----------|--------------|----------------|--------------|----------------|
| Type                       |           | Frequency    | (sq. meters)   |              | pool depth (m) |
| <b>Rapid</b>               | <b>2</b>  | <b>33.3</b>  | <b>412.1</b>   | <b>12.4</b>  |                |
| <b>Step pool cascade</b>   | <b>3</b>  | <b>50.0</b>  | <b>2,892.9</b> | <b>87.3</b>  |                |
| <b>Slip face cascade</b>   |           |              |                |              |                |
| Total Cascades             | <b>5</b>  | <b>83.3</b>  | <b>3,305.0</b> | <b>99.7</b>  |                |
| <b>Pocketwater</b>         |           |              |                |              |                |
| <b>Glide</b>               |           |              |                |              |                |
| <b>Run</b>                 |           |              |                |              |                |
| <b>Low gradient riffle</b> |           |              |                |              |                |
| Total Riffles              |           |              |                |              |                |
| <b>Dammed pool</b>         |           |              |                |              |                |
| <b>Eddy pool</b>           |           |              |                |              |                |
| <b>Plunge pool</b>         | 1         | 16.7         | 8.3            | 0.3          |                |
| <b>Scour pool</b>          |           |              |                |              |                |
| <b>Scour hole</b>          |           |              |                |              |                |
| <b>Beaver pond</b>         |           |              |                |              |                |
| Total Pools                | 1         | 16.7         | 8.3            | 0.3          |                |
| <b>Secondary channel</b>   |           |              |                |              |                |
| Grand Totals               | <b>6</b>  | <b>100.0</b> | <b>3,313.8</b> | <b>100.0</b> |                |

**Appendix D.24 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for West Fork Lake Creek during 1994.**

| Habitat                    | Frequency | %           | Total Area   | % Area      | Residual       |
|----------------------------|-----------|-------------|--------------|-------------|----------------|
| Type                       |           | Frequency   | (sq. meters) |             | pool depth (m) |
| <b>Rapid</b>               | <b>4</b>  | <b>22.2</b> | <b>517</b>   | <b>3.6</b>  |                |
| <b>Step pool cascade</b>   | <b>7</b>  | <b>38.9</b> | <b>5,308</b> | <b>36.5</b> |                |
| <b>Slip face cascade</b>   |           |             |              |             |                |
| Total Cascades             | 11        | 61.1        | 5,825        | 40.1        |                |
| <b>Pocketwater</b>         |           |             |              |             |                |
| <b>Glide</b>               |           |             |              |             |                |
| <b>Run</b>                 |           |             |              |             |                |
| <b>Low gradient riffle</b> | <b>2</b>  | <b>11.1</b> | <b>291</b>   | <b>2</b>    |                |
| Total Riffles              | 2         | 11.1        | 291          | 2           |                |
| <b>Dammed pool</b>         |           |             |              |             |                |
| <b>Eddy pool</b>           |           |             |              |             |                |
| <b>Plunge pool</b>         | 2         | 11.1        | 61           | 0.4         | 0.2            |
| <b>Scour pool</b>          | 2         | 11.1        | 213          | 1.5         |                |
| <b>Scour hole</b>          | 1         | 5.6         | 8,138        | 56.1        | 1.5            |
| <b>Beaver pond</b>         |           |             |              |             |                |
| Total Pools                | 5         | 67.9        | 8,412        | 58          |                |
| <b>Secondary channel</b>   |           |             |              |             |                |
| Grand Totals               | 18        |             | 14,528       |             |                |

**Appendix D.25 Summary report for Reach # 1 of the Bozard Creek  
Watershed data collected during 1993 and 1994.**

| Parameter  | 1993                    | 1994      |
|--|-------------------------|-----------|
| Elevation  | <b>2,540 - 2,590 ft</b> | <b>NA</b> |
| Total length   | <b>3,888.3 m</b>        |           |
| Stream order   | <b>3</b>                |           |
| Mean stream gradient                                   | <b>2.0</b>              |           |
| Riffle/pool ratio                                      | 12 : 1                  |           |
| Land use   |                         |           |
| <b>Forest</b>  | <b>50.5</b>             |           |
| <b>Agriculture</b>                                     | <b>49.5</b>             |           |
| <b>Livestock grazing</b>                               |                         |           |
| <b>Mining</b>  |                         |           |
| <b>Wetland</b>   |                         |           |
| <b>Floodplain</b>                                      |                         |           |
| <b>Other (includes residential right of way, etc.)</b> |                         |           |
| Vegetative type  |                         |           |
| <b>Deciduous</b>                                       |                         |           |
| <b>Coniferous</b>                                      |                         |           |
| <b>Mixed</b>   | <b>100%</b>             |           |
| Seral stage  |                         |           |
| <b>Grass/forb</b>                                      |                         |           |
| <b>Shrub</b>   | 92.3                    |           |
| <b>Pole</b>  |                         |           |
| <b>Young</b>   | <b>7.7</b>              |           |
| <b>Mature</b>  |                         |           |
| <b>Old growth</b>                                      |                         |           |
| <b>Other</b>   |                         |           |
| x Canopy cover   | 0.0                     |           |
| #Woody debris  |                         |           |
| Logs   |                         |           |
| <b>Root wads</b>                                       |                         |           |

**Appendix D.26 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for Bozard Creek during 1993.**

| Habitat                  | Frequency | %     | Total Area   | % Area | Residual |
|--------------------------|-----------|-------|--------------|--------|----------|
| Type                     | Frequency | %     | (sq. meters) |        | pool (m) |
| <b>Rapid</b>             |           |       |              |        |          |
| Step pool cascade        |           |       |              |        |          |
| Slip face cascade        |           |       |              |        |          |
| Total Cascades           |           |       |              |        |          |
| <b>Pocketwater</b>       |           |       |              |        |          |
| Glide                    | 5         | 38.5  | 9,672.5      | 37.5   |          |
| Run                      |           |       |              |        |          |
| Low gradient riffle      | 7         | 53.9  | 16,062.4     | 62.3   |          |
| Total Riffles            | 12        | 92.4  | 25,751.9     | 99.8   |          |
| <b>Dammed pool</b>       |           |       |              |        |          |
| Eddy pool                |           |       |              |        |          |
| Plunge pool              | 1         | 7.7   | 22.0         | 0.2    | 0.5      |
| Scour pool               |           |       |              |        |          |
| Scour hole               |           |       |              |        |          |
| Beaver pond              |           |       |              |        |          |
| Total Pools              | 1         | 7.7   | 22.0         | 0.2    |          |
| <b>Secondary channel</b> |           |       |              |        |          |
| Grand Totals             | 13        | 100.1 | 25,757.9     | 100.0  |          |

**Appendix D.27 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #for Creek during 1994.**

| Habit                    | Frequency | % | Total Area   | % Area | Residual       |
|--------------------------|-----------|---|--------------|--------|----------------|
| Type                     | Frequency | % | (sq. meters) |        | pool depth (m) |
| <b>Rapid</b>             |           |   |              |        |                |
| Step pool cascade        |           |   |              |        |                |
| Slip face cascade        |           |   |              |        |                |
| Total Cascades           |           |   |              |        |                |
| <b>Pocketwater</b>       |           |   |              |        |                |
| Glide                    |           |   |              |        |                |
| Run                      |           |   | NA           |        |                |
| Low gradient riffle      |           |   |              |        |                |
| Total Riffles            |           |   |              |        |                |
| <b>Dammed pool</b>       |           |   |              |        |                |
| Eddy pool                |           |   |              |        |                |
| Plunge pool              |           |   |              |        |                |
| Scour pool               |           |   |              |        |                |
| Scour hole               |           |   |              |        |                |
| Beaver pond              |           |   |              |        |                |
| Total Pools              |           |   |              |        |                |
| <b>Secondary channel</b> |           |   |              |        |                |
| Grand Totals             |           |   |              |        |                |

**Appendix D.28 Summary report for Reach # 1 of the Benawah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b>    |
|--|-------------------|----------------|
| <b>Elevation</b>                                 | 2, 250- 2,250 ft. | <b>same</b>    |
| <b>Total length</b>                              | <b>1,547 ft.</b>  | <b>1598 ft</b> |
| <b>Stream order</b>                              | 4                 | 4              |
| <b>Mean stream gradient</b>                      | <b>1.7</b>        | <b>1.0</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.7 ; 1</b>    | <b>2 ; 1</b>   |
| <b>Land use</b>                                  |                   |                |
| Forest   |                   |                |
| Agriculture                                      |                   |                |
| Livestock grazing                                |                   |                |
| Mining   |                   |                |
| Wetland  | 100 %             | 100 %          |
| Floodplain                                       |                   |                |
| Other (includes residential, right of way, etc.) |                   |                |
| <b>Vegetative type</b>                           |                   |                |
| Deciduous  |                   |                |
| Coniferous                                       |                   |                |
| Mixed  | 100 %             | 100 %          |
| <b>Seral stage</b>                               |                   |                |
| Grass/forb                                       |                   |                |
| Shrub  | 100 %             | 100 %          |
| Pole   |                   |                |
| Young  |                   |                |
| Mature   |                   |                |
| Old growth                                       |                   |                |
| Other  |                   |                |
| <b>x Canopy cover</b>                            | <b>0 %</b>        | <b>29.4 %</b>  |
| <b># Woody debris</b>                            |                   |                |
| Logs   | 8                 | 15             |
| Root wads  |                   | 1              |



**Appendix D.29 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b> | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|-------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. feet)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                   |               |                       |
| Step pool cascade     |                  |                  |                   |               |                       |
| Slip face cascade     |                  |                  |                   |               |                       |
| <b>Total Cascades</b> |                  |                  |                   |               |                       |
| Pocketwater           | 1                | 2.9              | 7795.3            | 27.5          |                       |
| Glide                 | 5                | 14.3             | 3600.6            | 12.7          |                       |
| Run                   |                  |                  |                   |               |                       |
| Low gradient riffle   | 14               | 40.0             | 9494.8            | 33.5          |                       |
| <b>Total Riffles</b>  | <b>20</b>        | <b>57.2</b>      | <b>20891</b>      | <b>73.7</b>   |                       |
| Dammed pool           |                  |                  |                   |               |                       |
| Eddy pool             |                  |                  |                   |               |                       |
| Plunge pool           | 2                | 5.7              | 507.5             | 1.8           | 0.8                   |
| Scour pool            | 13               | 37.14            | 5824.4            | 20.6          | 1.0                   |
| Scour hole            |                  |                  |                   |               |                       |
| Beaver pond           |                  |                  |                   |               |                       |
| <b>Total Pools</b>    | <b>15</b>        | <b>42.64</b>     | <b>6332</b>       | <b>22.4</b>   |                       |
| Secondary channel     |                  |                  |                   |               |                       |
| <b>Grand Totals</b>   | <b>35</b>        |                  | <b>27.223</b>     |               |                       |

**Appendix D.30. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #1 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 3                | 9.1              | 732                 | 21.8          |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>3</b>         | <b>9.1</b>       | <b>732</b>          | <b>21.8</b>   |                       |
| Pocketwater           | 3                | 9.1              | 933                 | 27.7          |                       |
| Glide                 | 4                | 12.1             | 573                 | 17.0          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 13               | 39.4             | 727                 | 21.6          |                       |
| <b>Total Riffles</b>  | <b>20</b>        | <b>60.6</b>      | <b>2,233</b>        | <b>66.3</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 10               | 30.3             | 401                 | 11.9          | 0.7                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>10</b>        | <b>30.3</b>      | <b>401</b>          | <b>11.9</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>33</b>        |                  | <b>3,366</b>        |               |                       |

**Appendix D.31 Summary report for Reach # 2 of the Benawah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b>    |
|--|-------------------|----------------|
| <b>Elevation</b>                                 | 2,250 - 2,330 ft. | <b>same</b>    |
| <b>Total length</b>                              | 4,402 ft.         | <b>4645 ft</b> |
| <b>Stream order</b>                              | 4                 | 4              |
| <b>Mean stream gradient</b>                      | 2.0               | 2.9            |
| <b>Riffle/pool ratio</b>                         | 4.3 ; 1           | 7 ; 1          |
| <b>Land use</b>                                  |                   |                |
| Forest   | 100 %             | 100 %          |
| Agriculture                                      |                   |                |
| Livestock grazing                                |                   |                |
| Mining   |                   |                |
| Wetland  |                   |                |
| Floodplain                                       |                   |                |
| Other (includes residential, right of way, etc.) |                   |                |
| <b>Vegetative type</b>                           |                   |                |
| Deciduous  |                   |                |
| Coniferous                                       |                   |                |
| Mixed  | 100 %             | 100 %          |
| <b>Seral stage</b>                               |                   |                |
| Grass/forb                                       |                   |                |
| Shrub  | 100 %             | 100 %          |
| Pole   |                   |                |
| Young  |                   |                |
| Mature   |                   |                |
| Old growth                                       |                   |                |
| Other  |                   |                |
| <b>x Canopy cover</b>                            | 80%               | 37%            |
| <b># Woody debris</b>                            |                   |                |
| Logs   | 5                 | 6              |
| Root wads  |                   |                |

**Appendix D.32 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 6                | 27.3             | 6790.6              | 13.3          | 0.3                   |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>6</b>         | <b>27.3</b>      | <b>6790.6</b>       | <b>13.3</b>   | <b>0.3</b>            |
| Pocketwater           | 5                | 22.7             | 11815.3             | 23.1          |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 8                | 36.4             | 32077.2             | 62.8          |                       |
| <b>Total Riffles</b>  | <b>13</b>        | <b>59.1</b>      | <b>43892.5</b>      | <b>85.9</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 1                | 4.5              | 108.8               | 0.2           | 0.4                   |
| Scour pool            | 2                | 9.1              | 271.3               | 0.5           | 0.3                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>3</b>         | <b>13.6</b>      | <b>380.1</b>        | <b>0.7</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>22</b>        |                  | <b>51.063</b>       |               |                       |

**Appendix D.33 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #2 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 4                | 33.3             | 5591                | 27.5          |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>4</b>         | <b>33.3</b>      | <b>5.59</b>         | <b>27.5</b>   |                       |
| Pocketwater           | 5                | 41.7             | 10,538              | 51.8          |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 2                | 16.6             | 3721                | 18.3          |                       |
| <b>Total Riffles</b>  | <b>7</b>         | <b>58.3</b>      | <b>14,259</b>       | <b>70.1</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 1                | 8.3              | 507                 | 2.5           | 0.4                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>1</b>         |                  |                     | <b>2.5</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>12</b>        |                  | <b>507</b>          |               |                       |

**Appendix D.34 Summary report for Reach # 3 of the Benawah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>      |
|--|--------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,330 - 2,330 ft.</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>2,376 ft.</b>         | <b>3,788 ft.</b> |
| <b>Stream order</b>                              | <b>4</b>                 | <b>4</b>         |
| <b>Mean stream gradient</b>                      | <b>2.5</b>               | <b>2.0</b>       |
| <b>Riffle/pool ratio</b>                         | <b>1.3 ; 1</b>           | <b>2.5 ; 1</b>   |
| <b>Land use</b>                                  |                          |                  |
| Forest   | 100%                     | 100%             |
| Agriculture                                      |                          |                  |
| Livestock grazing                                |                          |                  |
| Mining   |                          |                  |
| Wetland  |                          |                  |
| Floodplain                                       |                          |                  |
| Other (includes residential, right of way, etc.) |                          |                  |
| <b>Vegetative type</b>                           |                          |                  |
| Deciduous  |                          |                  |
| Coniferous                                       |                          |                  |
| Mixed  | 100%                     | 100%             |
| <b>Seral stage</b>                               |                          |                  |
| Grass/forb                                       |                          |                  |
| Shrub  | 100%                     | 100%             |
| Pole   |                          |                  |
| Young  |                          |                  |
| Mature   |                          |                  |
| Old growth                                       |                          |                  |
| Other  |                          |                  |
| <b>x Canopy cover</b>                            | <b>28.5 %</b>            | <b>40 %</b>      |
| <b># Woody debris</b>                            |                          |                  |
| Logs   | 3                        | 1                |
| Root wads  | 0                        | 0                |

**Appendix D.35 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           | 1                | 6.25             | 1408                | 4.8           |                       |
| Glide                 | 1                | 6.25             | 583                 | 1.97          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 7                | 43.75            | 24311               | 82.29         |                       |
| <b>Total Riffles</b>  | <b>9</b>         | <b>58.25</b>     | <b>26302</b>        | <b>89.06</b>  |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             | 1                | 6.3              | 109.7               | 0.4           | 0.79                  |
| Plunge pool           |                  |                  |                     |               |                       |
| scour pool            | 6                | 37.5             | 3129                | 10.6          | 0.56                  |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>7</b>         | <b>43.8</b>      | <b>3239</b>         | <b>11.0</b>   |                       |
| Sewn? ary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>16</b>        |                  | <b>29,541</b>       |               |                       |

**Appendix D.36 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #3 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 1                | 7.7              | 1746                | 7.7           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>1</b>         | <b>7.7</b>       | <b>1,748</b>        | <b>7.7</b>    |                       |
| Pocketwater           | 2                | 15.4             | 1597                | 7.1           |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient rime     | 6                | 46.2             | 18,429              | 81.6          |                       |
| <b>Total Riffles</b>  | <b>8</b>         | <b>61.6</b>      | <b>20,026</b>       | <b>88.7</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 4                | 44.2             | 814                 | 3.6           | 0.4                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>4</b>         | <b>44.2</b>      | <b>814</b>          | <b>3.6</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>13</b>        |                  |                     | <b>22,856</b> |                       |

**Appendix D.37 Summary report for Reach # 4 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b>    |
|--|-------------------|----------------|
| <b>Elevation</b>                                 | 2,330 - 2,360 ft. | <b>same</b>    |
| <b>Total length</b>                              | 4,425 ft.         | 5,670 ft.      |
| <b>Stream order</b>                              | 4                 |                |
| <b>Mean stream gradient</b>                      | 1.5               | 2.0            |
| <b>Riffle/pool ratio</b>                         | <b>1 ;1</b>       | <b>1.8 ; 1</b> |
| <b>Land use</b>                                  |                   |                |
| Forest   | 100%              | 100%           |
| Agriculture                                      |                   |                |
| Livestock grazing                                |                   |                |
| Mining   |                   |                |
| Wetland  |                   |                |
| Floodplain                                       |                   |                |
| Other (includes residential, right of way, etc.) |                   |                |
| <b>Vegetative type</b>                           |                   |                |
| Decidious  |                   |                |
| Coniferous                                       |                   |                |
| Mixed  | 100%              | 100%           |
| <b>Seral stage</b>                               |                   |                |
| Grass/forb                                       |                   |                |
| Shrub  | 100%              | 100%           |
| Pole   |                   |                |
| Young  |                   |                |
| Mature   |                   |                |
| Old growth                                       |                   |                |
| Other  |                   |                |
| <b>x Canopy cover</b>                            | <b>-ND-</b>       | <b>17.7 %</b>  |
| <b># Woody debris</b>                            |                   |                |
| Logs   | 3                 | 3              |
| Root wads  | 0                 | 0              |

**Appendix D.38 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # for Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 6                | 50.0             | 30768.3             | 95.3          |                       |
| <b>Total Riffles</b>  | 6                | 60.0             | 30768.3             | 95.3          |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 6                | 50.0             | 1506.9              | 4.7           | 0.33                  |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | 6                | 60.0             | 1506.9              | 4.7           |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | 12               | 100.0            | 32275.2             | 100.0         |                       |

**Appendix D.39 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 4                | 26.7             | 9845                | 22.3          |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | 4                | 26.7             | 9,845               | 22.3          |                       |
| Pocketwater           | 5                | 33.3             | 26,163              | 59.3          |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 2                | 13.3             | 6040                | 13.7          |                       |
| <b>Total Riffles</b>  | 7                | 46.6             | 32,203              | 73.0          |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             | 1                | 6.7              | 55                  | 0.1           | 0.2                   |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 3                | 20.0             | 1,990               | 4.5           | 1.1                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | 4                | 26.7             | 2,045               | 4.6           |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | 15               |                  | 44,093              |               |                       |

**Appendix D.40 Summary report for Reach # 5 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>      |
|--|-------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,380 -2,490 ft.</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>3,467 ft.</b>        | <b>4,081 ft.</b> |
| <b>Stream order</b>                              | <b>4</b>                | <b>4</b>         |
| <b>Mean stream gradient</b>                      | <b>3 %</b>              | <b>2.0</b>       |
| <b>Riffle/pool ratio</b>                         | <b>3 ; 1</b>            | <b>1 ; 0</b>     |
| <b>Land use</b>                                  |                         |                  |
| Forest   | 100%                    | 100%             |
| Agriculture                                      |                         |                  |
| Livestock grazing                                |                         |                  |
| Mining   |                         |                  |
| Wetland  |                         |                  |
| Floodplain                                       |                         |                  |
| Other (includes residential, right of way, etc.) |                         |                  |
| <b>Vegetative type</b>                           |                         |                  |
| Decidious  |                         |                  |
| Coniferous                                       |                         |                  |
| Mixed  | 100%                    | 100%             |
| <b>Seral stage</b>                               |                         |                  |
| Grass/forb                                       |                         |                  |
| Shrub  | 100 %                   | 100%             |
| Pole   |                         |                  |
| Young  |                         |                  |
| Mature   |                         |                  |
| Old growth                                       |                         |                  |
| Other  |                         |                  |
| <b>x Canopy cover</b>                            | <b>-ND-</b>             | <b>16.8 %</b>    |
| <b># Woody debris</b>                            |                         |                  |
| Logs   |                         | 3                |
| Root wads  | 0                       | 0                |



**Appendix D.41 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Benewah Creek during 1993.**

| Habitat               | Frequency | %           | Total Area    | % Area      | Residual       |
|-----------------------|-----------|-------------|---------------|-------------|----------------|
| Type                  |           | Frequency   | (sq. meters)  |             | pool depth (m) |
| Rapid                 |           |             |               |             |                |
| Step pool cascade     | 3         | 27.3        | 2114          | 6.8         | 0.3            |
| Slip face cascade     |           |             |               |             |                |
| <b>Total Cascades</b> | <b>3</b>  | <b>27.3</b> | <b>2114</b>   | <b>6.8</b>  | <b>0.3</b>     |
| Pocketwater           |           |             |               |             |                |
| Glide                 |           |             |               |             |                |
| Run                   |           |             |               |             |                |
| Low gradient riffle   | 6         | 54.6        | 28,530        | 92.2        |                |
| <b>Total Riffles</b>  | <b>6</b>  | <b>64.6</b> | <b>28630</b>  | <b>92.2</b> |                |
| Dammed pool           |           |             |               |             |                |
| Eddy pool             |           |             |               |             |                |
| Plunge pool           |           |             |               |             |                |
| Scour pool            | 2         | 18.2        | 312           | 1.0         | 0.21           |
| Scour hole            |           |             |               |             |                |
| Beaver pond           |           |             |               |             |                |
| <b>Total Pools</b>    | <b>2</b>  | <b>18.2</b> | <b>312</b>    | <b>1.0</b>  |                |
| Secondary channel     |           |             |               |             |                |
| <b>Grand Totals</b>   | <b>11</b> |             | <b>30,956</b> |             |                |

**Appendix D.42 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Benewah Creek during 1994.**

| Habitat               | Frequency | %           | Total Area    | % Area      | Residual       |
|-----------------------|-----------|-------------|---------------|-------------|----------------|
| Type                  |           | Frequency   | (sq. meters)  |             | pool depth (m) |
| Rapid                 |           |             |               |             |                |
| Step pool cascade     | 7         | 45.8        | 5,134         | 17.6        |                |
| Slip face cascade     |           |             |               |             |                |
| <b>Total Cascades</b> | <b>7</b>  | <b>43.8</b> | <b>5,134</b>  | <b>17.6</b> |                |
| Pocketwater           | 1         | 6.3         | 358           | 1           |                |
| Glide                 |           |             |               |             |                |
| Run                   |           |             |               |             |                |
| Low gradient riffle   | 8         | 50          | 23,664        | 81.2        |                |
| <b>Total Riffles</b>  | <b>9</b>  | <b>66.3</b> | <b>24,022</b> | <b>82.4</b> |                |
| Dammed pool           |           |             |               |             |                |
| Eddy pool             |           |             |               |             |                |
| Plunge pool           |           |             |               |             |                |
| Scour pool            |           |             |               |             |                |
| Scour hole            |           |             |               |             |                |
| Beaver pond           |           |             |               |             |                |
| <b>Total Pools</b>    |           |             |               |             |                |
| Secondary channel     |           |             |               |             |                |
| <b>Grand Totals</b>   | <b>16</b> |             | <b>29,156</b> |             |                |

**Appendix D.43 Summary report for Reach # 6 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b> |
|--|-------------------|-------------|
| <b>Elevation</b>                                 | 2,490 - 2,550 ft. | same        |
| <b>Total length</b>                              | 2,741 ft.         | 2,618 ft    |
| <b>Stream order</b>                              | 4                 | 4           |
| <b>Mean stream gradient</b>                      | 1.0 %             | 1.5 %       |
| <b>Riffle/pool ratio</b>                         | 1.5 ; 1           | 3 ; 1       |
| <b>Land use</b>                                  |                   |             |
| Forest   |                   |             |
| Agriculture                                      |                   |             |
| Livestock grazing                                | 100 %             | 100%        |
| Mining   |                   |             |
| Wetland  |                   |             |
| Floodplain                                       |                   |             |
| Other (includes residential, right of way, etc.) |                   |             |
| <b>Vegetative type</b>                           |                   |             |
| Deciduous  |                   |             |
| Coniferous                                       |                   |             |
| Mixed  | 100%              | 100 %       |
| <b>Seral stage</b>                               |                   |             |
| Grass/forb                                       |                   |             |
| Shrub  | 100 %             | 100%        |
| Pole   |                   |             |
| Young  |                   |             |
| Mature   |                   |             |
| Old growth                                       |                   |             |
| Other  |                   |             |
| <b>x Canopy cover</b>                            | -NA-              | 3 %         |
| <b># Woody debris</b>                            |                   |             |
| Logs   |                   | 0           |
| Root wads  | 0                 | 0           |

**Appendix D.44 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocket-water          |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient rime     | 3                | 60.0             | 6,651               | 95.1          |                       |
| <b>Total Riffles</b>  | <b>3</b>         | <b>60.0</b>      | <b>6,651</b>        | <b>95.1</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 2                | 40.0             | 345                 | 49            | 0.23                  |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>2</b>         | <b>40.0</b>      | <b>345</b>          | <b>49</b>     |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>5</b>         |                  | <b>6,995</b>        |               |                       |

**Appendix D.45 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 1                | 20               | 2,019               | 15.9          |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>1</b>         | <b>20</b>        | <b>2,019</b>        | <b>15.9</b>   |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 1                | 20               | 2,347               | 18.5          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 2                | 40               | 8,032               | 63.3          |                       |
| <b>Total Riffles</b>  | <b>3</b>         | <b>50</b>        | <b>10,379</b>       | <b>81.8</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 1                | 20               | 296                 | 23            | 0.6                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>1</b>         | <b>20</b>        | <b>296</b>          | <b>23</b>     |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>6</b>         |                  | <b>12,694</b>       |               |                       |

**Appendix D.46 Summary report for Reach # 7 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>          | <b>1994</b>      |
|--|----------------------|------------------|
| <b>Elevation</b>                                 | <b>2,550 - 2,660</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>10,477 ft.</b>    | <b>12,090 ft</b> |
| <b>Stream order</b>                              | <b>4</b>             | <b>4</b>         |
| <b>Mean stream gradient</b>                      | <b>1.4 %</b>         | <b>1.2 %</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.2 ; 1</b>       | <b>3.1 ; 1</b>   |
| <b>Land use</b>                                  |                      |                  |
| Forest   |                      |                  |
| Agriculture                                      |                      |                  |
| Livestock grazing                                | 100 %                | 100%             |
| Mining   |                      |                  |
| Wetland  |                      |                  |
| Floodplain                                       |                      |                  |
| Other (includes residential, right of way, etc.) |                      |                  |
| <b>Vegetative type</b>                           |                      |                  |
| Deciduous  |                      |                  |
| Coniferous                                       |                      |                  |
| Mixed  | 100%                 | 100%             |
| <b>Seral stage</b>                               |                      |                  |
| Grass/forb                                       |                      |                  |
| Shrub  | 100%                 | 100 %            |
| Pole   |                      |                  |
| Young  |                      |                  |
| Mature   |                      |                  |
| Old growth                                       |                      |                  |
| Other  |                      |                  |
| <b>x Canopy cover</b>                            | <b>48.7 %</b>        | <b>43 %</b>      |
| <b># Woody debris</b>                            |                      |                  |
| Logs   | 1                    | 10               |
| Root wads  | 0                    | 9                |

**Appendix D.47 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 7 for Benewah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 9                | 12.3             | 10,212              | 19.1          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 31               | 42.5             | 38,191              | 71.3          |                       |
| <b>Total Riffles</b>  | <b>40</b>        | <b>54.8</b>      | <b>48,403</b>       | <b>90.4</b>   |                       |
| Dammed pool           | 1                | 1.4              | 1,262               | 2.3           | 0.9                   |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 32               | 43.8             | 3,891               | 7.3           | 0.6                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>33</b>        | <b>45.2</b>      | <b>5,153</b>        | <b>9.6</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>73</b>        |                  | <b>53,557</b>       |               |                       |

**Appendix D.48 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #7 for Benewah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 2                | 2.6              | 917                 | 1.7           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>2</b>         | <b>2.6</b>       | <b>917</b>          | <b>1.7</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 23               | 29.9             | 19,339              | 35.7          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient rime     | 34               | 44.2             | 25,112              | 46.4          |                       |
| <b>Total Riffles</b>  | <b>57</b>        | <b>74.1</b>      | <b>44,451</b>       | <b>82.1</b>   |                       |
| Dammed pool           | 1                | 1.3              | 5,121               | 9.5           | 0.5                   |
| Eddy pool             | 1                | 1.3              | 292                 | 0.5           | 1.4                   |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 16               | 20.8             | 3,396               | 6.3           | 0.6                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>18</b>        | <b>23.4</b>      | <b>8,809</b>        | <b>16.3</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>77</b>        |                  | <b>54,177</b>       |               |                       |

**Appendix D.48 Summary report for Reach # 8 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>      |
|--|-------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,560 -2,640 ft.</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>9,035 ft.</b>        | <b>9,174 ft.</b> |
| <b>Stream order</b>                              | <b>4</b>                |                  |
| <b>Mean stream gradient</b>                      | <b>1.8 %</b>            | <b>1.8 %</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.3 ; 1</b>          |                  |
| <b>Land use</b>                                  |                         |                  |
| Forest   |                         |                  |
| Agriculture                                      |                         |                  |
| Livestock grazing                                | <b>100%</b>             | <b>100%</b>      |
| Mining   |                         |                  |
| Wetland  |                         |                  |
| Floodplain                                       |                         |                  |
| Other (includes residential, right of way, etc.) |                         |                  |
| <b>Vegetative type</b>                           |                         |                  |
| Deciduous  |                         |                  |
| Coniferous                                       |                         |                  |
| Mixed  | <b>100 %</b>            | <b>100 %</b>     |
| <b>Seral stage</b>                               |                         |                  |
| Grass/forb                                       |                         |                  |
| Shrub  | <b>100%</b>             | <b>100%</b>      |
| Pole   |                         |                  |
| Young  |                         |                  |
| Mature   |                         |                  |
| Old growth                                       |                         |                  |
| Other  |                         |                  |
| <b>x Canopy cover</b>                            | <b>21.4 %</b>           | <b>26.1 %</b>    |
| <b># Woody debris</b>                            |                         |                  |
| Logs   | <b>.21</b>              | <b>6</b>         |
| Root wads  | <b>2</b>                | <b>0</b>         |

**Appendix D.49** Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 8 for Benawah Creek during 1993.

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 17               | 19.8               | 13,162                         | 35.8          |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient rime     | 33               | 38.4               | 9,717                          | 26.4          |                                |
| <b>Total Riffles</b>  | <b>50</b>        | <b>58.2</b>        | <b>2,288</b>                   | <b>62</b>     |                                |
| Dammed pool           | 1                | 1.2                | 21,800                         | 5.9           | 1.13                           |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 35               | 40.7               | 11,705                         | 31.8          | 0.77                           |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>36</b>        | <b>41.9</b>        | <b>33,505</b>                  | <b>37.7</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>86</b>        |                    | <b>35,793</b>                  |               |                                |

**Appendix D.50** Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #8 for Benawah Creek during 1994.

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 16               | 19.0               | 12,482                         | 39            |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 40               | 47.6               | 9,646                          | 30.1          |                                |
| <b>Total Riffles</b>  | <b>56</b>        | <b>66.6</b>        | <b>22,128</b>                  | <b>69.1</b>   |                                |
| Dammed pool           | 1                | 1.2                | 1,668                          | 5.2           | 0.8                            |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 27               | 32.1               | 8,212                          | 25.7          | 0.7                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>28</b>        | <b>33.3</b>        | <b>9,880</b>                   | <b>30.9</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>84</b>        |                    | <b>32,007</b>                  |               |                                |

**Appendix D.51 Summary report for Reach # 9 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>      |
|--|--------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,640 - 2,640 ft.</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>3,386 ft.</b>         | <b>12,850 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                 | <b>0</b>         |
| <b>Mean stream gradient</b>                      | <b>1.75 %</b>            | <b>1.3 %</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.7 ; 1</b>           | <b>2.8 ; 1</b>   |
| <b>Land use</b>                                  |                          |                  |
| Forest   |                          |                  |
| Agriculture                                      |                          |                  |
| Livestock grazing                                | 100 %                    | 100 %            |
| Mining   |                          |                  |
| Wetland  |                          |                  |
| Floodplain                                       |                          |                  |
| Other (includes residential, right of way, etc.) |                          |                  |
| <b>Vegetative type</b>                           |                          |                  |
| Decidious  |                          |                  |
| Coniferous                                       |                          |                  |
| Mixed  | 100 %                    | 100 %            |
| <b>Seral stage</b>                               |                          |                  |
| Grass/forb                                       |                          |                  |
| Shrub  | 100%                     | 100 %            |
| Pole   |                          |                  |
| Young  |                          |                  |
| Mature   |                          |                  |
| Old growth                                       |                          |                  |
| Other  |                          |                  |
| <b>x Canopy cover</b>                            | <b>45 %</b>              | <b>10.5 %</b>    |
| <b># Woody debris</b>                            |                          |                  |
| Logs   | 29                       | 71               |
| Root wads  |                          | 3                |



**Appendix D.52 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 9 for Benewah Creek during 1993.**

| Habitat Type          | Frequency | % Frequency | Total Area (sq. meters) | % Area | Residual pool depth (m) |
|-----------------------|-----------|-------------|-------------------------|--------|-------------------------|
| Rapid                 |           |             |                         |        |                         |
| Step pool cascade     |           |             |                         |        |                         |
| Slip face cascade     |           |             |                         |        |                         |
| <b>Total Cascades</b> |           |             |                         |        |                         |
| Pocketwater           |           |             |                         |        |                         |
| Glide                 | 11        | 31.4        | 6,317                   | 23.3   |                         |
| Run                   |           |             |                         |        |                         |
| Low gradient riffle   | 11        | 31.4        | 26,650                  | 9.8    |                         |
| <b>Total Riffles</b>  | 22        | 62.8        | 8,982                   | 33.0   |                         |
| Dammed pool           | 3         | 8.6         | 14,715                  | 54.0   | 0.83                    |
| Eddy pool             |           |             |                         |        |                         |
| Plunge pool           |           |             |                         |        |                         |
| Scour pool            | 10        | 28.6        | 3,535.4                 | 12.9   | 0.88                    |
| Scour hole            |           |             |                         |        |                         |
| Beaver pond           |           |             |                         |        |                         |
| <b>Total Pools</b>    | 13        | 37.2        | 18,250                  | 66.9   |                         |
| Secondary channel     |           |             |                         |        |                         |
| <b>Grand Totals</b>   | 35        |             | 27,232                  |        |                         |

**Appendix D.53 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #9 for Benewah Creek during 1994.**

| Habitat Type          | Frequency | % Frequency | Total Area (sq. meters) | % Area | Residual pool depth (m) |
|-----------------------|-----------|-------------|-------------------------|--------|-------------------------|
| Rapid                 |           |             |                         |        |                         |
| Step pool cascade     |           |             |                         |        |                         |
| Slip face cascade     | 2         | 3.3         | 390                     | 0.7    |                         |
| <b>Total Cascades</b> | 2         | 3.3         | 390                     | 0.7    |                         |
| Pocketwater           |           |             |                         |        |                         |
| Glide                 | 17        | 28.3        | 13,149                  | 21.9   |                         |
| Run                   |           |             |                         |        |                         |
| Low gradient riffle   | 26        | 43.3        | 9346                    | 15.6   |                         |
| <b>Total Riffles</b>  | 43        | 71.6        | 22,495                  | 37.5   |                         |
| Dammed pool           | 4         | 6.7         | 35170                   | 58.6   | 1.0                     |
| Eddy pool             |           |             |                         |        |                         |
| Plunge pool           | 2         | 3.3         | 433                     | 0.7    | 0.7                     |
| Scour pool            | 9         | 15.0        | 1528                    | 2.5    | 0.7                     |
| Scour hole            |           |             |                         |        |                         |
| Beaver pond           |           |             |                         |        |                         |
| <b>Total Pools</b>    | 15        | 25          | 37,131                  | 61.8   |                         |
| Secondary channel     |           |             |                         |        |                         |
| <b>Grand Totals</b>   | 60        |             | 60,016                  |        |                         |

**Appendix D.54 Summary report for Reach # 10 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>     |
|--|--------------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,640 - 2,750 ft.</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>9,339 ft.</b>         | <b>2,275 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                 | <b>4</b>        |
| <b>Mean stream gradient</b>                      | <b>1.75 %</b>            | <b>1.2 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>3.5 ; 1</b>           | <b>3.5 ; 1</b>  |
| <b>Land use</b>                                  |                          |                 |
| Forest   |                          |                 |
| Agriculture                                      |                          |                 |
| Livestock grazing                                | <b>100%</b>              | <b>100%</b>     |
| Mining   |                          |                 |
| Wetland  |                          |                 |
| Floodplain                                       |                          |                 |
| Other (includes residential, right of way, etc.) |                          |                 |
| <b>Vegetative type</b>                           |                          |                 |
| Deciduous  |                          |                 |
| Coniferous                                       |                          |                 |
| Mixed  | <b>100%</b>              | <b>100%</b>     |
| <b>Seral stage</b>                               |                          |                 |
| Grass/forb                                       |                          |                 |
| Shrub  | <b>100%</b>              | <b>100%</b>     |
| Pole   |                          |                 |
| Young  |                          |                 |
| Mature   |                          |                 |
| Old growth                                       |                          |                 |
| Other  |                          |                 |
| <b>x Canopy cover</b>                            |                          |                 |
| <b># Woody debris</b>                            |                          |                 |
| Logs   | <b>24.2</b>              | <b>NA</b>       |
| Root wads  | <b>0</b>                 |                 |

**Appendix D.55 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 10 for Benawah Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Packetwater           |                  |                    |                                |               |                                |
| Glide                 | 22               | 34.9               | 13,253                         | 3.4           |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 27               | 42.8               | 8,144                          | 20.6          |                                |
| <b>Total Riffles</b>  | <b>49</b>        | <b>77.7</b>        | <b>21,397</b>                  | <b>64.0</b>   |                                |
| Dammed pool           | 2                | 3.2                | 16,070                         | 40.5          | 1.22                           |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 1.6                | 57                             | 0.1           | 0.76                           |
| Scour pool            | 11               | 17.5               | 2,113                          | 3.3           | 0.73                           |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>14</b>        | <b>22.3</b>        | <b>18,240</b>                  | <b>44</b>     |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>63</b>        |                    | <b>39,637</b>                  |               |                                |

**Appendix D.56 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #10 for Benawah Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 4                | 44.4               | 2337                           | 41.1          |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 3                | 33.3               | 3171                           | 55.7          |                                |
| <b>Total Riffles</b>  | <b>7</b>         | <b>77.7</b>        | <b>5,508</b>                   | <b>96.8</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 2                | 22.2               | 18.1                           | 3.2           | 0.5                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>2</b>         | <b>22.2</b>        | <b>18.1</b>                    | <b>3.2</b>    |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>9</b>         |                    | <b>5,689</b>                   |               |                                |

**Appendix D.57 Summary report for Reach # 11 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>    |
|--|--------------------------|----------------|
| <b>Elevation</b>                                 | <b>2,750 - 2,760 ft.</b> | <b>same</b>    |
| <b>Total length</b>                              | <b>1,935 ft.</b>         | <b>3758 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                 |                |
| <b>Mean stream gradient</b>                      |                          | <b>1.2 %</b>   |
| <b>Riffle/pool ratio</b>                         | <b>2.4 ; 1</b>           | <b>2.3 ; 1</b> |
| <b>Land use</b>                                  |                          |                |
| Forest   |                          |                |
| Agriculture                                      |                          |                |
| Livestock grazing                                | <b>100 %</b>             | <b>100%</b>    |
| Mining   |                          |                |
| Wetland  |                          |                |
| Floodplain                                       |                          |                |
| Other (includes residential, right of way, etc.) |                          |                |
| <b>Vegetative type</b>                           |                          |                |
| Decidious  |                          |                |
| Coniferous                                       |                          |                |
| Mixed  | <b>100 %</b>             | <b>100%</b>    |
| <b>Seral stage</b>                               |                          |                |
| Grass/forb                                       |                          |                |
| Shrub  | <b>100%</b>              | <b>100%</b>    |
| Pole   |                          |                |
| Young  |                          |                |
| Mature   |                          |                |
| Old growth                                       |                          |                |
| Other  |                          |                |
| <b>x Canopy cover</b>                            | <b>21.4 %</b>            | <b>35 %</b>    |
| <b># Woody debris</b>                            |                          |                |
| Logs   | <b>25</b>                | <b>32</b>      |
| Rwt wads   |                          |                |

**Appendix D.58 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 11 for Benawah Creek during 1993.**

| Habitat Type          | Frequency | % Frequency | Total Area (sq. meters) | % Area       | Residual pool depth (m) |
|-----------------------|-----------|-------------|-------------------------|--------------|-------------------------|
| Rapid                 |           |             |                         |              |                         |
| Step pool cascade     |           |             |                         |              |                         |
| Slip face cascade     |           |             |                         |              |                         |
| <b>Total Cascades</b> |           |             |                         |              |                         |
| Pocketwater           |           |             |                         |              |                         |
| Glide                 | 7         | 22.6        | 1,824                   | 29.3         |                         |
| Run                   |           |             |                         |              |                         |
| Low gradient riffle   | 15        | 48.4        | 2,677                   | 42.9         |                         |
| <b>Total Riffles</b>  | <b>22</b> | <b>71.0</b> | <b>4,501</b>            | <b>72.2</b>  |                         |
| Dammed pool           |           |             |                         |              |                         |
| Eddy pool             |           |             |                         |              |                         |
| Plunge pool           | 1         | 3.2         | 429                     | 6.9          | 1.15                    |
| Scour pool            | 8         | 25.8        | 1,302                   | 20.89        | 0.4                     |
| Scour hole            |           |             |                         |              |                         |
| Beaver pond           |           |             |                         |              |                         |
| <b>Total Pools</b>    | <b>9</b>  | <b>29.0</b> | <b>1,731</b>            | <b>27.79</b> |                         |
| Secondary channel     |           |             |                         |              |                         |
| <b>Grand Totals</b>   | <b>31</b> |             | <b>6,232</b>            |              |                         |

**Appendix D.59 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #11 for Benawah Creek during 1994.**

| Habitat Type          | Frequency | % Frequency | Total Area (sq. meters) | % Area      | Residual pool depth (m) |
|-----------------------|-----------|-------------|-------------------------|-------------|-------------------------|
| Rapid                 |           |             |                         |             |                         |
| Step pool cascade     |           |             |                         |             |                         |
| Slip face cascade     |           |             |                         |             |                         |
| <b>Total Cascades</b> |           |             |                         |             |                         |
| Pocketwater           |           |             |                         |             |                         |
| Glide                 | 2         | 20          | 587                     | 13.2        |                         |
| Run                   |           |             |                         |             |                         |
| Low gradient riffle   | 5         | 50          | 3,447                   | 77.7        |                         |
| <b>Total Riffles</b>  | <b>7</b>  | <b>70</b>   | <b>4,034</b>            | <b>90.9</b> |                         |
| Dammed pool           |           |             |                         |             |                         |
| Eddy pool             |           |             |                         |             |                         |
| Plunge pool           | 1         | 10          | 55                      | 1.2         | 0.4                     |
| Scour pool            | 2         | 20          | 350                     | 7.9         | 0.4                     |
| Scour hole            |           |             |                         |             |                         |
| Beaver pond           |           |             |                         |             |                         |
| <b>Total Pools</b>    | <b>3</b>  | <b>30</b>   | <b>495</b>              | <b>9.1</b>  |                         |
| Secondary channel     |           |             |                         |             |                         |
| <b>Grand Totals</b>   | <b>10</b> |             | <b>4,439</b>            |             |                         |

**Appendix D.60 Summary report for Reach # 12 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>     |
|--|-------------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,760 -2,940 ft.</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>4,127 ft.</b>        | <b>4,633 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                | <b>4</b>        |
| <b>Mean stream gradient</b>                      | <b>2.3 %</b>            | <b>1.3 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>1.3 ; 1</b>          | <b>2 :1</b>     |
| <b>Land use</b>                                  |                         |                 |
| Forest   | 100%                    | 100 %           |
| Agriculture                                      |                         |                 |
| Livestock grazing                                |                         |                 |
| Mining   |                         |                 |
| Wetland  |                         |                 |
| Floodplain                                       |                         |                 |
| Other (includes residential, right of way, etc.) |                         |                 |
| <b>Vegetative type</b>                           |                         |                 |
| Decidious  |                         |                 |
| Coniferous                                       |                         |                 |
| Mixed  | 100%                    | 100%            |
| <b>Seral stage</b>                               |                         |                 |
| Grass/forb                                       |                         |                 |
| Shrub  | 100%                    | 100%            |
| Pole   |                         |                 |
| Young  |                         |                 |
| Mature   |                         |                 |
| Old growth                                       |                         |                 |
| Other  |                         |                 |
| <b>x Canopy cover</b>                            | <b>51.6 %</b>           | <b>44 %</b>     |
| <b># Woody debris</b>                            |                         |                 |
| Logs   | 335                     | 186             |
| Root wads  |                         |                 |

**Appendix D.61 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 12 for Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>%</b>      | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     |                  |             |                     |               |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> |                  |             |                     |               |                       |
| Pocketwater           |                  |             |                     |               |                       |
| Glide                 | 2                | 9.5         | 525                 | 5.2           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 10               | 47.6        | 8530                | 83.65         |                       |
| <b>Total Riffles</b>  | <b>12</b>        | <b>57.1</b> | <b>9055</b>         | <b>88.85</b>  |                       |
| Dammed pool           |                  |             |                     |               |                       |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           |                  |             |                     |               |                       |
| Scour pool            | 9                | 42.9        | 1,142               | 11.2          | 0.4                   |
| Scour hole            |                  |             |                     |               |                       |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>9</b>         | <b>42.9</b> | <b>1,142</b>        | <b>11.2</b>   |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>21</b>        |             | <b>10,197</b>       |               |                       |

**Appendix D.62 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #12 for Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>%</b>      | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     |                  |             |                     |               |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> |                  |             |                     |               |                       |
| Pocketwater           |                  |             |                     |               |                       |
| Glide                 | 3                | 20          | 418                 | 3.4           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 7                | 46.7        | 11,204              | 90.4          |                       |
| <b>Total Riffles</b>  | <b>10</b>        | <b>66.7</b> | <b>11,622</b>       | <b>93.8</b>   |                       |
| Dammed pool           | 2                | 13.3        | 205                 | 1.7           | 0.4                   |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           | 1                | 6.7         | 20.1                | 0.2           | 0.2                   |
| Scour pool            | 2                | 13.3        | 547                 | 4.4           | 0.7                   |
| Scour hole            |                  |             |                     |               |                       |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>5</b>         | <b>33.3</b> | <b>772.1</b>        | <b>6.3</b>    |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>15</b>        |             | <b>12,394</b>       |               | <b>0.4</b>            |

**Appendix D.63 Summary report for Reach # 13 of the Benewah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>    |
|--|--------------------------|----------------|
| <b>Elevation</b>                                 | <b>2,940 - 3,060 ft.</b> | <b>same</b>    |
| <b>Total length</b>                              | <b>4,948 ft.</b>         | <b>2498 ft</b> |
| <b>Stream order</b>                              | <b>2</b>                 | <b>2</b>       |
| <b>Mean stream gradient</b>                      | <b>6.0 %</b>             | <b>3.8 %</b>   |
| <b>Riffle/pool ratio</b>                         | <b>1 : 1.2</b>           | <b>2 : 1</b>   |
| <b>Land use</b>                                  |                          |                |
| Forest   | <b>100%</b>              | <b>100%</b>    |
| Agriculture                                      |                          |                |
| Livestock grazing                                |                          |                |
| Mining   |                          |                |
| Wetland  |                          |                |
| Floodplain                                       |                          |                |
| Other (includes residential, right of way, etc.) |                          |                |
| <b>Vegetative type</b>                           |                          |                |
| Deciduous  |                          |                |
| Coniferous                                       |                          |                |
| Mixed  | <b>100%</b>              | <b>100%</b>    |
| <b>Seral stage</b>                               |                          |                |
| Grass/forb                                       |                          |                |
| Shrub  | <b>100%</b>              | <b>100%</b>    |
| Pole   |                          |                |
| Young  |                          |                |
| Mature   |                          |                |
| Old growth                                       |                          |                |
| Other  |                          |                |
| <b>x Canopy cover</b>                            | <b>75.3 %</b>            | <b>84 %</b>    |
| <b># Woody debris</b>                            |                          |                |
| Logs   | <b>48.8</b>              | <b>131</b>     |
| Root wads  | <b>0</b>                 |                |



**Appendix D.64 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 13 for Benawah Creek during 1993.**

| Habitat Type          | Frequency | % Frequency  | Total Area (sq. meters) | % Area       | Residual pool depth (m) |
|-----------------------|-----------|--------------|-------------------------|--------------|-------------------------|
| Rapid                 | 6         | 17.65        | 3,153                   | 22.87        |                         |
| Step pool cascade     | 4         | 11.76        | 885                     | 6.42         | 0.3                     |
| Slip face cascade     |           |              |                         |              |                         |
| <b>Total Cascades</b> | <b>10</b> | <b>29.41</b> | <b>4,038</b>            | <b>29.29</b> |                         |
| Pocketwater           |           |              |                         |              |                         |
| Glide                 | 1         | 2.94         | 293                     | 2.1          |                         |
| Run                   |           |              |                         |              |                         |
| Low gradient riffle   | 10        | 29.4         | 4,321                   | 31.35        |                         |
| <b>Total Riffles</b>  | <b>11</b> | <b>32.34</b> | <b>4,614</b>            | <b>33.45</b> |                         |
| Dammed pool           | 3         | 8.8          | 4666                    | 33.85        | 0.6                     |
| Eddy pool             |           |              |                         |              |                         |
| Plunge pool           | 5         | 14.7         | 152                     | 1.1          | 0.3                     |
| Scour pool            |           |              |                         |              |                         |
| Scour hole            | 5         | 14.7         | 314                     | 2.28         | 0.3                     |
| Beaver pond           |           |              |                         |              |                         |
| <b>Total Pools</b>    | <b>13</b> | <b>38.20</b> | <b>5,132</b>            | <b>37.23</b> |                         |
| Secondary channel     |           |              |                         |              |                         |
| <b>Grand Totals</b>   | <b>34</b> |              | <b>13,784</b>           |              |                         |

**Appendix D.65 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #13 for Benawah Creek during 1994.**

| Habitat Type         | Frequency | % Frequency | Total Area (sq. meters) | % Area      | Residual pool depth (m) |
|----------------------|-----------|-------------|-------------------------|-------------|-------------------------|
| Rapid                |           |             |                         |             |                         |
| Step po. de          | 4         | 40          | 1,946                   | 28.3        |                         |
| Slip face .e         |           |             |                         |             |                         |
| <b>Total Cas es</b>  | <b>4</b>  | <b>40</b>   | <b>1,946</b>            | <b>28.3</b> |                         |
| Pocketwater          | 1         | 10          | 245                     | 3.6         |                         |
| Glide                | 1         | 10          | 138                     | 2.0         |                         |
| Run                  |           |             |                         |             |                         |
| Low gradient riffle  | 2         | 20          | 1,042                   | 15.2        |                         |
| <b>Total Riffles</b> | <b>4</b>  | <b>40</b>   | <b>1,423</b>            | <b>20.8</b> |                         |
| Dammed po            | 1         | 10          | 3,414                   | 49.7        | 1.1                     |
| Eddy pool            |           |             |                         |             |                         |
| Plunge pool          | 1         | 10          | 91                      | 1.3         | 0.4                     |
| Scour pool           |           |             |                         |             |                         |
| Scour hole           |           |             |                         |             |                         |
| Beaver pond          |           |             |                         |             |                         |
| <b>Total Pools</b>   | <b>2</b>  | <b>20</b>   | <b>3,5</b>              | <b>51.0</b> |                         |
| Secondary channel    |           |             |                         |             |                         |
| <b>Grand Totals</b>  | <b>10</b> |             | <b>6,873</b>            |             | <b>0.7</b>              |

**Appendix D.66 Summary report for Reach # 1 of the S.E. Benawah Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>       |
|--|--------------------------|-------------------|
| <b>Elevation</b>                                 | <b>2,847 - 3,061 ft.</b> | <b>same</b>       |
| <b>Total length</b>                              | <b>10,807 ft.</b>        | <b>11,463 ft.</b> |
| <b>Stream order</b>                              | <b>3</b>                 | <b>3</b>          |
| <b>Mean stream gradient</b>                      | <b>7.5 %</b>             | <b>7.5 %</b>      |
| <b>Riffle/pool ratio</b>                         | <b>1.1 : 1</b>           | <b>2.7 : 1</b>    |
| <b>Land use</b>                                  |                          |                   |
| Forest   | 50 %                     | 50 %              |
| Agriculture                                      |                          |                   |
| Livestock grazing                                | 50 %                     | 50 %              |
| Mining   |                          |                   |
| Wetland  |                          |                   |
| Floodplain                                       |                          |                   |
| Other (includes residential, right of way, etc.) |                          |                   |
| <b>Vegetative type</b>                           |                          |                   |
| Deciduous  |                          |                   |
| Coniferous                                       |                          |                   |
| Mixed  | 100 %                    | 100 %             |
| <b>Seral stage</b>                               |                          |                   |
| Grass/forb                                       |                          |                   |
| Shrub  | 50 %                     | 50 %              |
| Pole   |                          |                   |
| Young  |                          |                   |
| Mature   | 50 %                     | 50 %              |
| Old growth                                       |                          |                   |
| Other  |                          |                   |
| <b>x Canopy cover</b>                            | <b>51.7 %</b>            | <b>76 %</b>       |
| <b># Woody debris</b>                            |                          |                   |
| Logs   | 392                      | 1 0 3             |
| Root wads  | 6                        |                   |

**Appendix D.67 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for S. E. Benawah Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 9                | 23.8             | 20,086              | 47.6          |                       |
| Step pool cascade     | 14               | 36.8             | 13,892              | 25.4          | 0.19                  |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>23</b>        | <b>60.4</b>      | <b>39,978</b>       | <b>73.0</b>   |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 8                | 21.1             | 12,268              | 22.4          |                       |
| <b>Total Riffles</b>  | <b>8</b>         | <b>21.1</b>      | <b>12,268</b>       | <b>22.4</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 4                | 10.5             | 1,008               | 1.8           | 0.28                  |
| Scour pool            | 3                | 7.9              | 1,541               | 2.8           | 0.24                  |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>7</b>         | <b>18.5</b>      | <b>2,549</b>        | <b>4.6</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>38</b>        |                  | <b>54,795.3</b>     |               |                       |

**Appendix D.68 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #1 for S. E. Benawah Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 2                | 4.3              | 2,831               | 11.5          |                       |
| Step pool cascade     | 15               | 32.6             | 11,568              | 47            |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>17</b>        | <b>36.9</b>      | <b>14,399</b>       | <b>58.5</b>   |                       |
| Pocketwater           | 2                | 4.2              | 296                 | 1.2           |                       |
| Glide                 | 1                | 2.1              | 47                  | 0.2           |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 19               | 40.4             | 9,515               | 38.7          |                       |
| <b>Total Riffles</b>  | <b>22</b>        | <b>46.7</b>      | <b>9,862</b>        | <b>40.1</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 8                | 17.4             | 348                 | 1.4           | 0.2                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>8</b>         | <b>17.4</b>      | <b>348</b>          | <b>1.4</b>    | <b>0.2</b>            |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>46</b>        |                  | <b>24,804</b>       |               |                       |

**Appendix D.69 Summary report for Reach # 1 of the Evans Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>     |
|--|--------------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,130 - 2,130 ft.</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>2,886 ft.</b>         | <b>5,194 ft</b> |
| <b>Stream order</b>                              | <b>3</b>                 | <b>3</b>        |
| <b>Mean stream gradient</b>                      | <b>1.0 %</b>             | <b>1.3 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>1 : 1.5</b>           | <b>1 : 1</b>    |
| <b>Land use</b>                                  |                          |                 |
| Forest   |                          |                 |
| Agriculture                                      |                          |                 |
| Livestock grazing                                | <b>100%</b>              | <b>100%</b>     |
| Mining   |                          |                 |
| Wetland  |                          |                 |
| Floodplain                                       |                          |                 |
| Other (includes residential, right of way, etc.) |                          |                 |
| <b>Vegetative type</b>                           |                          |                 |
| Deciduous  | <b>100%</b>              | <b>100 %</b>    |
| Coniferous                                       |                          |                 |
| Mixed  |                          |                 |
| <b>Seral stage</b>                               |                          |                 |
| Grass/forb                                       |                          |                 |
| Shrub  | <b>100%</b>              | <b>100 %</b>    |
| Pole   |                          |                 |
| Young  |                          |                 |
| Mature   | <b>59.5 %</b>            | <b>59.5 %</b>   |
| Old growth                                       |                          |                 |
| Other  |                          |                 |
| <b>x Canopy cover</b>                            | <b>28.3 %</b>            | <b>6.0%</b>     |
| <b># Woody debris</b>                            |                          |                 |
| Logs   | <b>0</b>                 | <b>5</b>        |
| Root wads  | <b>0</b>                 | <b>0</b>        |

**Appendix D.70 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for Evans Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 2                | 7.1              | 1,523               | 4.2           |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 11               | 39.3             | 23,951              | 66.6          |                       |
| <b>Total Riffles</b>  | <b>13</b>        | <b>46.4</b>      | <b>25,474</b>       | <b>70.8</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             | 5                | 17.9             | 1,553               | 4.3           | 0.5                   |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 10               | 35.7             | 8,930               | 24.8          | 0.4                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>15</b>        | <b>53.6</b>      | <b>10,483</b>       | <b>29.1</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>28</b>        |                  | <b>35,957</b>       |               |                       |

**Appendix D.71 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #1 for Evans Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 1                | 25               | 8323                | 87.1          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 1                | 25               | 858                 | 9.0           |                       |
| <b>Total Riffles</b>  | <b>2</b>         | <b>50</b>        | <b>9,181</b>        | <b>96.1</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 2                | 50               | 357                 | 3.9           | 0.5                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>2</b>         | <b>50</b>        | <b>357</b>          | <b>3.9</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>4</b>         |                  | <b>9,556</b>        |               |                       |

**Appendix D.72 Summary report for Reach # 2 of the Evans Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>          | <b>1994</b>     |
|--|----------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,130 - 2,130</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>5,189 ft.</b>     | <b>3,166 ft</b> |
| <b>Stream order</b>                              | <b>3</b>             | <b>3</b>        |
| <b>Mean stream gradient</b>                      | <b>1.5 %</b>         | <b>1.3 %</b>    |
| <b>Riffle/pool ratio</b>                         |                      | <b>1.2 : 1</b>  |
| <b>Land use</b>                                  |                      |                 |
| Forest   | 60.3 %               | 60.3 %          |
| Agriculture                                      |                      |                 |
| Livestock grazing                                | 39.7 %               | 39.7 %          |
| Mining   |                      |                 |
| Wetland  |                      |                 |
| Floodplain                                       |                      |                 |
| Other (includes residential, right of way, etc.) |                      |                 |
| <b>Vegetative type</b>                           |                      |                 |
| Decidious  | 60 %                 | 60 %            |
| Coniferous                                       |                      |                 |
| Mixed  | 40 %                 | 40 %            |
| <b>Seral stage</b>                               |                      |                 |
| Grass/forb                                       |                      |                 |
| Shrub  | 40.5 %               | 40.5 %          |
| Pole   |                      |                 |
| Young  |                      |                 |
| Mature   | 59.5 %               | 59.5 %          |
| Old growth                                       |                      |                 |
| Other  |                      |                 |
| <b>x Canopy cover</b>                            | <b>51.8 %</b>        | <b>34.1 %</b>   |
| <b># Woody debris</b>                            |                      |                 |
| Logs   | 59                   | 18              |
| Root wads  | 17                   | 8               |

**Appendix D.73 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for Evans Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 3                | 4.1              | 359                 |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     | 4                | 5.5              | 1,589               |               |                       |
| <b>Total Cascades</b> | <b>7</b>         | <b>9.6</b>       | <b>1,947</b>        | <b>2.6</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 34               | 46.6             | 84,956              |               |                       |
| <b>Total Riffles</b>  | <b>34</b>        | <b>46.6</b>      | <b>64,956</b>       | <b>86.5</b>   |                       |
| Dammed pool           | 3                | 4.1              | 701                 |               | 0.3                   |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 2                | 2.7              | 201                 |               | 0.2                   |
| Scour pool            | 22               | 30.1             | 6,425               |               | 0.1                   |
| Scour hole            | 5                | 6.8              | 834                 |               | 0.3                   |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>32</b>        | <b>43.7</b>      | <b>8,188</b>        | <b>10.9</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>73</b>        |                  | <b>75,090</b>       |               |                       |

**Appendix D.74 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #2 for Evans Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 5                | 7                | 1,900               | 7.6           |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 34               | 47.9             | 18,531              | 74.1          |                       |
| <b>Total Riffles</b>  | <b>39</b>        | <b>52.9</b>      | <b>20,431</b>       | <b>81.7</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 1                | 1.4              | 549                 | 2.2           | 0.7                   |
| Scour pool            | 24               | 33.8             | 2,949               | 11.8          | 0.5                   |
| Scour hole            | 7                | 9.9              | 1,077               | 4.3           | 0.6                   |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>32</b>        | <b>45.1</b>      | <b>4,575</b>        | <b>18.3</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>71</b>        |                  | <b>25,006</b>       |               |                       |

**Appendix D.75 Summary report for Reach # 3 of the Evans Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                | <b>1993</b>              | <b>1994</b>    |
|---|--------------------------|----------------|
| <b>Elevation</b>                                | <b>2,130 - 2,270 ft.</b> | <b>same</b>    |
| <b>Total length</b>                             | <b>5,652 ft.</b>         | <b>3,041 n</b> |
| <b>Stream order</b>                             | <b>3</b>                 | <b>3</b>       |
| <b>Mean stream gradient</b>                     | <b>2.0 %</b>             | <b>1.9 %</b>   |
| <b>Riffle/pool ratio</b>                        | <b>1 : 1.6</b>           | <b>1.2 : 1</b> |
| <b>Land use</b>                                 |                          |                |
| Forest  | 100 %                    | 100%           |
| Agriculture                                     |                          |                |
| Livestock grazing                               |                          |                |
| Mining  |                          |                |
| Wetland   |                          |                |
| Floodplain                                      |                          |                |
| Other (includes residential-right of way, etc.) |                          |                |
| <b>Vegetative type</b>                          |                          |                |
| Decidious                                       |                          |                |
| Coniferous                                      |                          |                |
| Mixed   | 100%                     | 100%           |
| <b>Seral stage</b>                              |                          |                |
| Grass/forb                                      |                          |                |
| Shrub   |                          |                |
| Pole  |                          |                |
| Young   |                          |                |
| Mature  | 100 %                    | 100%           |
| Old growth                                      |                          |                |
| Other   |                          |                |
| <b>x Canopy cover</b>                           | <b>65.2 %</b>            | <b>10.5 %</b>  |
| <b># Woody debris</b>                           |                          |                |
| Logs  | 49                       | 71             |
| Root wads                                       | 7                        | 19             |



**Appendix D.76 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Evans Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 20               | 17.5               | 4,643                          | 21.6          |                                |
| Step pool cascade     | 1                | 0.9                | 25                             | 0.1           |                                |
| Slip face cascade     | 1                | 0.9                | 43                             | 0.2           |                                |
| <b>Total Cascades</b> | <b>22</b>        | <b>19.3</b>        | <b>4,711</b>                   | <b>21.9</b>   |                                |
| Pocketwater<br>Glide  | .                |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 35               | 30.7               | 11,946                         | 55.6          |                                |
| <b>Total Riffles</b>  | <b>35</b>        | <b>30.7</b>        | <b>11,946</b>                  | <b>55.6</b>   |                                |
| Dammed pool           | 4                | 3.5                | 605                            | 2.8           | 1.4                            |
| Eddy pool             | 2                | 1.8                | 19                             | 0.08          | 1.1                            |
| Plunge pool           | 13               | 11.4               | 804                            | 3.7           |                                |
| Scour pool            | 36               | 31.6               | 3,299                          | 15.4          |                                |
| Scour hole            | 2                | 1.8                | 88                             | 0.4           | 1.2                            |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>57</b>        | <b>50.1</b>        | <b>4,815</b>                   | <b>22.4</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>114</b>       |                    | <b>21,471</b>                  |               |                                |

**Appendix D.77 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #3 for Evans Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     | 3                | 6.8                | 1,065                          | 7.4           |                                |
| Slip face cascade     | 1                | 2.3                | 12                             | 0.6           |                                |
| <b>Total Cascades</b> | <b>4</b>         | <b>9.1</b>         | <b>1,155</b>                   | <b>8.0</b>    |                                |
| Pocketwater<br>Glide  | 1                | 2.3                | 176                            | 1.2           |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 21               | 47.7               | 9,942                          | 68.7          |                                |
| <b>Total Riffles</b>  | <b>22</b>        | <b>60.0</b>        | <b>10,118</b>                  | <b>69.9</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 2                | 4.5                | 270                            | 1.9           | 0.6                            |
| Scour pool            | 9                | 20.4               | 2,061                          | 14.2          | 0.5                            |
| Scour hole            | 7                | 15.9               | 867                            | 6.0           | 0.5                            |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>18</b>        | <b>48.8</b>        | <b>3,198</b>                   | <b>22.1</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>44</b>        |                    | <b>14,471</b>                  |               |                                |

**Appendix D.78 Summary report for Reach # 4 of the Evans Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>    |
|--|--------------------------|----------------|
| <b>Elevation</b>                                 | <b>2,270 - 2,580 ft.</b> | <b>same</b>    |
| <b>Total length</b>                              | <b>5,906 ft.</b>         | <b>5,966</b>   |
| <b>Stream order</b>                              | <b>3</b>                 | <b>3</b>       |
| <b>Mean stream gradient</b>                      | <b>3.0 %</b>             | <b>2.3 %</b>   |
| <b>Riffle/pool ratio</b>                         | <b>1.3 : 1</b>           | <b>1.9 : 1</b> |
| <b>Land use</b>                                  |                          |                |
| Forest   | 100%                     | 100%           |
| Agriculture                                      |                          |                |
| Livestock grazing                                |                          |                |
| Mining   |                          |                |
| Wetland  |                          |                |
| Floodplain                                       |                          |                |
| Other (includes residential, right of way, etc.) |                          |                |
| <b>Vegetative type</b>                           |                          |                |
| Deciduous  |                          |                |
| Coniferous                                       |                          |                |
| Mixed  | 100%                     | 100%           |
| <b>Seral stage</b>                               |                          |                |
| Grass/forb                                       |                          |                |
| Shrub  |                          |                |
| Pole   |                          |                |
| Young  |                          |                |
| Mature   | 100%                     | 100%           |
| Old growth                                       |                          |                |
| Other  |                          |                |
| <b>x Canopy cover</b>                            | <b>69.4 %</b>            | <b>47.7 %</b>  |
| <b># Woody debris</b>                            |                          |                |
| Logs   | 16                       | 168            |
| Root wads  | 5                        | 18             |

**Appendix D.79 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Evans Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 34               | 45.9             | 21,698              | 84.9          |                       |
| Step pool cascade     | 8                | 10.8             | 1,651               | 6.6           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>42</b>        | <b>55.2</b>      | <b>23,348</b>       | <b>91.5</b>   |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 1                | 1.4              | 203.4               | 0.8           |                       |
| <b>Total Riffles</b>  | <b>1</b>         | <b>1.4</b>       | <b>203.4</b>        | <b>0.8</b>    |                       |
| Dammed pool           | 1                | 1.4              | 27.8                | 0.1           |                       |
| Eddy pool             | 2                | 2.8              | 51.6                | 0.2           |                       |
| Plunge pool           | 11               | 14.9             | 726                 | 2.8           |                       |
| Scour pool            | 17               | 22.9             | 1,192               | 4.7           |                       |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>32</b>        | <b>41.9</b>      | <b>1,997</b>        | <b>7.8</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>75</b>        | <b>99.5</b>      | <b>25,541</b>       | <b>100.1</b>  |                       |

**Appendix D.80 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #4 for Evans Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 4                | 4.4              | 1,771               | 5.8           |                       |
| Step pool cascade     | 18               | 20               | 12,733              | 41.7          |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>22</b>        | <b>24.4</b>      | <b>14,504</b>       | <b>47.5</b>   |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 36               | 40               | 12,574              | 41.2          |                       |
| <b>Total Riffles</b>  | <b>36</b>        | <b>40</b>        | <b>12,574</b>       | <b>41.2</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 12               | 13.3             | 1,038               | 3.4           | 0.3                   |
| Scour pool            | 18               | 20               | 2,182               | 7.2           | 0.3                   |
| Scour hole            | 1                | 1.1              | 98.8                | 0.3           | 0.4                   |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>31</b>        | <b>34.4</b>      | <b>3,319</b>        | <b>10.9</b>   |                       |
| Secondary channel     | 1                | 1.1              | 109.7               | 0.4           |                       |
| <b>Grand Totals</b>   | <b>90</b>        |                  | <b>30,505</b>       |               |                       |

**Appendix D.81 Summary report for Reach # 5 of the Evans Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>              | <b>1994</b>     |
|--|--------------------------|-----------------|
| <b>Elevation</b>                                 | <b>2,580 - 3,340 ft.</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>3,247 ft.</b>         | <b>2,412 ft</b> |
| <b>Stream order</b>                              | <b>2</b>                 | <b>2</b>        |
| <b>Mean stream gradient</b>                      | <b>10.8 %</b>            | <b>3.1 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>3.8 : 1</b>           | <b>4.1 : 1</b>  |
| <b>Land use</b>                                  |                          |                 |
| Forest   | 100%                     | 100%            |
| Agriculture                                      |                          |                 |
| Livestock grazing                                |                          |                 |
| Mining   |                          |                 |
| Wetland  |                          |                 |
| Floodplain                                       |                          |                 |
| Other (includes residential, right of way, etc.) |                          |                 |
| <b>Vegetative type</b>                           |                          |                 |
| Deciduous  |                          |                 |
| Coniferous                                       |                          |                 |
| Mixed  | 100 %                    | 100%            |
| <b>Seral stage</b>                               |                          |                 |
| Grass/forb                                       |                          |                 |
| Shrub  |                          |                 |
| Pole   |                          |                 |
| Young  |                          |                 |
| Mature   | 100 %                    | 100%            |
| Old growth                                       |                          |                 |
| Other  |                          |                 |
| <b>x Canopy cover</b>                            | <b>37.5 %</b>            | <b>59.7%</b>    |
| <b># Woody debris</b>                            |                          |                 |
| Logs   | 90                       | 191             |
| Root wads  |                          | 9               |

**Appendix D.82 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Evans Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 5                | 26.3               | 2,688                          | 25.1          |                                |
| Step pool cascade     | 7                | 36.8               | 4,095                          | 38.3          |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>12</b>        | <b>63.1</b>        | <b>6,783</b>                   | <b>63.4</b>   |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient rime     | 3                | 15.8               | 755                            | 7.1           |                                |
| <b>Total Riffles</b>  | <b>3</b>         | <b>15.8</b>        | <b>755</b>                     | <b>7.1</b>    |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 4                | 21.1               | 316                            | 29.5          | 0.5                            |
| Scour pool            |                  |                    |                                |               |                                |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>4</b>         | <b>21.1</b>        | <b>316</b>                     | <b>29.5</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>19</b>        | <b>100</b>         | <b>10,696</b>                  | <b>100</b>    |                                |

**Appendix D.83 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Evans Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 12               | 38.7               | 4,069                          | 22.9          |                                |
| Step pool cascade     | 12               | 38.7               | 12,937                         | 72.8          |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>24</b>        | <b>77.4</b>        | <b>17,006</b>                  | <b>95.7</b>   |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 1                | 3.2                | 185                            | 1.0           |                                |
| <b>Total Riffles</b>  | <b>1</b>         | <b>3.2</b>         | <b>185</b>                     | <b>1.0</b>    |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 3                | 9.7                | 333                            | 1.9           | 0.5                            |
| Scour pool            | 3                | 9                  | 254                            | 1.4           | 0.5                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>6</b>         | <b>19.4</b>        | <b>587</b>                     | <b>3.3</b>    |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>31</b>        |                    | <b>17,779</b>                  |               |                                |

**Appendix D.84 Summary report for Reach # 6 of the Evans Creek (right fork) Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b>     |
|--|-------------------|-----------------|
| <b>Elevation</b>                                 | 2,680 - 3,040 ft. | <b>same</b>     |
| <b>Total length</b>                              | 2,278 ft.         | <b>2,183 ft</b> |
| <b>Stream order</b>                              | 2                 | 2               |
| <b>Mean stream gradient</b>                      | <b>12.5 %</b>     | <b>8.1 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>NA</b>         | <b>5 : 1</b>    |
| <b>Land use</b>                                  |                   |                 |
| Forest   | 100%              | 100%            |
| Agriculture                                      |                   |                 |
| Livestock grazing                                |                   |                 |
| Mining   |                   |                 |
| Wetland  |                   |                 |
| Floodplain                                       |                   |                 |
| Other (includes residential, right of way, etc.) |                   |                 |
| <b>Vegetative type</b>                           |                   |                 |
| Deciduous  |                   |                 |
| Coniferous                                       |                   |                 |
| Mixed  | 100 %             | 100%            |
| <b>Seral stage</b>                               |                   |                 |
| Grass/forb                                       |                   |                 |
| Shrub  |                   |                 |
| Pole   |                   |                 |
| Young  |                   |                 |
| Mature   | 100 %             | 100%            |
| Old growth                                       |                   |                 |
| Other  |                   |                 |
| <b>x Canopy cover</b>                            | 70.0 %            | 66 %            |
| <b># Woody debris</b>                            |                   |                 |
| Logs   | 86                | 51              |
| Root wads  |                   | 3               |

**Appendix D.85 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Evans Creek (right fork) during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 5                | 50               | 728                 | 25.8          |                       |
| Step pool cascade     | 5                | 50               | 2,099               | 74.2          |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>10</b>        | <b>100</b>       | <b>2,827</b>        | <b>100</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   |                  |                  |                     |               |                       |
| <b>Total Riffles</b>  |                  |                  |                     |               |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            |                  |                  |                     |               |                       |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    |                  |                  |                     |               |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>10</b>        | <b>100</b>       | <b>2,827</b>        | <b>100</b>    |                       |

**Appendix D.86 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Evans Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 | 5                | 38.5             | 2,363               | 44.6          |                       |
| Step pool cascade     | 6                | 46.2             | 2,695               | 50.8          |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>11</b>        | <b>04.7</b>      | <b>5,058</b>        | <b>95.4</b>   |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 1                | 7.7              | 191                 | 3.4           |                       |
| <b>Total Riffles</b>  | <b>1</b>         | <b>7.7</b>       | <b>191</b>          | <b>3.4</b>    |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 1                | 7.7              | 55                  | 1.0           | 0.2                   |
| Scour pool            |                  |                  |                     |               |                       |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>1</b>         | <b>7.7</b>       | <b>55</b>           | <b>1.0</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>13</b>        |                  | <b>7,960</b>        |               |                       |

**Appendix D.87 Summary report for Reach # 1 of the Evans Creek (left fork) Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>       | <b>1994</b>      |
|--|-------------------|------------------|
| <b>Elevation</b>                                 | 2,580 - 3,170 ft. | <b>same</b>      |
| <b>Total length</b>                              | 3,205 ft.         | <b>3,331 ft.</b> |
| <b>Stream order</b>                              | 2                 | 2                |
| <b>Mean stream gradient</b>                      | <b>10.7 %</b>     | <b>10.8 %</b>    |
| <b>Riffle/pool ratio</b>                         | <b>3 : 1</b>      | <b>5 : 1</b>     |
| <b>Land use</b>                                  |                   |                  |
| Forest   | 91.2 %            | 91.2 %           |
| Agriculture                                      |                   |                  |
| Livestock grazing                                | 8.8 %             | 8.8 %            |
| Mining   |                   |                  |
| Wetland  |                   |                  |
| Floodplain                                       |                   |                  |
| Other (includes residential, right of way, etc.) |                   |                  |
| <b>Vegetative type</b>                           |                   |                  |
| Decidious  |                   |                  |
| Coniferous                                       |                   |                  |
| Mixed  | 100%              | 100%             |
| <b>Seral stage</b>                               |                   |                  |
| Grass/forb                                       |                   |                  |
| Shrub  |                   |                  |
| Pole   |                   |                  |
| Young  |                   |                  |
| Mature   | 100 %             | 100 %            |
| Old growth                                       |                   |                  |
| Other  |                   |                  |
| <b>x Canopy cover</b>                            | <b>34.0 %</b>     | <b>39 %</b>      |
| <b># Woody debris</b>                            |                   |                  |
| Logs   | 50                | 31               |
| Root wads  |                   | 4                |



**Appendix D.88 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for Evans Creek (left fork) during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 3                | 17.6               | 863                            | 14.4          |                                |
| Step pool cascade     | 6                | 35.3               | 2,184                          | 36.4          |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>9</b>         | <b>52.9</b>        | <b>3,047</b>                   | <b>50.8</b>   |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 6                | 35.3               | 1,781                          | 29.7          |                                |
| <b>Total Riffles</b>  | <b>6</b>         | <b>35.3</b>        | <b>1,781</b>                   | <b>29.7</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 2                | 11.8               | 1,169                          | 19.5          | 0.3                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>2</b>         | <b>11.8</b>        | <b>1,169</b>                   | <b>19.5</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>17</b>        | <b>100</b>         | <b>5,996</b>                   | <b>100</b>    |                                |

**Appendix D.89 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #5a for Evans Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 5                | 27.8               | 1,462                          | 22.4          |                                |
| Step pool cascade     | 7                | 38.9               | 3,356                          | 51.4          |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>12</b>        | <b>66.7</b>        | <b>4,818</b>                   | <b>73.8</b>   |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 5                | 27.8               | 1,665                          | 25.5          |                                |
| <b>Total Riffles</b>  | <b>5</b>         | <b>27.8</b>        | <b>1,665</b>                   | <b>25.5</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 5.6                | 43                             | 0.7           | 0.2                            |
| Scour pool            |                  |                    |                                |               |                                |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>1</b>         | <b>5.6</b>         | <b>43</b>                      | <b>0.7</b>    |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>18</b>        |                    | <b>6,526</b>                   |               |                                |

**Appendix D.90 Summary report for Reach # 1 of the Alder Creek  
Watershed, data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>      | <b>1994</b>  |
|--|------------------|--------------|
| <b>Elevation</b>                                 | 2,260 - 2,366 ft | <b>same</b>  |
| <b>Total length</b>                              | <b>2,892 ft</b>  | 2,874 ft     |
| <b>Stream order</b>                              | 4                | 4            |
| <b>Mean stream gradient</b>                      | <b>1.5 %</b>     | 5 %          |
| <b>Riffle/pool ratio</b>                         | <b>1.4 : 1</b>   | <b>1 : 1</b> |
| <b>Land use</b>                                  |                  |              |
| Forest   | 100 %            | same         |
| Agriculture                                      |                  |              |
| Livestock grazing                                |                  |              |
| Mining   |                  |              |
| Wetland  |                  |              |
| Floodplain                                       |                  |              |
| Other (includes residential, right of way, etc.) |                  |              |
| <b>Vegetative type</b>                           |                  |              |
| Deciduous  |                  |              |
| Coniferous                                       |                  |              |
| Mixed  | 100 %            | 100%         |
| <b>Seral stage</b>                               |                  |              |
| Grass/forb                                       | 7.1 %            | 7.1 %        |
| Shrub  | 35.7 %           | 35.7 %       |
| Pole   | 37.5 %           | 37.5 %       |
| Young  | 12.5 %           | 12.5 %       |
| Mature   | 7.1 %            | 7.1 %        |
| Old growth                                       |                  |              |
| Other  |                  |              |
| <b>x Canopy cover</b>                            | 32 %             | 22 %         |
| <b># Woody debris</b>                            |                  |              |
| Logs   | 14               | 28           |
| Root wads  | 0                | 0            |

**Appendix D.91 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for Alder Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| <b>Rapid</b>          |                  |                    |                                |               |                                |
| Step pool cascade     | 1                | 3.6                | 110                            |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>1</b>         | <b>3.6</b>         | <b>110</b>                     |               |                                |
| Pocketwater           | 4                | 14.3               | 10,269                         |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| <b>Run</b>            |                  |                    |                                |               |                                |
| Low gradient riffle   | 11               | 39.3               | 15,380                         |               |                                |
| <b>Total Riffles</b>  | <b>15</b>        | <b>53.6</b>        | <b>25,649</b>                  |               |                                |
| Dammed pool           | 1                | 3.6                | 274                            |               | 0.9                            |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 4                | 14.3               | 945                            |               | 0.4                            |
| Scour pool            | 6                | 21.4               | 1,047                          |               | 0.4                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           |                  |                    |                                |               |                                |
| <b>Total Pools</b>    | <b>1</b>         | <b>39.3</b>        | <b>2,266</b>                   |               |                                |
| Secondary channel     | 1                | 3.6                | 217                            |               |                                |
| <b>Grand Totals</b>   | <b>28</b>        | <b>100.1</b>       | <b>28,242</b>                  |               |                                |

**Appendix D.92 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #1 for Alder Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| <b>Rapid</b>          |                  |                    |                                |               |                                |
| Step pool cascade     | 2                | 14.3               | 1,064                          | 5.8           |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>2</b>         | <b>14.3</b>        | <b>1,064</b>                   | <b>5.8</b>    |                                |
| Pocketwater           | 4                | 28.6               | 14,073                         | 76.4          |                                |
| Glide                 |                  |                    |                                |               |                                |
| <b>Run</b>            |                  |                    |                                |               |                                |
| Low gradient riffle   | 2                | 14.3               | 1,606                          | 8.7           |                                |
| <b>Total Riffles</b>  | <b>6</b>         | <b>42.9</b>        | <b>15,679</b>                  | <b>85.1</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 2                | 14.3               | 750                            | 4.1           | 0.8                            |
| Scour pool            | 3                | 21.4               | 650                            | 3.5           | 0.4                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           | 1                | 7.1                | 268                            | 1.5           | 0.8                            |
| <b>Total Pools</b>    | <b>6</b>         | <b>42.8</b>        | <b>1,668</b>                   | <b>9.1</b>    |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>14</b>        |                    | <b>18,411</b>                  |               |                                |

**Appendix D.93 Summary report for Reach # 2 of the Alder Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>      |
|--|-------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,350 - 2,630 ft</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>13,739 ft</b>        | <b>13,803 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                | <b>4</b>         |
| <b>Mean stream gradient</b>                      | <b>1.5 %</b>            | <b>4.5 %</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.2 : 1</b>          | <b>4 : 1</b>     |
| <b>Land use</b>                                  |                         |                  |
| Forest   | 50 %                    | 50 %             |
| Agriculture                                      |                         |                  |
| Livestock grazing                                | 50 %                    | 50 %             |
| Mining   |                         |                  |
| Wetland  |                         |                  |
| Floodplain                                       |                         |                  |
| Other (includes residential, right of way, etc.) |                         |                  |
| <b>Vegetative type</b>                           |                         |                  |
| Deciduous  |                         |                  |
| Coniferous                                       |                         |                  |
| Mixed  | 100 %                   | 100 %            |
| <b>Seral stage</b>                               |                         |                  |
| Grass/forb                                       | 25.8 %                  | 25.8 %           |
| Shrub  | 9.7 %                   | 9.7 %            |
| Pole   |                         |                  |
| Young  | 6.4 %                   | 6.4 %            |
| Mature   | 56.5 %                  | 56.5 %           |
| Old growth                                       | 1.6 %                   | 1.6 %            |
| Other  |                         |                  |
| <b>x Canopy cover</b>                            | <b>30 %</b>             | <b>34 %</b>      |
| <b># Woody debris</b>                            |                         |                  |
| Logs   | 70                      | 49               |
| Root wads  | 0                       | 1                |

**Appendix D.94 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for Alder Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>% Area</b> | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     | 5                | 8.8         | 10,884              | 35.8          |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> | <b>5</b>         | <b>8.8</b>  | <b>10,884</b>       | <b>35.8</b>   |                       |
| Pocketwater           | 2                | 3.5         | 1,315               | 4.3           |                       |
| Glide                 | 1                | 1.7         | 373                 | 1.2           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 25               | 43.9        | 10,452              | 34.4          |                       |
| <b>Total Riffles</b>  | <b>28</b>        | <b>49.1</b> | <b>12,141</b>       | <b>39.9</b>   |                       |
| Dammed pool           | 2                | 3.5         | 274                 | 0.9           | 0.5                   |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           | 12               | 21.1        | 4,819               | 15.9          | 0.8                   |
| Scour pool            | 9                | 15.9        | 2,098               | 6.9           | 0.5                   |
| Scour hole            | 1                | 1.7         | 146                 | 0.5           | 0.7                   |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>24</b>        | <b>42.1</b> | <b>7,337</b>        | <b>24.2</b>   |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>57</b>        |             | <b>30,361</b>       |               |                       |

**Appendix D.95 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #2 for Alder Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>% Area</b> | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     | 17               | 40.5        | 34,185              | 35.5          |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> | <b>17</b>        | <b>40.5</b> | <b>34,185</b>       | <b>35.5</b>   |                       |
| Pocketwater           | 17               | 40.5        | 59,363              | 61.7          |                       |
| Glide                 | 2                | 4.8         | 7,476               | 0.8           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 1                | 2.4         | 166                 | 0.2           |                       |
| <b>Total Riffles</b>  | <b>20</b>        | <b>90.9</b> | <b>67,005</b>       | <b>62.7</b>   |                       |
| Dammed pool           |                  |             |                     |               |                       |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           | 4                | 9.5         | 1,540               | 1.6           | 1.3                   |
| Scour pool            | 1                | 2.4         | 260                 | 0.3           | 0.2                   |
| Scour hole            |                  |             |                     |               |                       |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>5</b>         | <b>11.9</b> | <b>1,800</b>        | <b>1.9</b>    |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>42</b>        |             | <b>96,260</b>       |               |                       |

**Appendix D.96 Summary report for Reach # 3 of the Alder Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>     |
|--|-------------------------|-----------------|
| <b>Elevation</b>                                 | 2,630 - 2,700 <b>ft</b> | <b>same</b>     |
| <b>Total length</b>                              | <b>8,172 ft</b>         | 8,003 <b>ft</b> |
| <b>Stream order</b>                              | 4                       | 4               |
| <b>Mean stream gradient</b>                      | <b>1.6 %</b>            | 2.5 <b>%</b>    |
| <b>Riffle/pool ratio</b>                         | <b>1.3 : 1</b>          | <b>4.4 ; 1</b>  |
| <b>Land use</b>                                  |                         |                 |
| Forest   |                         |                 |
| Agriculture                                      |                         |                 |
| Livestock grazing                                | 100 %                   | 100%            |
| Mining   |                         |                 |
| Wetland  |                         |                 |
| Floodplain                                       |                         |                 |
| Other (includes residential, right of way, etc.) |                         |                 |
| <b>Vegetative type</b>                           |                         |                 |
| Deciduous  |                         |                 |
| Coniferous                                       |                         |                 |
| Mixed  | 100%                    | 100%            |
| <b>Seral stage</b>                               |                         |                 |
| Grass/forb                                       | 24.7 %                  | 24.7 %          |
| Shrub  | 62.4 %                  | 62.4 %          |
| Pole   |                         |                 |
| Young  |                         |                 |
| Mature   | 12.9 %                  | 12.9 %          |
| Old growth                                       |                         |                 |
| Other  |                         |                 |
| <b>x Canopy cover</b>                            | <b>32 %</b>             | <b>32 %</b>     |
| <b># Woody debris</b>                            |                         |                 |
| Logs   | 13                      | 15              |
| Root wads  | 1                       | 0               |

**Appendix D.97 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for Alder Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>% Area</b> | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     | 3                | 6.4         | 1,454               | 3.4           |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> | <b>3</b>         | <b>6.4</b>  | <b>1,454</b>        | <b>3.4</b>    |                       |
| Pocketwater           |                  |             |                     |               |                       |
| Glide                 | 1                | 2.1         | 376                 | 0.9           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 24               | 51.1        | 36,032              | 84.9          |                       |
| <b>Total Riffles</b>  | <b>25</b>        | <b>53.2</b> | <b>36,401</b>       | <b>85.8</b>   |                       |
| Dammed pool           |                  |             |                     |               |                       |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           | 5                | 10.6        | 1,368               | 3.2           | 1.2                   |
| Scour pool            | 14               | 29.8        | 3,200               | 7.5           | 1.2                   |
| Scour hole            |                  |             |                     |               |                       |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>19</b>        | <b>48.4</b> | <b>4,588</b>        | <b>10.7</b>   |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>47</b>        |             | <b>42,423</b>       |               |                       |

**Appendix D.98 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach #3 for Alder Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>    | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|-------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>%</b>    | <b>(sq. meters)</b> | <b>% Area</b> | <b>pool depth (m)</b> |
| Rapid                 |                  |             |                     |               |                       |
| Step pool cascade     | 3                | 10          | 1,293               | 2.4           |                       |
| Slip face cascade     |                  |             |                     |               |                       |
| <b>Total Cascades</b> | <b>3</b>         | <b>10</b>   | <b>1,293</b>        | <b>2.4</b>    |                       |
| Pocketwater           | 3                | 10.0        | 16,185              | 29.4          |                       |
| Glide                 | 5                | 16.7        | 3,039               | 5.5           |                       |
| Run                   |                  |             |                     |               |                       |
| Low gradient riffle   | 14               | 46.7        | 33,567              | 61.0          |                       |
| <b>Total Riffles</b>  | <b>22</b>        | <b>73.4</b> | <b>52,791</b>       | <b>95.9</b>   |                       |
| Dammed pool           |                  |             |                     |               |                       |
| Eddy pool             |                  |             |                     |               |                       |
| Plunge pool           | 1                | 3.3         | 307                 | 0.6           | 0.3                   |
| Scour pool            | 4                | 13.3        | 605                 | 1.1           | 0.2                   |
| Scour hole            |                  |             |                     |               |                       |
| Beaver pond           |                  |             |                     |               |                       |
| <b>Total Pools</b>    | <b>5</b>         | <b>16.6</b> | <b>912</b>          | <b>1.7</b>    |                       |
| Secondary channel     |                  |             |                     |               |                       |
| <b>Grand Totals</b>   | <b>30</b>        |             | <b>55,005</b>       |               |                       |

**Appendix D.99 Summary report for Reach # 4 of the Alder Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>      |
|--|-------------------------|------------------|
| <b>Elevation</b>                                 | <b>2,700 - 2,900 ft</b> | <b>same</b>      |
| <b>Total length</b>                              | <b>16,579 ft</b>        | <b>19,049 ft</b> |
| <b>Stream order</b>                              | <b>4</b>                | <b>4</b>         |
| <b>Mean stream gradient</b>                      | <b>1.8 %</b>            | <b>2.6 %</b>     |
| <b>Riffle/pool ratio</b>                         | <b>1.2 : 1</b>          | <b>2.6 : 1</b>   |
| <b>Land use</b>                                  |                         |                  |
| Forest   |                         |                  |
| Agriculture                                      |                         |                  |
| Livestock grazing                                | <b>100%</b>             | <b>100%</b>      |
| Mining   |                         |                  |
| Wetland  |                         |                  |
| Floodplain                                       |                         |                  |
| Other (includes residential, right of way, etc.) |                         |                  |
| <b>Vegetative type</b>                           |                         |                  |
| Deciduous  |                         |                  |
| Coniferous                                       |                         |                  |
| Mixed  | <b>100%</b>             | <b>100%</b>      |
| <b>Seral stage</b>                               |                         |                  |
| Grass/forb                                       | <b>5.9 %</b>            | <b>5.9 %</b>     |
| Shrub  | <b>26.7 %</b>           | <b>26.7 %</b>    |
| Pole   |                         |                  |
| Young  |                         |                  |
| Mature   | <b>67.3 %</b>           | <b>63.7 %</b>    |
| Old growth                                       |                         |                  |
| Other  |                         |                  |
| <b>x Canopy cover</b>                            | <b>6.2 %</b>            | <b>46.6 %</b>    |
| <b># Woody debris</b>                            |                         |                  |
| Logs   | <b>71</b>               | <b>106</b>       |
| Root wads  | <b>6</b>                | <b>10</b>        |



**Appendix D.100 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Alder Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 1                | 1.0              | 1,723               | 1.8           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>1</b>         | <b>1.0</b>       | <b>1,723</b>        | <b>1.8</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 2                | 2.0              | 1,033               | 1.1           |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient rime     |                  |                  |                     |               |                       |
| <b>Total Riffles</b>  | <b>52</b>        | <b>54</b>        | <b>76,823</b>       | <b>79.4</b>   |                       |
| Dammed pool           | 2                | 2.0              | 4,750               | 4.9           |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 5                | 5.0              | 1,292               | 1.3           | 0.4                   |
| Scour pool            | 37               | 37.0             | 12,150              | 12.5          | 0.4                   |
| Scour hole            | 1                | 1                | 107                 | 0.1           | 0.7                   |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>45</b>        | <b>45</b>        | <b>18,299</b>       | <b>18.8</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>100</b>       | <b>100</b>       | <b>96,846</b>       | <b>100</b>    |                       |

**Appendix D.101 Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 4 for Alder Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 10               | 6.4              | 6,217               | 7.7           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>10</b>        | <b>6.4</b>       | <b>6,217</b>        | <b>7.7</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 39               | 24.8             | 14,391              | 17.9          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffie   | 67               | 42.7             | 46,686              | 58.1          |                       |
| <b>Total Riffles</b>  | <b>106</b>       | <b>67.5</b>      | <b>61,077</b>       | <b>76.0</b>   |                       |
| Dammed pool           | 1                | 0.6              | 493                 | 0.6           | 0.5                   |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 3                | 1.9              | 683                 | 0.9           | 0.4                   |
| Scour pool            | 33               | 21.0             | 6,086               | 7.8           | 0.4                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           | 3                | 1.9              | 5,863               | 7.7           | 0.6                   |
| <b>Total Pools</b>    | <b>40</b>        | <b>25.4</b>      | <b>13,125</b>       | <b>17.0</b>   |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>157</b>       |                  | <b>80,418</b>       |               |                       |

**Appendix D.102 Summary report for Reach # 5 of the Alder Creek  
Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                | <b>1993</b>     | <b>1994</b> |
|---|-----------------|-------------|
| <b>Elevation</b>                                | 2,900 - 2,940   | <b>same</b> |
| <b>Total length</b>                             | <b>5,768 ft</b> | 4,807 ft    |
| <b>Stream order</b>                             | 4               | 4           |
| <b>Mean stream gradient</b>                     | <b>1.4 %</b>    | 3.6 %       |
| <b>Riffle/pool ratio</b>                        | <b>1 : 1</b>    | 6.8 : 1     |
| <b>Land use</b>                                 |                 |             |
| Forest  | 83.8 %          | 83.8 %      |
| Agriculture                                     |                 |             |
| Livestock grazing                               | 16.2 %          | 16.2 %      |
| Mining  |                 |             |
| Wetland   |                 |             |
| Floodplain                                      |                 |             |
| Other (includes residential.right of way, etc.) |                 |             |
| <b>Vegetative type</b>                          |                 |             |
| Deciduous                                       |                 |             |
| Coniferous                                      |                 |             |
| Mixed   | 100 %           | 100%        |
| <b>Seral stage</b>                              |                 |             |
| Grass/forb                                      | 29.6 %          | 29.6 %      |
| Shrub   | 28.4 %          | 28.4 %      |
| Pole  |                 |             |
| Young   |                 |             |
| Mature  | 6.2 %           | 6.2 %       |
| Old growth                                      | 35.8 %          | 35.8 %      |
| Other   |                 |             |
| <b>x Canopy cover</b>                           | <b>17.7 %</b>   | 43.7 %      |
| <b># Woody debris</b>                           |                 |             |
| Logs  | 64              | 76          |
| Root wads                                       | 1               | 0           |

**Appendix D.103. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Alder Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 1                | 2.5                | 218                            | 0.6           |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient rime     | 19               | 47.5               | 19,627                         | 52.6          |                                |
| <b>Total Riffles</b>  | <b>20</b>        | <b>50.0</b>        | <b>19,844</b>                  | <b>53.2</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 2.5                | 214                            | 0.6           | 0.4                            |
| Scour pool            | 14               | 35                 | 1,902                          | 5.1           | 0.4                            |
| Scour hole            | 1                | 2.5                | 101                            | 0.3           | 0.3                            |
| Beaver pond           | 4                | 10                 | 15,221                         | 40.8          | 0.8                            |
| <b>Total Pools</b>    | <b>20</b>        | <b>50.0</b>        | <b>17,483</b>                  | <b>46.8</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>40</b>        |                    | <b>37,282</b>                  |               |                                |

**Appendix D.104. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 5 for Alder Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     | 1                | 3.1                | 259                            | 0.6           |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>1</b>         | <b>3.1</b>         | <b>259</b>                     | <b>0.6</b>    |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 12               | 37.5               | 3,635                          | 7.9           |                                |
| Run                   | 15               | 46.9               | 6,036                          | 13.2          |                                |
| Low gradient riffle   |                  |                    |                                |               |                                |
| <b>Total Riffles</b>  | <b>27</b>        | <b>84.4</b>        | <b>9,671</b>                   | <b>21.1</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            | 3                | 9.4                | 41                             | 0.9           | 0.7                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           | 1                | 3.1                | 35,454                         | 77.4          | 0.8                            |
| <b>Total Pools</b>    | <b>4</b>         | <b>12.5</b>        | <b>35,495</b>                  | <b>78.3</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>32</b>        |                    | <b>45,789</b>                  |               |                                |

**Appendix D.105. Summary report for Reach # 6 of the Alder Creek Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>      | <b>1994</b>    |
|--|------------------|----------------|
| <b>Elevation</b>                                 | 2,946 - 3,330 ft | same           |
| <b>Total length</b>                              | <b>21,754 ft</b> | 20,596 ft      |
| <b>Stream order</b>                              | 2                | 2              |
| <b>Mean stream gradient</b>                      | <b>1.8 %</b>     | <b>5.3 %</b>   |
| <b>Riffle/pool ratio</b>                         | <b>1 : 1.7</b>   | <b>2.4 : 1</b> |
| <b>Land use</b>                                  |                  |                |
| Forest   | 95.1 %           | 95.1 %         |
| Agriculture                                      |                  |                |
| Livestock grazing                                | 4.9 %            | 4.9 %          |
| Mining   |                  |                |
| Wetland  |                  |                |
| Floodplain                                       |                  |                |
| Other (includes residential, right of way, etc.) |                  |                |
| <b>Vegetative type</b>                           |                  |                |
| Deciduous  |                  |                |
| Coniferous                                       |                  |                |
| Mixed  | 100%             | 100%           |
| <b>Seral stage</b>                               |                  |                |
| Grass/forb                                       | 21.0 %           | 21.0 %         |
| Shrub  | 3.7 %            | 3.7 %          |
| Pole   |                  |                |
| Young  |                  |                |
| Mature   | 75.3 %           | 75.3 %         |
| Old growth                                       |                  |                |
| Other  |                  |                |
| <b>x Canopy cover</b>                            | 25 %             | 68 %           |
| <b># Woody debris</b>                            |                  |                |
| Logs   | 127              | 307            |
| Root wads  | 0                | 0              |

**Appendix D.106. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Alder Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 30               | 36.6               | 20,923                         | 13.0          |                                |
| <b>Total Riffles</b>  | <b>30</b>        | <b>36.6</b>        | <b>20,923</b>                  | <b>13.0</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 1.2                | 91.4                           | 0.5           | 0.8                            |
| scour pool            | 9                | 11.0               | 757                            | 0.5           | 0.3                            |
| Scour hole            | 12               | 14.6               | 1,674                          | 1.0           | 0.3                            |
| Beaver pond           | 30               | 36.6               | 136,992                        | 85.0          | 0.9                            |
| <b>Total Pools</b>    | <b>52</b>        | <b>63.4</b>        | <b>140,311</b>                 | <b>87.0</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>82</b>        | <b>100</b>         | <b>161,234</b>                 | <b>100</b>    |                                |

**Appendix D.107. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 6 for Alder Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 3                | 13.0               | 1,217                          | 1.7           |                                |
| Step pool cascade     | 3                | 13.0               | 3,890                          | 5.3           |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>6</b>         | <b>26.0</b>        | <b>5,107</b>                   | <b>7.0</b>    |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 | 5                | 21.7               | 1,542                          | 2.1           |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 7                | 30.4               | 2,809                          | 3.8           |                                |
| <b>Total Riffles</b>  | <b>12</b>        | <b>52.1</b>        | <b>4,351</b>                   | <b>5.9</b>    |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 4.4                | 67                             | 0.09          | 0.4                            |
| Scour pool            | 1                | 4.4                | 61                             | 0.08          | 0.2                            |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           | 3                | 13.0               | 63,534                         | 86.9          | 0.5                            |
| <b>Total Pools</b>    | <b>5</b>         | <b>21.8</b>        | <b>63,662</b>                  | <b>87.1</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>23</b>        |                    | <b>73,119.2</b>                |               |                                |

**Appendix D.108. Summary report for Reach # 1 of the North Fork Alder Creek Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b>        |
|--|-------------------------|--------------------|
| <b>Elevation</b>                                 | <b>2,940 - 3,070 ft</b> | <b>same</b>        |
| <b>Total length</b>                              | <b>4,134 ft</b>         | <b>5,502.00 ft</b> |
| <b>Stream order</b>                              | <b>2</b>                | <b>2</b>           |
| <b>Mean stream gradient</b>                      | <b>3 %</b>              | <b>3.1 %</b>       |
| <b>Riffle/pool ratio</b>                         | <b>1.3 : 1</b>          | <b>16 : 1</b>      |
| <b>Land use</b>                                  |                         |                    |
| Forest   |                         |                    |
| Agriculture                                      |                         |                    |
| Livestock grazing                                | 100 %                   | 100%               |
| Mining   |                         |                    |
| Wetland  |                         |                    |
| Floodplairf                                      |                         |                    |
| Other (includes residential, right of way, etc.) |                         |                    |
| <b>Vegetative type</b>                           |                         |                    |
| Decidious  |                         |                    |
| Coniferous                                       |                         |                    |
| Mixed  | 100%                    | 100 %              |
| <b>Seral stage</b>                               |                         |                    |
| Grass/forb                                       | 100 %                   | 100%               |
| Shrub  |                         |                    |
| Pole   |                         |                    |
| Young  |                         |                    |
| Mature   |                         |                    |
| Old growth                                       |                         |                    |
| Other  |                         |                    |
| <b>x Canopy cover</b>                            | <b>36.3 %</b>           | <b>68 %</b>        |
| <b># Woody debris</b>                            |                         |                    |
| Logs   | 24                      | 22                 |
| Root wads  | 0                       | 0                  |

**Appendix D.109. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for North Fork Alder Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 9                | 56.3             | 10,666              | 95.3          |                       |
| <b>Total Riffles</b>  | <b>9</b>         | <b>66.3</b>      | <b>10,666</b>       | <b>95.2</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            | 4                | 25               | 385                 | 3.4           | 0.5                   |
| Scour hole            | 3                | 18.8             | 155                 | 1.4           | 0.4                   |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>7</b>         | <b>43.8</b>      | <b>540</b>          | <b>4.8</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>16</b>        | <b>100</b>       | <b>11,206</b>       | <b>100</b>    |                       |

**Appendix D.110. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 1 for North Fork Alder Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           | <b>Frequency</b> | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 | 8                | 47.1             | 1,122               | 19.9          |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 8                | 47.1             | 4,441               | 78.6          |                       |
| <b>Total Riffles</b>  | <b>16</b>        | <b>94.2</b>      | <b>5,563</b>        | <b>98.5</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 1                | 5.9              | 87.2                | 1.5           |                       |
| Scour pool            |                  |                  |                     |               |                       |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>1</b>         | <b>5.9</b>       | <b>87.2</b>         | <b>1.5</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>17</b>        |                  | <b>5,651</b>        |               |                       |

**Appendix D.111. Summary report for Reach # 2 of the North Fork Alder Creek Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>      | <b>1994</b>      |
|--|------------------|------------------|
| <b>Elevation</b>                                 | 3,070 - 3,280 ft | <b>same</b>      |
| <b>Total length</b>                              | <b>11,355 ft</b> | <b>13,946.00</b> |
| <b>Stream order</b>                              | 2                | 2                |
| <b>Mean stream gradient</b>                      | <b>1.9 %</b>     | <b>8.7 %</b>     |
| <b>Riffle/pool ratio</b>                         |                  |                  |
| <b>Land use</b>                                  |                  |                  |
| Forest   | 100%             | 100%             |
| Agriculture                                      |                  |                  |
| Livestock grazing                                |                  |                  |
| Mining   |                  |                  |
| Wetland  |                  |                  |
| Flood plain                                      |                  |                  |
| Other (includes residential, right of way, etc.) |                  |                  |
| <b>Vegetative type</b>                           |                  |                  |
| Decidious  |                  |                  |
| Coniferous                                       |                  |                  |
| Mixed  | 100%             | 100%             |
| <b>Seral stage</b>                               |                  |                  |
| Grass/forb                                       | 1.5 %            | 11.5 %           |
| Shrub  |                  |                  |
| Pole   |                  |                  |
| Young  |                  |                  |
| Mature   | 88.5 %           | 88.5 %           |
| Old growth                                       |                  |                  |
| Other  |                  |                  |
| <b>x Canopy cover</b>                            | 35.2             | <b>87.2 %</b>    |
| <b># Woody debris</b>                            |                  |                  |
| Logs   | 46               | 663              |
| Root wads  | 0                |                  |



**Appendix D.112. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for North Fork Alder Creek during 1993.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 |                  |                    |                                |               |                                |
| Step pool cascade     |                  |                    |                                |               |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> |                  |                    |                                |               |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 16               | 62                 | 32,824                         | 32.9          |                                |
| <b>Total Riffles</b>  | <b>16</b>        | <b>62</b>          | <b>32,824</b>                  | <b>32.9</b>   |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           | 1                | 3.8                | 14.6                           | 0.1           | 0.2                            |
| Scour pool            | 2                | 7.7                | 221                            | 0.2           | 0.3                            |
| Scour hole            | 2                | 7.7                | 218                            | 0.2           | 0.5                            |
| Beaver pond           | 5                | 19.3               | 66,548                         | 66.7          | 0.6                            |
| <b>Total Pools</b>    | <b>10</b>        | <b>36.4</b>        | <b>67,002</b>                  | <b>67.1</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>26</b>        | <b>99.9</b>        | <b>9,826</b>                   | <b>100</b>    |                                |

**Appendix D.113. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 2 for North Fork Alder Creek during 1994.**

| <b>Habitat Type</b>   | <b>Frequency</b> | <b>% Frequency</b> | <b>Total Area (sq. meters)</b> | <b>% Area</b> | <b>Residual pool depth (m)</b> |
|-----------------------|------------------|--------------------|--------------------------------|---------------|--------------------------------|
| Rapid                 | 8                | 40                 | 3,251                          | 4.7           |                                |
| Step pool cascade     | 6                | 30                 | 3,986                          | 5.8           |                                |
| Slip face cascade     |                  |                    |                                |               |                                |
| <b>Total Cascades</b> | <b>14</b>        | <b>70</b>          | <b>7,237</b>                   | <b>10.6</b>   |                                |
| Pocketwater           |                  |                    |                                |               |                                |
| Glide                 |                  |                    |                                |               |                                |
| Run                   |                  |                    |                                |               |                                |
| Low gradient riffle   | 3                | 15                 | 997                            | 1.4           |                                |
| <b>Total Riffles</b>  | <b>3</b>         | <b>15</b>          | <b>997</b>                     | <b>1.4</b>    |                                |
| Dammed pool           |                  |                    |                                |               |                                |
| Eddy pool             |                  |                    |                                |               |                                |
| Plunge pool           |                  |                    |                                |               |                                |
| Scour pool            |                  |                    |                                |               |                                |
| Scour hole            |                  |                    |                                |               |                                |
| Beaver pond           | 3                | 15                 | 61,054                         | 88.1          | 0.5                            |
| <b>Total Pools</b>    | <b>3</b>         | <b>15</b>          | <b>61,064</b>                  | <b>88.1</b>   |                                |
| Secondary channel     |                  |                    |                                |               |                                |
| <b>Grand Totals</b>   | <b>20</b>        |                    | <b>59,288</b>                  |               |                                |

**Appendix D.114. Summary report for Reach # 3 of the North Fork Alder Creek Watershed data collected during 1993 and 1994.**

| <b>Parameter</b>                                 | <b>1993</b>             | <b>1994</b> |
|--|-------------------------|-------------|
| <b>Elevation</b>                                 | <b>3,280 - 3,620 ft</b> |             |
| <b>Total length</b>                              | <b>10,271</b>           |             |
| <b>Stream order</b>                              | <b>2</b>                |             |
| <b>Mean stream gradient</b>                      | <b>7.1 %</b>            |             |
| <b>Riffle/pool ratio</b>                         | <b>2.3 : 1</b>          |             |
| <b>Land use</b>                                  |                         |             |
| Forest   | <b>100%</b>             |             |
| Agriculture                                      |                         |             |
| Livestock grazing                                |                         |             |
| Mining   |                         |             |
| Wetland  |                         |             |
| Floodplain                                       |                         |             |
| Other (includes residential, right of way, etc.) |                         |             |
| <b>Vegetative type</b>                           |                         |             |
| Deciduous  |                         |             |
| Coniferous                                       |                         |             |
| Mixed  | <b>100%</b>             |             |
| <b>Seral stage</b>                               |                         |             |
| Grass/forb                                       |                         |             |
| Shrub  | <b>11.5 %</b>           |             |
| Pole   |                         |             |
| Young  | <b>29.3 %</b>           |             |
| Mature   | <b>70.8 %</b>           |             |
| Old growth                                       |                         |             |
| Other  |                         |             |
| <b>x Canopy cover</b>                            | <b>39.0 %</b>           |             |
| <b># Woody debris</b>                            |                         |             |
| Logs   | <b>47</b>               |             |
| Root wads  | <b>0</b>                |             |

**Appendix D.115. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for North Fork Alder Creek during 1993.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     | 1                | 7.1              | 17.0                | 0.4           |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> | <b>1</b>         | <b>7.1</b>       | <b>17.0</b>         | <b>0.4</b>    |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   | 9                | 64.3             | 4,580               | 97.4          |                       |
| <b>Total Riffles</b>  | <b>9</b>         | <b>64.3</b>      | <b>4,560</b>        | <b>97.4</b>   |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           | 1                | 7.1              | 17.1                | 0.4           | 0.2                   |
| Scour pool            | 3                | 21.4             | 87.5                | 1.9           | 0.2                   |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    | <b>4</b>         | <b>28.5</b>      | <b>105</b>          | <b>2.2</b>    |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   | <b>14</b>        | <b>99.9</b>      | <b>4,682</b>        | <b>100</b>    |                       |

**Appendix D.116. Frequency of occurrence, total percent occurrence, total area, percent area, residual pool depth and volume values for Reach # 3 for North Fork Alder Creek during 1994.**

| <b>Habitat</b>        | <b>Frequency</b> | <b>%</b>         | <b>Total Area</b>   | <b>% Area</b> | <b>Residual</b>       |
|-----------------------|------------------|------------------|---------------------|---------------|-----------------------|
| <b>Type</b>           |                  | <b>Frequency</b> | <b>(sq. meters)</b> |               | <b>pool depth (m)</b> |
| Rapid                 |                  |                  |                     |               |                       |
| Step pool cascade     |                  |                  |                     |               |                       |
| Slip face cascade     |                  |                  |                     |               |                       |
| <b>Total Cascades</b> |                  |                  |                     |               |                       |
| Pocketwater           |                  |                  |                     |               |                       |
| Glide                 |                  |                  |                     |               |                       |
| Run                   |                  |                  |                     |               |                       |
| Low gradient riffle   |                  |                  |                     |               |                       |
| <b>Total Riffles</b>  |                  |                  |                     |               |                       |
| Dammed pool           |                  |                  |                     |               |                       |
| Eddy pool             |                  |                  |                     |               |                       |
| Plunge pool           |                  |                  |                     |               |                       |
| Scour pool            |                  |                  |                     |               |                       |
| Scour hole            |                  |                  |                     |               |                       |
| Beaver pond           |                  |                  |                     |               |                       |
| <b>Total Pools</b>    |                  |                  |                     |               |                       |
| Secondary channel     |                  |                  |                     |               |                       |
| <b>Grand Totals</b>   |                  |                  |                     |               |                       |

## Appendix E

**Appendix E.1. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during Spring (May, June), 1993.**

| Site                    | Lower            | Middle          | Upper          |
|-------------------------|------------------|-----------------|----------------|
| <b>Shock time (min)</b> | <b>24.3</b>      | <b>13.6</b>     | <b>18.5</b>    |
| Cutthroat trout         |                  | <b>8 (44.4)</b> |                |
| Eastern brook trout     |                  | <b>2 (11.1)</b> | <b>9 (100)</b> |
| Sculpin spp.            | <b>14 (87.5)</b> |                 |                |
| Longnosesucker          |                  | <b>5 (27.8)</b> |                |
| Longnosedace            | <b>2 (11.1)</b>  |                 |                |
| Brown bullhead          |                  | <b>3 (16.7)</b> |                |
| <b>TOTAL</b>            | <b>16</b>        | <b>18</b>       | <b>9</b>       |

**Appendix E.2. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during Spring (May, June), 1993.**

| Site                    | Lower | Middle          | Upper          |
|-------------------------|-------|-----------------|----------------|
| <b>Shock time (min)</b> |       | <b>11.6</b>     | <b>27.8</b>    |
| Cutthroat trout         |       | <b>2 (66.7)</b> |                |
| Sculpin spp.            |       | <b>1 (33.3)</b> | <b>4 (100)</b> |
| <b>TOTAL</b>            |       | <b>3</b>        | <b>4</b>       |

**Appendix E.3 Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during Spring (May, June), 1993.**

| Site                    | Lower | Middle   | Upper           |
|-------------------------|-------|----------|-----------------|
| <b>Shock time (min)</b> |       | <b>•</b> | <b>12.4</b>     |
| Sculpin spp.            |       |          | <b>2 (66.7)</b> |
| Western speckled date   |       |          | <b>1 (33.3)</b> |
| <b>TOTAL</b>            |       |          | <b>3</b>        |

**Appendix E.4. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benewah Creek during Spring (May, June), 1993.**

| Site                  | Lower     | Middle | Upper     |
|-----------------------|-----------|--------|-----------|
| Shock time (min)      |           | 16.4   | 16.7      |
|                       |           |        | 21.7      |
| Cutthroat trout       | 4 (16.0)  |        | 2 (6.3)   |
| Sculpin spp.          |           |        | 1 (3.1)   |
| Longnose sucker       | 15 (60.0) |        | 7 (21.9)  |
| Longnose dace         |           |        | 12 (8.4)  |
| Western speckled dace |           |        | 14 (43.8) |
| Redside shiner        |           |        | 1 (0.7)   |
| Northern Squawfish    | 6 (24.0)  |        | 8 (25.0)  |
| TOTAL                 | 25        |        | 32        |
|                       |           |        | 142       |

**Appendix E5. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Coon Creek<sup>1</sup> during Spring (May, June), 1993.**

| Site                  | Lower     | Upper     |
|-----------------------|-----------|-----------|
| Shock time (min)      |           | 13.5      |
|                       |           | 11.8      |
| Cutthroat trout       | 3 (12.0)  | 7 (100.0) |
| Longnose sucker       | 3 (12.0)  |           |
| Western speckled dace | 1 (4.0)   |           |
| Redside shiner        | 18 (72.0) |           |
| TOTAL                 | 25        | 7         |

**Appendix E.6. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Cow Creek<sup>1</sup> during Spring (May, June), 1993.**

| Site             | Lower     |
|------------------|-----------|
| Shock time (min) | 6.4       |
| Longnose sucker  | 5 (15.2)  |
| Redside shiner   | 27 (81.8) |
| Brown bullhead   | 1 (3.0)   |
| TOTAL            | 33        |

**Appendix E.7. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Harmony Creek<sup>1</sup> during Spring (May, June), 1993.**

|                   |           |
|-------------------|-----------|
| <b>Site Lower</b> |           |
| Shock time (min)  | 7.4       |
| Cutthroat trout   | 6 (42.9)  |
| Sculpin spp.      | 8 (57.1)  |
| <b>TOTAL</b>      | <b>14</b> |

**Appendix E.8. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Whitetail Creek<sup>1</sup> during Spring (May, June), 1993.**

|                     |              |          |
|---------------------|--------------|----------|
| <b>Site Lower</b>   | <b>Upper</b> |          |
| Shock time (min)    | 10.6         | 5.1      |
| Cutthroat trout     | 2 (66.7)     | 2 (100)  |
| Eastern brook trout | 1 (33.3)     |          |
| <b>TOTAL</b>        | <b>3</b>     | <b>2</b> |

**Appendix E.9. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Windfall Creek<sup>1</sup> during Spring (May, June), 1993.**

|                   |              |          |
|-------------------|--------------|----------|
| <b>Site Lower</b> | <b>Upper</b> |          |
| Shock time (min)  | 13.9         | 6.2      |
| Cutthroat trout   | 2 (100)      | 2 (100)  |
| <b>TOTAL</b>      | <b>2</b>     | <b>2</b> |

\* site could not be accessed because of deep water

<sup>1</sup> Tributaries to Benewah Creek

**Table E.10. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during Summer (August), 1993.**

|                     |               |              |           |
|---------------------|---------------|--------------|-----------|
| <b>Site Lower</b>   | <b>Middle</b> | <b>Upper</b> |           |
| Shock time (min)    | 14.0          | 14.9         | 22.3      |
| Cutthroat trout     | 2 (5.9)       | 12 (48.0)    | 6 (15.8)  |
| Eastern brook trout |               | 7 (28.0)     | 28 (73.7) |
| Sculpin spp.        | 30 (88.2)     |              | 4 (10.5)  |
| Longnose sucker     |               | 6 (28.0)     |           |
| longnose dace       | 2 (5.9)       |              |           |
| Brown bullhead      |               |              |           |
| <b>TOTAL</b>        | <b>34</b>     | <b>25</b>    | <b>38</b> |

**Table E.11. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during Summer (August), 1993.**

| Site             | Lower | Middle    | Upper     |           |
|------------------|-------|-----------|-----------|-----------|
| Shock time (min) |       | 20.0      | 20.0      | 18.8      |
| Cutthroat trout  |       | 1 (4.5)   | 30 (48.4) | 56 (57.1) |
| Sculpin spp.     |       | 18 (81.8) | 32 (51.6) | 42 (42.9) |
| Longnose sucker  |       | 3 (13.6)  |           |           |
| <b>TOTAL</b>     |       |           | 62        | 98        |

**Table E.12. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during Summer (August), 1992.**

| Site             | Lower | Middle    | Upper    |          |
|------------------|-------|-----------|----------|----------|
| Shock time (min) |       | 12.6      | 17.1     | 11.6     |
| Cutthroat trout  |       | 6 (10.3)  | 9 (37.5) | 6 (60.0) |
| Sculpin spp.     |       | 17 (29.3) | 9 (37.5) | 2 (20.0) |
| longnose sucker  |       | 1 (1.7)   |          | 2 (20.0) |
| Redside shiner   |       |           | 1 (4.2)  |          |
| Longnosedace     |       | 34 (58.6) |          |          |
| <b>TOTAL</b>     |       | 58        | 24       | 19       |

**Table E.13. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during Summer (August), 1993.**

| Site             | Lower | Middle    | Upper     |           |
|------------------|-------|-----------|-----------|-----------|
| Shock time (min) |       | 10.4      | 22.3      | 26.8      |
| Cutthroat trout  |       |           | 4 (6.4)   | 11 (8.0)  |
| Sculpin spp.     |       | 7 (53.1)  | 3 (4.8)   |           |
| Longnose sucker  |       | 3 (10.3)  | 12 (19.4) | 10 (7.3)  |
| longnose dace    |       | 5 (17.2)  | 24 (38.7) | 32 (23.4) |
| Redside shiner   |       | 14 (48.3) | 19 (30.6) | 84 (61.3) |
| <b>TOTAL</b>     |       | 29        | 62        | 137       |

**Table E.14. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during fall (October), 1993.**

| Site                | Lower    | Middle   | Upper    |
|---------------------|----------|----------|----------|
| Shock time (min)    | 15.5     | 18.7     | 32.0     |
| Cutthroat trout     | 1 (13.3) | 6 (31.6) | 9 (25.0) |
| Eastern brook trout |          |          | 1 (2.8)  |
| Rainbow trout       | 1 (6.7)  |          |          |
| Sculpin spp.        | 3 (33.3) | 3 (15.8) | 9 (25.0) |



**Table E.14. (continued): Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during fall (October), 1993.**

|                  | Lower     | Middle    | Upper     |
|------------------|-----------|-----------|-----------|
| Shock time (min) | 15.5      | 18.7      | 32.0      |
| Longnose sucker  | 8 (26.7)  | 6 (31.6)  | 6 (16.7)  |
| Largemouth bass  | 2 (6.7)   |           |           |
| Redside shiner   | 4 (13.3)  | 4 (21.1)  | 11 (30.6) |
| <b>TOTAL</b>     | <b>30</b> | <b>19</b> | <b>36</b> |

**Table E.15. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during fall (October), 1993.**

| Site             | Lower | Middle    | Upper      |
|------------------|-------|-----------|------------|
| Shock time (min) |       | 29.7      | 15.2       |
| Cutthroat trout  |       | 11 (31.4) |            |
| Sculpin spp.     |       | 14 (40.0) | 22 (100.0) |
| Bull trout       |       | 1 (2.9)   | 7 (53.8)   |
| Longnosesucker   |       | 9 (25.7)  | 6 (46.2)   |
| <b>TOTAL</b>     |       | <b>35</b> | <b>22</b>  |

**Table E.16. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during fall (October), 1993.**

| Site             | Lower | Middle    | Upper     |
|------------------|-------|-----------|-----------|
| Shock time (min) |       | 8.4       | 33.2      |
| Cutthroat trout  |       |           | 38 (71.7) |
| Sculpin spp.     |       | 6 (100.0) | 15 (28.3) |
| <b>TOTAL</b>     |       | <b>6</b>  | <b>53</b> |

**Table E.17. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during fall (October), 1993.**

| Site                | Lower | Middle    | Upper     |
|---------------------|-------|-----------|-----------|
| Shock time (min)    |       | 11.4      | 28.1      |
| Cutthroat trout     |       |           | 14 (56.0) |
| Eastern brook trout |       |           | 2 (8.0)   |
| Sculpin spp.        |       | 11 (64.6) | 5 (20.0)  |
| Longnosesucker      |       | 2 (15.4)  | 4 (16.0)  |
| <b>TOTAL</b>        |       | <b>13</b> | <b>25</b> |

**Table E.18. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during Spring (May, June), 1994.**

| Site                | Lower | Middle     | Upper      |             |
|---------------------|-------|------------|------------|-------------|
| Shock time (min)    |       | 11.4       | 13.6       | 13.3        |
| Cutthroat trout     |       | 3 (12.0%)  | 13 (43.3%) |             |
| Eastern brook trout |       | 4 (16.0%)  | 13 (43.3%) | 11 (100.0%) |
| Sculpin spp.        |       | 11 (44.0%) | 3 (10.0%)  |             |
| Longnose sucker     |       | 7 (28.0%)  | 1 (3.3%)   |             |
| <b>TOTAL</b>        |       | <b>25</b>  | <b>30</b>  | <b>11</b>   |

**Table E.19. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during Spring (May, June), 1994.**

| Site             | Lower | Middle    | Upper      |            |
|------------------|-------|-----------|------------|------------|
| Shock time (min) |       | 9.9       | 13.3       | 11.7       |
| Cutthroat trout  |       |           | 8 (34.8%)  | 24 (82.8%) |
| Sculpin spp.     |       | 11 (100%) | 15 (65.2%) | 5 (17.2%)  |
| <b>TOTAL</b>     |       | <b>11</b> | <b>23</b>  | <b>29</b>  |

**Table E.20. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during Spring (May, June), 1994.**

| Site             | Lower | Middle     | Upper      |           |
|------------------|-------|------------|------------|-----------|
| Shock time (min) |       | 13.0       | *          | 12.7      |
| Cutthroat trout  |       | 5 (16.7%)  | 1 (7.1%)   | 2 (22.2%) |
| Sculpin          |       | 14 (46.7%) | 11 (78.6%) | 7 (77.8%) |
| Redside shiner   |       |            | 2 (14.3%)  |           |
| Longnose sucker  |       | 11 (36.7%) |            |           |
| <b>TOTAL</b>     |       | <b>30</b>  | <b>14</b>  | <b>9</b>  |

**Table E.21. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during Spring (May, June), 1994.**

| Site               | Lower | Middle     | Upper     |            |
|--------------------|-------|------------|-----------|------------|
| Shock time (min)   |       | 13.7       | 14.2      | 12.2       |
| Cutthroat trout    |       | 6 (15.8%)  | 4 (16.7%) | 3 (9.1%)   |
| Sculpin spp.       |       | 12 (31.6%) | 3 (12.5%) | 2 (6.1%)   |
| Longnose sucker    |       | 16 (42.1%) | 9 (37.5%) | 21 (63.6%) |
| Redside shiner     |       | 3 (7.9%)   | 8 (33.3%) | 7 (21.2%)  |
| Northern Squawfish |       | 1 (2.6%)   |           |            |
| <b>TOTAL</b>       |       | <b>38</b>  | <b>24</b> | <b>33</b>  |

**Table E.22. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during Summer (August), 1994.**

| Site                | Lower | Middle     | Upper      |            |
|---------------------|-------|------------|------------|------------|
| Shock time (min)    |       | 14.7       | 14.8       | 14.1       |
| Cutthroat trout     |       | 2 (6.4%)   | 21 (44.7%) | 3 (6.4%)   |
| Eastern brook trout |       |            | 18 (36.3%) | 43 (93.5%) |
| Sculpin spp.        |       | 25 (80.7%) |            | 1 (2.1%)   |
| Longnose sucker     |       | 4 (12.9%)  | 8 (17.0%)  |            |
| <b>TOTAL</b>        |       | <b>31</b>  | <b>47</b>  | <b>47</b>  |

**Table E.23. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during Summer (August), 1994.**

| Site             | Lower | Middle     | Upper     |           |
|------------------|-------|------------|-----------|-----------|
| Shock time (min) |       | 10.7       | 6.8       | 5.2       |
| Cutthroat trout  |       | 0          | 17 (100%) | 1 (11.1%) |
| Sculpin spp.     |       | 21 (65.6%) |           | 8 (88.9%) |
| Longnose sucker  |       | 4 (12.5%)  |           |           |
| Redside shiner   |       | 6 (18.8%)  |           |           |
| Brown bullhead   |       | 1 (3.1%)   |           |           |
| <b>TOTAL</b>     |       | <b>32</b>  | <b>17</b> | <b>9</b>  |

**Table E.24. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during Summer (August), 1994.**

| Site                  | Lower | Middle    | Upper      |           |
|-----------------------|-------|-----------|------------|-----------|
| Shock time (min)      |       | 7.9       | 10.9       | 10.8      |
| Cutthroat trout       |       |           | 18 (51.4%) |           |
| Sculpin spp.          |       |           | 2 (5.7%)   | 8 (61.5%) |
| longnose sucker       |       | 4 (57.1%) | 5 (14.3%)  | 5 (38.5%) |
| Redside shiner        |       | 3 (42.9%) | 6 (17.1%)  |           |
| Western speckled dace |       |           | 4 (11.4%)  |           |
| <b>TOTAL</b>          |       | <b>7</b>  | <b>35</b>  | <b>13</b> |

**Table E.25. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during Summer (August), 1994.**

| Site Lower<br>Shock time (min) | Middle<br>8.0 | Upper<br>9.2 | 9.1          |
|--------------------------------|---------------|--------------|--------------|
| Cutthroat trout                |               | 1.0 (3.3%)   |              |
| Sculpin spp.                   | 15 (51.7%)    | 6 (20.0%)    |              |
| Longnose sucker                | 4 (13.8%)     | 2 (6.7%)     |              |
| Redside shiner                 | 10 (34.5%)    | 21 (70.0%)   | 100 (100.0%) |
| <b>TOTAL</b>                   | <b>29</b>     | <b>30</b>    | <b>100</b>   |

**Table E.26. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Benawah Creek during fall (October), 1994.**

| Site Lower<br>Shock time (min) | Middle<br>16.5 | Upper<br>22.9 | 23.0        |
|--------------------------------|----------------|---------------|-------------|
| Cutthroat trout                | 3 (7.7%)       | 1 (2.1%)      |             |
| Sculpin s p.                   | 6 (15.4%)      | 15 (31.2%)    | 12 (4.1%)   |
| Longnose sucker                | 7 (18.0%)      | 13 (27.1%)    | 15 (5.1%)   |
| Redside shiner                 | 23 (59.0%)     | 19 (39.6%)    | 268 (90.9%) |
| <b>TOTAL</b>                   | <b>39</b>      | <b>48</b>     | <b>295</b>  |

**Table E.27. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Lake Creek during fall (October), 1994.**

| Site Lower<br>Shock time (min) | Middle<br>17.8 | Upper<br>* | 19.4       |
|--------------------------------|----------------|------------|------------|
| Cutthroat trout                |                | 8 (18.2%)  |            |
| Sculpin spp.                   | 9 (22.5%)      | 27 (61.4%) | 5 (15.6%)  |
| Longnose sucker                | 16 (40.0%)     | 9 (20.5%)  | 4 (12.5%)  |
| Redside shiner                 | 15 (37.5%)     |            | 23 (71.9%) |
| <b>TOTAL</b>                   | <b>40</b>      | <b>44</b>  | <b>32</b>  |

**Table E.28. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Evans Creek during fall (October), 1994.**

| Site Lower<br>Shock time (min) | Middle<br>12.0 | Upper<br>* | 26.6       |
|--------------------------------|----------------|------------|------------|
| Cutthroat trout                | 1 (12.5%)      |            | 26 (53.1%) |
| Sculpin spp.                   | 2 (25.0%)      |            | 23 (46.9%) |
| Largemouth bass                | 5 (62.5%)      |            |            |
| <b>TOTAL</b>                   | <b>8</b>       |            | <b>49</b>  |

**Table E.29. Total number and relative abundance (%) of each species caught during relative abundance electrofishing surveys on Alder Creek during fall (October), 1994.**

| Site                | Lower | Middle     | Upper      |
|---------------------|-------|------------|------------|
| Shock time (min)    |       | 23.9       | 25.6       |
|                     |       |            | 21.1       |
| Cutthroat trout     |       | 2 (10.5%)  | 21 (63.6%) |
| Eastern brook trout |       | 2 (10.5%)  | 12 (36.4%) |
| Sculpin spp.        |       | 15 (78.9%) | 32 (48.5%) |
| <b>TOTAL</b>        |       | <b>19</b>  | <b>33</b>  |
|                     |       |            | <b>66</b>  |

## Appendix F

**APPENDIX F.I: Water Quality Data collected using a price pygmy meter (discharge) and Hydrolab in selected tributaries during 1993.**

255

| <b>LOCATIO</b> | <b>DATE</b> | <b>pH</b> | <b>Q (cfs)</b> | <b>(°C)</b> | <b>K (mmhoslcm)</b> | <b>DO (ppm)</b> | <b>REDOX (mV)</b> |
|----------------|-------------|-----------|----------------|-------------|---------------------|-----------------|-------------------|
| ALDER          | 3/11/93     |           | 11.62          | 2.3         | 0.880               | 16.89           | 298               |
| ALDER          | 3/16/93     | 7.2       | 38.16          | 2.6         | 0.073               | 16.39           | 289               |
| ALDER          | 3/23/93     |           | 168.10         | 36          |                     |                 |                   |
| ALDER          | 4/14/93     | 7.2       | 44.70          | 7.5         | 0.069               | 10.93           |                   |
| ALDER          | 5/4/93      | 6.8       | 289.00         | 41F         | 0.041               |                 |                   |
| ALDER          | 6/28/93     | 7.2       | 2.82           | 12.5        | 0.081               | 8.88            | 288               |
| ALDER          | 7/27/93     | 6.9       | 4.69           | 13.0        | 0.071               | 8.29            | 273               |
| ALDER          | 8/24/93     | 6.7       | 1.68           | 12.8        | 0.092               | 1.54            | 270               |
| ALDER          | 9/20/93     | 7.2       | 1.09           | 7.8         | 0.092               | 9.57            | 313               |
| ALDER          | 10/28/93    | 7.2       |                | 2.9         | 0.091               | 14.66           | 842               |
| ALDER          | 11/19/93    | 6.3       | 1.94           | 0.5         | 0.091               | 15.68           |                   |
| BENEWAI        | 1/20/93     |           |                | 31          |                     |                 |                   |
| BENEWAI        | 3/11/93     |           | 31.98          | 3.1         | 0.840               | 14.24           | 309               |
| BENEWAI        | 3/16/93     | 7.2       | 68.88          | 4.2         | 0.080               | 13.01           | 308               |
| BENEWAI        | 3/23/93     |           | 350.10         | 37          |                     |                 |                   |
| BENEWAI        | 4/14/93     | 7.2       | 85.94          | 41          | 0.060               | 13.47           |                   |
| BENEWAI        | 5/4/93      | 7.0       | 676.00         | 44F         | 0.042               |                 |                   |
| BENEWAI        | 6/28/93     | 7.4       | 2.60           | 15.6        | 0.057               | 11.02           | 278               |
| BENEWAI        | 7/28/93     | 7.4       | 11.48          | 18.8        | 0.060               | 8.23            | 260               |
| BENEWAI        | 8/24/93     | 6.9       | 0.83           | 11.9        | 0.071               | 2.15            | 286               |
| BENEWAI        | 9/20/93     | 7.3       | 0.35           | 10.8        | 0.069               | 10.60           | 415               |
| BENEWAI        | 10/28/93    | 7.5       | 1.98           | 7.2         | 0.090               | 11.45           | 874               |
| BENEWAI        | 11/19/93    | 6.4       | 3.78           | 0.7         | 0.070               | 14.77           | 939               |
| EVANS          | 1/20/93     |           | 6.47           | 30          |                     |                 |                   |
| EVANS          | 3/11/93     |           | 19.60          | 2.8         | 0.054               | 14.48           | 302               |
| EVANS          | 3/16/93     | 7.0       | 26.55          | 3.9         | 0.068               | 14.45           | 302               |
| EVANS          | 3/23/93     |           | 181.70         | 38          |                     |                 |                   |
| EVANS          | 4/15/93     | 7.2       | 47.85          | 39          | 0.050               | 22.00           | 291               |
| EVANS          | 5/4/93      | 7.0       | 456.00         | 42F         | 0.026               |                 |                   |
| EVANS          | 6/28/93     | 6.6       | 5.11           | 10.6        | 0.030               | 9.93            | 313               |
| EVANS          | 7/28/93     | 6.6       | 6.80           | 11.8        | 0.034               | 8.45            | 312               |
| EVANS          | 8/24/93     | 7.0       | 3.56           | 11.6        | 0.039               | 1.33            | 266               |
| EVANS          | 9/20/93     | 7.0       | 3.05           | 7.9         | 0.000               | 10.79           | 446               |
| EVANS          | 10/28/93    | 6.9       | 2.27           | 6.2         | 0.009               | 10.30           | 910               |
| EVANS          | 11/19/93    | 6.7       | 6.49           | 1.8         | 0.035               | 13.80           | 133               |

Appendix F.I: (cont.)

|                 |          |         |        |      |       |       |     |
|-----------------|----------|---------|--------|------|-------|-------|-----|
| LAKE            | 1/25/93  | 6.7     | 4.51   | 18.0 | 0.107 | 14.60 |     |
| LAKE            | 3/11/93  | 7.0     | 41.17  | 0.8  | 0.136 | 13.07 | 311 |
| LAKE            | 3/16/93  | 6.8     | 57.06  | 3.3  | 0.104 | 11.78 | 332 |
| LAKE            | 3/23/93  |         | 700.00 | 36   |       |       |     |
| LAKE            | 4/14/93  | 7.2     | 51.35  | 5.9  | 0.091 | 13.02 |     |
| LAKE            | 5/4/93   | 7.0     | 840.00 |      | 0.070 |       |     |
| LAKE            | 6/28/93  | 7.2     | 3.97   | 13.4 | 0.045 | 15.00 | 240 |
| LAKE            | 7/28/93  | 7.0     | 6.18   | 15.6 | 0.056 | 7.36  | 275 |
| LAKE            | 8/24/93  | 7.1     | 1.34   | 14.1 | 0.059 | 1.21  | 299 |
| LAKE            | 9/20/93  | 7.2     | 1.48   | 6.9  | 0.060 | 10.38 | 469 |
| LAKE            | 10/28/93 | 7.0     | 0.82   | 2.4  | 0.061 | 15.09 | 942 |
| LAKE CONTROL    | 3/23/93  |         |        |      |       |       |     |
| LAKE-DOWNSTREAM |          | 3/16/93 |        |      |       |       |     |
| LAKE-DOWNSTREAM |          | 3/23/93 |        |      |       |       |     |



**APPENDIX F.2: Monthly Water Quality Parameters for selected Coeur d'Alene tributaries during 1993.**

| LOCATION | DATE     | URB<br>(NTU) | TSS<br>(mg/l) | NO3<br>(ppm) | NO3<br>(ppm) | NO2<br>(ppm) | NO2<br>(ppm) | TOT N<br>(mg/l) | TKN<br>(mg/l) | PO4<br>(ppm) | IRON<br>(ppm) | ALK<br>(ppm) | Cl<br>(mg/l) | Total P<br>(mg/l) |
|----------|----------|--------------|---------------|--------------|--------------|--------------|--------------|-----------------|---------------|--------------|---------------|--------------|--------------|-------------------|
| ALDER    | 3/11/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| ALDER    | 3/16/93  | 27           | 18            | 0.29         | 1.28         | 0.07         | 0.23         |                 |               | 0.70         | 0.94          | 30           |              |                   |
| ALDER    | 3/23/93  | 75           | 100           |              |              |              |              |                 |               |              |               |              |              |                   |
| ALDER    | 4/14/93  | 9            | 7.5           | 0.04         | 0.18         | 0.02         | 0.07         |                 |               | 0.01         | 0.29          | 40           |              |                   |
| ALDER    | 5/4/93   | 320          | 5940          |              |              |              |              | 0.01            | 6.50          |              |               |              | 2            | 1.50              |
| ALDER    | 6/28/93  | 18           | 2.75          | 0.00         |              | 0.03         |              |                 |               | 0.01         | 0.60          |              |              |                   |
| ALDER    | 7/27/93  | 24           | 66            | 0.08         |              | 0.01         |              |                 |               | 0.81         |               | 30           |              |                   |
| ALDER    | 8/24/93  | 24           | 6             | 0.00         |              | 0.03         |              |                 |               | <b>3.00</b>  |               | 60           |              |                   |
| ALDER    | 9/20/93  | 38           | 8             |              |              |              |              |                 |               | 0.07         |               | 70           |              |                   |
| ALDER    | 10/28/93 | 18           |               | 0.04         |              | 0.03         |              |                 |               | 1.18         |               | 50           |              |                   |
| ALDER    | 11/19/93 | 172          |               | 0.00         |              | 0.00         |              |                 |               | 0.01         |               | 70           |              |                   |
| BENEWAH  | 1/20/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| BENEWAH  | 3/11/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| BENEWAH  | 3/16/93  | 46           | 26.7          | 0.29         | 1.28         | 0.01         | 0.03         |                 |               | 0.44         | 1.98          | 30           |              |                   |
| BENEWAH  | 3/23/93  | 164          | 1093.3        |              |              |              |              |                 |               |              |               |              |              |                   |
| BENEWAH  | 4/14/93  | 40           | 12            | 0.00         |              | 0.01         | 0.03         |                 |               | 0.07         | 0.80          | 50           |              |                   |
| BENEWAH  | 5/4/93   | 87           | 640           |              |              |              |              | 0.04            | 0.87          |              |               |              | 1            | 0.72              |
| BENEWAH  | 6/28/93  | 12           | 1.25          | 0.00         |              | 0.00         |              |                 |               | 1.43         | 0.60          |              |              |                   |
| BENEWAH  | 7/28/93  | 24           | 8             | 0.00         |              | 0.01         |              |                 |               | 0.12         |               | 30           |              |                   |
| BENEWAH  | 8/24/93  | 12           | 2             | 0.00         |              | 0.06         |              |                 |               | 1.11         |               | 50           |              |                   |
| BENEWAH  | 9/20/93  | 24           | 30            |              |              |              |              |                 |               | 0.00         |               | 50           |              |                   |
| BENEWAH  | 10/28/93 | 12           |               | 0.00         |              | 0.03         |              |                 |               | 0.52         |               | <b>50</b>    |              |                   |
| BENEWAH  | 11/19/93 | 12           |               | 0.00         |              | 0.00         |              |                 |               | 0.00         |               | 50           |              |                   |
| EVANS    | 1/20/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| EVANS    | 3/11/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| EVANS    | 3/16/93  | >500         | 6.7           | 0.00         |              | 0.00         |              |                 |               | 0.18         | 0.60          | 30           |              |                   |
| EVANS    | 3/23/93  | 35           | 106.7         |              |              |              |              |                 |               |              |               |              |              |                   |
| EVANS    | 4/15/93  | 30           | 2.25          | 0.00         |              | 0.04         | 0.13         |                 |               | 0.91         | 0.29          | 20           |              |                   |
| EVANS    | 5/4/93   | 30           | 460           |              |              |              |              | 0.01            | 0.25          |              |               |              | 2            | 0.18              |
| EVANS    | 6/28/93  | 0            | 2.5           | <b>0.00</b>  |              | 0.03         |              |                 |               | 0.00         | 0.34          |              |              |                   |
| EVANS    | 7/28/93  | 0            | 2             | <b>0.00</b>  |              | 0.00         |              |                 |               | 0.00         |               | 30           |              |                   |
| EVANS    | 8/24/93  | 5            | 10            | <b>0.00</b>  |              | 0.08         |              |                 |               | 0.18         |               | 30           |              |                   |
| EVANS    | 9/20/93  | 5            | 2             |              |              |              |              |                 |               | 0.07         |               | 40           |              |                   |
| EVANS    | 10/28/93 | 12           |               | 0.00         |              | 0.03         |              |                 |               | 1.68         |               | 40           |              |                   |
| EVANS    | 11/19/93 | 0            |               | 0.00         |              | 0.00         |              |                 |               | 0.40         |               | 50           |              |                   |
| LAKE     | 1/25/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |
| LAKE     | 3/11/93  |              |               |              |              |              |              |                 |               |              |               |              |              |                   |

Appendix F.2. (cont.)

| LOCATION        | DATE     | T URB<br>(NTU) | TSS<br>(mg/l) | NO3<br>(ppm) | <i>NO3</i><br>(ppm) | NO2<br>(ppm) | NO2<br>(ppm) | TOT N<br>(mg/l) | TKN<br>(mg/l) | PO4<br>(ppm) | IRON<br>(ppm) | ALK<br>(ppm) | Cl<br>(mg/l) | Total P<br>(mg/l) |
|-----------------|----------|----------------|---------------|--------------|---------------------|--------------|--------------|-----------------|---------------|--------------|---------------|--------------|--------------|-------------------|
| LAKE            | 3/16/93  | 35             |               | >3.58        | 15.75               | 0.04         | 0.13         |                 |               | 0.36         | 0.85          | 30           |              |                   |
| LAKE            | 3/23/93  |                | 286.7         |              |                     |              |              |                 |               |              |               |              |              |                   |
| LAKE            | 4/14/93  | 195            | 265           | 1.03         | 4.53                | 0.03         | 0.10         |                 |               | 2.30         | 3.12          | 40           |              |                   |
| LAKE            | 5/4/93   | 75             | 280           |              |                     |              |              | 1.00            | 0.56          |              |               |              | 2            | 0.22              |
| LAKE            | 6/28/93  | 9              | 9.25          | 0.00         |                     | 0.01         |              |                 |               | 0.23         | 0.94          |              |              |                   |
| LAKE            | 7/28/93  | 24             | 19            | 0.00         |                     | 0.04         |              |                 |               | 0.44         |               | 30           |              |                   |
| LAKE            | 8/24/93  | 18             | 5             | 0.00         |                     | 0.08         |              |                 |               | 0.27         |               | 30           |              |                   |
| LAKE            | 9/20/93  | 21             | 10            |              |                     |              |              |                 |               | 0.07         |               | 40           |              |                   |
| LAKE            | 10/28/93 | 18             |               | 0            |                     | 0.03         |              |                 |               | 1.92         |               | 50           |              |                   |
| LAKE CONTROL    | 3/23/93  | 286            |               |              |                     |              |              |                 |               |              |               |              |              |                   |
| LAKE-DOWNSTREAM | 3/16/93  |                | 86.7          |              |                     |              |              |                 |               |              |               |              |              |                   |
| LAKE-DOWNSTREAM | 3/23/93  | >500           |               |              |                     |              |              |                 |               |              |               |              |              |                   |

**Appendix F.3: Water Quality data collected using price pygmy meter (discharge) and Hydrolab multi-meter in selected tributaries during 1994.**

| LOCATION | DATE     | Q (cfs)      | (C <sup>o</sup> ) | CONDUCTIVITY (mmhos/cm) | DO (ppm)    | REDOX (mV) | STAGE       |
|----------|----------|--------------|-------------------|-------------------------|-------------|------------|-------------|
| ALDER    | 3/22/94  | 19.01        | 1.13              | <b>0.0386</b>           | 13.98       | <b>379</b> | 1.48        |
| ALDER    | 4/22/94  | 14.48        | 5                 |                         |             |            | 1.4         |
| ALDER    | 5/18/94  | <b>9.03</b>  | <b>9.76</b>       | 0.0671                  | 11.31       | <b>278</b> | 1.2         |
| ALDER    | 6/22/94  | 1.8          |                   | 0.0761                  | 9.15        | <b>277</b> | 1.1         |
| ALDER    | 8/29/94  | <.5          | 16.48             | 0.0964                  | <b>7.38</b> | <b>493</b> |             |
| ALDER    | 9/20/94  | <.5          | 13.33             | <b>0.095</b>            | <b>8.32</b> | <b>277</b> |             |
| ALDER    | 10/18/94 | 1.46         | 5.7               | <b>0.0935</b>           | <b>8.75</b> | <b>347</b> |             |
| ALDER    | 11116194 | <b>2.79</b>  | <b>0.87</b>       | <b>0.0963</b>           | <b>9.93</b> | <b>354</b> | 1.16        |
| BENEWAH  | 3/22/04  | <b>32.09</b> | <b>2.04</b>       | <b>0.0408</b>           | 14.08       | 416        | <b>0.57</b> |
| BENEWAH  | 4/18/94  | <b>2.93</b>  | 16.83             | <b>0.0424</b>           | <b>8.82</b> | <b>336</b> | <b>0.68</b> |
| BENEWAH  | 5/16/94  | 7.18         | 17.35             | <b>0.0569</b>           | <b>9.45</b> | <b>345</b> | 0.1         |
| BENEWAH  | 6/22/94  | <b>2.38</b>  | <b>28.07</b>      | <b>0.059</b>            | 8.18        | 177        |             |
| BENEWAH  | 7/18/94  | <b>2.24</b>  | 23.31             | <b>0.069</b>            | 9.14        | 160        |             |
| BENEWAH  | 8/29/94  | <.5          | 17.07             | <b>0.0938</b>           | <b>9.62</b> | 368        |             |
| BENEWAH  | 9/19/94  | <b>0.56</b>  | 10.2              | <b>0.0904</b>           | 10.2        | <b>378</b> | <b>0</b>    |
| BENEWAH  | 10/18/94 | 1.34         | <b>7.79</b>       | <b>0.0758</b>           | <b>9.33</b> | <b>394</b> | <b>0</b>    |
| BENEWAH  | 11/16/94 | <b>7.48</b>  | 1.72              | 0.0615                  | 10.81       | <b>338</b> | <b>0.24</b> |
| EVANS    | 3/21/94  | 15.68        | <b>3.66</b>       | 0.0221                  | 13.56       | <b>407</b> |             |
| EVANS    | 4/19/94  | <b>32.79</b> | <b>6.97</b>       | 0.0156                  | 10.94       | <b>328</b> |             |
| EVANS    | 5/16/94  | 15.74        | <b>8.24</b>       | 0.0117                  | 11.5        | <b>336</b> | 3.11        |
| EVANS    | 6/20/94  | 2.61         | 12.03             | <b>0.0248</b>           | 10.45       | <b>326</b> | <b>3.05</b> |
| EVANS    | 7/18/94  | <b>3.49</b>  | 13.99             | <b>0.0282</b>           | 10.02       | <b>279</b> | 3.1         |
| EVANS    | 8/29/94  | <b>1.77</b>  | 12.67             | <b>0.0289</b>           | 9.18        | <b>360</b> |             |
| EVANS    | 9/19/94  | 1.4          | 11.6              | <b>0.0396</b>           | <b>9.6</b>  | <b>356</b> |             |
| EVANS    | 10119194 | 1.25         | 6.41              | 0.0314                  | <b>9.44</b> | <b>466</b> |             |
| EVANS    | 11115194 | <b>3.57</b>  | <b>2.74</b>       | <b>0.0303</b>           | 10.55       | 412        |             |
| LAKE     | 1/4/94   | <b>63.82</b> | 0.17              | <b>0.0544</b>           | 8.15        | <b>339</b> |             |
| LAKE     | 2/18/94  | 12.26        | 0                 |                         |             |            |             |
| LAKE     | 3/21/94  | 38.41        | <b>2.35</b>       | <b>0.0553</b>           | 13.8        | <b>385</b> | 1           |
| LAKE     | 4/18/94  | 18.96        | <b>9.94</b>       | <b>0.0376</b>           | <b>9.43</b> | <b>367</b> |             |
| LAKE     | 5/16/94  | <b>6.02</b>  | <b>9.63</b>       | <b>0.0404</b>           | 11.05       | <b>357</b> | <b>0.3</b>  |
| LAKE     | 6/22/94  | 1.78         | 15.92             | 0.9431                  | 9.71        | <b>227</b> |             |
| LAKE     | 7/18/94  | <b>0.32</b>  | <b>20.06</b>      | 0.0912                  | <b>8.58</b> | 176        |             |
| LAKE     | 8/29/94  | <.01         | 11.33             | <b>0.0233</b>           | <b>7.67</b> | <b>337</b> |             |
| LAKE     | 9/19/94  | <b>0.5</b>   | 14.86             | <b>0.0833</b>           | <b>8.8</b>  | <b>337</b> |             |
| LAKE     | 10117/94 | <b>0.57</b>  | <b>5.19</b>       | 0.0613                  | <b>9.95</b> | <b>366</b> |             |
| LAKE     | 11/15/94 | <b>2.88</b>  | 1.02              | 0.0514                  | 1118        | <b>455</b> |             |

**Appendix F.4: Monthly Water Quality Parameters for selected Coeur d' Alene tributaries in 1994.**

| LOCATION | DATE     | pH    | TUR (NTU) | TSS (mg/l) | NO <sup>3</sup> (ppm) | NO <sup>2</sup> (ppm) | PHOSPHATE (ppm) | IRON (ppm) | ALKALINITY (ppm) |
|----------|----------|-------|-----------|------------|-----------------------|-----------------------|-----------------|------------|------------------|
| ALDER    | 3/22/94  | 7.48  | 15        |            | 0                     | 0.03                  | 0               |            | 55               |
| ALDER    | 4/22/94  |       | 5         |            |                       | 0.01                  |                 |            | 5                |
| ALDER    | 5/18/94  | 7.75  | 18        |            | 0                     | 0.01                  | 0.07            |            | 60               |
| ALDER    | 6/22/94  | 8.24  | 21        |            | 0.04                  | 0.08                  | 0.12            |            | 75               |
| ALDER    | 8/29/94  | 6.69  | 24        |            |                       | 0.05                  | 0.07            |            | 50               |
| ALDER    | 9/20/94  | 7     | 24        |            |                       | 0.08                  | 0.18            |            | 70               |
| ALDER    | 10/18/94 | 6.99  | 15        |            | 0.05                  |                       | 0.07            |            | 70               |
| ALDER    | 11/16/94 | 7.16  |           |            |                       |                       |                 |            |                  |
| BENEWAH  | 3/22/04  | 7.69  | 30        |            | 0                     | 0.03                  | 0.07            |            | 40               |
| BENEWAH  | 4/18/94  | 7.78  | 9         | 10         |                       | 0.01                  | 0.01            |            | 5                |
| BENEWAH  | 5/16/94  | 7.08  | 24        |            | 0                     | 0                     | 0.07            |            | 60               |
| BENEWAH  | 6/22/94  | 10.33 | 15        |            |                       | 0.08                  | 0.18            |            | 50               |
| BENEWAH  | 7/18/94  | 10.68 | 50        |            |                       | 6                     | 0.12            |            | 50               |
| BENEWAH  | 8/29/94  | 8.26  | 12        |            |                       | 0.01                  | 0.07            |            | 35               |
| BENEWAH  | 9/19/94  | 7.86  | 12        |            |                       | 0.08                  | 0.12            |            | 70               |
| BENEWAH  | 10/18/94 | 7.69  | 9         |            |                       | 0.04                  | 0.07            |            | 70               |
| BENEWAH  | 11/16/94 | 7.35  |           |            |                       |                       |                 |            |                  |
| EVANS    | 3/21/94  | 7.2   | 5         |            | 0                     | 0.05                  | 0.01            |            | 40               |
| EVANS    | 4/19/94  | 6.92  | 3         | 6          |                       |                       | 0.07            |            | 4                |
| EVANS    | 5/16/94  | 6.88  | 0         |            | 0                     | 0.02                  | 0.01            |            | 50               |
| EVANS    | 6/20/94  | 8.07  | 5         |            |                       | 0.07                  | 0.07            |            | 70               |
| EVANS    | 7/18/94  | 8.51  | 5         |            |                       | 0.04                  | 0.01            |            | 50               |
| EVANS    | 8/29/94  | 7.01  | 0         |            |                       | 0.03                  | 0.01            |            | 35               |
| EVANS    | 9/19/94  | 7.05  | 5         |            | 0                     | 0.03                  | 0.01            |            | 50               |
| EVANS    | 10/19/94 | 6.87  | 5         |            | 0                     | 0.08                  | 0.01            |            | 50               |
| EVANS    | 11/15/94 | 7.01  |           |            |                       |                       |                 |            |                  |
| LAKE     | 1/4/94   | 8     | 54        |            | 0.98                  | 0.03                  | 0.27            |            | 40               |
| LAKE     | 2/18/94  |       |           | 43         |                       |                       |                 |            |                  |
| LAKE     | 3/21/94  | 7.36  | 40        |            | 0.61                  | 0.03                  | 0.07            |            | 30               |
| LAKE     | 4/18/94  | 7.21  | 9         | 7          |                       |                       | 0.01            |            | 3                |
| LAKE     | 5/16/94  | 6.54  | 21        |            | 0                     | 0.03                  | 0.07            |            | 40               |
| LAKE     | 6/22/94  | 8.93  | 9         |            | 0.09                  | 0.09                  | 0.32            |            | 45               |
| LAKE     | 7/18/94  | 8.51  | 15        |            |                       | 0.07                  | 0.4             |            | 50               |
| LAKE     | 8/29/94  | 7.33  | 12        |            |                       | 0.05                  | 0.12            |            | 40               |
| LAKE     | 9/19/94  | 7.24  | 18        |            |                       | 0.03                  | 0.27            |            | 50               |
| LAKE     | 10/17/94 | 7.1   | 9         |            |                       | 0.08                  | 0.18            |            | 60               |
| LAKE     | 11/15/94 | 7.09  |           |            |                       |                       |                 |            |                  |

# **Coeur d'Alene Tribe Fish, Water, and Wildlife Program**

## **Supplementation Feasibility Report on the Coeur d'Alene Indian Reservation**



# **Coeur d'Alene Tribe Fish, Water, and Wildlife Program**

## **Supplementation Feasibility Report on the Coeur d'Alene Indian Reservation**

**Prepared By**

**Ron Peters  
Angelo J. Vitale  
and  
Kelly Lillengreen  
Program Manager**

**August 1998**

**Coeur d'Alene Tribe Department of Natural Resources  
Fish, Water, and Wildlife Program  
850 A Street, P.O. Box 408  
Plummer, ID 83851-0408**

**PHONE: (208) 686-5302  
FAX: (208) 686-3021**

# Table of Contents

|  |    |
|--|----|
| <b>1.0 Introduction</b> .....  | 1  |
| 1.1 Study Area .....   | 3  |
| Coeur d'Alene Lake   |    |
| Target Tributaries   |    |
| <b>2.0 Methods and Materials</b> .....   | 4  |
| 2.1 Lake Studies .....   | 4  |
| 2.1.1 Water Quality .....  | 4  |
| 2.1.2 Fisheries .....  | 8  |
| 2.2 Stream Studies .....   | 10 |
| 2.2.1 Water Quality .....  | 10 |
| 2.2.2 Habitat Suitability Index .....  | 11 |
| 2.2.3 Fisheries .....  | 12 |
| <b>3.0 Results</b> .....   | 18 |
| 3.1 Lake Studies .....   | 18 |
| 3.1.1 Water Quality .....  | 18 |
| 3.1.2 Fisheries .....  | 37 |
| 3.2 Stream Studies .....   | 38 |
| 3.2.1 Water Quality .....  | 38 |
| 3.2.2 Habitat Suitability Indices .....  | 43 |
| 3.2.3 Fisheries .....  | 44 |
| <b>4.0 Discussion</b> .....  | 54 |
| 4.1 Limiting Factors .....   | 54 |
| 4.2 Supplementation Feasibility .....  | 58 |
| <b>5.0 Bibliography</b> .....  | 62 |
| <br>   |    |
| <b>Appendix A. Vertical hydrolab profiles of thirteen stations on Coeur d'Alene Lake, 1997.</b>  |    |
| <b>Appendix B. Secchi readings and empirically derived estimates of euphotic zone depth for thirteen stations on Coeur d'Alene Lake, 1997.</b> |    |
| <b>Appendix C. Cohort analysis of growth for cutthroat trout and brook trout, 1996-1997.</b>   |    |

## List of Tables

|   |    |
|---|----|
| Table 1.1 Basin morphometry of the Lake Creek, Benewah Creek, Alder Creek and Evans Creek watersheds.         | 4  |
| Table 2.1 Geomorphologically similar water quality sample sites on Coeur d’Alene Lake.                        | 6  |
| Table 2.2 Streams sampled for water quality parameters  | 11 |
| Table 2.3 Juvenile and adult fish traps were operated at 11 different locations in the four target drainages. | 18 |
| Table 3.1 Sample stations on Coeur d’Alene Lake that had a dissolved oxygen reading less than 6.0 mg/L.       | 19 |
| Table 3.2 Total suspended solids (mg/L) results from epilimnion of Coeur d’Alene Lake.                        | 29 |
| Table 3.3 Total suspended solids (mg/L) results from hypolimnion of Coeur d’Alene Lake                        | 29 |
| Table 3.4 Turbidity (NTU) results from epilimnion of Coeur d’Alene Lake.                                      | 31 |
| Table 3.5 Turbidity (NTU) results from hypolimnion of Coeur d’Alene Lake                                      | 31 |
| Table 3.6 Ortho-Phosphate (mg/L) results from epilimnion of Coeur d’Alene Lake.                               | 32 |
| Table 3.7 Ortho-Phosphate (mg/L) results from hypolimnion of Coeur d’Alene Lake.                              | 32 |
| Table 3.8 Nitrate (mg/L) results from epilimnion of Coeur d’Alene Lake.                                       | 33 |
| Table 3.9 Nitrate (mg/L) results from hypolimnion of Coeur d’Alene Lake                                       | 33 |
| Table 3.10 Nitrite (mg/L) results from epilimnion of Coeur d’Alene Lake.                                      | 34 |
| Table 3.11 Nitrite (mg/L) results from hypolimnion of Coeur d’Alene Lake.                                     | 34 |
| Table 3.12 Average annual and seasonal Secchi and euphotic zone measurements on Coeur d’Alene Lake.           | 35 |
| Table 3.13 Chlorophyll <sub>a</sub> (µg/L) results from epilimnion of Coeur d’Alene Lake.                     | 36 |
| Table 3.14 Chlorophyll <sub>a</sub> (µg/L) results from hypolimnion of Coeur d’Alene Lake.                    | 36 |
| Table 3.15 Electrofishing relative abundance from 1994-1997.  | 37 |
| Table 3.16 Electrofishing catch per unit effort (CPUE) from 1994-1997.  | 39 |
| Table 3.17 Gillnetting relative abundance from 1994-1997.   | 40 |
| List of tables continued-   |    |



|   |    |
|---|----|
| Table 3.18 Gillnetting catch per unit effort (CPUE) from 1994-1997. ....          | 41 |
| Table 3.19 Habitat suitability index for lacustrine cutthroat trout.....          | 42 |
| Table 3.20 Habitat suitability index for riverine cutthroat trout.....            | 43 |
| Table 3.21 Cutthroat trout abundance and distribution 1996. ....                  | 45 |
| Table 3.22 Cutthroat trout abundance and distribution 1997. ....                  | 46 |
| Table 3.23 Age frequency of cutthroat trout 1997. ....                            | 47 |
| Table 3.24 Brook trout abundance and distribution 1997. ....                      | 48 |
| Table 3.25 Mean back calculated lengths (mm) at age for cutthroat trout 1996..... | 48 |
| Table 3.26 Mean back calculated lengths (mm) at age for cutthroat trout 1997..... | 48 |

## List of Figures

|  |    |
|--|----|
| Figure 2.1 Coeur d'Alene Lake water quality sample sites. ....                                     | 5  |
| Figure 2.2 Coeur d'Alene Lake fisheries sample sites. ....   | 9  |
| Figure 2.3 Lake Creek shock sites. ....  | 13 |
| Figure 2.4 Benewah Creek shock sites. ....   | 14 |
| Figure 2.5 Evans Creek shock sites. ....   | 15 |
| Figure 2.6 Alder Creek shock sites. ....   | 16 |
| Figure 3.1 pH profiles for three similar locations on Coeur d'Alene Lake, 1997 .....               | 20 |
| Figure 3.2 Conductivity profile Coeur d'Alene River, 1997. ....                                    | 21 |
| Figure 3.3 Temperature profiles for Round Lake and Chatcolet shallow sampling stations, 1997. .... | 23 |
| Figure 3.4 Temperature profiles for Benewah Lake, Chatcolet Lake and Hidden Lake, 1997. ....       | 24 |
| Figure 3.5 Temperature profiles for Rockford, Carey, Windy Bays, 1997. ....                        | 25 |
| Figure 3.6 Temperature profiles for Windy Bay deep, Mid Lake and Conkling Point, 1997. ....        | 26 |
| Figure 3.7 Temperature profiles for Coeur d'Alene River, 1997. ....                                | 27 |
| Figure 3.8 Temperature profiles for St. Joe River, 1997. ....                                      | 28 |
| Figure 3.9 Regression equations of body length and age for cutthroat and brook trout, 1996. ....   | 49 |
| Figure 3.10 Regression equations of body length and age for cutthroat and brook trout, 1997. ....  | 50 |
| Figure 3.11 Cutthroat trout migration timing in Lake Creek, 1996. ....                             | 52 |
| Figure 3.12 Cutthroat trout migration timing in Lake Creek, 1997. ....                             | 53 |
| Figure 4.1 Supplementation flow chart. ....  | 61 |

## Abstract

This study was initiated to respond to concerns regarding recent declines in native salmonid fish populations, particularly westslope cutthroat trout (*Oncorhynchus clarki*) and bull trout (*Salvelinus confluentus*), in the Coeur d'Alene basin. These declines are a result of anthropogenic disturbances that include construction of Post Falls Dam, major alterations in land cover types, changing land use patterns, and introduction of exotic fish species. This report 1) examines the effects of lake and stream habitat, species interactions and water quality on cutthroat trout population dynamics; 2) provides population estimates for westslope cutthroat trout in target tributaries; 3) defines supplementation and evaluates the role that supplementation may play in rebuilding cutthroat trout stocks on the reservation; and 4) develops the framework for constructing a hatchery and implementing a supplementation program.

Water quality in Coeur d'Alene Lake has a detrimental effect on habitat suitability for cutthroat trout. Temperatures recorded in the upper 10 meters of the water column are generally outside of the optimal range for rearing. Furthermore, the euphotic zone rarely drops below 10 meters so foraging runs require fish to enter areas with potentially stressful conditions. Indirect effects related to low dissolved oxygen concentrations, total suspended solids, and large masses of aquatic macrophytes effectively limit the cutthroat trout population in the areas where these conditions exist. Specifically, these areas include the littoral zones less than 10 meters deep, which are most affected by construction of Post Falls Dam.

Competition occurs between cutthroat trout and other fish species in the lake. The introduction of non-native species has altered the historic behavioral pattern of cutthroat trout in both the littoral and limnetic zones. Introduced species comprised over 68 percent of the catch in relative abundance studies from 1994-1997 while cutthroat trout comprised less than 1 percent of the catch. Five introduced species have been demonstrated to actively feed on other fish species, including adult and juvenile cutthroat trout. Historically, bull trout and northern squawfish were the only predators of cutthroat trout in the lake.

The availability of summer rearing habitat for cutthroat trout is a primary constraint on tributary populations. The range of suitable summer rearing habitat in each of the target watershed is significantly reduced when compared with the historic range of the fish. Population estimates have consistently shown that abundance of juvenile cutthroat trout is greatest in first and second order tributaries, where water quality conditions are most favorable. Typical base flow conditions, however, force juvenile trout into small pools where competition for space and food may occur. Furthermore, displacement into water quality limited reaches may be a significant source of mortality. Application of habitat suitability indices indicates that improving habitat condition through restoration and protection has the potential to partially compensate for short-term degradation in water quality.

Due to the persistence of adverse conditions in natal streams and Coeur d'Alene Lake, cutthroat trout populations are thought to be at least moderately damaged. Based on the analysis of limiting factors, gains from habitat restoration and protection alone will not achieve the goals set forth by the Coeur d'Alene Tribal Council to provide for self-sustaining populations and harvestable numbers of cutthroat trout. Supplementation is proposed as an additional means to boost the population density above a certain minimum viable population size. The minimum viable spawning escapement size and the minimum effective breeding number are values that are currently under investigation and will need to be refined to develop an initial management strategy. The most effective management strategy would employ a supplementation program to a level that minimizes the risk of extirpation.

## 1.0 Introduction

Recent declines in native salmonid fish populations, particularly westslope cutthroat trout (*Oncorhynchus clarki*) and bull trout (*Salvelinus confluentus*), in the Coeur d' Alene basin have been the focus of study by the Coeur d' Alene Fish, Water, and Wildlife program since 1990. In fact, early studies on Coeur d'Alene Lake showed that significant declines had occurred as early as 1932. It appears that there are a number of factors that contributed to the decline of resident fish stocks within Coeur d'Alene Lake and its tributaries (Ellis 1932; Oien 1957; Mallet 1969; Scholz et. al. 1985, Lillengreen et. al. 1993). These factors include: construction of Post Falls Dam in 1906; major changes in land cover types from primarily forested areas to forests with recent and recovering clearcuts, agricultural and pasture lands, urban development, mining and open range land; and introduction of exotic fish species.

Over 100 years of mining activities in the Silver Valley have had devastating effects on the quality of the water in the Coeur d'Alene River drainage and Coeur d'Alene Lake. Effluent from tailings and mining waste have contributed vast quantities of trace heavy metals to the system. Poor agricultural and forest practices have also contributed to the degradation of water quality and habitat suitability for resident salmonids. Increased sediment loads from agricultural runoff and recent and recovering clearcuts, and increases in water temperature due to riparian canopy removal may be two of the most important problems currently affecting westslope cutthroat trout. Increases in water temperature have reduced the range of resident salmonids to a fraction of its historic extent. Within this new range, sediment has reduced the quality of both spawning and rearing habitats. Historically, municipal waste contributed large quantities of phosphates and nitrogen that accelerated the eutrophication process in Coeur d'Alene Lake. However, over the last 25 years work has been completed to reduce the annual load of these materials. Wastewater treatment facilities have been established near all major municipalities in and around the basin.

Species interactions with introduced exotics as well as native species are also acting to limit cutthroat trout populations. Two mechanisms are at work: interspecific competition, and species replacement. Competition occurs when two species utilize common resources, the supply of which is short; or if the resources are not in short supply, they harm each other in the process of seeking these resources. Replacement occurs when some environmental or anthropogenic change (e.g., habitat degradation, fishing pressure, etc.) causes the decline or elimination of one species and another species, either native or introduced, fills the void left by the other.

Within their historic range, cutthroat trout evolved with few other fish species. Thus, they developed as “generalists” and are more susceptible to interspecific interactions than other fish which evolved in the presence of multiple species (Griffith, 1988). In the Coeur d' Alene basin, it is most likely that environmental conditions, rather than competition through interspecific interactions, shaped cutthroat trout behavior and morphology. Rapid changing environmental conditions in the Coeur d' Alene Basin has allowed introduced exotics, and native species other than cutthroat trout to proliferate. This has been shown to be true in the Columbia River system where dam construction has altered the fish species composition and allowed northern squawfish (*Ptychocheilus oregonensis* Richardson) to effectively reduce the numbers of juvenile salmon migrating downstream (Beamesderfer and Rieman 1991).

There are twenty-one identified species of fish inhabiting the Coeur d'Alene Lake study area, of which only seven are indigenous species. The indigenous species are cutthroat trout, bull trout, northern squawfish, largescale sucker (*Catostomus macrocheilus* Girard), longnose sucker (*Catostomus catostomus* Forster), mountain whitefish (*Prosopium williamsoni* Girard), and sculpin (*Cottus* sp.). The introduced exotic species include yellow perch (*Perca flavescens* Mitchill), pumpkinseed (*Lepomis gibbosus* Rafinesque), largemouth bass (*Micropterus salmoides* Lacepede), black crappie (*Pomoxis*

*nigromaculatus* Lesuerur), brown bullhead (*Ictalurus nebulosus* Lesuerur), black bullhead (*Ictalurus melas* Rafinesque), channel catfish (*Ictalurus punctatus* Rafinesque), tench (*Tinca tinca* Linnaeus), northern pike (*Esox lucius* Linnaeus), kokanee salmon (*Oncorhynchus nerka* Walbaum), chinook salmon (*Oncorhynchus tshawytscha* Walbaum), smallmouth bass (*Micropterus dolomieu* Lacepede), rainbow trout (*Oncorhynchus mykiss*), and lake superior whitefish (*Coregonis clupeaformis* Mitchell). The Idaho Department of Fish and Game has been stocking and/or managing for exotic species in Coeur d'Alene Lake for over 50 years, most notably kokanee salmon (1937), chinook salmon (1982), and northern pike (unknown). Other studies have indicated that most of these species can competitively exclude or replace cutthroat trout in both stream and lake environments (Griffith, 1972, 1974, 1988; Moyle and Vondracek, 1985; and Marnell, 1986, 1987, 1988).

Westslope cutthroat trout and bull trout exhibit three primary life history forms: adfluvial, fluvial, and resident. Of these forms, adfluvial production is considered the most important in the Coeur d'Alene basin. The reasons for this are that they attain the largest size and played an important role in the subsistence economy of the Coeur d'Alene Tribe. Adfluvial salmonids move through many different habitats during their life history, thus production is governed by a wide variety of physical, chemical, and biological influences. In particular, they differ from fluvial and resident forms in that they are dependent on both the tributary and lake environments for completion of their life cycle. Because of this dependency on multiple habitats, they are more sensitive to widespread changes in habitat quality and complexity than the fluvial and resident forms. Furthermore, the maximum potential for interaction with introduced exotic species is realized in the adfluvial life history strategies of the westslope cutthroat trout and bull trout. During the course of their lifecycle, they interact with all other species either through predation by other piscivorous fishes or through competition for living space and food.

Work conducted by the Coeur d'Alene Tribe Fish, Water and Wildlife Program has helped determine that habitat components utilized in each of the three critical life history phases, as well as interactions with introduced species, potentially limit production of adfluvial fishes. These components include spawning habitat and juvenile rearing habitat in tributary streams, and adult rearing habitat in the lake. In order to effectively increase populations of westslope cutthroat trout, habitat restoration must take place in natal streams. However, restoration of the critical tributary habitat does not guarantee increases in adfluvial trout production because adfluvial westslope cutthroat trout reside in Coeur d'Alene Lake for two-thirds of their life cycle. Evidence suggests that production of cutthroat trout is indirectly limited by lake habitat features, but the extent of this limitation is not fully understood.

It is the intent of the Coeur d'Alene Tribe Fish, Water and Wildlife program to increase native populations of bull trout and cutthroat trout to sustainable and harvestable levels. This will be achieved through restoration of critical tributary habitat and management of limiting factors in the lake. This report will examine, in as much detail as possible, the specific abiotic and biotic factors affecting the cutthroat trout population in Coeur d'Alene Lake and its tributaries. Options will be provided for the effective restoration of adfluvial cutthroat trout populations which will best serve the Coeur d'Alene Tribe's goal of self-sustaining and harvestable stocks of bull trout and westslope cutthroat trout.

In order to determine what steps are necessary to rebuild existing trout populations to self-sustaining and harvestable levels on the Coeur d'Alene Reservation, the following goals and objectives have been identified:

- Determine the effect lake habitat, species interactions, and water quality have on cutthroat trout populations.
- Evaluate the effects of habitat, water quality and interspecific interactions on cutthroat trout population dynamics in target watersheds.

- Provide population estimates of westslope cutthroat trout in target watersheds.
- Define supplementation and evaluate the role that supplementation may play in rebuilding cutthroat trout stocks in target watershed on the Coeur d'Alene Indian Reservation.
- Develop the framework for constructing a hatchery and implementing a supplementation program.

These objectives are designed to facilitate the decision making process regarding the need and/or feasibility of supplementing the adfluvial cutthroat trout population in Coeur d'Alene Lake.

## 1.1 Study Area

### **Coeur d'Alene Lake**

Coeur d'Alene Lake is the second largest lake in Idaho and is located in the panhandle section of northern Idaho. Population centers are located in the Northern most shoreline of Coeur d'Alene Lake (Coeur d'Alene) and at the mouth of the Coeur d'Alene River (Harrison). The lake is located in two Idaho counties: Kootenai and Benewah. The city of Coeur d'Alene is the largest in Kootenai County and Harrison is the second largest in Benewah County. The largest town in Benewah County (St. Maries) lies about 19 kilometers upstream of Coeur d'Alene Lake on the St. Joe River.

Coeur d'Alene Lake is within the 17,300 square kilometer Spokane River drainage basin. The lake lies in a naturally dammed river valley with the outflow currently controlled by Post Falls Dam. Post Falls Dam controls the level of the lake from the dam to the town of St. Maries on the St. Joe River. At full pool (lake elevation 648.7 meters) the lake covers 129 square kilometers and at minimum pool level (lake elevation of 646.2 meters) the lake covers 122 square kilometers with a mean depth of 22 meters and a maximum depth of 63.7 meters. The lake is 42 kilometers long and anywhere from 1.6 to 9 kilometers wide. Morphometric data was taken from Woods and Berenbrock (1994).

Many tributaries feed the lake. The two main tributaries of Coeur d'Alene Lake, Coeur d'Alene and St. Joe Rivers, drain the Coeur d'Alene and St. Joe Mountains. These mountains are composed of primarily metasedimentary rocks of the belt group with local intrusions of granitics. Lower elevations are composed primarily of glaciofluvial deposits. The southern end of Coeur d'Alene Lake is made up of four shallow lakes (Hidden, Round, Chatcolet, and Benewah) flooded as a result of construction of Post Falls Dam.

The regional climate is subhumid temperate with cool, wet winters and warm, dry summers. The lake receives about 64.5 centimeters of precipitation annually with more in the higher elevations around the lake (97.2 centimeters Wallace, ID). A distinct precipitation season typically begins in October or November and continues through April or May. Approximately two-thirds of annual precipitation occurs during this period. The average daily maximum temperature in July is 30° C, the average daily minimum in January is -5° C. Moist, Pacific air masses that enter the area in late winter and early spring often generate rain-on-snow events. Geological data was taken from U.S. Department of Agriculture (1984).

## Target Tributaries

Four target tributaries, including Alder, Evans, Benewah, and Lake Creeks have been identified and described in previous reports (Lillengreen 1993; Lillengreen, et. al. 1996; Kootenai-Shoshone Soil Conservation District 1991; Krueger 1998). Basin morphometrics were derived from the Tribal GIS database following the definitions of Gardiner (1990), and are given in Table 1.1

Table 1.1 Basin morphometry of the Lake Creek, Benewah Creek, Alder Creek and Evans Creek watersheds.

| Characteristic   | Lake Creek                      | Benewah Creek                    | Alder Creek                     | Evans Creek                   |
|------------------|---------------------------------|----------------------------------|---------------------------------|-------------------------------|
| Basin Area       | 93 km <sup>2</sup> (23,117 ac.) | 152 km <sup>2</sup> (37,447 ac.) | 69 km <sup>2</sup> (17,047 ac.) | 34 km <sup>2</sup> (8,512 ac) |
| Basin Length     | 16.2 km                         | 22.2 km.                         | 20.2 km.                        | 10.4 km.                      |
| Basin Relief     | 938 m                           | 772 m                            | 820 m                           | 999 m                         |
| Basin Perimeter  | 55.2 km                         | 82.3 km                          | 43.2 km                         | 27.8 km                       |
| Relief Ratio     | 0.057                           | 0.034                            | 0.040                           | 0.095                         |
| Channel Length*  | 152.8 km                        | 219.6 km                         | 110.0 km                        | 46.6 km                       |
| Drainage Density | 1.64 km/km <sup>2</sup>         | 1.44 km/km <sup>2</sup>          | 1.59 km/km <sup>2</sup>         | 1.37 km/km <sup>2</sup>       |

\*Includes intermittent tributaries.

## 2.0 Materials and Methods

### 2.1 Lake Studies

#### 2.1.1 Water Quality

Coeur d'Alene Lake is a multi-basin lake with variable water quality conditions. Lake level fluctuations and water retention times in the various sub-basins cause water quality to change during the course of the year and from year to year. This compounds the problem associated with development of a lake wide management strategy because an observed problem may not be recognized in successive years. An appropriate monitoring strategy must account for these fluctuations. The Coeur d'Alene Lake Management Plan (1994) attempted to address this issue by dividing the lake into four water quality management zones. These zones include, nearshore (water depths less than 20 feet), shallow southern lake (Coeur d'Alene Lake south of the confluence with the Coeur d'Alene River), lower rivers (Coeur d'Alene and St. Joe Rivers affected by back water from Coeur d'Alene Lake), and deep open water (north of the mouth of the Coeur d'Alene River).

Coeur d'Alene Tribe Fish, Water, and Wildlife Program staff monitored stations in the southern section of Coeur d'Alene Lake from Rockford Bay south to the St. Joe River. Selection of sample stations was based on geomorphology, visual habitat characteristics, and the potential for changing water quality conditions during the course of the year. Thirteen sample stations were selected to encompass all four water quality management zones identified in the Coeur d'Alene Lake Management Plan (Figure 2.1). These sites, however, do not include a majority of the deep open water zone, which is a major factor in controlling the water quality of the outflow leaving Coeur d'Alene Lake.

The sample stations can be categorized into five very distinct habitat areas. The first distinct habitat area is comprised of two shallow water stations created entirely by inundation from Post Falls Dam. This

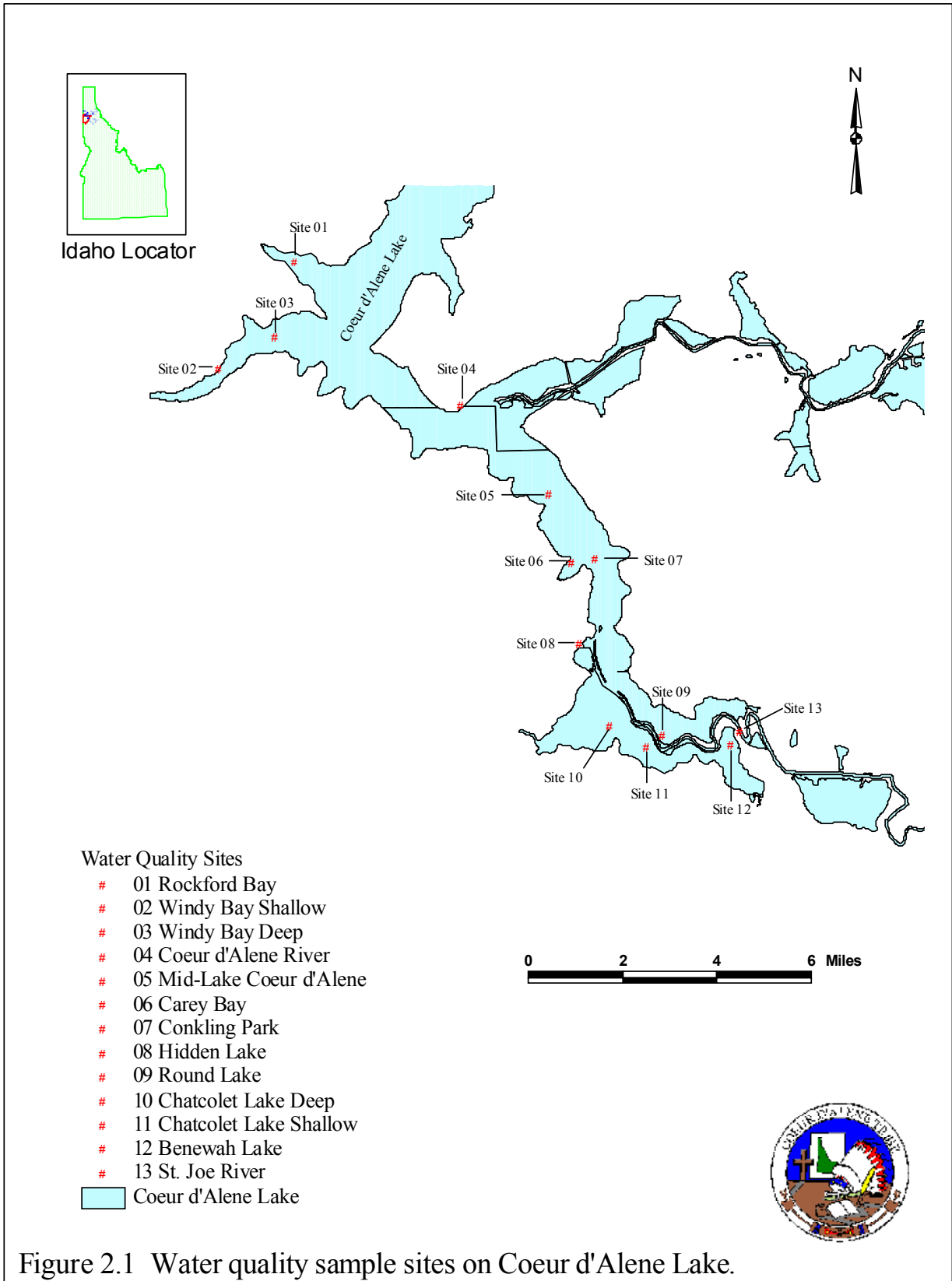


Figure 2.1 Water quality sample sites on Coeur d'Alene Lake.



area is dry during the drawdown period and wetted at full pool. The second habitat area is comprised of the three shallow southern chain lakes of the St. Joe River. These lakes were separated from the Coeur d'Alene Lake system until the completion of Post Falls Dam. The third area consists of three deep, open water sections in the southern part of Coeur d'Alene Lake. These areas are considered pelagic in nature. The fourth habitat area consists of three shallow bays located in the main part of Coeur d'Alene Lake. The fifth area is riverine habitat inundated by waters from Post Falls Dam. Monitoring stations have been established in each of the separate locations (Table 2.1).

Table 2.1 Water quality sample stations grouped by habitat area.

| Habitat Area         | Stations                                       |
|----------------------|--|
| Shallow Water        | Round Lake<br>Chatcolet Lake Shallow           |
| Southern Chain Lakes | Benewah Lake<br>Chatcolet Lake<br>Hidden Lake  |
| Interior Bays        | Carey Bay<br>Windy Bay Shallow<br>Rockford Bay |
| Open Water           | Conkling Point<br>Mid Lake<br>Windy Bay Deep   |
| Inundated Rivers     | Coeur d' Alene River<br>St. Joe River          |

**Parameters**

**Physical/Chemical**

Temperature, dissolved oxygen, pH, and conductivity were monitored at each station using a Hydrolab H20 multi-probe transmitter. Quality control was maintained through strict adherence to the standard operating procedures outlined in the Hydrolab manual (1991). Instrument calibration took place prior to every use. A calibration log was used to record the date and time of calibration, analyst performing calibration, calibration parameters, and other comments. At the end of the sampling run the instrument was checked for drift. All readings were recorded in the calibration log. All standards used for calibration were traceable to NIST or other comparable standards. Reagents used for calibration was accompanied by the following documentation: manufacturer, lot numbers, expiration dates, date opened. A logbook was kept which contains all information related to preparation of reagents and standards. Field measurements were completed by lowering the instrument by cable to the bottom and bringing it back up 1-2 meters at a time pausing to allow the instrument to stabilize then recording the values. This step was repeated until the instrument reached the surface.

A Secchi disk (20 cm diameter) was used to estimate the transparency of water. Transparency measured in this way is the mean of the depth at which the Secchi disk disappears when viewed from the shaded side of the boat and at which it reappears upon raising after it has been lowered beyond visibility. Secchi readings are used to empirically estimate euphotic zone depth; the depth at which 99% of the surface radiation has dissipated. Secchi disk readings only represent a portion of the total euphotic zone depth that is more rigorously defined by a submersible photometer. The euphotic zone can also be defined as the depth at which gross photosynthesis exceeds respiration. The relationship between euphotic zone depth and secchi disk readings varies from lake to lake depending on factors which influencing euphotic zone depth. However, a general equation ( $EZD = 2.2302 + 1.4914(SD) r^2 = .78$ ), which we use, has been

developed by the FRED division of Alaska Department of Fish and Game (1987). Secchi disk readings are taken at all sites during the sampling run.

Water samples submitted for laboratory analysis were collected using a certified water collection device and transferred to the appropriate containers for transportation to the contract laboratory. All samples were handled according to Standard methods for the examination of water and wastewater, 18<sup>th</sup> Ed. (APHA), 1992, procedure 1060: Collection and preservation of samples. Strict chain of custody procedures as outlined in section 1060.B.1: Chain of custody procedures (APHA) was followed. All containers used were specially cleaned and prepared by the contract laboratory.

Total Suspended Solids was analyzed using EPA method 160.2: Gravimetric determination of Total Suspended Solids. TSS is defined as the residue left on a filter paper of 2 $\mu$ m or smaller after a portion of sample has been filtered through (APHA, 1992). A qualified analytical laboratory completed turbidity analysis in accordance with (APHA, 1992) standard method 2130B: Nephelometric determination of turbidity and/or EPA method 180.1. Turbidity is an expression of the optical property that causes light to be scattered and absorbed rather than transmitted in straight lines (APHA, 1992). Turbidity in water is caused by suspended matter including clay, silt, and finely divided organic and inorganic matter. The clarity of a natural body of water is a major determinant of the condition and productivity of that system (APHA, 1992).

All metals samples were handled as described previously for collection of water for laboratory analysis. Metals samples were preserved by acidification to 2% HNO<sub>3</sub> as soon as possible after collection. Metals samples were analyzed using EPA method 200.7/200.8 Inductively Coupled Plasma scan by a qualified contract analytical laboratory. The following trace elements were analyzed: zinc, silica, antimony, barium, beryllium, magnesium, arsenic, sodium, aluminum, calcium, copper, silver, lead, cadmium, cobalt, nickel, manganese, iron, chromium.

#### Nutrients

Nutrient samples were collected in the same manner as turbidity and metals. Nutrient sampling consisted of a euphotic zone composite sample determined by secchi disk and temperature analysis, and a hypolimnetic composite sample with the upper portion of the stratum determined by the temperature profile. Composite sampling was in accordance with APHA method 1060.A.3.B: Composite sample collection. All samples were handled according to Standard methods for the examination of water and wastewater 18<sup>th</sup> Ed. (APHA) 1992 procedure 1060 collection and preservation of samples. The contract laboratory analyzed nutrient samples with an ion chromatograph using EPA method 300.0. The following ions were tested for: ortho-phosphate, nitrate, and nitrite. Other ions looked at were fluoride, chloride, and sulfate.

#### Primary Productivity

Chlorophyll<sub>a</sub> samples were collected in amber colored bottles and placed directly on ice. Samples were collected at the same locations as nutrients. A contract laboratory completed sample analysis. The method used was Standard methods for the examination of water and wastewater 18<sup>th</sup> Ed. (APHA) 1992 procedure 10200 parts 1: pigment extraction and 2: spectrophotometric determination of chlorophyll.

#### Sample Timing

The sampling schedule was designed to capture data related to significant changes in the water quality throughout the year. This included physical/chemical characteristics, nutrient characteristics, and phytoplankton and macrophyte growth. Sampling was initiated just prior to the onset of the growing season in the spring and continued until the lake turned over in the fall, marking the end of the growing season. Limited sampling was completed after fall turnover because weather usually prohibited any

extensive sampling. During this time, however, very little natural change is occurring. A representative sample was taken during the winter and applied to the rest of the winter season.

#### Sampling Schedule

##### Physical/ Chemical sampling

Sampling was scheduled to begin the last week in February, however, foul weather postponed the start until mid-March. Only one sample was taken between November and February when the lake was isothermal. Ice formed on the surface of the lake within the study area thus, inhibiting sampling during that time frame. The following parameters were monitored at all sites on a bi-weekly basis: temperature, pH, dissolved oxygen, and conductivity. Turbidity was monitored at all sites on a monthly basis. Trace heavy metals were monitored at only three sites on a monthly basis. Surface to bottom depth profiles were taken for temperature, pH, dissolved oxygen, and conductivity. Composite samples were taken in the euphotic zone and the hypolimnion for turbidity and trace heavy metals.

##### Nutrient Sampling

Five nutrient samples were taken at all sites on a monthly basis from July to November. The following parameters were monitored: Phosphate, Nitrate, Nitrite, Sulfate, Chloride, and Fluoride. Composite samples were taken in the euphotic zone and the hypolimnion.

##### Phytoplankton (primary productivity)

Chlorophyll<sub>a</sub> samples were taken at the same time and frequency as the nutrients. Composite samples were taken in the euphotic zone and the hypolimnion.

#### **2.1.2 Fisheries**

Fish abundance surveys were conducted on a monthly basis from July through October 1997 on Coeur d'Alene Lake. The section of Coeur d'Alene Lake sampled extended from Windy Bay south and was divided into 12 different transect areas (Figure 2.2) along a longitudinal axis in an attempt to decrease overall sampling bias. Each transect area contained 2-4 sample reaches. The reach locations were determined by visual habitat characteristics and transect size. These reaches were chosen in order to best represent the shoreline habitat within the transect area. The collective submerged habitat within all reaches spans the ranges of conditions within each transect area.

Fish were collected using horizontal gillnets and electrofishing. Three twisted nylon strand horizontal nets: 12X200, 10X200, and 8X200 feet each with 4 panels of 1-4 inches stretched mesh were set in the deep open water or pelagic zones within each transect area. Nets were used only in locations not susceptible to electrofishing sampling techniques, following the methods described by Huber (1983). One net would be set in the transect area and it would be fished for approximately 12 hours after which it would be pulled and the fish would be removed. Once the net was cleared it would be reset in the same place and fished another 12 hours. This was completed one time each month from July to October.

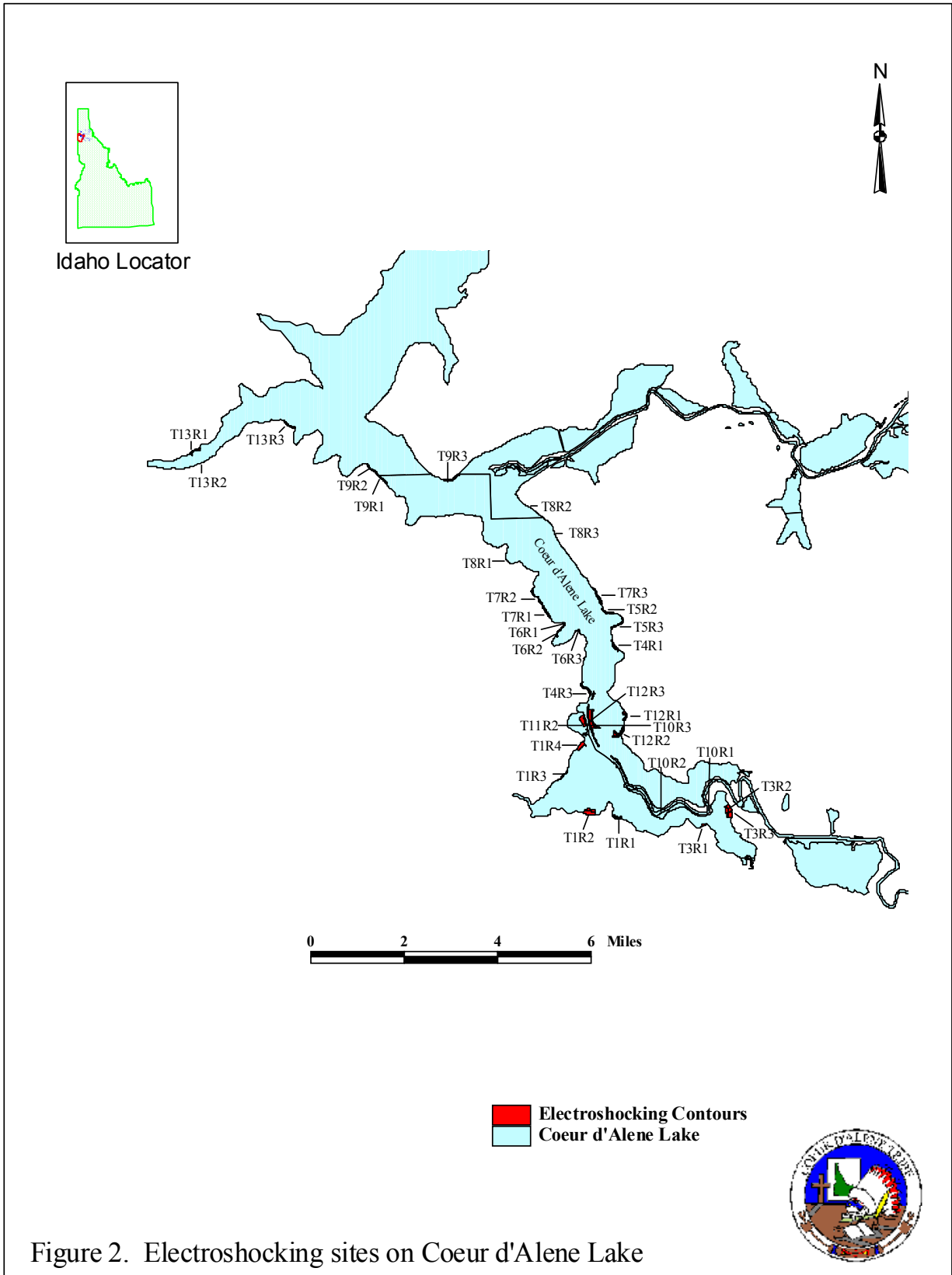


Figure 2. Electroshocking sites on Coeur d'Alene Lake

A custom made aluminum boat equipped with a Smith-Root 7.5 GPP electrofishing unit was used to sample the near-shore or littoral areas within the designated reaches according to methods adapted from Reynolds (1983) and Novoteny and Priegle (1974). The voltage of the electrofishing unit was adjusted to produce a current between 5.0 and 8.0 Amps. A current of 5.0-8.0 seemed to produce the best results in the low conductivity waters (35  $\mu$ mhos- 80  $\mu$ mhos) of Coeur d'Alene Lake. Each reach was sampled two times per month; once during the daylight hours and once during the nighttime hours. Netters were to net all sizes and species of fish with equal effort in an attempt to limit the amount of sampling bias introduced into the final data set. Duration of the sampling effort varied from 5-10 minutes. Work completed by Miranda, *et al.* (1996) demonstrated that variations in the duration of the sampling period did not affect overall catch per unit effort. They showed that in areas where densities are high shorter sampling times gives equal results when compared to longer sampling times.

Relative abundance is an index of population density that assumes that catch per unit effort (CPUE) is proportional to stock density. However, several physical and behavioral characteristics of each species of interest govern both distribution and susceptibility to sampling techniques used. Thus, large fluctuations in CPUE can occur which will hamper interpretation of the CPUE data. In order to use CPUE, special attention must be given to reducing variability by standardizing gear, methods and sampling design. Relative abundance was determined by calculating the percentage of the total catch for each species for a given sample period and sample area. CPUE was calculated by dividing the number of fish captured for each individual species by the total sample time.

The capture location and total length (mm) was recorded for all fish sampled. All game fish were weighed and scaled for age determination. Age determinations were completed using methods described by Jearld (1983). For yellow perch, scales were collected only from the first 20 fish captured in a single sampling run. All adult cutthroat trout, rainbow trout, largemouth bass, northern pike, and channel catfish were tagged with a numbered floy tag and released. Weight (grams) was recorded only for the first 20 fish of each non-game species collected in a single sampling run.

A habitat based model developed by Hickman and Raleigh (1982) was used to evaluate the suitability of lacustrine habitat types for cutthroat trout. The lacustrine model consists of two components: water quality and reproduction. The water quality component takes three variables into consideration, including temperature, dissolved oxygen and pH. Water quality data collected in 1997 were used to calculate the individual suitability index (SI) values using published curves. The reproduction component was not examined in this report.

## **2.2 Stream Studies**

### **2.2.1 Water Quality**

Water quality monitoring was conducted on 13 streams in 1997 (Table 2.2). Each stream was sampled for the same parameters as described above for lake studies, except chlorophyll<sub>a</sub>. Metals analyses were only completed at the two Fighting Creek sample sites. Additional monitoring parameters are described below.

A stage/discharge relationship was developed for each stream based on a linear regression of staff gauge height vs. stream discharge. The rating curve was used to determine the annual water budget for each stream sampled. Staff gauge heights were recorded to the nearest 0.002 of a foot. Discharge measurements were taken at low, medium, and high flows in order to complete the rating curve. Discharge measurements were taken in accordance with standard IFIM methodologies (Bovee and Milhouse *et al.*, 1984). The wetted stream channel was divided into 20 equal cells and water velocity was measured in each cell using a Price model 622 digital flow meter. Discharge for each cell was

calculated by multiplying the cell width by depth and velocity. All individual cell discharges were summed to determine total discharge in cubic feet per second. Channel profiles were measured to evaluate changing flow dynamics over time.

Table 2.2 Stream water quality sites and monitoring parameters.

| Location                   | Discharge | Temperature | DO | pH | Conductivity | Turbidity | TSS | Metals | Nutrients |
|----------------------------|-----------|-------------|----|----|--------------|-----------|-----|--------|-----------|
| Lower Plummer Creek        | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Lower Lake Creek           | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Upper Lake Creek           | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Lower Fighting Creek       | X         | X           | X  | X  | X            | X         | X   | X      | X         |
| Upper Fighting Creek       | X         | X           | X  | X  | X            | X         | X   | X      | X         |
| Evans Creek                | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Benewah Creek <sup>a</sup> | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Windfall Creek             | X         | X           | X  | X  | X            | X         | X   |        | X         |
| School House Creek         | X         | X           | X  | X  | X            | X         | X   |        | X         |
| West Fork Benewah Creek    | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Alder Creek                | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Hangman Creek              | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Little Hangman Creek       | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Indian Creek               | X         | X           | X  | X  | X            | X         | X   |        | X         |

<sup>a</sup>Benewah Creek had three sampling stations, Three Mile, Nine Mile and Upper Benewah.

### 2.2.2 Habitat Suitability Index Model

A modified habitat suitability index (HSI) model was used to evaluate the effect of water quality parameters on cutthroat trout populations within and among the target watersheds. A HSI was calculated for the water quality subcomponent of the model described by Hickman and Raleigh (1982). Model variables included: average maximum water temperature ( $V_1$ ); average minimum dissolved oxygen ( $V_3$ ); annual maximal or minimal pH ( $V_{13}$ ); and average annual base flow as a percentage of the average annual daily flow ( $V_{14}$ ). Individual suitability index (SI) values were calculated for each variable using curves published in Hickman and Raleigh (1982). The following equation was used to calculate the final HSI score:

$$C_{OQ} = (V_1 \times V_3 \times V_{13} \times V_{14})^{1/4}$$

Where;  $C_{OQ}$  = HSI for water quality component, and

$V_n$  = suitability index for water quality parameters.

Water quality data collected in 1997 and in 1998, when available, were used as input variables. The following modifications were made to address site specific conditions: a seven-day running average of maximum temperature was used; and average minimum dissolved oxygen was calculated for the period of greatest average water temperatures. Continuous discharge measurements were only available for the two sample sites on Lake Creek. For the remaining sites, average annual daily flow was calculated based on a minimum of 12 discharge measurements taken during the year, and average annual base flow was calculated for the period of low flow which corresponded to the greatest average water temperatures.

The final HSI was calculated using both a compensatory and a non-compensatory method. The compensatory method assumes that moderately degraded water quality conditions can be partially compensated for by good physical habitat conditions. The non-compensatory method assumes that degraded water quality conditions cannot be compensated for, and variables with suitability indices (SI) < 0.4 become limiting factors on habitat suitability. For purposes of interpretation, HSI with values ranging from 0 - 0.25 were considered very poor; 0.25 – 0.4 were poor; 0.4 – 0.6 were good; and 0.6 – 1.0 were very good.

### 2.2.3 Fisheries

#### Sample Site Selection

The channel types delineated during previous surveys (Lillengreen, 1996) served as the basic geomorphic units for selecting sample sites for conducting fish population surveys. In these early channel type surveys, stream reaches were classified into relatively homogeneous types according to broad geomorphologic characteristics of stream morphology as defined by Rosgen (1994). Sample sites within each reach were selected to include habitat types representative of the reach as a whole (Figures 2.3 – 2.6). The length of each sample unit was defined as twenty times the average stream width, with a minimum sample distance of 100 meters. Longer stream reaches were sampled more intensively than shorter reaches. Sample sites were also selected in each of the tributaries known to have spawning activity, regardless of whether channel type surveys had been completed. In these cases, sample sites were distributed evenly at approximately 1,500 meter intervals.

Sites were sampled in the summer to quantify the abundance and distribution of fishes during base flow conditions (July 30 – August 22, 1996 and June 6 – July 11, 1997). An additional sampling effort in the fall (September 11 – October 28, 1996 and August 11 – September 11, 1997) attempted to capture young of the year fish that had been missed during the summer sampling period and to document fish migration in response to changing water quality conditions.

#### Population Estimates

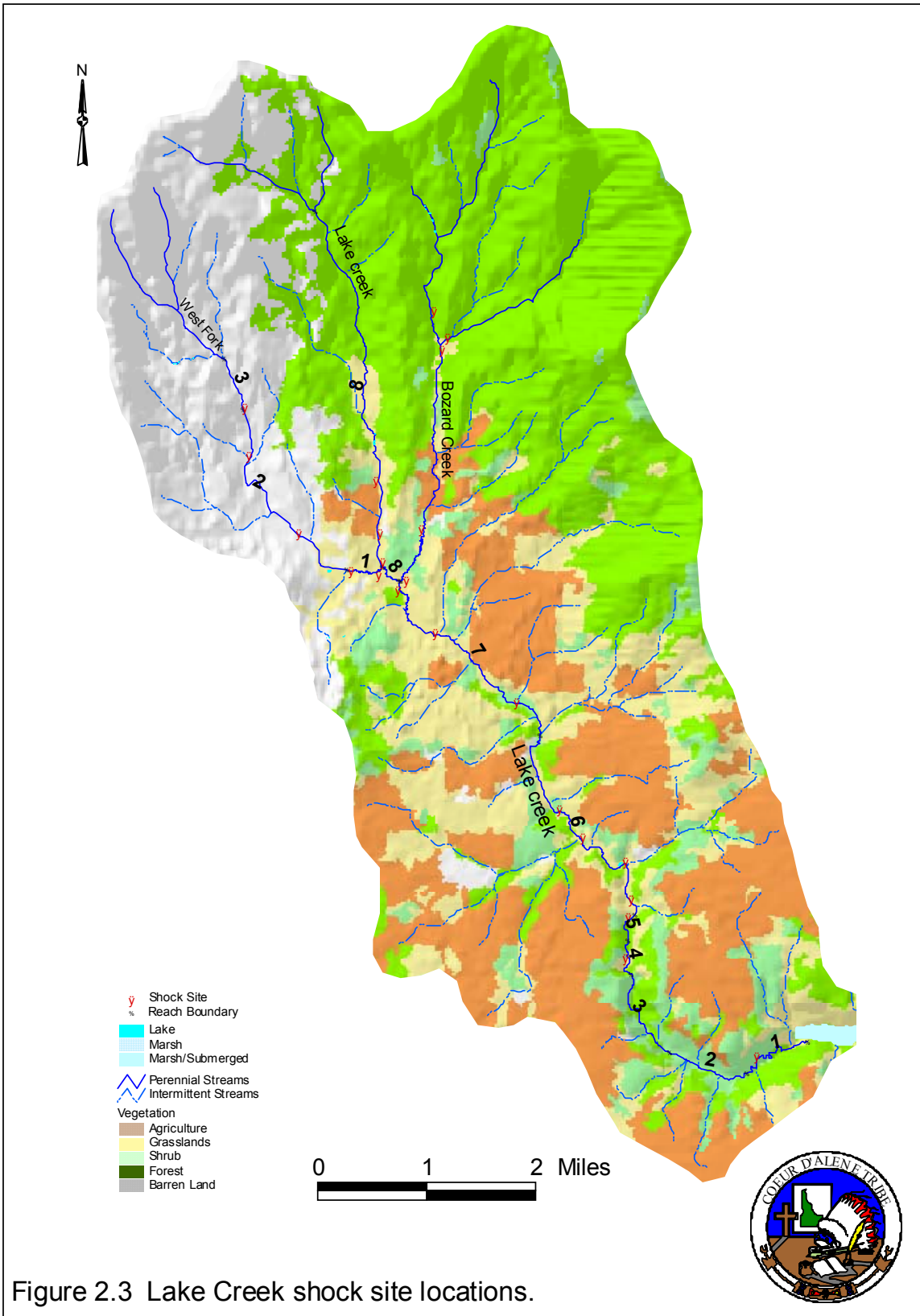
Trout populations were estimated in 1996 and 1997 using the removal-depletion method (Seber and LeCren 1967, Zippen 1958). Blocknets were placed at the upstream and downstream boundaries to prevent immigration and emigration. Each sample site was electrofished using the standard guidelines and procedures described by Reynolds (1983). Fish were collected by spot shocking using a Smith-Root Type VII pulsed-DC backpack electrofisher. Two electrofishing passes were made for each sample site. Salmonid species, including cutthroat trout, brook trout, and bull trout, were the target species for this study. Captured fish were identified, enumerated, measured (TL to nearest mm), and weighed. Cutthroat trout greater than 200 mm in length were tagged with a Floy FD-6B numbered anchor tag. Other species such as longnose dace, redbreast shiner, longnose sucker, and sculpin (sp.) were considered incidental catch and were only counted.

Population estimates were calculated using the following equation (Armour et. al. 1983):

$$N = \frac{U_1}{1 - (U_2 / U_1)}$$

Where:

N = estimated population size;





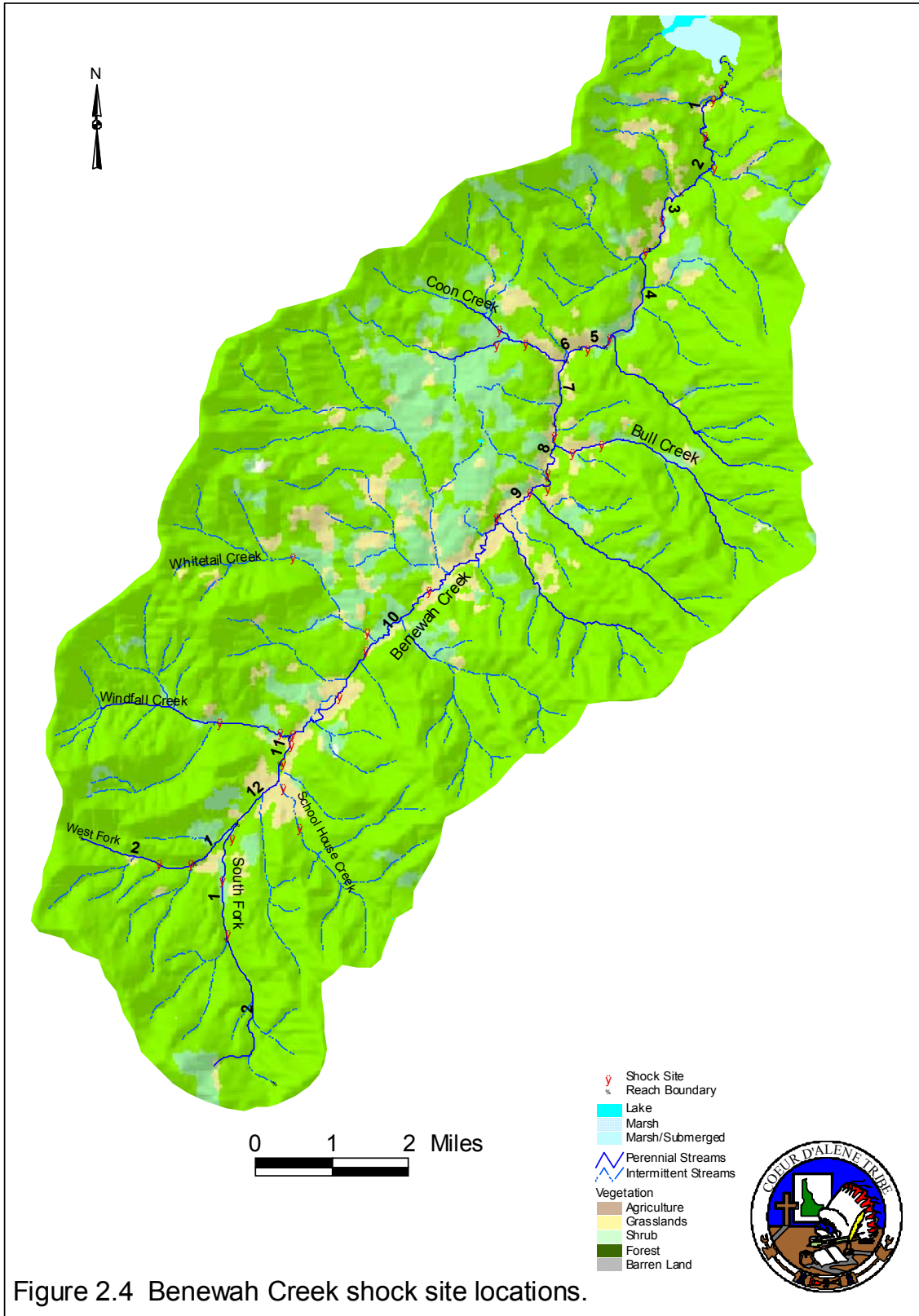
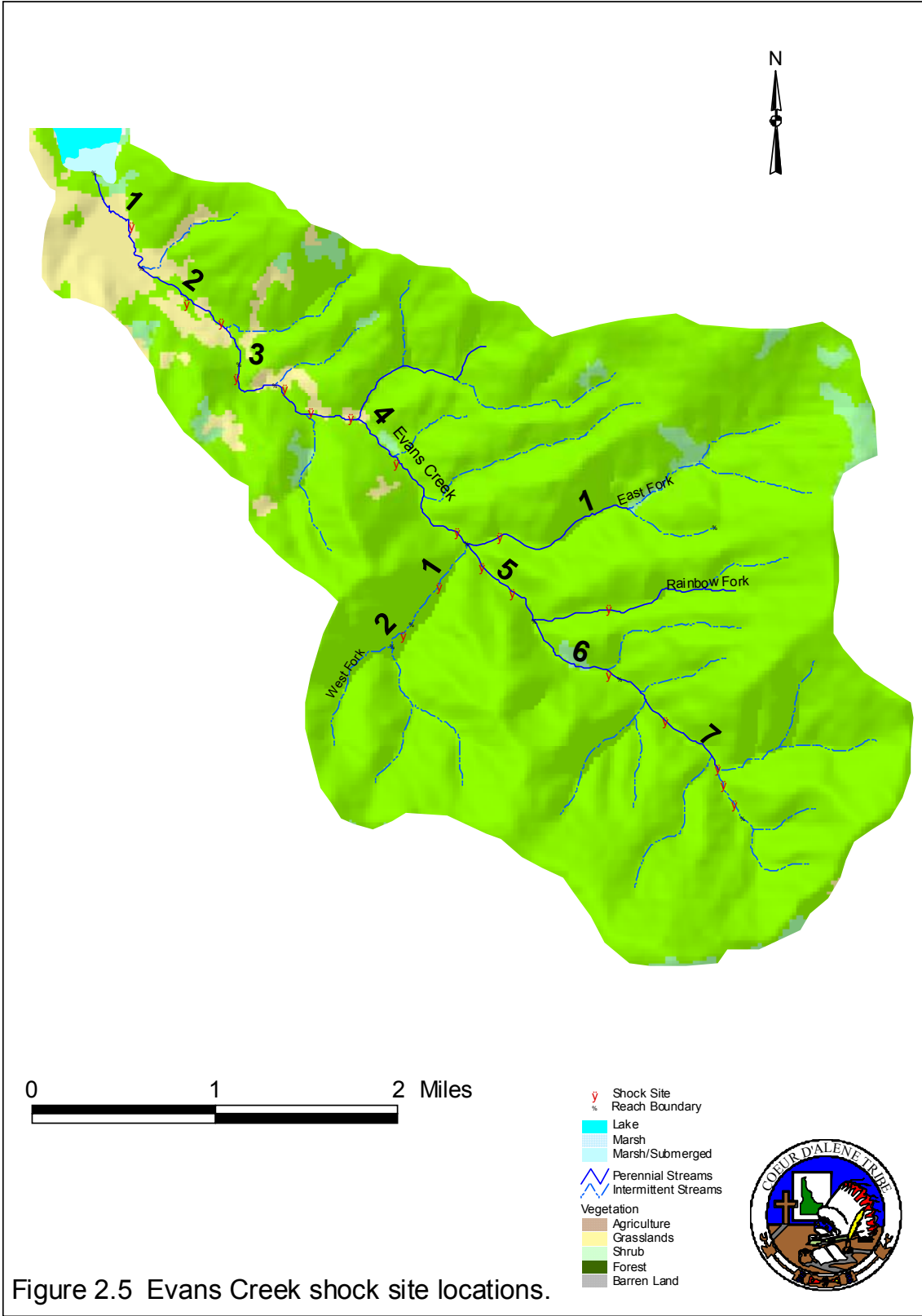
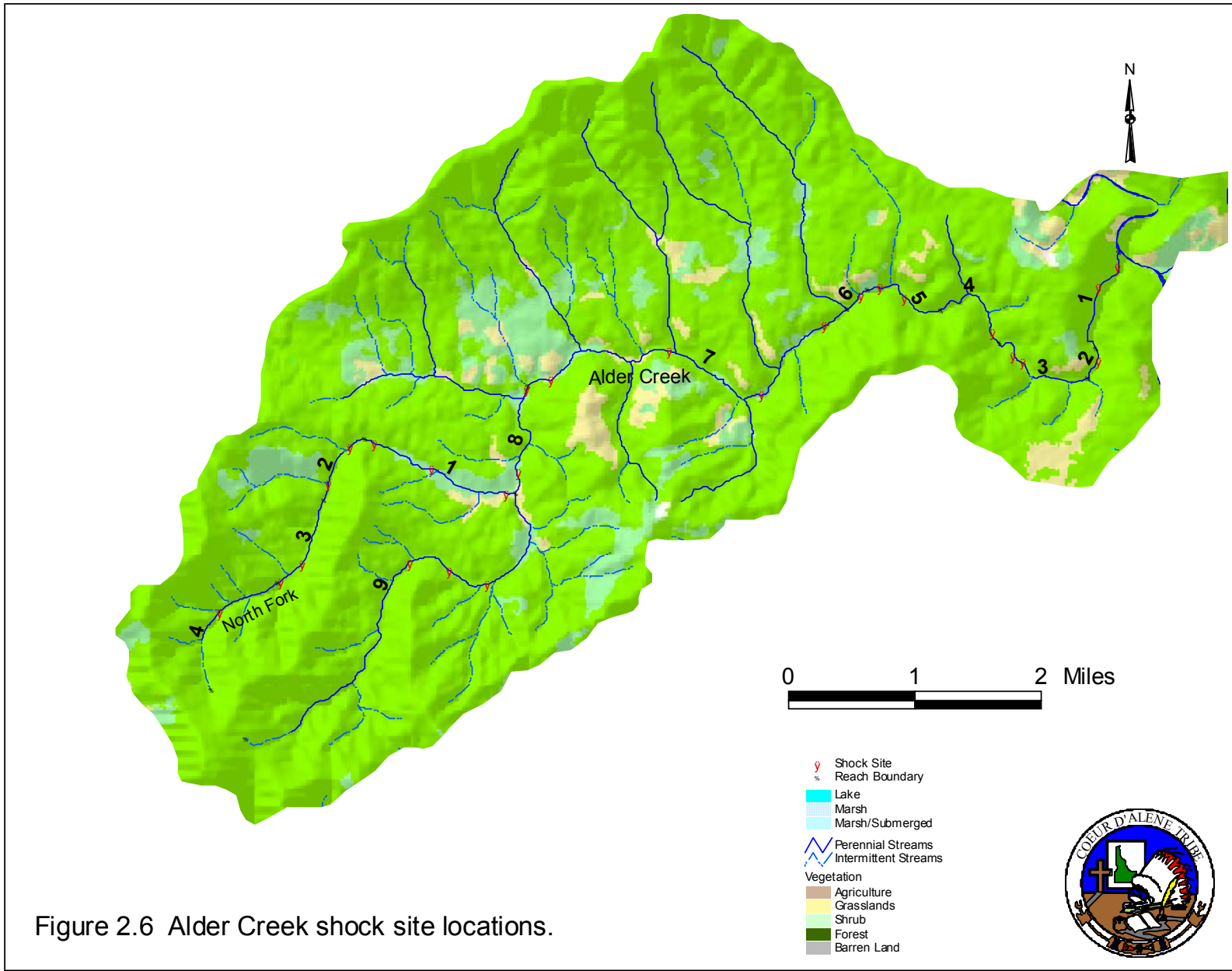


Figure 2.4 Benewah Creek shock site locations.





U<sub>1</sub>= number of fish collected in the first pass; and  
 U<sub>2</sub>= number of fish collected in the second pass.

The standard error of the estimate was calculated as:

$$se(N) = \sqrt{\frac{M(1 - M / N)}{A - [(2p)^2 (U_2 / U_1)]}}$$

where:

se(N) = standard error of the population estimate;

M= U<sub>1</sub> + U<sub>2</sub>;

A= (M/N)<sup>2</sup>; and

p=  $1 - \frac{U_2}{U_1}$ .

The population estimates were converted into density values (# fish/100 meters) for each sample site then extrapolated to the reach in which the samples were collected. The confidence intervals were converted in the same manner.

### Age and Growth

Raw scales were used for age determination and calculating growth rates. Salmonid scales were taken from the side of the body just behind the dorsal fin and above the lateral line (Jearld 1983). Scales samples were sorted by watershed to allow for independent determination of age and growth rate. In the laboratory, several dried scales were mounted between two glass microscope slides and viewed using a Realist, Inc., Vantage 5 microfiche reader. Age was determined by counting the number of annuli (Lux 1971, Jearld 1983). Simultaneous to age determination, a measurement was made from the center of the focus to the furthest edge of the scale. Along this line, measurements were made to each annulus under a constant magnification. Annual growth was then back calculated using the Lee method as described by Carlander (1981). The formula used:

$$L_i = a + \left( \frac{L_c - a}{S_c} \right) S_i$$

where:

L<sub>i</sub> = Length of fish (in mm) at each annulus;

a = intercept of the body scale regression line;

L<sub>c</sub> = length of fish (in mm) at time of capture;

S<sub>c</sub> = distance (in mm) from the focus to the edge of the scale; and

S<sub>i</sub> = scale measurement to each annulus.

The intercept (a) was obtained from the linear regression of body length versus scale length at time of capture. The proportional method of back-calculation was used for species with small sample sizes due to poor regression results. The following equation was used:

$$L_i = \left( \frac{S_i}{S_c} \right) L_c$$

This formula does not take into account the size of fish at scale formation as does the Lee method.

A linear regression of body length versus age was calculated independently for fish from each subject watershed and the resulting equation was used to determine the age of fish for which scale samples were not taken.

### Trout Migration

Migration traps were installed in Lake Creek, Benewah Creek, Evans Creek, and Alder Creek to assess migratory patterns, reproductive cycles, distribution, and relative abundance. Traps were functional from March 24 – June 8, 1996 and March 6 – June 5, 1997, except during periods of high stream flow. The traps consisted of a weir, runway and a holding box. The design was a modification of the juvenile downstream trap found in Conlin and Tuty (1979). Traps were installed in pairs at some locations to allow monitoring of both upstream and downstream movements of fish (Table 2.3). Paired traps were placed approximately 20 meters apart.

Traps were checked at least once daily during peak spawning periods from mid-March through the middle of May and once daily afterwards until late June, when traps were removed. Fishes captured in the traps were identified, counted, measured, and weighed. A scale sample was taken to assess the age, growth, and condition of the fish. Trap efficiency was calculated to allow for comparisons among years and was determined as catch per unit effort (CPUE), where one unit effort was defined as one 24 hour period.

Table 2.3. Fish trap location in the four target watersheds during 1996-1997.

| Lake Creek                          | Benewah Creek         | Alder Creek       | Evans Creek       |
|-------------------------------------|-----------------------|-------------------|-------------------|
| At Hyw 95 crossing (p)              | Mouth Benewah (p)     | At mouth of creek | At mouth of creek |
| Upstream of Bozard Creek confluence | Mouth W.F. Benewah    |                   |                   |
| Mouth Bozard Creek                  | Mouth S.F. Benewah    |                   |                   |
|                                     | Mouth Coon Creek      |                   |                   |
|                                     | Mouth Whitetail Creek |                   |                   |
|                                     | Mouth Windfall Creek  |                   |                   |

(p) Indicates a paired trap location.

## 3.0 Results

### 3.1 Lake Studies

#### 3.1.1 Water Quality

Seven physical/chemical properties of water, two essential nutrients, and 11 dissolved metals were tested for on Coeur d'Alene Lake in 1997. The testing program was directed primarily at the properties of water that most affected fish production and distribution within the lake. The pH of the water in Coeur d'Alene Lake was tested at 13 different sites. Vertical profiles were taken at two-week intervals from April to November. The pH of Coeur d'Alene Lake does not change very much from season to season or on an annual basis. Observed values ranged from 6.8 to 8.0 (Appendix A), all within the optimal tolerance limits for cutthroat production. The only variation occurred at three stations in the southern lakes section (Chatcolet Lake, Benewah Lake, and Hidden Lake) where the pH rose to 9.4 (Figure 3.1) in

the upper 2 meters of the water column. The three stations that showed similar changes in pH were in geomorphologically similar areas.

Vertical profiles of the specific conductance were taken concurrently with the other physical/chemical parameters (Appendix A). All conductivity measurements were within the tolerance limits for cutthroat trout production and ranged from 26  $\mu$ mhos to 82  $\mu$ mhos. Any variability to the conductivity was due to natural environmental conditions except at the mouth of the Coeur d'Alene River (Figure 3.2) where higher than normal levels of dissolved metals are flowing into the lake. Values remained very stable during the fall, winter, and spring months. Only during the peak summer conditions did variation occur.

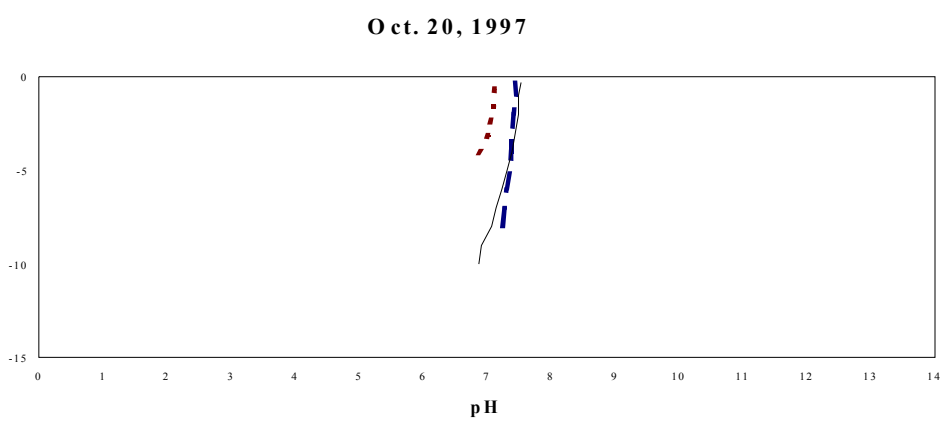
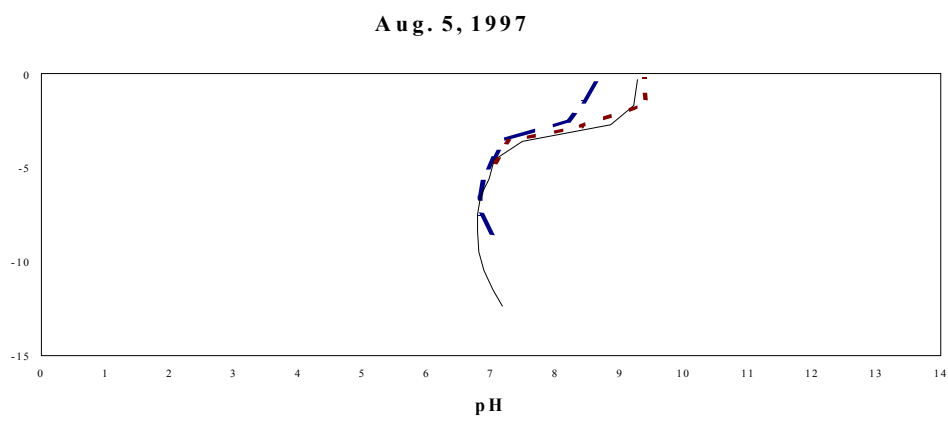
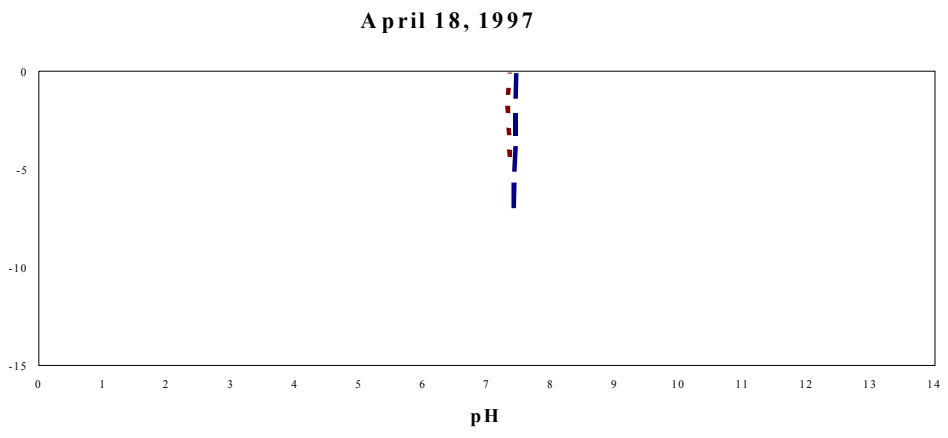
Vertical profiles of dissolved oxygen were taken at each of the thirteen stations (Appendix A). Dissolved oxygen, when at low concentrations is an indication of high levels of organic decomposition. The surface water of every sampling station remained at acceptable levels (>6.0 mg/L) over the course of the entire year. However, four of the thirteen sampling stations showed D.O. concentrations of less than 6.0 mg/L (Table 3.1). Three of the stations, which violated the dissolved oxygen standard, were located in geomorphologically similar areas (Benewah Lake, Chatcolet Lake, and Hidden Lake) the fourth was located in the mid-lake sampling station. The drop in D.O. is most likely related to the decomposition of large quantities of aquatic macrophytes growing in these areas.

Table 3.1 Sample stations on Coeur d' Alene Lake that had a dissolved oxygen reading less than 6.0 mg/L.

| Location       | Depth (m) | Dissolved Oxygen (mg/L) |
|----------------|-----------|-------------------------|
| Mid Lake       | 13        | 4.5                     |
| Hidden Lake    | 7         | 0.25                    |
| Chatcolet Lake | 9         | 0.50                    |
| Benewah Lake   | 4.5       | 1.25                    |

The lowest dissolved oxygen concentrations were located in Hidden Lake where a reading of 0.25 mg/L was recorded in the lowest one meter of the lake. Dissolved oxygen concentrations in Hidden Lake were in violation of the 6.0 mg/L standard in the lowest 2.5 meters of the water column. Chatcolet Lake had the second lowest reading (0.50 mg/L) and was in violation of the 6.0 mg/L standard in the lowest 6 meters of the water column. Benewah Lake was also in violation of the 6.0 mg/L standard in the bottom 1 meter with a low of 1.25 mg/l reading. The only station found in violation of the 6.0 mg/L standard within the main Coeur d'Alene Lake was the mid-lake station where the 6.0 mg/L standard was violated in the bottom 1 meter with a reading of 4.5 mg/L. This drop in DO levels is a general indicator of increasing trophic status or eutrophication. The general trend was an increase in trophic status in a north to south direction. However, Conkling Point sample station which is between mid-lake and Hidden Lake did not violate the 6.0 mg/L dissolved oxygen standard.

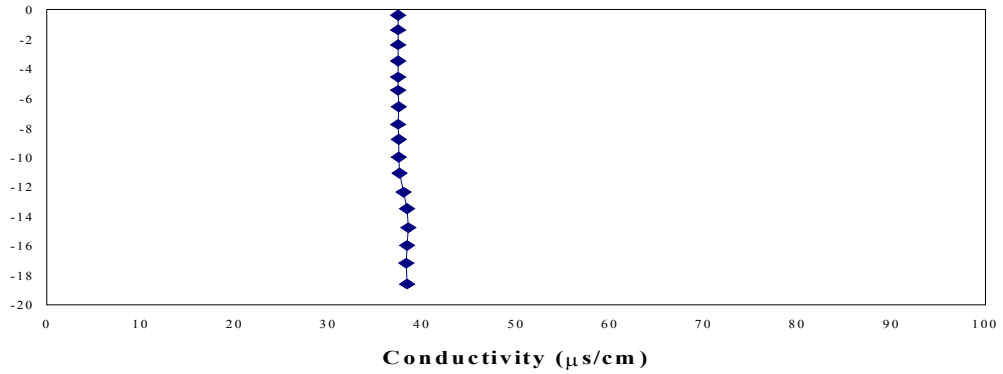
Vertical profiles of temperature were taken at all thirteen stations during the course of the year (Appendix A). Geomorphologically similar stations showed similarities in the temperature profiles in both timing of stratification and magnitude of the warming. Shallow stations heated up sooner than deeper water stations. The shallow southern lakes had more variability in the timing and magnitude of



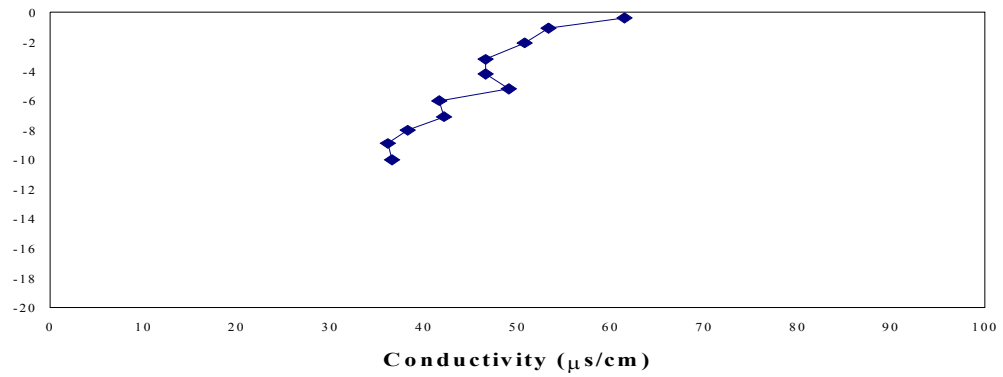
Hidden Lake ----- Chatcolet Lake ————— Benewah Lake - - - - -

Figure 3.1 Peak spring, summer and fall pH profiles for three geomorphologically similar sampling locations on Coeur d'Alene Lake during 1997.

**April 18, 1997**



**August 5, 1997**



**October 21, 1997**

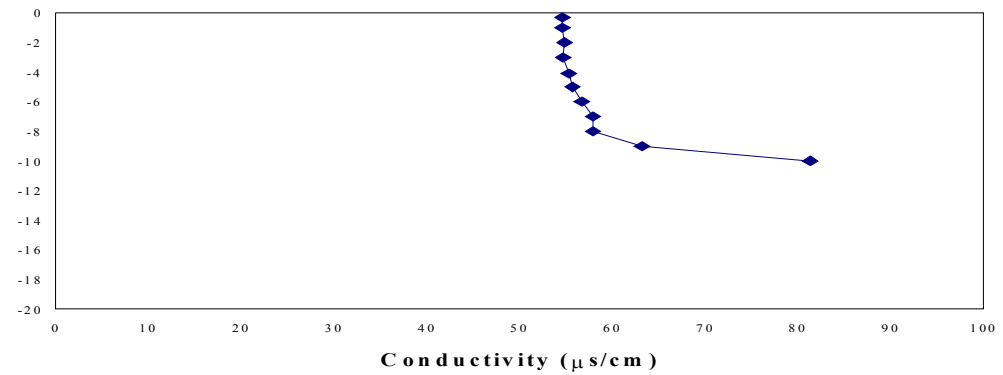


Figure 3.2 Peak spring, summer and fall conductivity profiles for the Coeur d'Alene River during 1997.



the change. All stations were isothermal (5.5°C) on the April 18<sup>th</sup> sample date with formation of a thermocline by May 16<sup>th</sup>.

In the shallowest stations (Round lake and Chatcolet Shallow), 1.5m deep or less, the epilimnion extended to the bottom before May 16<sup>th</sup> (Figure 3.3). In the shallow southern lake stations (Chatcolet Lake, Benewah Lake, and Hidden Lake) the thermocline was present by May 16<sup>th</sup> and reached its deepest point on September 17<sup>th</sup> however, by September 29<sup>th</sup> the stations were isothermal and the lakes had turned over (Figure 3.4). At turnover the isothermal temperature was 15.5° C.

Vertical profiles for the interior bay stations (Windy Bay Shallow, Rockford bay, and Carry Bay) showed that the thermocline had started to build in by May 16<sup>th</sup> and the water nearest to the bottom had warmed from 5.5° C to 7.0° C (Figure 3.5). By June 26<sup>th</sup> the thermocline had reached the bottom. By September 17<sup>th</sup> all of the interior bay sample stations were isothermal at 15.5° C.

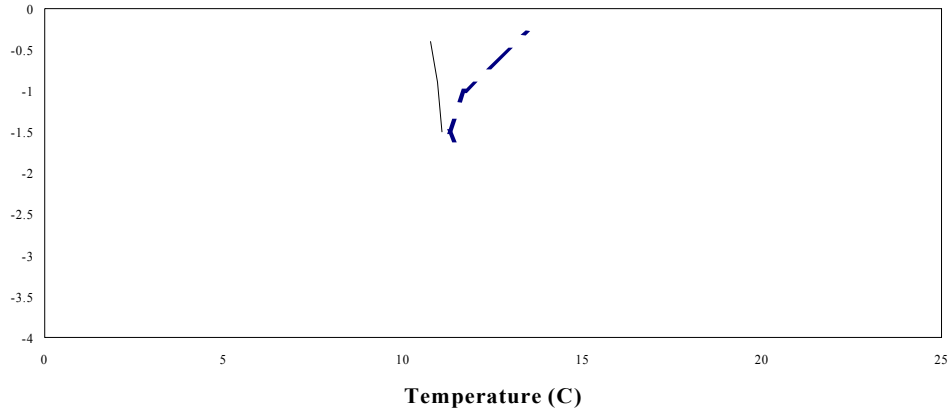
Temperature profiles for the deep-water stations (Conkling Point, Windy Bay Deep, Mid-Lake) showed a definite thermocline by May 16<sup>th</sup> (Figure 3.6) with complete thermal stratification by August 5<sup>th</sup>. The thermocline was still present in the deepest sampling station on October 22<sup>nd</sup>, however, it was nearly broken down in the other two deep water stations. By November 4<sup>th</sup> the deep water stations were isothermal but had only cooled to about 10.5° C while the interior bay sample stations had cooled to 10.0° C and the 3 shallow southern lakes sample stations and 2 shallow stations had cooled to 7.0° C (Appendix A).

Vertical temperature profiles for the riverine sample stations showed similar timing in the formation of the thermocline (Figures 3.7 and 3.8). The St. Joe sample station did show slightly higher temperatures on the average during the spring. The peak temperatures for the Coeur d'Alene River sample station during the other times of year was slightly higher than the peak temperatures for the St. Joe River.

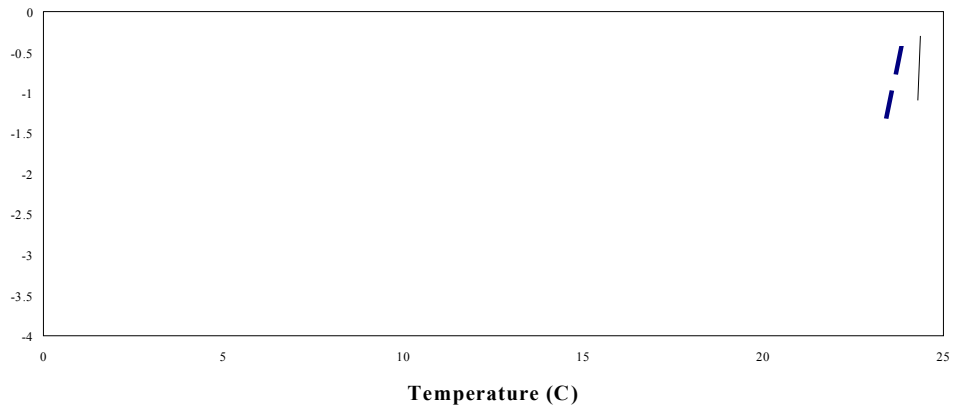
Optimal temperatures for adult salmonid rearing are around 15.0<sup>0</sup> C. In the shallowest sampling stations temperatures greater than the 15.0<sup>0</sup> C temperature standard existed from July 8<sup>th</sup> through September 29<sup>th</sup> from surface to bottom. The temperature peaked at 26.5<sup>0</sup> C on August 8<sup>th</sup>. In the shallow southern lakes sampling stations the 15.00 C temperature criteria was exceeded from August 5<sup>th</sup> to September 29<sup>th</sup> surface to bottom however, between June 11<sup>th</sup> and August 5<sup>th</sup> the temperature criteria was exceeded and habitat was limited in increasingly larger portions of the water column. The temperature peaked at 26.5<sup>0</sup> C on August 8<sup>th</sup>. In the interior bays sample stations the temperature criteria was exceeded from September 17<sup>th</sup> to September 29<sup>th</sup> surface to bottom however, the temperature criteria was exceeded in increasingly larger portions of the water column between May 29<sup>th</sup> and September 17<sup>th</sup>. The temperature peaked on August 5<sup>th</sup> at 23.5<sup>0</sup> C. In the deep water sampling stations the temperature criteria were never exceed from surface to bottom however, the temperature standard was exceed in a portion of the water column and habitat was limited in that area. The criteria were exceeded between June 11<sup>th</sup> and September 29<sup>th</sup> except in the Windy Bay deep station where it was exceeded by May 29<sup>th</sup>. The depth of 15.0<sup>0</sup> C water increased from June 11<sup>th</sup> to August 27<sup>th</sup> then it gradually decreased to September 29<sup>th</sup> reaching a maximum depth of 13 meters on August 27<sup>th</sup>.

Total suspended solids were fairly uniform throughout the epilimnion of the lake (Table 3.2) with only a few differences found. The shallow stations ran slightly higher than the other stations with Round Lake reaching a high of 16.0 mg/l. The southern lakes also ran slightly higher with Hidden Lake and Benewah Lake reaching a high of 10 mg/l. The hypolimnion was quite variable throughout the lake (Table 3.3). The Windy Bay deep reached a high of 27.0 mg/l. Drinking water standards are set a 500 mg/l however, levels much lower can impart a foul taste to the water. No analysis of the composition of the suspended

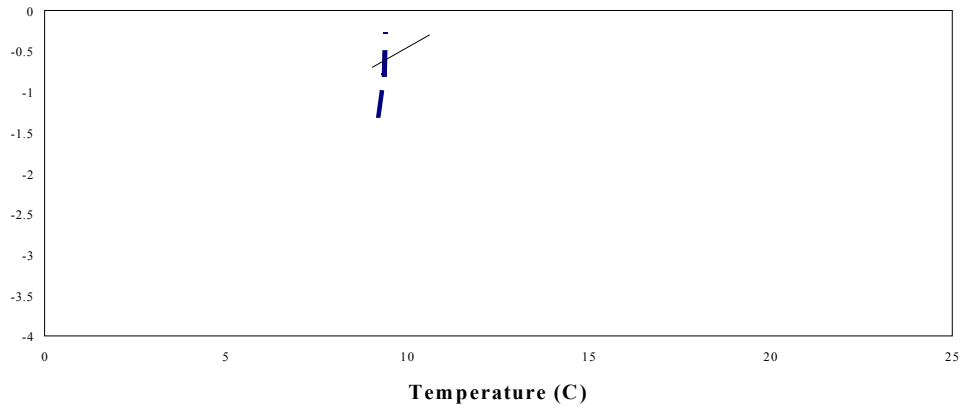
**June 11, 1997**



**Aug. 14, 1997**



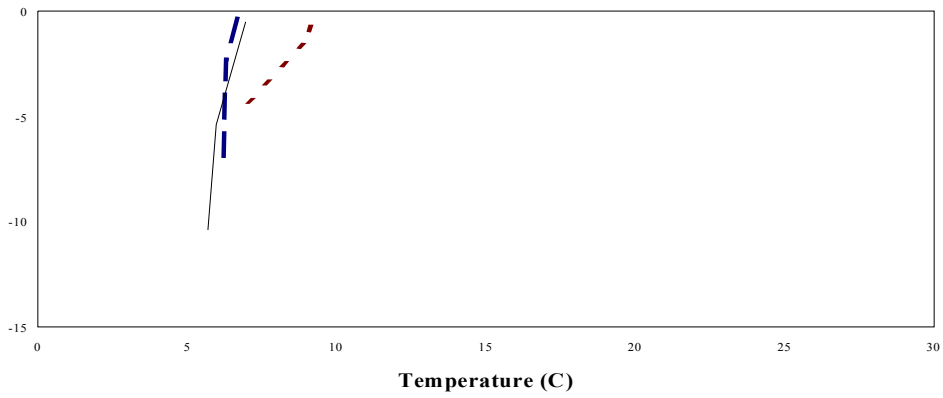
**Oct. 20, 1997**



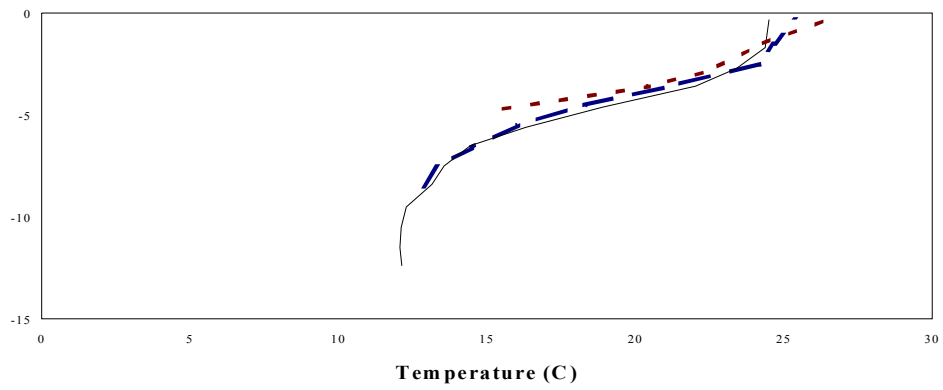
Round Lake - - - - - Chatcolet Shallow \_\_\_\_\_

Figure 3.3 Peak spring, summer and fall temperature profiles vs depth for Two geomorphologically similar sampling locations on Coeur d'Alene lake during 1997.

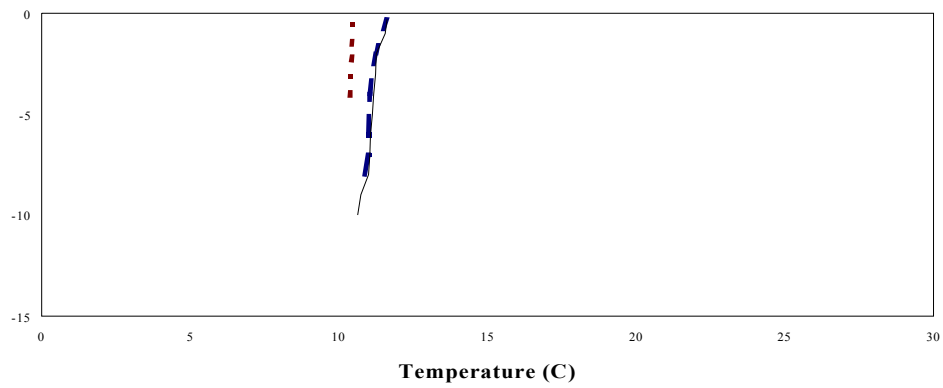
**April 18, 1997**



**Aug. 5, 1997**



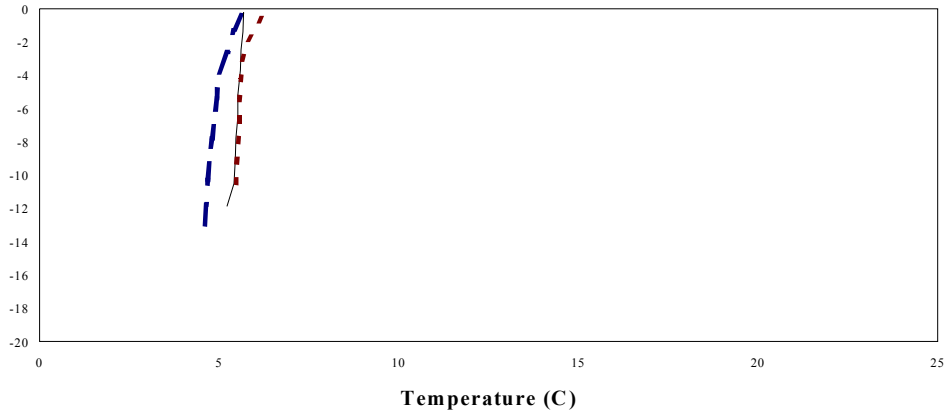
**Oct.20, 1997**



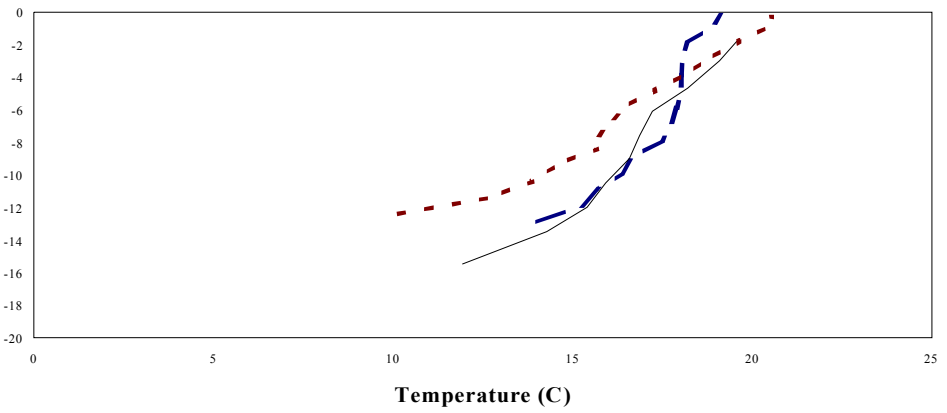
Hidden Lake - - - - - Chatcolet Lake \_\_\_\_\_ Benewah Lake - - - - -

Figure 3.4 Peak spring, summer and fall temperature profiles vs depth for three morphologically similar sampling locations on Coeur d'Alene Lake during 1997.

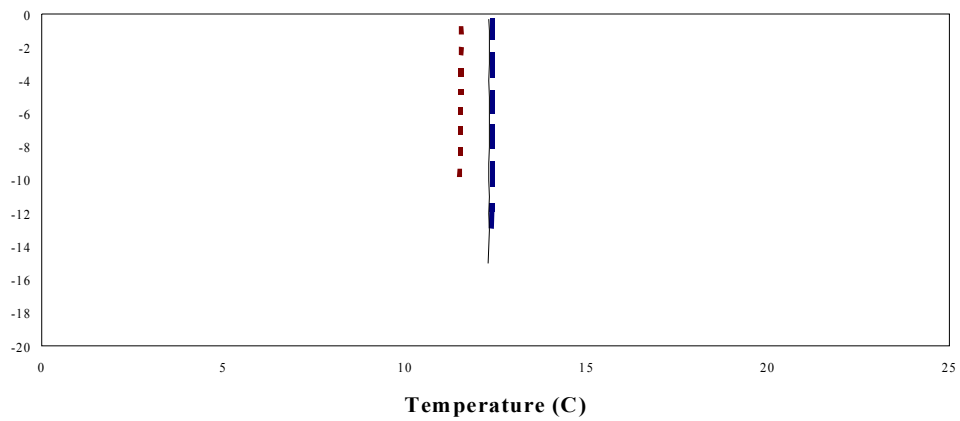
April 18, 1997



July 23, 1997



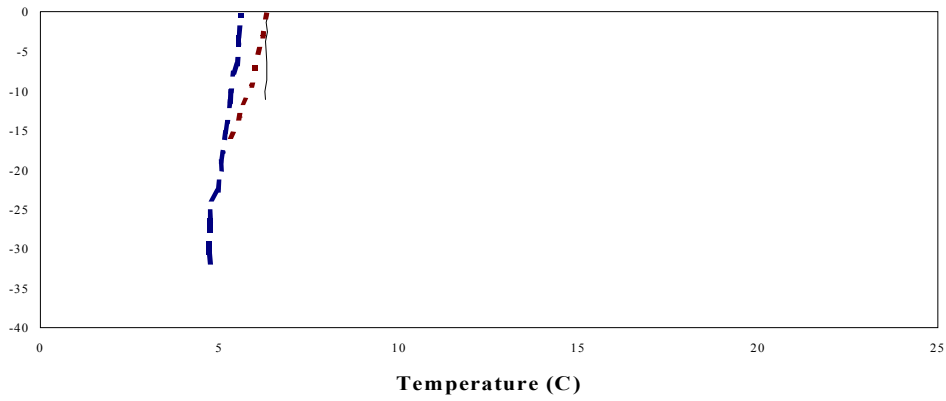
Oct. 21, 1997



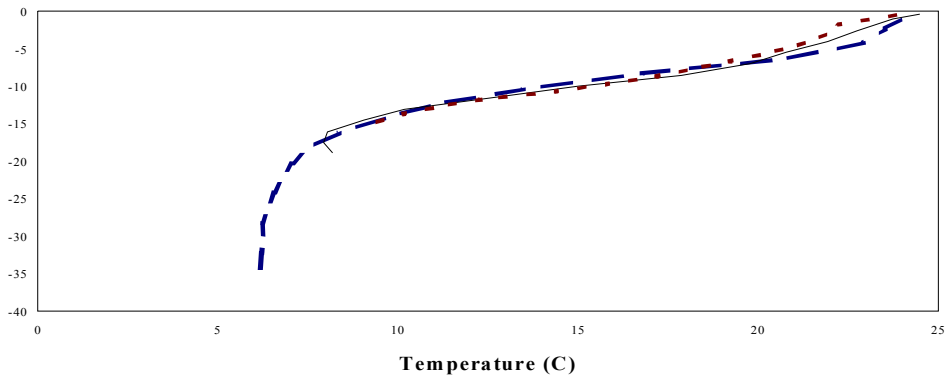
Rockford Bay - - - - - Windy Bay Shallow ——— Carey Bay - - - - -

Figure 3.5 Peak spring, summer and fall temperature profiles vs depth for Three geomorphologically similar sampling locations on Coeur d'Alene Lake during 1997.

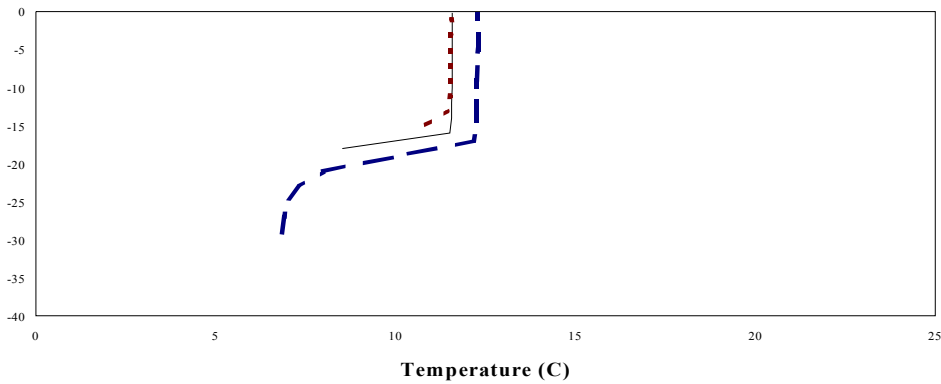
April 18, 1997



Aug. 5, 1997



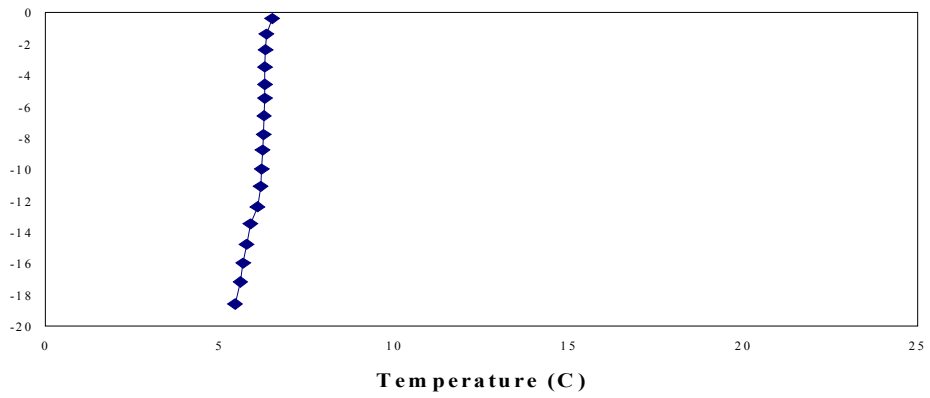
Oct. 22, 1997



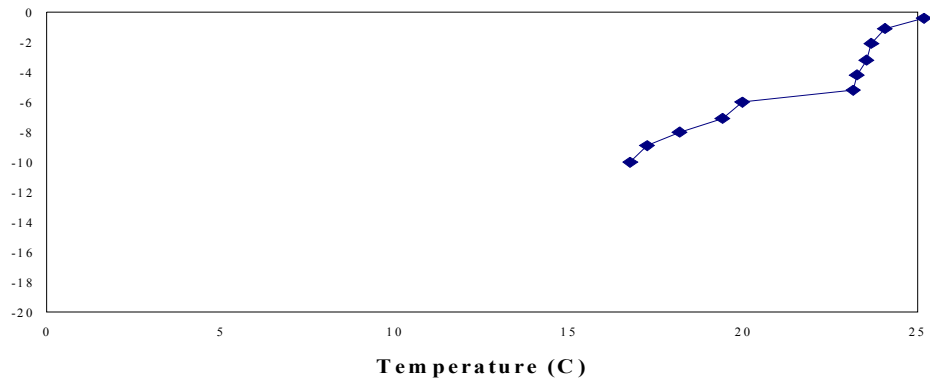
Windy Bay Deep - - - - - Mid Lake - - - - - Conkling Point - - - - -

Figure 3.6 Peak spring, summer and fall temperature profiles vs depth for Three geomorphologically similar sampling locations on Coeur d'Alene Lake during 1997.

**April 18, 1997**



**August 5, 1997**



**October 21, 1997**

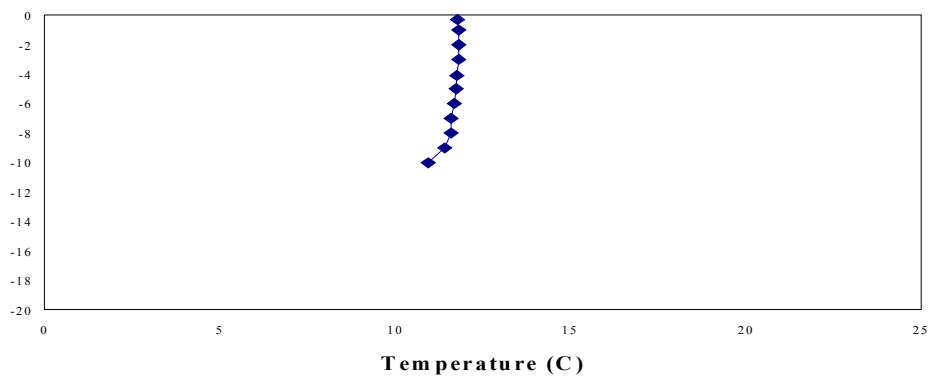
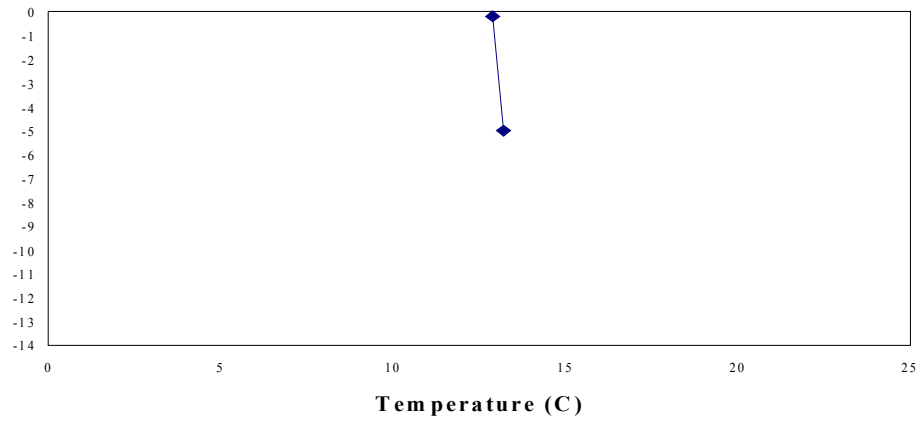
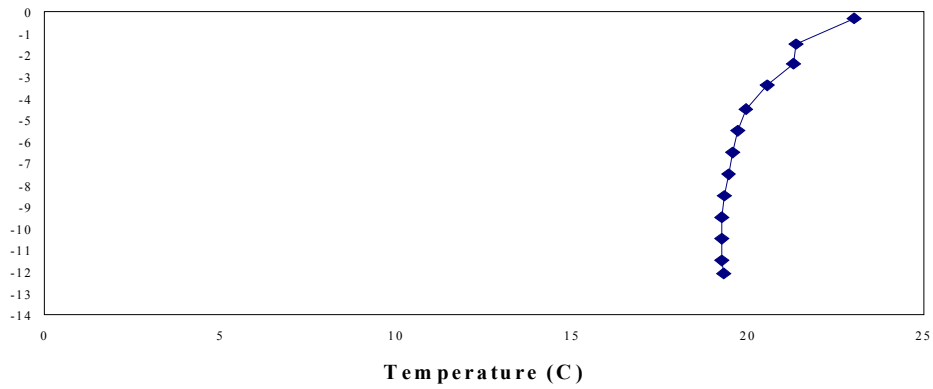


Figure 3.7 Peak spring, summer and fall temperature profiles vs depth for the Coeur d'Alene River during 1997.

**April 18, 1997**



**August 5, 1997**



**October 20, 1997**

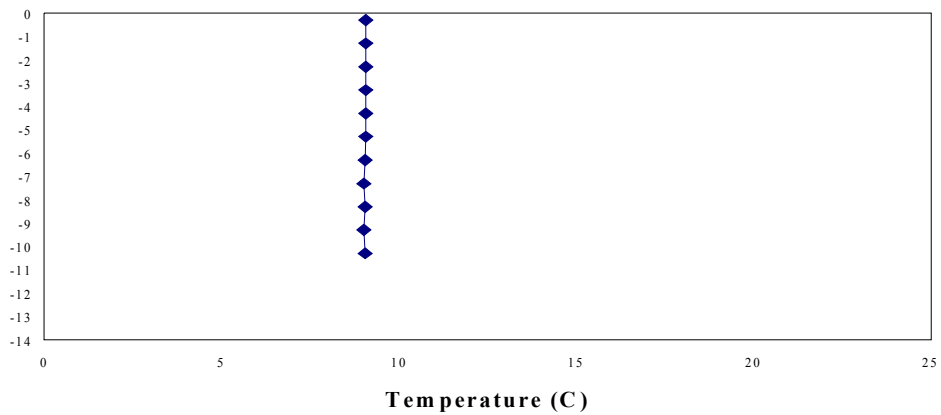


Figure 3.8 St. Joe River peak spring, summer and fall temperature profiles vs depth for 1997.

Table 3.2 Total suspended solids (mg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date    |         |          |         |
|-----------|---------------------------|---------|---------|----------|---------|
|           |                           | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| TSS       | Rockford Bay Upper        | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Windy Bay Shallow Upper   | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Windy Bay Deep Upper      | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Coeur d'Alene River Upper | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Mid Lake Upper            | 2.000   | 2.000   | 0.250    | 1.500   |
|           | Carey Bay Upper           | 2.000   | 2.000   | 0.250    | 4.000   |
|           | Conkling Point Upper      | 2.000   | 2.000   | 0.250    | 6.500   |
|           | Hidden Lake Upper         | 10.000  | 97.000  | 2.500    | 0.250   |
|           | Round Lake                | 2.000   | 2.000   | 1.500    | 16.000  |
|           | Chatcolet Lake Upper      | 2.000   | 2.000   | 2.000    | 0.250   |
|           | Chatcolet Lake Shallow    | 10.000  | 16.000  | 6.500    | 4.000   |
|           | Benewah Lake Upper        | 10.000  | 3.000   | 8.000    | -       |
|           | St. Joe River Upper       | -       | 2.000   | 5.000    | 3.000   |

- No sample taken to lab.

Table 3.3 Total suspended solids (mg/L) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date    |         |          |         |
|-----------|---------------------------|---------|---------|----------|---------|
|           |                           | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| TSS       | Rockford Bay Lower        | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Windy Bay Shallow Lower   | 2.000   | 2.000   | 0.250    | 0.250   |
|           | Windy Bay Deep Lower      | 2.000   | 2.000   | 0.250    | 29.000  |
|           | Coeur d'Alene River Lower | 3.000   | 2.000   | 0.250    | 1.500   |
|           | Mid Lake Lower            | 2.000   | 2.000   | 1.000    | 3.000   |
|           | Carey Bay Lower           | 2.000   | 2.000   | 0.250    | 4.000   |
|           | Conkling Point Lower      | 2.000   | 2.000   | 0.250    | 7.500   |
|           | Hidden Lake Lower         | 10.000  | 2.000   | 3.000    | 0.250   |
|           | Chatcolet Lake Lower      | 20.000  | 5.000   | 12.500   | 0.250   |
|           | St. Joe River Lower       | -       | 2.000   | 4.500    | 0.250   |

- No sample taken to lab.



solids was completed. However, it was noted that high levels of suspended sediment were present in the southern lakes sample station and that most likely the high suspended solids present in the Windy Bay deep station was related to decomposing algae not sediment.

Turbidity in the epilimnion and hypolimnion showed the same trend as suspended solids (Tables 3.4 and 3.5) with a general increase on a North to South axis. The highest turbidity reading was recorded at the Chatcolet shallow sample station (11.2 NTU). The next highest was Benewah Lake (5.810 NTU) followed by Hidden Lake (4.760 NTU). The lowest value was recorded at the Windy Bay Deep sample station (0.230 NTU). The high turbidity readings were due to suspended sediments flowing in from Plummer Creek a tributary to Chatcolet Lake.

The following nutrients were tested for on two dates, one in October and one in November, in Coeur d'Alene Lake: ortho-phosphate, nitrate, and nitrite. Ortho-phosphate concentrations were below the detection limit (0.026 mg/l) at all sites in both the epilimnion and hypolimnion (Tables 3.6 and 3.7). The data were reported as half of the detection limit so data point loss would not occur. For the most part nitrate concentrations sampled at the 13 stations (Table 3.8) were below the detection limit (0.005 mg/l) thus, all data below detection limits were reported as half detection limit. In only a few instances (both stations in Windy Bay, Rockford Bay, and the two shallow stations) did nitrate concentrations exceed the detection limit in the epilimnion, reaching a high of 0.113 mg/l at the Rockford Bay sample station in the October sample period. However, in the hypolimnion concentrations of nitrate generally increased in the November sample (Table 3.9) over the October sample period reaching a high of 0.131 mg/l in the Windy Bay Deep sample station. Nitrite, the form of nitrogen found in the smallest quantities were all below the detection limit in the epilimnion and hypolimnion for both samples (Tables 3.10 and 3.11) at all thirteen stations. All data measured below the detection limit were recorded as half the detection limit.

Secchi disk readings were taken at each of the thirteen stations throughout the year and they were used to determine the euphotic zone depth. Graphs showing the measured secchi disk readings and the empirically derived euphotic zone depth are located in Appendix B. The empirically derived euphotic zone depth was variable throughout the lake (Table 3.12) with each station having different depths based on variable site-specific conditions. Similarities existed between the five distinct habitat areas, however, the general trend was decreasing secchi and euphotic zone depths in a north to south direction within the lake.

Chlorophyll<sub>a</sub> values ranged from 0.005 µg/l to 25.790 µg/l (Table 3.13) in the epilimnion while values ranged from 0.005 µg/l to 34.14 µg/l (Table 3.14) in the hypolimnion. The general trend is increasing values in a north to south direction. Higher levels of chlorophyll<sub>a</sub> is an indicator of increasing trophic status.

Table 3.4 Turbidity (NTU) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date    |         |          |         |
|-----------|---------------------------|---------|---------|----------|---------|
|           |                           | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| Turbidity | Rockford Bay Upper        | 0.350   | 0.300   | 0.270    | 0.240   |
|           | Windy Bay Shallow Upper   | 0.630   | 0.330   | 0.270    | 0.250   |
|           | Windy Bay Deep Upper      | 0.280   | 0.400   | 0.230    | 0.270   |
|           | Coeur d'Alene River Upper | 0.870   | 0.510   | 0.510    | 0.430   |
|           | Mid Lake Upper            | 0.330   | 1.990   | 0.840    | 0.380   |
|           | Carey Bay Upper           | 0.560   | 0.840   | 0.880    | 0.490   |
|           | Conkling Point Upper      | 0.540   | 0.760   | 0.760    | 0.860   |
|           | Hidden Lake Upper         | 4.760   | 1.560   | 0.960    | 1.420   |
|           | Round Lake                | 3.820   | 0.450   | 1.880    | 0.930   |
|           | Chatcolet Lake Upper      | 1.010   | 1.390   | 1.080    | 1.610   |
|           | Chatcolet Lake Shallow    | 3.690   | 0.770   | 2.010    | 11.200  |
|           | Benewah Lake Upper        | 5.810   | 1.550   | 1.600    | -       |
|           | St. Joe River Upper       | -       | 0.620   | 1.87     | 7.860   |

- No sample taken to lab.

Table 3.5 Turbidity (NTU) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date    |         |          |         |
|-----------|---------------------------|---------|---------|----------|---------|
|           |                           | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| Turbidity | Rockford Bay Lower        | 0.600   | 0.370   | 0.240    | 0.270   |
|           | Windy Bay Shallow Lower   | 0.920   | 0.420   | 0.280    | 0.850   |
|           | Windy Bay Deep Lower      | 0.860   | 0.870   | 0.670    | 0.740   |
|           | Coeur d'Alene River Lower | 1.020   | 0.510   | 0.850    | 0.880   |
|           | Mid Lake Lower            | 1.300   | 0.820   | 1.130    | 0.760   |
|           | Carey Bay Lower           | 0.560   | 0.840   | 0.880    | 0.490   |
|           | Conkling Point Lower      | 1.870   | 2.000   | 0.920    | 1.010   |
|           | Hidden Lake Lower         | 1.310   | 3.60    | 1.060    | 1.390   |
|           | Chatcolet Lake Lower      | 3.520   | 1.900   | 2.630    | 3.680   |
|           | St. Joe River Lower       | -       | 0.760   | 1.830    | 7.680   |

- No sample taken to lab.

Table 3.6 Ortho-Phosphate (mg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients       | Location                  | Date     |         |
|-----------------|---------------------------|----------|---------|
|                 |                           | 10/20/97 | 11/4/97 |
| Ortho-Phosphate | Rockford Bay Upper        | 0.013    | 0.013   |
|                 | Windy Bay Shallow Upper   | 0.013    | 0.013   |
|                 | Windy Bay Deep Upper      | 0.013    | 0.013   |
|                 | Coeur d'Alene River Upper | 0.013    | 0.013   |
|                 | Mid Lake Upper            | 0.013    | 0.013   |
|                 | Carey Bay Upper           | 0.013    | 0.013   |
|                 | Conkling Point Upper      | 0.013    | 0.013   |
|                 | Hidden Lake Upper         | 0.013    | 0.013   |
|                 | Round Lake                | 0.013    | 0.013   |
|                 | Chatcolet Lake Upper      | 0.013    | 0.013   |
|                 | Chatcolet Lake Shallow    | 0.013    | 0.013   |
|                 | Benewah Lake Upper        | 0.013    | -       |
|                 | St. Joe River Upper       | 0.013    | 0.013   |

- No sample taken to lab.

Table 3.7 Phosphate (mg/L) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date     |         |
|-----------|---------------------------|----------|---------|
|           |                           | 10/20/97 | 11/4/97 |
| Phosphate | Rockford Bay Lower        | 0.013    | 0.013   |
|           | Windy Bay Shallow Lower   | 0.0013   | 0.013   |
|           | Windy Bay Deep Lower      | 0.013    | 0.013   |
|           | Coeur d'Alene River Lower | 0.013    | 0.013   |
|           | Mid Lake Lower            | 0.013    | 0.013   |
|           | Carey Bay Lower           | 0.013    | 0.013   |
|           | Conkling Point Lower      | 0.013    | 0.013   |
|           | Hidden Lake Lower         | 0.013    | 0.013   |
|           | Chatcolet Lake Lower      | 0.013    | 0.013   |
|           | St. Joe River Lower       | 0.013    | 0.013   |

Table 3.8 Nitrate (mg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date     |         |
|-----------|---------------------------|----------|---------|
|           |                           | 10/20/97 | 11/4/97 |
| Nitrate   | Rockford Bay Upper        | 0.113    | 0.053   |
|           | Windy Bay Shallow Upper   | 0.003    | 0.059   |
|           | Windy Bay Deep Upper      | 0.003    | 0.053   |
|           | Coeur d'Alene River Upper | 0.003    | 0.003   |
|           | Mid Lake Upper            | 0.003    | 0.003   |
|           | Carey Bay Upper           | 0.003    | 0.003   |
|           | Conkling Point Upper      | 0.003    | 0.003   |
|           | Hidden Lake Upper         | 0.003    | 0.003   |
|           | Round Lake                | 0.003    | 0.067   |
|           | Chatcolet Lake Upper      | 0.003    | 0.003   |
|           | Chatcolet Lake Shallow    | 0.003    | 0.043   |
|           | Benewah Lake Upper        | 0.003    | -       |
|           | St. Joe River Upper       | 0.003    | 0.066   |

- No sample taken to lab.

Table 3.9 Nitrate (mg/L) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date     |         |
|-----------|---------------------------|----------|---------|
|           |                           | 10/20/97 | 11/4/97 |
| Nitrate   | Rockford Bay Lower        | 0.003    | 0.050   |
|           | Windy Bay Shallow Lower   | 0.003    | 0.063   |
|           | Windy Bay Deep Lower      | 0.131    | 0.075   |
|           | Coeur d'Alene River Lower | 0.003    | 0.101   |
|           | Mid Lake Lower            | 0.054    | 0.128   |
|           | Carey Bay Lower           | 0.003    | 0.003   |
|           | Conkling Point Lower      | 0.003    | 0.125   |
|           | Hidden Lake Lower         | 0.003    | 0.003   |
|           | Chatcolet Lake Lower      | 0.003    | 0.003   |
|           | St. Joe River Lower       | 0.003    | 0.003   |

Table 3.10 Nitrite results (mg/L) from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date     |         |
|-----------|---------------------------|----------|---------|
|           |                           | 10/20/97 | 11/4/97 |
| Nitrite   | Rockford Bay Upper        | 0.013    | 0.013   |
|           | Windy Bay Shallow Upper   | 0.013    | 0.013   |
|           | Windy Bay Deep Upper      | 0.013    | 0.013   |
|           | Coeur d'Alene River Upper | 0.013    | 0.013   |
|           | Mid Lake Upper            | 0.013    | 0.013   |
|           | Carey Bay Upper           | 0.013    | 0.013   |
|           | Conkling Point Upper      | 0.013    | 0.013   |
|           | Hidden Lake Upper         | 0.013    | 0.013   |
|           | Round Lake                | 0.013    | 0.013   |
|           | Chatcolet Lake Upper      | 0.013    | 0.013   |
|           | Chatcolet Lake Shallow    | 0.013    | 0.013   |
|           | Benewah Lake Upper        | 0.013    | -       |
|           | St. Joe River Upper       | 0.013    | 0.013   |

- No sample taken to lab.

Table 3.11 Nitrite (mg/L) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients | Location                  | Date     |         |
|-----------|---------------------------|----------|---------|
|           |                           | 10/20/97 | 11/4/97 |
| Nitrite   | Rockford Bay Lower        | 0.013    | 0.013   |
|           | Windy Bay Shallow Lower   | 0.013    | 0.013   |
|           | Windy Bay Deep Lower      | 0.013    | 0.013   |
|           | Coeur d'Alene River Lower | 0.013    | 0.013   |
|           | Mid Lake Lower            | 0.013    | 0.013   |
|           | Carey Bay Lower           | 0.013    | 0.013   |
|           | Conkling Point Lower      | 0.013    | 0.013   |
|           | Hidden Lake Lower         | 0.013    | 0.013   |
|           | Chatcolet Lake Lower      | 0.013    | 0.013   |
|           | St. Joe River Lower       | 0.013    | 0.013   |

Table 3.12 Average annual and seasonal secchi measurements taken at thirteen stations on Coeur d' Alene Lake in 1997. Euphotic zone depths were empirically derived using the regression equation  $EZD=2.2302+1.4914(SD)R^2=.78$  published by Alaska Fish and Game 1987. All measurements are in meters.

| Location             | Station Depth | Average                    |                           | Spring                       |                        | Summer                       |                        | Fall                         |                        |
|----------------------|---------------|----------------------------|---------------------------|------------------------------|------------------------|------------------------------|------------------------|------------------------------|------------------------|
|                      |               | Annual Secchi <sup>a</sup> | Annual Euphotic Zone      | Seasonal Secchi <sup>b</sup> | Seasonal Euphotic Zone | Seasonal Secchi <sup>c</sup> | Seasonal Euphotic Zone | Seasonal Secchi <sup>d</sup> | Seasonal Euphotic Zone |
| Round Lake           | 1.50          | 1.27                       | Bottom (1.5) <sup>e</sup> | 1.20                         | Bottom (1.5)           | 1.31                         | Bottom (1.5)           | 1.23                         | Bottom (1.5)           |
| Chatcolet Shallow    | 1.50          | .89                        | Bottom (1.5)              | 1.00                         | Bottom (1.5)           | 0.87                         | Bottom (1.5)           | 0.77                         | Bottom (1.5)           |
| Rockford Bay         | 13.00         | 5.29                       | 10.11                     | 153                          | 4.50                   | 6.43                         | 11.81                  | 7.63                         | 13.61                  |
| Windy Bay Shallow    | 14.00         | 4.58                       | 9.05                      | 1.66                         | 4.71                   | 5.37                         | 10.24                  | 6.60                         | 12.07                  |
| Carey Bay            | 12.00         | 4.10                       | 8.34                      | 1.73                         | 4.80                   | 5.13                         | 9.87                   | 4.87                         | 9.48                   |
| Windy Bay Deep       | 33.00         | 5.44                       | 10.35                     | 1.75                         | 4.84                   | 6.77                         | 12.33                  | 7.27                         | 13.07                  |
| Mid Lake             | 18.00         | 4.08                       | 8.31                      | 1.74                         | 4.82                   | 5.00                         | 9.68                   | 5.03                         | 9.73                   |
| Conkling Point       | 16.00         | 3.44                       | 7.36                      | 1.43                         | 4.36                   | 4.16                         | 8.43                   | 4.47                         | 8.89                   |
| Hidden Lake          | 8.00          | 3.02                       | 6.81                      | 1.58                         | 4.65                   | 3.94                         | 8.18                   | 2.80                         | 6.48                   |
| Chatcolet Lake       | 11.50         | 2.56                       | 6.04                      | 1.65                         | 4.69                   | 3.01                         | 6.73                   | 2.70                         | 6.26                   |
| Benewah Lake         | 4.50          | 2.42                       | 5.84                      | 1.60                         | 4.62                   | 2.96                         | 6.64                   | 2.20                         | 5.51                   |
| Coeur d' Alene River | 10.00         | 3.92                       | 8.08                      | 1.50                         | 4.47                   | 4.73                         | 9.28                   | 5.27                         | 10.08                  |
| St. Joe River        | 12.50         | 2.25                       | 5.58                      | 1.08                         | 3.83                   | 3.18                         | 6.98                   | 1.93                         | 5.11                   |

<sup>a</sup> Average annual secchi was calculated from April 18, 1997 to November 4, 1997.

<sup>b</sup> Annual Spring secchi is from April 18, 1997 to June 11, 1997.

<sup>c</sup> Annual Summer secchi is from June 26, 1997 to September 17, 1997.

<sup>d</sup> Annual Fall secchi is from September 29, 1997 to November 4, 1997.

<sup>e</sup> Numbers in parenthesis represent the bottom in meters.

Table 3.13 Chlorophyll<sub>a</sub> (µg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients                | Location                  | Date    |              |          |                      |
|--------------------------|---------------------------|---------|--------------|----------|----------------------|
|                          |                           | 8/13/97 | 9/16/97      | 10/20/97 | 11/4/97 <sup>a</sup> |
| Chlorophyll <sub>a</sub> | Rockford Bay Upper        | 0.005   | 0.700        | 0.005    | 3.300                |
|                          | Windy Bay Shallow Upper   | 0.005   | 0.005        | 0.740    | 2.140                |
|                          | Windy Bay Deep Upper      | 0.005   | 1.340        | 2.230    | 1.380                |
|                          | Coeur d'Alene River Upper | 0.650   | 0.005        | -        | 2.960                |
|                          | Mid Lake Upper            | 1.420   | 0.640        | 1.900    | 2.910                |
|                          | Carey Bay Upper           | 0.640   | 0.005        | -        | 1.280                |
|                          | Conkling Point Upper      | 0.005   | 2.160        | 1.380    | 3.390                |
|                          | Hidden Lake Upper         | 0.005   | Contaminated | 3.620    | 6.890                |
|                          | Round Lake                | 0.005   | 1.540        | 0.670    | 0.640                |
|                          | Chatcolet Lake Upper      | 3.140   | 2.810        | 4.260    | 7.660                |
|                          | Chatcolet Lake Shallow    | 5.320   | 4.890        | 2.780    | 0.690                |
|                          | Benewah Lake Upper        | 0.640   | 25.790       | 3.540    | -                    |
|                          | St. Joe River Upper       | -       | Contaminated | 1.400    | 0.005                |

- No sample taken to lab.

<sup>a</sup> Coeur d'Alene Lake was isothermal.

Table 3.14 Chlorophyll<sub>a</sub> (µg/L) results from the hypolimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho.

| Nutrients                | Location                  | Date    |              |          |                      |
|--------------------------|---------------------------|---------|--------------|----------|----------------------|
|                          |                           | 8/13/97 | 9/16/97      | 10/20/97 | 11/4/97 <sup>a</sup> |
| Chlorophyll <sub>a</sub> | Rockford Bay Lower        | 1.400   | 2.090        | -        | -                    |
|                          | Windy Bay Shallow Lower   | 2.140   | 0.680        | -        | -                    |
|                          | Windy Bay Deep Lower      | 0.640   | 0.680        | 0.005    | -                    |
|                          | Coeur d'Alene River Lower | 0.690   | 1.320        | 0.005    | -                    |
|                          | Mid Lake Lower            | 0.690   | 2.070        | 1.380    | -                    |
|                          | Carey Bay Lower           | 0.005   | 1.330        | 1.400    | -                    |
|                          | Conkling Point Lower      | 0.005   | 1.370        | 0.005    | -                    |
|                          | Hidden Lake Lower         | 11.920  | 20.250       | 1.980    | -                    |
|                          | Chatcolet Lake Lower      | 6.030   | Contaminated | 14.04    | -                    |
|                          | Benewah Lake Lower        | 34.14   | 28.78        | -        | -                    |
|                          | St. Joe River Lower       | -       | 0.700        | 0.005    | -                    |

- No sample taken to lab.

<sup>a</sup> Coeur d'Alene Lake was isothermal.

### 3.1.2 Fisheries

From 1994-1997 Coeur d'Alene Lake was sampled using electrofishing (n= 8280) and gillnetting (n=1211) methods, primarily from April- October each year. Both the littoral and limnetic zones were sampled. Eighteen of the twenty-one species of fish known to occur in the lake were captured. Electrofishing efforts were focused in the littoral zone while gillnetting efforts were focused primarily in the limnetic zone, however, some gillnetting effort was focused in the littoral zones as well.

Relative abundance estimates derived from electrofishing data show that 61.89% of the catch consisted of introduced species, with yellow perch and largemouth bass being the most abundant (Table 3.15). Yellow perch were the most abundant introduced species in each of the four years sampled. Native fishes comprised only 38.11% of the catch, with largescale suckers being the most abundant. Largescale suckers were the most abundant native species in each of the individual sample seasons. Cutthroat trout only comprised 0.83% of the catch from 1994-1997. Of the 69 cutthroat trout caught in the littoral zone between 1994-1997, 47 were captured at night and only 22 were captured during the day.

Table 3.15 Electrofishing relative abundance from 1994-1997.

|         | 1994         |         | 1995         |         | 1996         |         | 1997          |         | 1994-1997     |
|---------|--------------|---------|--------------|---------|--------------|---------|---------------|---------|---------------|
| Species | (n=1418)     | Species | (n=1727)     | Species | (n=536)      | Species | (n=4599)      | Species | (n=8280)      |
| YP      | 25.88% (367) | LSS     | 27.74% (479) | LSS     | 28.92% (155) | LSS     | 28.55% (1313) | LSS     | 27.27% (2258) |
| LSS     | 21.93% (311) | YP      | 26.52% (458) | YP      | 22.01% (118) | YP      | 22.94% (1055) | YP      | 24.13% (1998) |
| LMB     | 18.55% (263) | LMB     | 13.20% (228) | PSS     | 15.30% (82)  | PSS     | 10.55% (485)  | LMB     | 11.92% (987)  |
| SQW     | 10.65% (151) | PSS     | 10.48% (181) | SQW     | 15.11% (81)  | LMB     | 09.92% (456)  | PSS     | 10.28% (851)  |
| PSS     | 07.26% (103) | SQW     | 06.89% (119) | LMB     | 07.46% (40)  | SQW     | 09.35% (430)  | SQW     | 09.43% (781)  |
| BC      | 06.63% (94)  | BBH     | 05.19% (102) | BC      | 04.66% (25)  | BC      | 08.52% (392)  | BC      | 06.87% (569)  |
| TCH     | 02.40% (34)  | TCH     | 04.17% (72)  | TCH     | 02.99% (16)  | BBH     | 04.02% (185)  | BBH     | 04.01% (332)  |
| BBH     | 02.19% (31)  | BC      | 03.36% (58)  | BBH     | 02.61% (14)  | TCH     | 03.50% (161)  | TCH     | 03.42% (283)  |
| PIK     | 01.48% (21)  | CTT     | 0.75% (13)   | PIK     | 0.56% (3)    | CTT     | 0.78% (36)    | CTT     | 0.83% (69)    |
| CTT     | 01.27% (18)  | MWF     | 0.41% (7)    | CTT     | 0.37% (2)    | PIK     | 0.50% (23)    | PIK     | 0.58% (48)    |
| KOK     | 01.13% (16)  | SCP     | 0.29% (5)    | BLT     | 0            | SCP     | 0.43% (20)    | KOK     | 0.41% (34)    |
| SCP     | 0.49% (7)    | KOK     | 0.17% (3)    | CCF     | 0            | KOK     | 0.33% (15)    | SCP     | 0.39% (32)    |
| MWF     | 0.07% (1)    | BLT     | 0.06% (1)    | CHN     | 0            | CHN     | 0.30% (14)    | MWF     | 0.18% (15)    |
| CCF     | 0.07% (1)    | PIK     | 0.06% (1)    | KOK     | 0            | MWF     | 0.15% (7)     | CHN     | 0.17% (14)    |
| SMB     | 0            | CCF     | 0            | MWF     | 0            | CCF     | 0.09% (4)     | CCF     | 0.06% (5)     |
| RBT     | 0            | CHN     | 0            | RBT     | 0            | SMB     | 0.04% (2)     | SMB     | 0.02% (2)     |
| CHN     | 0            | RBT     | 0            | SCP     | 0            | RBT     | 0.02% (1)     | RBT     | 0.01% (1)     |
| BLT     | 0            | SMB     | 0            | SMB     | 0            | BLT     | 0             | BLT     | 0.01% (1)     |

Relative abundance estimates derived from gillnetting data show 57.89% of the fish captured were introduced species (Table 3.16). Native fishes comprised only 42.11% of the catch of which cutthroat trout comprised 0.91%. These data show that nine species of fish were more abundant than cutthroat trout lakewide. However, cutthroat trout were the fourth most abundant species in locations where they were captured. In nearly all transects where cutthroat trout were captured, kokanee salmon were the most abundant species, followed by northern squawfish and largescale suckers.



Largescale suckers and northern squawfish thrive throughout the lake. They do just as well in both the limnetic and littoral zones of the lake. Most of the other introduced species are primarily limited to the littoral zone of the lake, the area most influenced by construction of Post Falls dam. Kokanee salmon appear to be the introduced species that most effectively thrive in the limnetic zones of the lake. Chinook salmon are also found primarily in the limnetic zones, however, they are less susceptible to the capture methods that we utilized for this report.

Cutthroat trout captured in the littoral zone were found primarily in habitat associated with steep, rocky shorelines where few aquatic macrophytes exist. Of the thirty-six cutthroat trout captured in 1997, sixteen were caught in inundated riverine habitat or in reaches near mouths of tributaries where spawning is known to occur. Cutthroat trout caught in the limnetic zone were captured in nets set in locations greater than ten meters deep.

Analysis of the CPUE (Tables 3.17 and 3.18) data shows wide fluctuations from year to year. Every attempt was made to standardize gear and methods in 1997. Future estimates should contain less bias and more accurately reflect actual stock densities.

Results from the water quality component of the HSI model indicated that there is suitable habitat for cutthroat trout in the lake (Table 3.19). The quantity of suitable habitat, however, decreases as water temperature increases during the year. The suitability index was poor or very poor (<0.25) in the shallow portion of the water column at all sample stations. While water quality does not directly exclude cutthroat trout from these shallow areas, unsuitable habitat exerts added stress on cutthroat trout making foraging runs into the upper 10 meters of the water column.

## **3.2 Stream Studies**

### **3.2.1 Water Quality**

#### **Lake Creek**

Water temperature measured at the lower Lake Creek station indicated a maximum of 24° C on July 21, 1997. The maximum 7-day moving average at the same station exceeded 16° C through most of June and all of July. Daily temperature fluctuations ranged from 4.7° C to 7.8° C during July, indicating that substantial cooling did take place. A gap in data from late July to October due to equipment failure probably missed continued temperature pollution problems, as stream temperatures generally do not begin cooling until late August. Temperatures recorded at the upper Lake Creek station are considerably lower. The maximum 7-day moving average exceeded 16° C from August 1-20, while the daily maximum exceeded 18° C only once (8-6-97). Dissolved oxygen did not drop below 9 mg/l at the lower station, where the highest temperatures were recorded. Measured base flows ranged from 3.8 to 7.9 cfs from July 1 through August 9.

#### **Benewah Creek**

A maximum temperature of 24° C was recorded at the 9 mile station on August 6, 1997. The maximum 7-day moving average exceeded 16° C from mid June through mid September and minimum temperature did not fall below 16° C for 15 days during this period. Dissolved oxygen was less than 9 mg/l during the period of highest water temperatures, but did not drop below 7.6 mg/l. Discharge at the 3 mile station ranged from 2.9 to 3.9 cfs from mid August through September.

Water quality conditions in the tributaries varied greatly, but in general provided more suitable water temperatures than mainstem reaches. Temperature did not exceed 16° C in S.F. Benewah Creek and dissolved oxygen did not drop below 9 mg/l. Discharge, however, was less than 1.0 cfs from mid August

Table 3.16 Electroshocking catch per unit effort from 1994-1997.

| 1994    |          | 1995         |         | 1996     |              | 1997    |         | 1994-1997   |         |          |              |         |          |               |
|---------|----------|--------------|---------|----------|--------------|---------|---------|-------------|---------|----------|--------------|---------|----------|---------------|
| Species | (n=1418) | Time/hr      | Species | (n=1727) | Time/hr      | Species | (n=536) | Time/hr     | Species | (n=4599) | Time/Hr      | Species | (n=8280) | Time/Hr       |
|         |          | <b>20.93</b> |         |          | <b>24.89</b> |         |         | <b>12.0</b> |         |          | <b>42.45</b> |         |          | <b>100.27</b> |
| YP      | 367      | 17.53        | LSS     | 479      | 19.24        | LSS     | 157     | 13.08       | LSS     | 1313     | 30.93        | LSS     | 2258     | 22.52         |
| LSS     | 311      | 14.86        | YP      | 458      | 18.40        | YP      | 123     | 10.25       | YP      | 1055     | 24.85        | YP      | 1998     | 19.93         |
| LMB     | 263      | 12.57        | LMB     | 228      | 9.16         | SQW     | 86      | 7.17        | PSS     | 485      | 11.43        | PSS     | 987      | 9.84          |
| SQW     | 151      | 7.21         | PSS     | 181      | 7.27         | PSS     | 82      | 6.83        | LMB     | 456      | 10.74        | LMB     | 851      | 8.49          |
| PSS     | 103      | 4.92         | SQW     | 119      | 4.78         | LMB     | 41      | 3.42        | SQW     | 430      | 10.13        | SQW     | 781      | 7.79          |
| BC      | 94       | 4.49         | BBH     | 102      | 4.10         | BC      | 25      | 2.08        | BC      | 392      | 9.23         | BC      | 569      | 5.67          |
| TCH     | 34       | 1.62         | TCH     | 72       | 2.89         | BBH     | 19      | 1.58        | BBH     | 185      | 4.36         | BBH     | 332      | 3.31          |
| BBH     | 31       | 1.48         | BC      | 58       | 2.33         | TCH     | 16      | 1.33        | TCH     | 161      | 3.79         | TCH     | 283      | 2.82          |
| PIK     | 21       | 1.00         | CTT     | 13       | 0.52         | PIK     | 3       | 0.25        | CTT     | 36       | 0.85         | CTT     | 69       | 0.69          |
| CTT     | 18       | 0.86         | MWF     | 7        | 0.28         | CTT     | 2       | 0.17        | PIK     | 23       | 0.54         | PIK     | 48       | 0.48          |
| KOK     | 16       | 0.76         | SCP     | 5        | 0.20         | BLT     | 0       | 0.00        | SCP     | 20       | 0.47         | KOK     | 34       | 0.34          |
| SCP     | 7        | 0.33         | KOK     | 3        | 0.12         | CCF     | 0       | 0.00        | KOK     | 15       | 0.35         | SCP     | 32       | 0.32          |
| CCF     | 1        | 0.05         | BLT     | 1        | 0.04         | CHN     | 0       | 0.00        | CHN     | 14       | 0.33         | MWF     | 15       | 0.15          |
| MWF     | 1        | 0.05         | PIK     | 1        | 0.04         | KOK     | 0       | 0.00        | MWF     | 7        | 0.16         | CHN     | 14       | 0.14          |
| BLT     | 0        | 0.00         | CCF     | 0        | 0.00         | MWF     | 0       | 0.00        | CCF     | 4        | 0.09         | CCF     | 5        | 0.05          |
| CHN     | 0        | 0.00         | CHN     | 0        | 0.00         | RBT     | 0       | 0.00        | SMB     | 2        | 0.05         | SMB     | 2        | 0.02          |
| RBT     | 0        | 0.00         | RBT     | 0        | 0.00         | SCP     | 0       | 0.00        | RBT     | 1        | 0.02         | RBT     | 1        | 0.01          |
| SMB     | 0        | 0.00         | SMB     | 0        | 0.00         | SMB     | 0       | 0.00        | BLT     | 0        | 0.00         | BLT     | 1        | 0.01          |

Table 3.17 Gillnetting relative abundance from 1994-1997.

| 1994    |              | 1995    |             | 1996    |              | 1997    |              | 1994-1997 |              |
|---------|--------------|---------|-------------|---------|--------------|---------|--------------|-----------|--------------|
| Species | (n=211)      | Species | (n=78)      | Species | (n=286)      | Species | (n=636)      | Species   | (n=1211)     |
| YP      | 63.51% (134) | YP      | 65.38% (51) | LSS     | 41.96% (120) | YP      | 34.28% (218) | YP        | 40.71% (493) |
| SQW     | 16.11% (34)  | SQW     | 20.51% (16) | YP      | 31.47% (90)  | LSS     | 21.23% (135) | LSS       | 23.45% (284) |
| LSS     | 09.00% (19)  | LSS     | 12.82% (10) | SQW     | 16.08% (46)  | SQW     | 18.55% (118) | SQW       | 17.67% (214) |
| BBH     | 02.84% (6)   | BBH     | 01.28% (1)  | BBH     | 02.45% (7)   | KOK     | 10.06% (64)  | KOK       | 06.28% (76)  |
| KOK     | 02.84% (6)   | BC      | 0           | KOK     | 02.10% (6)   | BBH     | 07.39% (47)  | BBH       | 05.04% (61)  |
| PSS     | 01.90% (4)   | BLT     | 0           | PIK     | 01.40% (4)   | BC      | 02.52% (16)  | BC        | 01.57% (19)  |
| PIK     | 01.42% (3)   | CCF     | 0           | TCH     | 01.40% (4)   | CTT     | 01.42% (9)   | TCH       | 01.16% (14)  |
| TCH     | 0.95% (2)    | CHN     | 0           | CCF     | 01.05% (3)   | TCH     | 01.26% (8)   | PSS       | 01.07% (13)  |
| BC      | 0.47% (1)    | CTT     | 0           | PSS     | 01.05% (3)   | PSS     | 0.94% (6)    | PIK       | 0.99% (12)   |
| CTT     | 0.47% (1)    | KOK     | 0           | BC      | 0.70% (2)    | CHN     | 0.79% (5)    | CTT       | 0.91% (11)   |
| LMB     | 0.47% (1)    | LMB     | 0           | CTT     | 0.35% (1)    | PIK     | 0.79% (5)    | CCF       | 0.58% (7)    |
| BLT     | 0            | MWF     | 0           | BLT     | 0            | CCF     | 0.63% (4)    | CHN       | 0.41% (5)    |
| CCF     | 0            | PIK     | 0           | CHN     | 0            | MWF     | 0.16% (1)    | MWF       | 0.08% (1)    |
| CHN     | 0            | PSS     | 0           | LMB     | 0            | BLT     | 0            | LMB       | 0.08% (1)    |
| MWF     | 0            | RBT     | 0           | MWF     | 0            | LMB     | 0            | SMB       |              |
| RBT     | 0            | SCP     | 0           | RBT     | 0            | RBT     | 0            | SCP       |              |
| SCP     | 0            | SMB     | 0           | SCP     | 0            | SCP     | 0            | RBT       |              |
| SMB     | 0            | TCH     | 0           | SMB     | 0            | SMB     | 0            | BLT       |              |

Table 3.18 Gillnetting catch per unit effort from 1994-1997.

| 1994    |         |           | 1995    |        |           | 1996    |         |               | 1997    |         |            | 1994-1997 |          |               |
|---------|---------|-----------|---------|--------|-----------|---------|---------|---------------|---------|---------|------------|-----------|----------|---------------|
| Species | (n=211) | <b>69</b> | Species | (n=78) | <b>79</b> | Species | (n=286) | <b>170.50</b> | Species | (n=636) | <b>677</b> | Species   | (n=1211) | <b>995.50</b> |
| YP      | 134     | 1.94      | YP      | 51     | 0.65      | LSS     | 120     | 0.70          | YP      | 218     | 0.32       | YP        | 493      | 0.50          |
| SQW     | 34      | 0.49      | SQW     | 16     | 0.20      | YP      | 90      | 0.53          | LSS     | 135     | 0.20       | LSS       | 284      | 0.29          |
| LSS     | 19      | 0.28      | LSS     | 10     | 0.13      | SQW     | 46      | 0.27          | SQW     | 118     | 0.17       | SQW       | 214      | 0.21          |
| BBH     | 6       | 0.09      | BBH     | 1      | 0.01      | BBH     | 7       | 0.04          | KOK     | 64      | 0.09       | KOK       | 76       | 0.08          |
| KOK     | 6       | 0.09      | BC      | 0      | 0.00      | KOK     | 6       | 0.04          | BBH     | 47      | 0.07       | BBH       | 61       | 0.06          |
| PSS     | 4       | 0.06      | BLT     | 0      | 0.00      | PIK     | 4       | 0.02          | BC      | 16      | 0.02       | BC        | 19       | 0.02          |
| PIK     | 3       | 0.04      | CCF     | 0      | 0.00      | TCH     | 4       | 0.02          | CTT     | 9       | 0.01       | CTT       | 11       | 0.01          |
| TCH     | 2       | 0.03      | CHN     | 0      | 0.00      | CCF     | 3       | 0.02          | TCH     | 8       | 0.01       | TCH       | 14       | 0.01          |
| BC      | 1       | 0.01      | CTT     | 0      | 0.00      | PSS     | 3       | 0.02          | PSS     | 6       | 0.01       | PSS       | 13       | 0.01          |
| CTT     | 1       | 0.01      | KOK     | 0      | 0.00      | BC      | 2       | 0.01          | CHN     | 5       | 0.01       | CHN       | 5        | 0.01          |
| LMB     | 1       | 0.01      | LMB     | 0      | 0.00      | CTT     | 1       | 0.01          | PIK     | 5       | 0.01       | PIK       | 12       | 0.01          |
| BLT     | 0       | 0.00      | MWF     | 0      | 0.00      | BLT     | 0       | 0.00          | CCF     | 4       | 0.01       | CCF       | 7        | 0.01          |
| CCF     | 0       | 0.00      | PIK     | 0      | 0.00      | CHN     | 0       | 0.00          | MWF     | 1       | 0.00       | MWF       | 1        | 0.00          |
| CHN     | 0       | 0.00      | PSS     | 0      | 0.00      | LMB     | 0       | 0.00          | BLT     | 0       | 0.00       | BLT       | 0        | 0.00          |
| MWF     | 0       | 0.00      | RBT     | 0      | 0.00      | MWF     | 0       | 0.00          | LMB     | 0       | 0.00       | LMB       | 1        | 0.00          |
| RBT     | 0       | 0.00      | SCP     | 0      | 0.00      | RBT     | 0       | 0.00          | RBT     | 0       | 0.00       | RBT       | 0        | 0.00          |
| SCP     | 0       | 0.00      | SMB     | 0      | 0.00      | SCP     | 0       | 0.00          | SCP     | 0       | 0.00       | SCP       | 0        | 0.00          |
| SMB     | 0       | 0.00      | TCH     | 0      | 0.00      | SMB     | 0       | 0.00          | SMB     | 0       | 0.00       | SMB       | 0        | 0.00          |

Table 3.19 Habitat suitability index for lucustrine Cutthroat Trout based on water quality.

| Location               | Depth                       | HSI <sup>a</sup>                 |   | Suitability Index |
|------------------------|-----------------------------|----------------------------------|---|-------------------|
| Rockford Bay           | 0-7 Meters                  | $(0.25 \times 1 \times 1)^{1/3}$ | = | 0.25 SI           |
|                        | 7-11 Meters                 | $(0.60 \times 1 \times 1)^{1/3}$ | = | 0.845 SI          |
|                        | 11-Bottom (14) <sup>b</sup> | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Windy Bay Shallow      | 0-7 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 7-10 Meters                 | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 10-Bottom (15)              | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Windy Bay Deep         | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 10-15 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 15-Bottom (33)              | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Coeur d'Alene River    | 0-Bottom                    | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
| Mid-Lake Coeur d'Alene | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 10-13 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 13-Bottom (17)              | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Carey Bay              | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 10-12 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 12-Bottom (13)              | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Conkling Park          | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 10-13 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 13-Bottom (16)              | $(1 \times 1 \times 1)^{1/3}$    | = | 1.0 SI            |
| Hidden Lake            | 0-5 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 5-7 Meters                  | $(0.8 \times 1 \times 1)^{1/3}$  | = | 0.92 SI           |
|                        | 7-Bottom (10)               | $(1 \times 0.0 \times 1)^{1/3}$  | = | 0.0 SI            |
| Round Lake             | 0-Bottom (1.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
| Chatcolet Lake         | 0-6 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
|                        | 6-9 Meters                  | $(0.85 \times 1 \times 1)^{1/3}$ | = | 0.94 SI           |
|                        | 9-Bottom (11)               | $(1 \times 0.0 \times 1)^{1/3}$  | = | 0.0 SI            |
| Chatcolet Shallow      | 0-Bottom (1.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
| Benewah Lake           | 0-Bottom (4.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = | 0.0 SI            |
| St. Joe River          | 0-Bottom (12.5)             | $(0.4 \times 1 \times 1)^{1/3}$  | = | 0.4 SI            |

<sup>a</sup> Habitat Suitability Index (HSI).

<sup>b</sup> Numbers in parenthesis represent the bottom in meters.

through early October. Measured water temperatures in W.F. Benewah Creek exceeded 16° C only once in early August and dissolved oxygen did not decrease below 9 mg/l. School House Creek was also a source of cold water and the maximum 7-day moving average exceeded 16° C for only 5 days in August. There were periods where habitat consisted of stagnant pools from July through August and dissolved oxygen decreased from 6.8 to 3.2 mg/l during this time. The maximum 7-day moving average in Windfall Creek exceeded 16° C from mid June through mid September and exceeded 20° C for 40 days. A maximum temperature of 27° C was recorded on August 6. Daily temperature fluctuations were as high as 9.4° C in August and minimum temperature did not drop below 16° C for 9 days. There was no flow in Windfall Creek from late July through August and dissolved oxygen dropped below 9 mg/l during this period.

### Evans Creek

Maximum water temperature was 17° C on August 6, 1997. The maximum 7-day moving average exceeded 16° C for 8 days in August. Dissolved oxygen did not drop below 9 mg/l during the period of highest water temperatures. Measured base flows ranged from 2.7 to 4.5 cfs from mid-August through early October.

### Alder Creek

Maximum water temperature was 22° C on August 6, 1997. The maximum 7-day moving average exceeded 16° C from mid July through early September. Dissolved oxygen did not drop below 9 mg/l during the period of highest water temperatures. Measured base flows ranged from 1.8 to 0.3 cfs from July through September.

### 3.2.2 Habitat Suitability Indices

The suitability index (SI) values for individual water quality parameters vary considerably between sample locations (Table 3.20). The greatest variability occurs for the temperature parameter (V<sub>1</sub>), where the SI ranges from 0 to 1.0. Water temperatures are limiting for the mainstem of Benewah Creek, lower Lake Creek, and lower Windfall Creek. The SI for the base flow parameter (SI<sub>14</sub>) is < 0.4 for all sample locations except for Evans Creek and mainstem Benewah Creek, indicating that base flow is also a limiting factor at most locations. The SI for dissolved oxygen (SI<sub>3</sub>) and pH (SI<sub>13</sub>) are generally greater than 0.8, and therefore are not considered limiting. The exception occurs in School House Creek where dissolved oxygen is limiting (SI<sub>3</sub>=0.3) during the period of warmest water temperatures.

Table 3.20 Habitat Suitability Index (HSI) calculations for riverine cutthroat trout.

| Location     |      |      |     |      |         |      |     |      | Comp | Non-Comp |
|--------------|------|------|-----|------|---------|------|-----|------|------|----------|
|              | V1   | SI1  | V3  | SI3  | V13     | SI13 | V14 | SI14 | HSI  | HSI      |
| L. Lake      | 22.6 | 0    | 9.5 | 1    | 6.8/7.7 | 1    | 9   | 0.2  | 0.00 | 0        |
| U. Lake      | 17.9 | 0.78 | 7.9 | 0.9  | 6.5/7.5 | 1    | 13  | 0.25 | 0.65 | 0.25     |
| L. Benewah   | 23   | 0    | 8.9 | 1    | 7.0/8.3 | 1    | 18  | 0.4  | 0.00 | 0        |
| U. Benewah   | 22.8 | 0    | 7.7 | 0.87 | 6.7/7.6 | 1    | 18  | 0.4  | 0.00 | 0        |
| S.E. Benewah | 14.7 | 1    | 9.7 | 1    | 6.6/7.6 | 1    | 16  | 0.32 | 0.75 | 0.32     |
| School House | 16.4 | 0.92 | 5.7 | 0.3  | 6.8/7.4 | 1    | 6   | 0.15 | 0.45 | 0.15     |
| W.F. Benewah | 16.6 | 0.9  | 9.3 | 1    | 6.7/7.5 | 1    | 11  | 0.25 | 0.69 | 0.25     |
| Windfall     | 25.1 | 0    | 7.8 | 0.89 | 6.7/7.6 | 1    | 13  | 0.25 | 0.00 | 0        |
| Evans        | 16.4 | 0.92 | 9.6 | 1    | 6.3/7.7 | 0.95 | 28  | 0.6  | 0.85 | 0.6      |
| Alder        | 20.6 | 0.45 | 9.6 | 1    | 6.8/7.8 | 1    | 16  | 0.32 | 0.62 | 0.32     |

HSI scores that are calculated using the non-compensatory method show a very poor to poor rating for all sample locations, with the exception of Evans Creek, which is considered good. In other words, when

habitat suitability is rated based on water quality parameters alone, then all sample locations, with the exception of Evans Creek, are rated very poor to poor with regard to cutthroat trout preferences. In six of ten locations, however, differences between HSI calculations using the compensatory versus non-compensatory method indicate that good habitat conditions have the potential to partially compensate for short-term degradation in water quality. These sites include upper Lake Creek, S.F. Benewah Creek, School House Creek, W.F. Benewah Creek, Evans Creek, and Alder Creek. Lower Lake Creek, the mainstem of Benewah Creek, and lower Windfall Creek are considered very poor regardless of the method used.

### **3.2.3 Fisheries**

#### **Abundance and Distribution**

General patterns of cutthroat trout abundance and distribution vary among the target watersheds and among years, but seem to be highly correlated to seasonal changes in water quality and quantity. Cutthroat trout are sporadically distributed in the Lake Creek, Benewah Creek, and Alder Creek watersheds during both the summer and fall seasons (Tables 3.21 and 3.22). Abundance in the second order tributaries of Lake Creek and Benewah Creek are consistently much higher than in adjacent mainstem reaches, despite the effects of low flow conditions. During base flow conditions, for example, cutthroat trout have been known to crowd into small, isolated pools ( $>15$  fish/m<sup>2</sup>) located in cool tributaries, rather than face conditions of high water temperatures in mainstem reaches. In contrast, favorable water quality conditions in Evans Creek result in a relatively even distribution of cutthroat trout. Cutthroat trout abundance is consistently lowest in Alder Creek.

Surveys conducted in 1997 showed that cutthroat trout abundance increased dramatically during the fall sample period. Fall surveys were conducted following fry emergence, which occurs in late June to early July, and young of the year fish accounted for most of the seasonal variation in abundance within sites (Table 3.23). Young of the year fish were found principally in small tributaries, which supports the hypothesis that the majority of spawning activity takes place in second order streams in these watersheds.

Brook trout have been found only in Alder Creek and Benewah Creek; the respective dates of introduction are unknown. Fish are distributed in fairly high numbers (up to 30/100m<sup>2</sup>) throughout the upper reaches of the Alder Creek watershed. Distribution in the Benewah Creek watershed is limited to the upper mainstem and a few of the primary tributaries and abundance is typically much lower than for cutthroat trout. In Alder Creek, however, brook trout are found in greater numbers than cutthroat trout in all but the lowermost stream reaches (Table 3.24).

#### **Age and Growth**

A total of 680 cutthroat trout scales were examined for age and growth determination in 1996 and 162 additional scales were examined in 1997. Growth and potential maximum size of cutthroat trout varies from stock to stock (Table 3.25). The adfluvial stocks found in Benewah Creek and Lake Creek exhibit the maximum growth potential for this species. Record size for Lake Creek cutthroat trout is 17.9 inches (456 mm TL) and 1.5 pounds (698 grams), caught in 1997. Record size for Benewah Creek cutthroat trout is 16.8 inches (427 mm TL) and 1.5 pounds (670 grams), caught in 1994. Maximum growth for resident fish stocks ranges from 222 to 280 mm total length and 102 to 180 grams. A complete tabular analysis of growth for cutthroat trout is provided in Appendix C.

Table 3.21 Cutthroat trout abundance and distribution by watershed, 1996.

| Cutthroat Trout Lake Creek |       | Summer 96 |         | Fall 96 |         |
|----------------------------|-------|-----------|---------|---------|---------|
| Tributary                  | Reach | N±(SE)    | #/100m2 | N±(SE)  | #/100m2 |
| Mainstem                   | 1     | 1         | 0.5     | 1       | 0.3     |
|                            | 2     | --        | --      | --      | --      |
|                            | 3     | --        | --      | --      | --      |
|                            | 4     | 23        | 7.7     | 36(0.7) | 12.1    |
|                            | 5     | 18        | 4.8     | 2       | 0.5     |
|                            | 6     | 31(0.4)   | 4.9     | 33(1.2) | 3.9     |
|                            | 7     | 16        | 2.4     | 6       | 0.4     |
|                            | 8     | 7         | 4.2     | 0       | 0.0     |
| West Fork                  |       | 91(25.7)  | 16.9    | 6(0.8)  | 6.5     |
| Bozard                     |       | 25(0.3)   | 10.8    | 11      | 5.1     |
| Totals                     |       | 212(25.7) |         | 83(2.5) |         |

| Cutthroat Trout Benewah Creek |            | Summer 96 |         | Fall 96 |         |
|-------------------------------|------------|-----------|---------|---------|---------|
| Tributary                     | Reach      | N±(SE)    | #/100m2 | N±(SE)  | #/100m2 |
| Mainstem                      | 1          | 1         | 0.4     | 2.0     | 0.5     |
|                               | 2          | 5(1.5)    | 0.8     | 4.0     | 1.8     |
|                               | 3          | 14(0.3)   | 6.3     | 10      | 1.6     |
|                               | 4          | 5(1.0)    | 1.1     | 5(1.5)  | 1.5     |
|                               | 5          | 3         | 1.2     | 0       | 0.0     |
|                               | 7          | --        | --      | 7       | 0.9     |
|                               | 8          | 4         | 1.4     | 7       | 2.5     |
|                               | 9          | 12(4.2)   | 1.9     | 2       | 0.5     |
|                               | 10         | 14(0.6)   | 1.4     | 9       | 0.9     |
|                               | 11         | 3         | 1.2     | 0       | 0.0     |
|                               | South Fork |           | 6       | 5.4     | 9(0.77) |
| Bull                          |            | 25(0.6)   | 19.2    | --      | --      |
| West Fork                     |            | 28(1.4)   | 18.8    | 23(4.2) | 10.3    |
| Whitetail                     |            | 33(0.4)   | 22.2    | --      | --      |
| Windfall                      |            | 27(0.85)  | 14.5    | --      | --      |
| Totals                        |            | 180(6.4)  |         | 78(4.5) |         |

| Cutthroat Trout Evans Creek |       | Summer 96 |         |
|-----------------------------|-------|-----------|---------|
| Tributary                   | Reach | N±(SE)    | #/100m2 |
| Mainstem                    | 1     | 0         | 0.0     |
|                             | 2     | 18(0.9)   | 4.0     |
|                             | 3     | 4         | 1.5     |
|                             | 4     | 33(4.31)  | 4.9     |
|                             | 5     | 23(0.6)   | 10.3    |
|                             | 6     | 5         | 2.2     |
|                             | 7     | 30(4.1)   | 4.9     |
| East Fork                   |       | 6         | 10.8    |
| Rainbow Fork                |       | 36(3.6)   | 19.4    |
| South Fork                  |       | 34(6.9)   | 12.2    |
| Totals                      |       | 189(9.9)  |         |

| Cutthroat Trout Alder Creek |       | Summer 96 |         |
|-----------------------------|-------|-----------|---------|
| Tributary                   | Reach | N±(SE)    | #/100m2 |
| Mainstem                    | 1     | 3         | 0.5     |
|                             | 2     | 0         | 0.0     |
|                             | 3     | 3         | 0.9     |
|                             | 4     | 1         | 0.3     |
|                             | 5     | 14(0.5)   | 2.1     |
|                             | 6     | 6         | 1.8     |
|                             | 7     | 21        | 2.7     |
|                             | 8     | 2         | 0.5     |
|                             | 9     | 1         | 0.2     |
| North Fork                  | 1     | 0         | 0.0     |
|                             | 2     | 0         | 0.0     |
|                             | 3     | 0         | 0.0     |
|                             | 4     | 2         | 2.7     |
| Totals                      |       | 53(0.5)   |         |

Population estimates were not completed for Evans Creek or Alder Creek in 1996.



Table 3.22 Cutthroat trout abundance and distribution, 1997.

| Lake Creek |       |           |         |          |         | Evans Creek |           |           |           |         |         |
|------------|-------|-----------|---------|----------|---------|-------------|-----------|-----------|-----------|---------|---------|
|            |       | Summer 97 |         | Fall 97  |         |             |           | Summer 97 |           | Fall 97 |         |
| Tributary  | Reach | N±(SE)    | #/100m2 | N±(SE)   | #/100m2 | Tributary   | Reach     | N±(SE)    | #/100m2   | N±(SE)  | #/100m2 |
| Mainstem   | 1     | 0.0       | 0.0     | 2.0      | 1.8     | Mainstem    | 1         | 0         | 0.0       | 0       | 0.0     |
|            | 2     | --        | --      | --       | --      |             | 2         | 6(1.0)    | 1.1       | 7       | 1.4     |
|            | 3     | --        | --      | --       | --      |             | 3         | 2         | 1.0       | 7       | 1.9     |
|            | 4     | 8.0       | 2.1     | 14(1.4)  | 3.7     |             | 4         | 32(0.6)   | 6.6       | 68(6.8) | 9.1     |
|            | 5     | 2.0       | 0.5     | 8.0      | 1.9     |             | 5         | 26(1.4)   | 7.8       | 23(3.2) | 5.4     |
|            | 6     | 3.0       | 0.4     | 8.0      | 1.4     |             | 6         | 23(1.9)   | 8.3       | 30(3.3) | 20.2    |
|            | 7     | 0.0       | 0.0     | 3.0      | 1.1     |             | 7         | 30(0.9)   | 12.4      | 12(4.2) | 25.8    |
|            | 8     | 0.0       | 0.0     | 20.0     | 9.0     |             | E.F.Evans | 5(1.0)    | 2.2       | 22(0.6) | 29.6    |
| Bozard     |       | 36(2.1)   | 12.1    | 78(1.7)  | 32.3    | R.F.Evans   | --        | --        | 21(5.1)   | 32.3    |         |
| West Fork  |       | 14.0      | 4.7     | 93(3.6)  | 19.3    | S.F.Evans   | 12(0.6)   | 7.2       | 26(3.0)   | 10.8    |         |
| Totals     |       | 63(2.1)   |         | 224(6.7) |         | Totals      | 136(3.0)  |           | 216(11.0) |         |         |

| Benewah Creek |            |           |          |          |          | Alder Creek |           |           |         |         |         |     |
|---------------|------------|-----------|----------|----------|----------|-------------|-----------|-----------|---------|---------|---------|-----|
|               |            | Summer 97 |          | Fall 97  |          |             |           | Summer 97 |         | Fall 97 |         |     |
| Tributary     | Reach      | N±(SE)    | #/100m2  | N±(SE)   | #/100m2  | Tributary   | Reach     | N±(SE)    | #/100m2 | N±(SE)  | #/100m2 |     |
| Mainstem      | 1          | 0         | 0.0      | 0        | 0.0      | Mainstem    | 1         | 3         | 0.5     | 1       | 0.1     |     |
|               | 2          | 0         | 0.0      | 6        | 0.9      |             | 2         | 1         | 0.3     | 0       | 0.0     |     |
|               | 3          | 1         | 0.2      | 19(0.69) | 3.8      |             | 3         | 3         | 0.8     | 0       | 0.0     |     |
|               | 4          | 0         | 0.0      | 2        | 0.5      |             | 4         | --        | --      | 42(3.4) | 10.8    |     |
|               | 5          | 3         | 0.8      | 1        | 0.5      |             | 5         | 6         | 1.0     | 13      | 1.7     |     |
|               | 8          | 0         | 0.0      | 2        | 0.4      |             | 6         | 3         | 1.6     | 0       | 0.0     |     |
|               | 9          | 8         | 2.2      | 6        | 1.5      |             | 7         | 17        | 1.8     | 15(1.7) | 1.3     |     |
|               | 10         | 3         | 2.7      | 4        | 1.1      |             | 8         | 5         | 1.8     | 2       | 0.5     |     |
|               | 11         | 4         | 1.1      | 0        | 0.0      |             | 9         | 11(2.9)   | 4.2     | 1       | 0.2     |     |
|               | South Fork | 1         | 0.5      | 13(1.41) | 7.8      |             | N.F.Alder | 1         | 0       | 0.0     | 0       | 0.0 |
|               | Bull       |           | 16(5.94) | 8.6      | 18(1.43) |             | 17.6      | 2         | 0       | 0.0     | 0       | 0.0 |
| Coon          |            | 23(1.72)  | 11.8     | 51       | 34.3     | 3           | 0         | 0.0       | 0       | 0.0     |         |     |
| School House  |            | 8(0.77)   | 6.6      | 6        | 5.4      | 4           | 0         | 0.0       | 0       | 0.0     |         |     |
| West Fork     |            | 13(0.49)  | 10.0     | 48(1.26) | 43.1     | Totals      | 49(2.9)   |           | 74(3.7) |         |         |     |
| Whitetail     |            | 6.0       | 5.9      | 17       | 5.1      |             |           |           |         |         |         |     |
| Windfall      |            | 20(0.73)  | 10.8     | 79(3.4)  | 42.5     |             |           |           |         |         |         |     |
| Totals        |            | 106(2.7)  |          | 272(4.2) |          |             |           |           |         |         |         |     |

Table 3.23 Age frequency of cutthroat trout comparing summer and fall shocking results, 1997.

| SEASON              | LOCATION        | AGE |     |     |     |    | TOTAL |      |
|---------------------|-----------------|-----|-----|-----|-----|----|-------|------|
|                     |                 | 0   | 1   | 2   | 3   | 4  |       | 5    |
| SUMMER              | ALDER           | 4   | 4   | 24  | 14  | 2  |       | 48   |
|                     | BENEWAH         |     | 14  | 5   | 2   |    |       | 21   |
|                     | BOZARD          |     | 10  | 4   | 9   |    |       | 23   |
|                     | BULL            |     | 9   | 3   | 1   |    |       | 13   |
|                     | COON            | 1   | 5   | 9   | 2   | 1  | 4     | 22   |
|                     | E.F. BOZARD     |     | 7   |     | 4   |    |       | 11   |
|                     | E.F.EVANS       | 2   | 3   |     |     |    |       | 5    |
|                     | EVANS           | 2   | 26  | 43  | 38  | 7  | 1     | 117  |
|                     | LAKE            |     |     |     | 10  | 3  |       | 13   |
|                     | S.F.EVANS       | 1   | 4   | 7   |     |    |       | 12   |
|                     | SCHOOL HOUSE    |     | 2   | 5   | 1   |    |       | 8    |
|                     | W.F. BENEWAH    |     | 5   | 6   | 2   |    |       | 13   |
|                     | W.F. LAKE       |     | 5   | 2   | 3   | 3  | 1     | 14   |
|                     | WHITETAIL       |     | 5   | 1   |     |    |       | 6    |
|                     | WINDFALL        | 4   | 14  |     | 2   |    |       | 20   |
| <b>Summer Total</b> |                 | 14  | 113 | 109 | 88  | 16 | 6     | 346  |
| FALL                | ALDER           | 7   | 21  | 19  | 17  | 6  |       | 70   |
|                     | BENEWAH         |     | 7   | 34  | 8   | 3  |       | 52   |
|                     | BOZARD          | 27  | 13  | 8   | 7   |    |       | 55   |
|                     | BULL            | 10  | 1   | 5   | 1   |    |       | 17   |
|                     | COON            | 23  | 12  | 13  | 3   |    |       | 51   |
|                     | E.F. BOZARD     | 1   | 8   | 7   | 5   |    |       | 21   |
|                     | E.F.EVANS       | 14  | 1   | 7   |     |    |       | 22   |
|                     | EVANS           | 11  | 25  | 51  | 24  | 13 | 9     | 133  |
|                     | LAKE            | 1   | 11  | 7   | 14  | 13 | 8     | 54   |
|                     | RAINBOW F.EVANS |     | 6   | 8   | 3   |    |       | 17   |
|                     | S.F.EVANS       | 1   | 13  | 5   | 4   | 1  |       | 24   |
|                     | SCHOOL HOUSE    |     |     | 2   | 4   |    |       | 6    |
|                     | W.F. BENEWAH    | 32  | 5   | 6   | 4   |    |       | 47   |
|                     | W.F. LAKE       | 50  | 20  | 8   | 9   | 1  | 1     | 89   |
|                     | WHITETAIL       | 5   | 2   | 7   | 3   |    |       | 17   |
| WINDFALL            | 52              | 17  | 6   |     |     |    | 75    |      |
| <b>Fall Total</b>   |                 | 234 | 162 | 193 | 106 | 37 | 18    | 750  |
| <b>Grand Total</b>  |                 | 248 | 275 | 302 | 194 | 53 | 24    | 1096 |

Table 3.24 Brook trout abundance and distribution in Benewah Creek and Alder Creek, summer 1997.

| Benewah Creek |            |         |         | Alder Creek |         |           |         |         |      |
|---------------|------------|---------|---------|-------------|---------|-----------|---------|---------|------|
| Tributary     | Reach      | N±(SE)  | #/100m2 | Tributary   | Reach   | N±(SE)    | #/100m2 |         |      |
| Mainstem      | 1          | 0       | 0.0     | Mainstem    | 1       | 2         | 0.3     |         |      |
|               | 2          | 0       | 0.0     |             | 2       | 1         | 0.3     |         |      |
|               | 3          | 0       | 0.0     |             | 3       | 1         | 0.3     |         |      |
|               | 4          | 0       | 0.0     |             | 4       | --        | --      |         |      |
|               | 5          | 0       | 0.0     |             | 5       | 6         | 1.0     |         |      |
|               | 8          | 0       | 0.0     |             | 6       | 8(2.9)    | 4.3     |         |      |
|               | 9          | 0       | 0.0     |             | 7       | 42(3.6)   | 4.4     |         |      |
|               | 10         | 2       | 1.8     |             | 8       | 22(3.1)   | 7.9     |         |      |
|               | 11         | 2       | 0.5     |             | 9       | 27(1.1)   | 10.4    |         |      |
|               | South Fork |         | 2       |             | 1.0     | N.F.Alder | 1       | 45(5.7) | 10.5 |
|               | Bull       |         | 0       |             | 0.0     |           | 2       | 37(4.0) | 15.3 |
| Coon          |            | 0.0     | 0.0     | 3           | 44(5.1) |           | 18.2    |         |      |
| School House  |            | 0.0     | 0.0     | 4           | 9(2.1)  |           | 6.9     |         |      |
| West Fork     |            | 5(1.5)  | 3.8     | Totals      |         | 244(10.5) |         |         |      |
| Whitetail     |            | 0       | 0.0     |             |         |           |         |         |      |
| Windfall      |            | 0       | 0.0     |             |         |           |         |         |      |
| Totals        |            | 11(1.5) |         |             |         |           |         |         |      |

Table 3.25. Mean back calculated lengths (mm) at age for cutthroat trout, 1996.

| Location | Age |     |     |     |     |     |     |     |
|----------|-----|-----|-----|-----|-----|-----|-----|-----|
|          | 1+  | 2+  | 3+  | 4+  | 5+  | 6+  | 7+  | 8+  |
| Alder    | 83  | 118 | 148 | 192 |     |     |     |     |
| Benewah  | 83  | 121 | 158 | 207 | 250 | 299 | 318 |     |
| Cherry   | 86  | 135 | 176 | 214 | 256 | 298 |     |     |
| Evans    | 84  | 124 | 163 | 207 |     |     |     |     |
| Lake     | 85  | 131 | 171 | 222 | 264 | 308 | 335 | 366 |

A total of 139 brook trout scales were examined for age and growth in 1996 and an additional 51 scales were examined in 1997. The growth potential for brook trout in Alder Creek and Benewah Creek is similar (Table 3.26). Record size for Alder Creek brook trout is 13.4 inches (340 mm TL) and 0.7 pounds (306 grams), caught in 1997. Record size for Benewah Creek is 14.9 inches (380 mm TL) and 1.5 pounds (700 grams). Brook trout have not been identified in the Lake Creek or Evans Creek watersheds. Complete tabular analyses of growth for brook trout are provided in Appendix C.

Table 3.26. Mean back calculated lengths (mm) at age for brook trout, 1996.

| Location | Age |     |     |     |     |
|----------|-----|-----|-----|-----|-----|
|          | 1+  | 2+  | 3+  | 4+  | 5+  |
| Alder    | 84  | 120 | 151 | 198 | 227 |
| Benewah  | 88  | 127 | 166 | 215 | 268 |

The linear regressions of body length versus age for cutthroat trout and brook trout are shown in figures 3.9 and 3.10. For the subject watersheds, between 75 and 96 percent of the total variation in body length is explained by age.

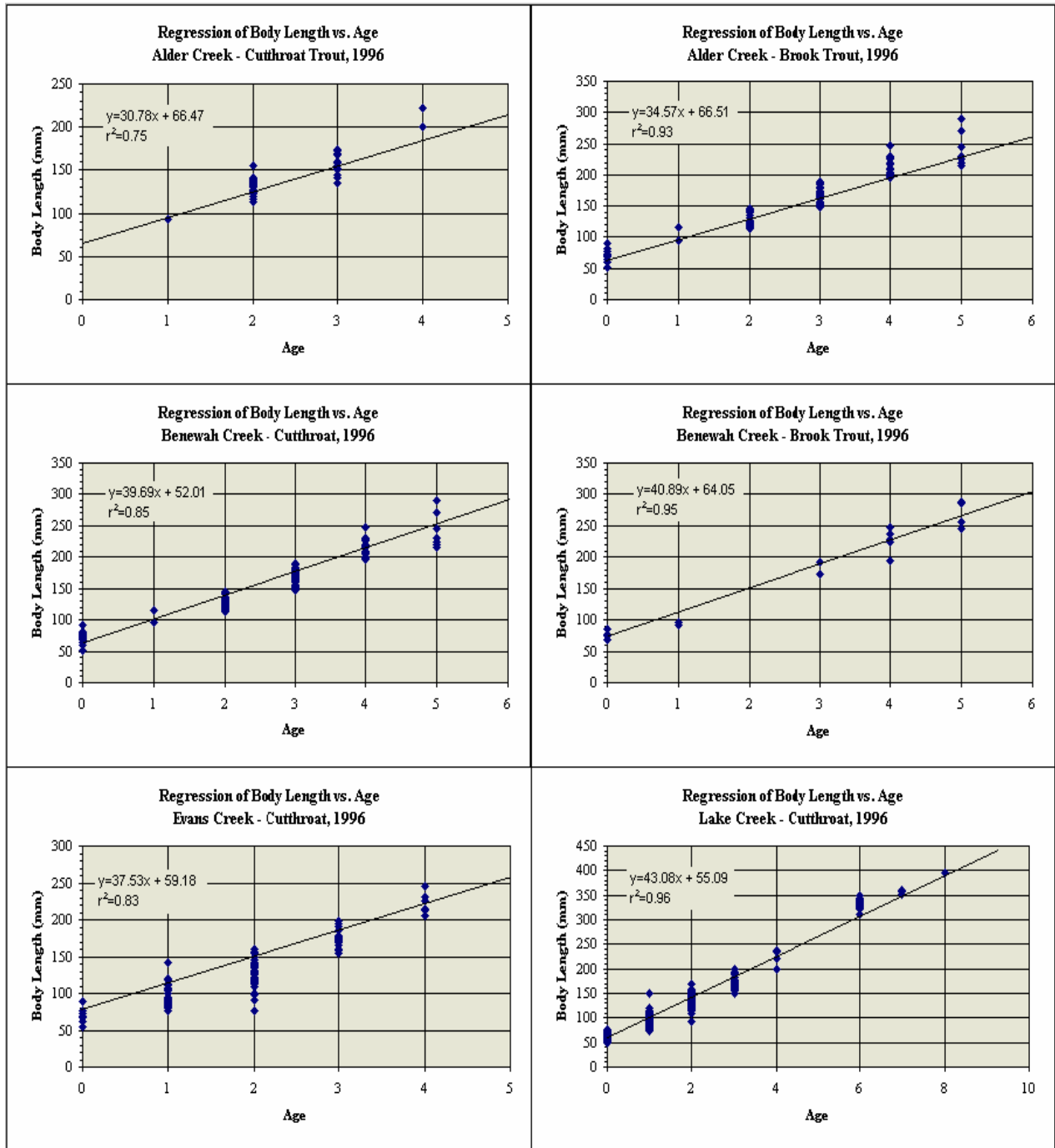


Figure 3.9 Regression equations of body length versus age for cutthroat and brook trout in the target watersheds, 1996.

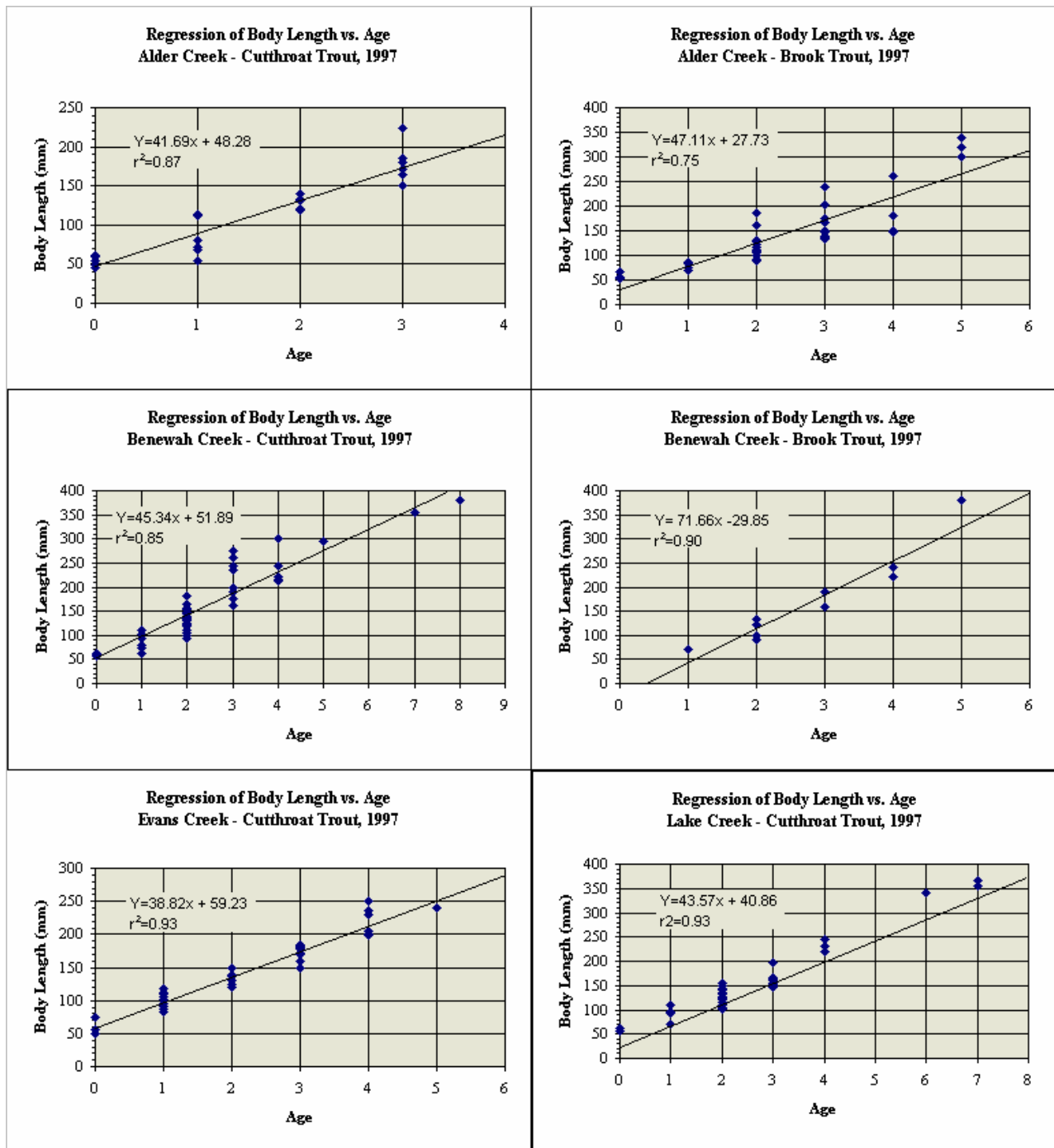


Figure 3.10 Regression equations of body length versus age for cutthroat and brook trout in the target watersheds, 1997.

## Trout Migration

A total of 907 cutthroat trout were caught in the lower Lake Creek trap in 1996, while only 273 were caught in 1997 (Figure 3.11-3.12). Adult fish (age IV or older) accounted for 3 percent of the catch each year (Figure 3.11-3.12). Although total numbers varied considerably among years, catch per unit effort was similar (12.4 fish/day and 7.8 fish/day, respectively). Measured stream flows in excess of 150 cubic feet/second during the period April 16-30, 1997 reduced the effectiveness of trapping efforts and, in part, account for the lower numbers of trout during that year.

Trapping success in Benewah Creek for the years 1996-97 was considerably lower than in Lake Creek. Only one cutthroat trout was caught in 1996 (0.04 fish/day) while a total of 26 were caught in 1997 (0.7 fish/day). Adult fish (age IV or older) accounted for 27 percent of the catch in 1997. Above normal precipitation and runoff greatly reduced the effectiveness of trapping efforts.

Upstream migration of adult fish into Benewah Creek and Lake Creek was documented for the period March 6 through April 20, 1997. Average daily water temperature in Lake Creek increased from 1° C to 8° C (mean = 3.8° C) during the documented period of upstream migration and stream discharge ranged from 57 cfs to 784 cfs. Measurements of suspended sediment indicated that concentrations rarely exceeded 1,000 mg/L during the migratory period for cutthroat trout in Lake Creek (Bauer, 1998). These concentrations are considerably lower than those reported to cause avoidance in migrating salmonids (Cordone and Kelley 1961, Bell 1986). Residence time for adults in both watersheds varied from 21 to 47 days, based on mark-recapture data (n=7) and radio telemetry data (n=1). Observed variation is likely a result of several factors, including spawning site selection and the availability of suitable spawning partners, as well as natural changes in flow, temperature, and turbidity.

The adaptation to migration in Coeur d'Alene basin westslope cutthroat trout appears to be more closely related to size rather than age. This has also been reported for coastal cutthroat (Johnson, 1982). Outmigrant behavior is observed predominantly in age II fish, with more than 70 percent of juvenile outmigrants consisting of age II cutthroat trout (Figures 3.11 and 3.12). These fish range in size from 9 to 54 grams and are 120 to 162 mm long. More than 70 percent of these fish are greater than 18 grams and 125mm long. Migration by smaller fish has been observed, however, mortality both from predation and physical damage probably increases with decreasing size and weight. The smallest migrating cutthroats weigh from 8 to 18 grams and are 90 to 119 mm long. It is postulated that frequent rain on snow events during early spring may be partially responsible for displacing higher proportions of age I fish, as was observed in Lake Creek in 1996.

Smolt movement to Lake Coeur d'Alene begins in late April to early May and is completed by mid-June. Peak outmigration is often correlated with small spikes in the hydrograph (Figures 3.11 and 3.12). Daily fluctuations in water temperature may also affect downstream migration. For example, downstream movement into traps has been shown to increase during the night when daytime water temperatures rise above 16°C (George Aripa, personal communication). Upstream migration into small tributaries by juvenile cutthroat was documented to occur as late as mid May in the Benewah Creek watershed. These migrants may be resident forms that were displaced from preferred habitat during spring runoff. Alternatively, the onset of increasing temperatures in the mainstem may elicit a migratory response in resident fish.

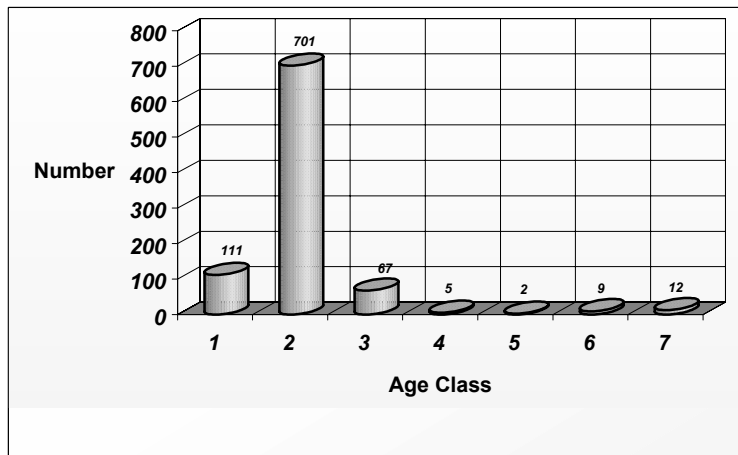
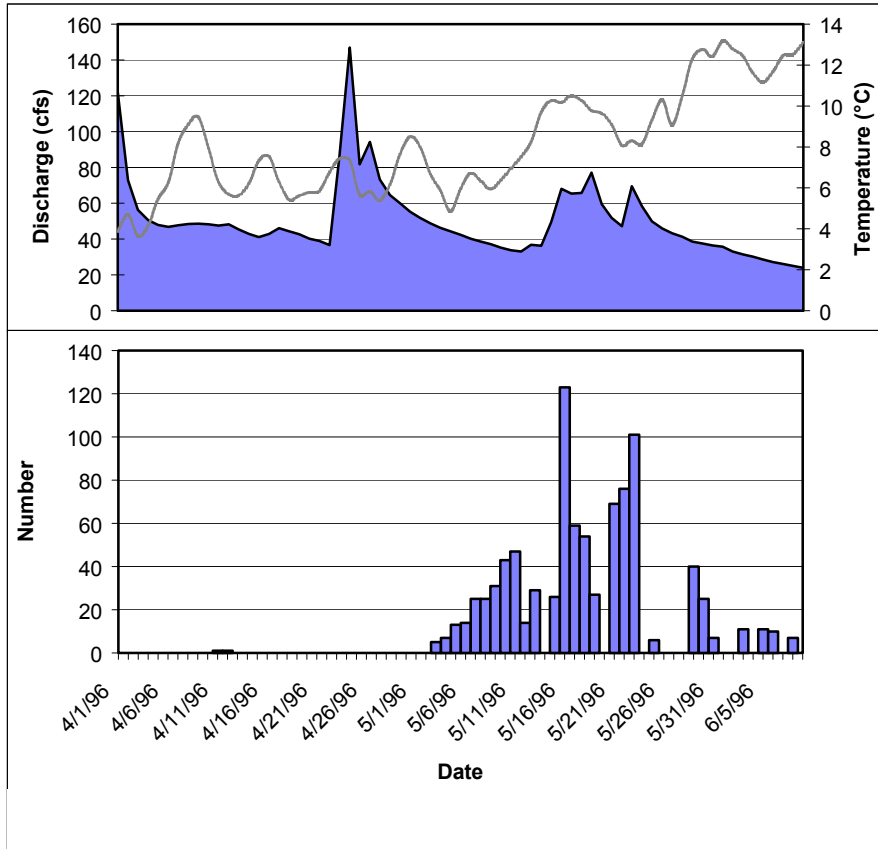


Figure 3.11 Analysis of cutthroat trout migration showing migration timing versus discharge and water temperature and age frequency of migrants in Lake Creek, 1996.

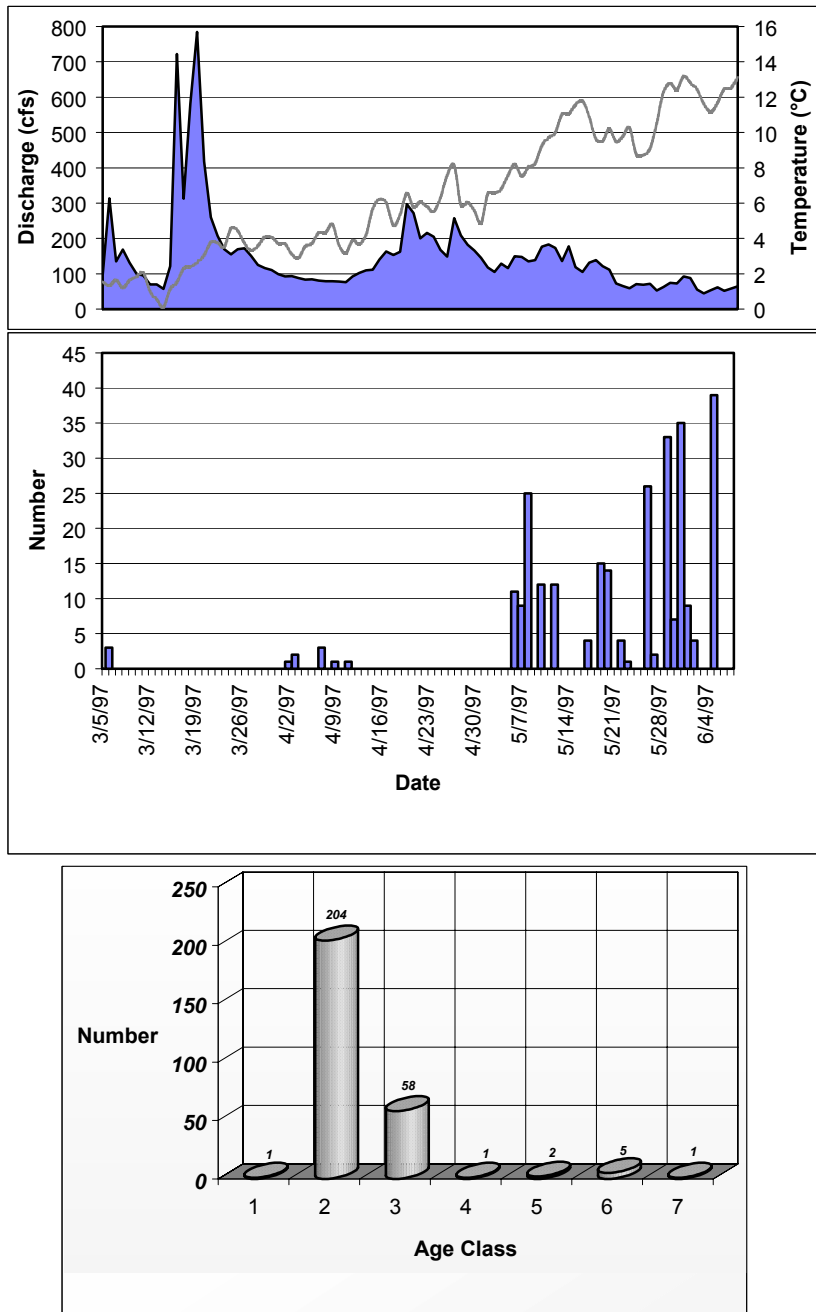


Figure 3.12 Analysis of cutthroat trout migration showing migration timing versus discharge and water temperature and age frequency of migrants in Lake Creek, 1997.



## 4.0 Discussion

### 4.1 Limiting Factors

#### Lake Studies

Water quality monitoring was completed on Coeur d'Alene Lake in order to assess its effect on cutthroat trout production. Earlier work has shown (Funk *et. al.*, 1973 and 1975; Horowitz *et. al.*, 1994 and Woods and Beckwith, 1995) that impacts from many different sources have caused a general decline in the water quality in Coeur d'Alene Lake. This, in turn, has had a detrimental effect on the cutthroat trout population in the lake. There is some evidence that the water quality may be improving (Woods, 1994) however, the possible legacy effects have yet to be determined.

Water Quality monitoring was completed at thirteen sites on Coeur d'Alene Lake. Multiple physical/chemical parameters were looked at in 1997. Of the parameters looked at it appears that only temperature is directly influencing the distribution of cutthroat trout in the lake. Dissolved oxygen, pH, conductivity, turbidity, total suspended solids, and dissolved heavy metals appeared to be within acceptable levels for cutthroat trout survival. Dissolved oxygen and suspended sediments though not directly impairing cutthroat suitability does affect overall water quality and impairs some use by cutthroat trout.

Low quantities of dissolved oxygen did occur at some of the sample sites, however, it is not considered limiting for cutthroat trout suitability. Low dissolved oxygen, however, is thought to have indirect affects on cutthroat trout suitability in the southern lakes area. Low dissolved oxygen values most likely are occurring from decomposition of organic matter from allochthonous sources as well as from aquatic macrophytes. Reiman (1980) and Woods (1989) noted hypolimnetic oxygen deficits in Coeur d'Alene Lake in 1979 and 1987 as well.

Increased loading of sediments from agricultural runoff does affect cutthroat suitability, though not directly, in areas near the mouths of streams in and around the lake. Sediment is accumulating at the mouth of Plummer creek in Chatcolet Lake at a rate of 2.4 cm/year (Breithaupt, 1990). This, in turn, increases the surface area where large masses of aquatic macrophytes can grow. These masses of aquatic plants can impair juvenile and adult migrations and serve as the primary foraging areas for larger piscivorous fishes.

The deposition of trace elements in the sediments of Coeur d'Alene Lake is well documented (Funk, 1973; Reiman 1980, Woods, 1989). Measured levels of dissolved metals did not appear to be directly affecting cutthroat suitability in Coeur d'Alene Lake. However, when coupled with increases in hypolimnetic oxygen deficits the potential for release of large quantities of the trace heavy metals becomes a real possibility. Currently, the principle means of controlling the levels of dissolved heavy metals in the waters of Coeur d'Alene Lake is keeping them bound up in the sediments of the lakebed. This means that managing the nutrient and sediment income in order to curtail the development of anaerobic conditions that would facilitate the release of these metals from the sediments is of paramount importance.

Nutrient and chlorophyll<sub>a</sub> levels did not appear to be significantly high when compared to values reported by Woods and Beckwith (1995) in 1991 and 1992. Analysis of the species composition of the algae by Woods and Beckwith in 1991 and 1992 showed very few, if any, species which would indicate a problem with eutrophication or a trend towards a change in trophic status in the main parts of Coeur d'Alene Lake. However, the presence of some blue-green algae in the southern lakes sample stations indicates that there is a change in trophic status towards eutrophic as you move along a southerly axis in the lake. This

would most likely have a detrimental effect on the distribution of cutthroat trout in those areas. Also, large masses of aquatic macrophytes are present in the southern chain lakes area which, in all probability, are utilizing large quantities of the available nutrients and keeping chlorophyll values lower than they would normally be with the same amount of nutrients. Breithaupt (1990) completed work in Chatcolet Lake that showed the highest peaks of total phosphorous occurred shortly after peak runoff, however, these events were not accompanied by corresponding changes in algae biomass.

Trophic state indices calculated in 1975 (U.S. EPA, 1977) classified Coeur d'Alene Lake as mesotrophic lakewide. Data collected in 1989 (Breithaupt, 1990) classified the southern lakes area as eutrophic during the peak runoff period and mesotrophic for the other times of the year. Woods (1994) classified Coeur d'Alene Lake as oligotrophic for all parameters except secchi disk transparency, which classified the lake as mesotrophic. Our data classified the lake as oligotrophic in the north and meso-eutrophic in the south with water quality parameters associated with eutrophic conditions increasing in a southerly direction.

Changes in the trophic state of the lake can also have dramatic effects on the succession of fish species within a lake. Hayward and Margraf (1987) showed rapid successional changes in the species structure of fish as a result of trophic status changes in Lake Erie. Leach *et al.* (1977) showed successional changes in response to trophic status changes, but at a much slower rate. As a lake becomes more oligotrophic following clean up or restoration it can be expected that some change in the fish species abundance will occur. Some species will be able to take advantage of changing conditions, while others will not. It is not surprising that salmonid populations have declined as a result of the eutrophication of Coeur d'Alene Lake from 1950-1970's. However, evidence suggests that in some portions of the lake this process may be reversing itself and slowly going towards oligotrophy again. This may mean that conditions which favored salmonid populations historically may return, however, the successional changes may not mirror the response to eutrophication. For example, other authors have shown that once cutthroat trout are replaced by another salmonid species (i.e. kokanee salmon), it is unlikely that space will be regained by cutthroat trout under natural conditions (Moyle and Vondracek, 1985).

Based on the water quality HSI's calculated for cutthroat trout, the upper 10 meters of the water column generally is not suitable habitat. At only one location was the HSI value higher than 0.0. This does not mean that cutthroat trout will not be found there, but they will have trouble sustaining themselves over a long period of time. Furthermore, the euphotic zone rarely drops below 10 meters so any foraging runs into that zone will take them into unsuitable habitat, which results in added stress. Thus, all areas represented by sample stations less than ten meters in depth would be considered unsuitable cutthroat trout habitat with deeper stations showing limited distribution during certain times of the year.

It does appear that improvements in the water quality of Coeur d'Alene Lake are occurring, however, water quality is still having a detrimental effect on habitat suitability for cutthroat trout. In addition to the direct affects of temperature, indirect effects related to low dissolved oxygen concentrations, total suspended solids, and large masses of aquatic macrophytes effectively limit the cutthroat trout population in the areas where these conditions exist. Specifically, these areas include the sample locations most affected by construction of Post Falls Dam; the two shallow stations, the three southern chain lake stations, and the three interior bay stations.

## **Fisheries**

There is no doubt that inter-specific species competition occurs between cutthroat trout and other fish species, especially the introduced ones (Griffith 1974, 1988; Marnell 1986, 1987, 1988; and others). Two mechanisms are controlling the population of cutthroat trout competitive exclusion and species replacement due to rapid changes in the environmental conditions within the lake. The extent that each individual mechanism controls the population has yet to be worked out. However, the fact that the

adfluvial population has not been extirpated from the lake shows that these fish have some resiliency to the detrimental effects from interactions with the introduced species. Petroskey and Bjornn (1985) demonstrated that cutthroat in the St. Joe River system show little detrimental effects from the introduction of hatchery reared rainbow trout. Griffith (1988) postulated that this resiliency may be attributed to the fact the cutthroat trout are not existing in habitat that is optimal for them but existing in habitat that is sub-optimal for the other species.

It is thought that juvenile cutthroat trout spend some time in the littoral zone just after they enter the lake from the tributaries where they would be subject to predation by these larger piscivorous fishes. It can be expected that some mortality will occur during this life stage. Insufficient data has been collected to determine exactly what effect this life stage mortality has on the overall population dynamics of the cutthroat trout in Coeur d'Alene Lake.

Of the introduced species the following have been shown to have the ability to actively feed on other fish species including adult and juvenile cutthroat trout: northern pike, largemouth bass, smallmouth bass, chinook salmon, and channel catfish. Historically, bull trout and northern squawfish were the only predators of cutthroat trout in the lake. Electrofishing data shows that these predators are associated primarily with the shoreline littoral zone. The relative abundance data shows that five species of piscivorous fishes (four introduced) have relative abundances higher than cutthroat trout. This would indicate that cutthroat trout are probably being competitively excluded from this littoral zone habitat by these other fishes.

Historically, cutthroat trout in Coeur d'Alene Lake probably utilized the littoral zone of the lake until they were large enough to move offshore and feed, most likely, on mid water prey and fish when available. Nilsson and Northcote (1981) noted that cutthroat trout in allopatry with other salmonids were found throughout the lake and in sympatry, they were located primarily in the littoral zone. It has been shown that introduction of kokanee salmon will also have detrimental effects on the cutthroat trout population (Gerstung, 1988; Marnell, 1988). Marnell (1988) determined that declines in westslope cutthroat trout populations in lakes in Glacier National Park where kokanee were introduced were caused by competition for planktivorous food. Thus, the introduction of non-native species into Coeur d'Alene Lake, at the minimum, altered the normal behavioral pattern of the cutthroat trout in both the littoral and limnetic zones of Coeur d'Alene Lake.

Based on the relative abundance information from 1994-1997 it appears that cutthroat trout are more successful in the limnetic zone than the littoral zone. In the limnetic zones with depths greater than 10 meters cutthroat trout were the third most abundant fish species caught. In the littoral zones of these same areas cutthroat trout were one of the least abundant species caught. Introduced species made up over 68% of the catch in relative abundance studies from 1994-1997 while cutthroat trout comprised less than 1% of the catch. In the littoral zones problems associated with temperature and inter-specific interactions are maximized. In the limnetic zone there is some relief from the effects of temperature however, problems associated with introduced species still exist. In relative abundance studies completed in the limnetic zones greater than ten meters deep from 1994-1997 introduced species (kokanee salmon) made up only 32% of the catch. There appears to be some association with the locations where cutthroat trout are caught in both the littoral and limnetic zones. It appears that in areas where fish are found in the limnetic zones they are also found in the littoral zones located nearby. This could mean that these fish are avoiding high temperatures in the upper waters by making foraging runs into the littoral zones during times when the water temperatures cool slightly at night. This could also be a predator avoidance mechanism as well.

## **Stream Studies**

Peak flows in Lake Creek and Benewah Creek have been identified in previous reports as a potential limiting factor for trout production (Lillengreen 1996). Generally, increased flows during egg incubation will be favorable until they reach the point when scouring and other flood damage may take place (Allen 1969). Spikes in stream discharge during the early spring, as is characteristic of the Lake Creek and Benewah Creek watersheds, may cause redd scouring and egg damage, although no attempt has been made to quantify this source of mortality. For example, stream flows in upper Lake Creek during spring of 1997 exceeded the sheer stress of spawning gravels (5 cm geometric mean particle diameter) for 4 consecutive days during the incubation period. It is conceivable that flow events of this magnitude could scour trout redds and result in complete year class failures. Although flood damage is a natural source of mortality, canopy reduction in each of the target watersheds has probably contributed to higher storm runoff peaks. Scouring of trout redds is certainly a more frequent occurrence than in the recent past.

Habitat availability in the target watersheds is not likely to be limiting for fry life stages. It has been demonstrated that young-of-the-year cutthroat trout are conspicuous inhabitants of slow-water areas near the margins of streams. Low velocity and low flow characterize these habitats. Moore and Gregory (1988) studied Cascade mountain streams and found that only about 35% of the total number of fry were observed at velocities greater than 1 cm/second, and no fry were observed at velocities greater than 15 cm/second. Focal depth for these fish increased rapidly from mid-August to early September, but the average depth was always less than 35 cm. Griffith (1972) reported similar focal points for cutthroat trout in north Idaho streams. These preferences of fry for depth and velocity are easily met during base flow conditions in the target tributaries.

This report demonstrates that abundance of juvenile cutthroat is greatest in first and second order tributaries, suggesting a close link to the most heavily utilized spawning areas. Downstream displacement, however, has been recognized as a common occurrence when stream flows approach zero in the principle spawning tributaries. While not being unique, this mechanism has not been commonly reported for most salmonid populations in the Pacific Northwest. For most salmonid species, it has been demonstrated that instream movement is minimal; individuals may remain in limited areas for several weeks or months and may return to the same locations in successive years (Edmundson et al. 1968; Bachman 1984). Limiting migration in this way is thought to confer an adaptive advantage by maximizing the net energy intake of individuals (Puckett and Dill 1985).

Typical base flow conditions in the target watersheds force juvenile trout into small pools where competition for limited space and food may occur. Other authors have suggested that at high densities, competition for space among juveniles may lead to dispersal, downstream displacement or mortality in salmonids (Chapman 1962; Mason and Chapman 1965; Everest 1971; Erman and Leidy 1975; LeCren 1973). In water quality limited systems, such as Lake Creek, Benewah Creek, and Alder Creek, dispersal to downstream areas exposes juvenile cutthroat to suboptimal temperature conditions that increase stress, weaken individuals and may result in mortality.

Cutthroat trout did not evolve with brook trout in the Benewah Creek and Alder Creek watersheds. Therefore, mechanisms that promote coexistence and resource partitioning have likely not developed. Griffith (1972) demonstrated that cutthroat trout fry emerge from the gravel later in the year than brook trout and, thus, age-0 cutthroat trout acquire a statistically significant length disadvantage that may continue throughout their lifetime. Such a size discrepancy may enhance resource partitioning, but in times of habitat shortage cutthroat trout may be at a disadvantage if they cannot hold territories against larger competitors. Competitive exclusion is a likely cause of decline for cutthroat trout in Alder Creek. Replacement of this kind, at least in stream environments, may be an irreversible process (Moyle and Vondracek 1985). This was found to be the case in Yellowstone National Park, where the introduction of brook trout has nearly always resulted in the disappearance of the cutthroat trout (Varley and Gresswell

1988). Implications for Benewah Creek are that cutthroat trout may have a difficult time recovering given continued water quality degradation and the persistence of brook trout.

Anecdotal evidence cited by residents (Ness personal communication; Hodgson personal communication) suggests that historic base flows in Lake Creek and Benewah Creek were often much lower than those considered optimal for trout production (Hickman and Raleigh, 1982; Binns and Eiserman 1979). This is most likely true for Alder Creek as well. This is not surprising considering the relatively small sizes of the target watersheds, the elevations, and the regional climate (Table 1.1). It is conceivable that base flow conditions provided the selective pressure that encouraged genetic differentiation of adfluvial stocks in the Coeur d'Alene basin.

In addition, much of the conversion of forested land to agricultural or pasture land, and removal of riparian canopy occurred prior to 1950. This suggests that stream temperatures have played a part in limiting cutthroat trout abundance and distribution within the target watersheds for at least 50 years. The range of suitable summer rearing habitat for cutthroat trout in each target watershed is significantly reduced when compared with the historic range of these fish. HSI calculations published in this report, however, indicate that improving habitat condition through restoration and protection has the potential to partially compensate for short-term degradation in water quality. In considering this information, it appears that the native strains of westslope cutthroat trout are adapted to local conditions and display a high degree of resiliency.

## **4.2     Supplementation Feasibility**

### **Need For Supplementation**

Due to the persistence of adverse conditions in natal streams and Coeur d'Alene Lake, cutthroat trout populations are thought to be at least moderately damaged (i.e. average spawning escapements fall between the minimum viable population and the number of adults needed to produce 50% of the carrying capacity of the stream environment). The following factors effectively limit the population of cutthroat trout in natal streams and in Coeur d'Alene Lake:

- Stochastic events that result in increased mortality of embryo, fry, and juvenile lifestages (e.g. peak flow events) have been exacerbated by land use practices during the last 60 years;
- Competition for limited space and food during base flow conditions cause displacement of juveniles into water quality limited stream reaches;
- Competitive interactions with introduced salmonids may result in replacement of native trout in Alder Creek and Benewah Creek;
- Water temperatures in the upper ten meters of the water column in Coeur d'Alene Lake exceed the optimum as described in the HSI for cutthroat trout;
- Sediment loading from tributaries in combination with large quantities of aquatic macrophyte growth and low dissolved oxygen concentrations in the hypolimnion promote conditions more favorable for introduced fish species in the lake; and
- Competitive interactions with introduced species for food, living space, and through predation limit cutthroat trout in both the littoral and limnetic zones of Coeur d'Alene Lake.

These conclusions are supported by work that was conducted previously by other investigators, as well as, work conducted by the Coeur d'Alene Tribe Fish/Water/Wildlife Program. Based on the analysis of

these limiting factors, gains from habitat restoration and protection alone will not achieve the goals set forth by the Coeur d'Alene Tribal Council to provide for self-sustaining populations and harvestable numbers of cutthroat trout.

### **Definition**

For the purposes of this document, supplementation is defined as the stocking of fish into the natural habitat to increase the abundance of naturally reproducing fish populations. Maintaining the long-term genetic fitness of the target population, while keeping the ecological and genetic impacts on non-target populations within acceptable limits, is inherent in this working definition.

On the Coeur d'Alene Indian Reservation, supplementation activities would involve stocking fish into habitats that contain depressed but existing natural fish populations. Unlike many traditional hatchery programs, the objective of supplementation here is to increase the abundance of a naturally reproducing fish populations and therefore, is oriented toward maintaining the natural biological characteristics of the population and reliance on the rearing capabilities of the natural habitat. Supplementation measures will not obviate the need to concurrently pursue other necessary actions such as habitat protection and improvement, and harvest management to rebuild stocks.

### **Proposed Use of Supplementation**

The objective of supplementation will be to boost the population density above a certain minimum viable population size as quickly as possible. Minimum viable spawning escapement size for each stock will be calculated from the minimum effective breeding number by a transfer function, whose elements include the amount of spawning and rearing habitat available and the average total mortality. These values are currently under investigation and will be needed to develop an initial management strategy. The concept is to employ a supplementation program to a level that minimizes the risk of extirpation.

The primary role of supplementation in this case is to increase the survival rate of the population during its early life history (egg through smolt) relative to its survival rate under natural conditions. It is anticipated that this effort will result in increased adult returns to seed sparsely populated habitats and provide for limited harvest opportunities.

For depressed stocks that exist in tributaries of Coeur d'Alene Lake, the question of how many or what proportion of the natural stock to intercept for broodstock can only be resolved by careful evaluation of the impact of initially taking a small fraction of the depressed population for broodstock. An important consideration will be to ensure that capture of broodstock does not reduce the number of naturally spawning adults required to maintain the minimum viable population.

### **Approach To Supplementation**

Supplementation has been a common strategy for increasing natural fish production in the Columbia River Basin. However, there is not yet a detailed understanding of which techniques work best under which circumstances (Cuenco, et. al. 1993). A phased implementation and adaptive management approach will be used to identify uncertainties and design experiments to facilitate the development of construction alternatives. A flow chart identifying critical implementation steps and unmet needs has been developed (Figure 4.1). Three distinct levels comprise this phased approach. Level I identifies the need for supplementation through limiting factor analysis and projections of stock status. This report completes the level I tasks and describes the purpose and need for supplementation.

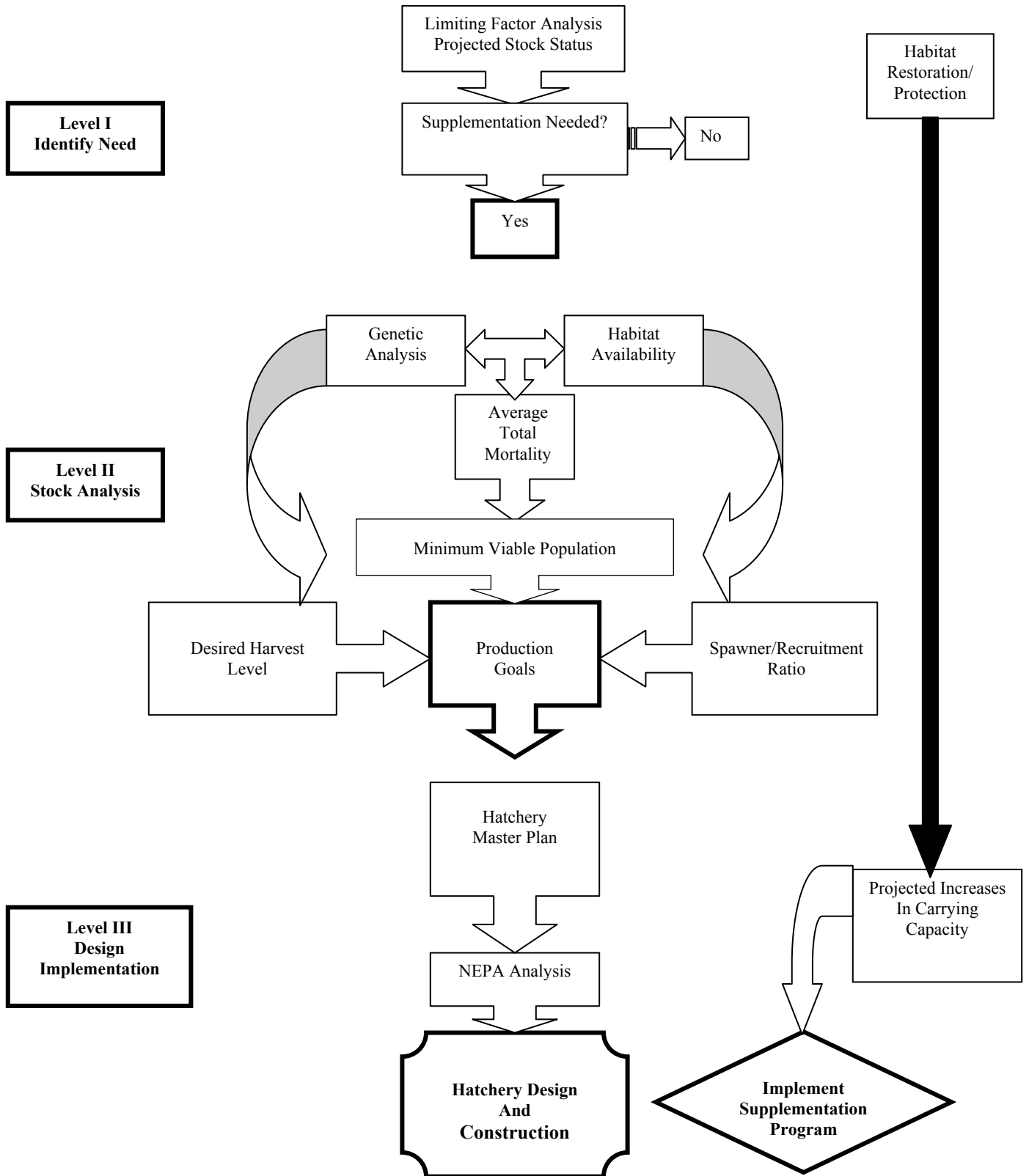
Level II is a comprehensive stock analysis with the ultimate purpose of developing production goals that will maintain the minimum viable population, while providing for some harvest opportunities.

Information gained through genetic analysis, habitat quantification, and estimates of average total mortality throughout the lifecycle will determine minimum viable populations for the target watersheds. Several critical questions must be answered to determine overall production goals. First, does sufficient

natural habitat exist, capable of supporting a viable, self-perpetuating population in the face of natural stochastic events (e.g., floods, droughts, etc.) demographic factors (e.g., ability to find a mate, sex ratio, age structure), and genetic considerations (prevent loss of genetic variation). This is important to ensure that the carrying capacity of the habitat does not become a limiting factor in population recovery. Secondly, are suitable stocks available for supplementation. To maximize the chances of success, indigenous stocks should be used for broodstock in their own enhancement program. This will ensure that important biological traits, such as spawning, incubation, and emergence timing are synchronized with favorable environmental conditions. Finally, are sufficient numbers of spawning adults available for maintenance of the minimum viable population and subsequent supplementation needs, which will provide for desired levels of harvest.

Level III outlines interim steps from master plan development to hatchery construction and implementation of the supplementation program. Information from the stock analysis will be used to develop a master plan that will identify supplementation strategies, project future harvest returns, provide a cost-benefit analysis, locate an appropriate construction site, and address the requirements of the National Environmental Policy Act (NEPA) for hatchery construction. Ongoing habitat restoration and protection will provide for increases in carrying capacity that will facilitate this phased implementation approach.

The concept of supplementation is still relatively new and uncertainties still exist about its effectiveness and safety. Monitoring and evaluation will be used to assess the performance and degree of success for each supplemented stock. The results will be used to guide other proposed supplementation projects. The following elements will be included in the monitoring and evaluation program: clearly defined, quantifiable objectives; performance measures for each objective; an experimental design which will facilitate the decision making process by allowing for adaptive management.



**Figure 4.1 Supplementation flow chart.**



## 5.0 Bibliography

- American Public Health Association. 1992. Standard methods for the examination of water and wastewater. Victor Graphics, Inc. Baltimore Maryland. 18<sup>th</sup> Ed.
- Alaska Department of Fish and Game. 1987. Limnology field and laboratory manual: Methods for assessing aquatic production. Juneau, Alaska. P. 212.
- Allen, K.R. 1969. Limitations on production in salmonid populations in streams. In: T.G. Northcote (ed.) Symposium on salmon and trout in streams. University of British Columbia, Vancouver. Pp 3-18.
- Aripa, G. 1997. Personal communication. Fisheries technician supervisor, Coeur d'Alene Tribe.
- Armour, C.L., K.P. Burnham, and W.S. Platts. 1983. Field methods and statistical analyses for monitoring small salmonid streams. USDI, Fish and Wildlife Service. FWS/OBS-83/33.
- Bachman, R.A. 1984. Foraging behavior of free-ranging wild and hatchery brown trout in a stream. Transactions of the American Fisheries Society 113:1-32.
- Bauer, S.B. 1998. Lake Creek 1996 interim monitoring report, Kootenai County, Idaho. Kootenai-Shoshone Soil Conservation District. Coeur d'Alene, Idaho. 27p.
- Beamesderfer, R.C. and B.E. Rieman. 1991. Abundance and distribution of northern squawfish, walleyes, and smallmouth bass in John Day reservoir, Columbia River. Transactions of the American Fisheries Society 120:439-447.
- Bell, M.C. 1986. Fisheries handbook of engineering requirements and biological criteria. U.S. Army Corps of Engineers, Office of the Chief of Engineers, Fish Passage Development and Evaluation Program, Portland, Oregon.
- Binns, N.A. and F.M. Eiserman. 1979. Quantification of fluvial trout habitat in Wyoming. Transactions of the American Fisheries Society 108:215-228.
- Breithaupt, S.A. 1990. Plummer Creek and Chatcolet Lake Benewah and Kootenai Counties. Water Quality Status Report No. 94. Idaho Dept. of Health and welfare Division of Environmental Quality Water Quality Bureau.
- Carlander, K.D. 1981. Caution on the use of the regression method of back-calculating lengths from scale measurements. Fisheries 6:2-4.
- Chapman, D.W. 1962. Aggressive behavior in juvenile coho salmon as a cause of emigration. Journal of the Fisheries Research Board of Canada 19:1047-1080.
- Conlin, K. and B.D. Tutty. 1979. Juvenile salmon field trapping manual. Dept. of Fisheries and Oceans. Fisheries and Marine Service Resource Services Branch, Habitat Protection Division, Vancouver, B.C. 136p.
- Cordone, A.J. and D.W. Kelley. 1961. The influences of inorganic sediment on the aquatic life of streams. California Fish and Game 47:189-228.
- Cuenca, M.L., T.W.H. Backman and P.R. Mundy. 1993. *in*: J.G. Cloud and G.H. Thorgaard (eds.), Genetic Conservation of Salmonid Fishes. Plenum Press, New York, New York.
- Edmundson, E., F.E. Everest, and D.W. Chapman. 1968. Permanence of station in juvenile chinook salmon and steelhead trout. Journal of the Fisheries Research Board of Canada 25:1453-1464.
- Ellis, M.N. 1932. Pollution of the Coeur d'Alene River and adjacent waters by mine wastes. U.S. Bureau of Fisheries. Mimeo Report. 55p.
- Environmental Protection Agency 1977. Report on Coeur d'Alene Lake Benewah and Kootenai Counties Idaho. National Eutrophication Survey Working Paper no. 778. 20p. 5 appendices.
- Erman, D.C. and G.R. Leidy. 1975. Downstream movement of rainbow trout fry in a tributary of Sagehen Creek, under permanent and intermittent flow. Transactions of the American Fisheries Society 104:467-474.

- Everest, F.H., Jr. 1971. An ecological and fish cultural study of summer steelhead in the Rogue River, Oregon. Annual Progress Report Anadromous Fish Project. AFS-31. Oregon State Game Commission.
- Funk, W.H., Rabe, F.W., and Filby, Royston. 1973. The biological impact of combined metallic and organic pollution in the Coeur d'Alene-Spokane River drainage system. Water and Energy Resources Research Institute. Moscow, Idaho. 202p.
- Funk, W.H., Rabe, F.W., Filby, Royston, Bailey, Gary, Bennet, Paul, Shah, Kisher, Sheppard, J.C., Savage, N.L. Bauer, S.B., Bourg, A.C.M., Bannon, Gerald, Edwards, George, Anderson, Dale, Syms, Pat, Rother, Jane, and Seamster, Alan. 1975. An integrated study on the impact of metallic trace element pollution in the Coeur d'Alene- Spokane River drainage system. Water Research Institute. Moscow, Idaho. 332p.
- Gardiner, V. 1990. Drainage basin morphometry. In: Goudie, A.S. (ed.), Geomorphological Techniques. London, Unwin Hyman.
- Gerstung, E.R. 1988. Status, life history, and management of Lahontan cutthroat trout. American Fisheries Society Symposium 4:93-106.
- Griffith, J.S. 1972. Comparative behavior and habitat utilization of brook trout (*Salvelinus fontinalis*) and cutthroat trout (*Salmo clarki*) in small streams in northern Idaho. Journal of the Fisheries Research Board of Canada 29:265-273.
- \_\_\_\_\_. 1974. Utilization of invertebrate drift by brook trout (*Salvelinus fontinalis*) and cutthroat trout (*Salmo clarki*) in small streams in Idaho. Transactions of the American Fisheries Society 103:440-447.
- \_\_\_\_\_. 1988. Review of competition between cutthroat trout and other salmonids. American Fisheries society Symposium 4:134-140.
- Hayward, R.S. and F.J. Margraf. 1987. Eutrophication effects on prey size and food available to yellow perch in Lake Erie. Transactions of the American Fisheries Society 116:210-223.
- Hickman, T. and R.F. Raleigh. 1982. Habitat suitability index models: Cutthroat trout. USDI, Fish and Wildlife Service. FWS/OBS-82/10.5. 38 pp.
- Hodgson, T. 1998. Personal communication. Board member, Benewah Valley Association.
- Horowitz, A.J., Elrick, K.A. Robbins, J.A., and Cook, R.B. 1994. Effect of mining and related activities on the sediment trace element geochemistry of Coeur d'Alene Lake, Idaho, USA. Part II subsurface sediments Hydrological Processes. V. 8 pg.
- Huber, W.A. 1983. Passive capture techniques. pgs. 95-122 in L.A. Nielson and D.L. Johnson editor. Fisheries Techniques, American Fisheries Society. Southern Printing Co., Inc. Blacksburg, Virginia. pgs. 468.
- Hydrolab Corporation. 1991. Hydrolab Multiparameter Water Quality Monitoring Instruments, 63 pages.
- Jearld, T. 1983. Age determination. in: Nielsen, L.A. and D.L. Johnson (eds.), Fisheries Techniques. American Fisheries Society, Bethesda, MD. 468p.
- Johnston, J.M. 1982. Life histories of anadromous cutthroat with emphasis on migratory behaviour. Pages 123-127 in Brannon and Salo (1982).
- Krueger, E. 1998. Lake Creek watershed assessment (Draft). Coeur d'Alene Tribe Fish, Water, and Wildlife Program. 27p.
- LeCren, E.D. 1973. The population dynamics of young trout in relation to density and territorial behavior. Rapports P.V. Reun. Cons. Perm. Int. Explor. Mer. 164:241-246.
- Lillengreen, K.L., Tami Skillingstad, and Allen T. Scholz. 1993. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation. Bonneville Power Administration. Division of Fish and Wildlife. Portland Or. Project # 90-44. 218p.
- Lillengreen, K.L., A.J. Vitale, and R. Peters. 1996. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian reservation, 1993-1994 annual report. USDE, Bonneville Power Administration, Portland, OR. 260p.

- Lux, F.E. 1971. Age determination of fishes (revised). NMFS, Fishery leaflet 637. 7p.
- Mallet, J. 1969. The Coeur d' Alene Lake fishery. Idaho Wildlife Review. May-June, 1969. 3-6 p.
- Marnell, L.F. 1986. Impacts of hatchery stocks on wild fish populations. Pages 339-347 in R.H. Stroud, editor. Fish culture in fisheries management. American Fisheries Society, Fish Culture Section and Fisheries Management Section, Bethesda, Maryland.
- \_\_\_\_\_. 1988. Status of westslope cutthroat trout on Glacier National Park, Montana. American Fisheries Society Symposium 4:61-70.
- Marnell, L.F., R.J. Behnke, and F.W. Allendorf. 1987. Genetic identification of cutthroat trout (*Salmo clarki*) in Glacier National Park, Montana. Canadian Journal of Fisheries and Aquatic Sciences 44:1830-1839.
- Mason, J.C. and D.W. Chapman. 1965. Significance of early emergence, environmental rearing capacity, and behavioral ecology of juvenile coho salmon in stream channels. Journal of the Fisheries Research Board of Canada 22:173-190.
- Miranda, L.E., W.D. Hubbard, S. Sangare, and T. Holman. 1996. Optimizing electrofishing sample duration for estimating relative abundance of largemouth bass in reservoirs. North American Journal of Fisheries Management. 16:324-331.
- Moyle, P.B. and B. Vondracek. 1985. Persistence and structure of the fish assemblage in a small California stream. Ecology 66:1-13.
- Ness, E. 1998. Personal Communication. Lake Creek watershed resident.
- Nisson, N.A. and T.G. Northcote. 1981. Rainbow trout (*Salmo gairdneri*) and cutthroat trout (*S. clarki*) interactions in coastal British Columbia lakes. Canadian Journal of Fisheries and Aquatic Sciences. 38:1228-1246.
- Novoteny, R.G. and G.R. Priegal. 1974. Electrofishing boats: Improved designs and operational guidelines to increase effectiveness of boom shockers. Wisc. Dept. Nat. Resour. Tech. Bull. No 73. 48p.
- Oien, W.E. 1957. A pre-logging inventory of four trout streams in northern Idaho. M.S. Thesis. University of Idaho. Moscow, Idaho. 92p.
- Petrosky, C.E. and T.C. Bjornn. 1985. Competition from catchables- a second look. Pg. 63-68 in F. Richardson and R.H. Hamre, editors. Wild trout II. Federation of Fly Fisherers and Trout Unlimited, West Yellowstone, Montana.
- Puckett, K.J. and L.M. Dill. 1985. The energetics of feeding territoriality in juvenile coho salmon (*Oncorhynchus kisutch*). Behavior 92:97-110.
- Reynolds, J.B. 1983. Electrofishing. in: Nielsen, L.A. and D.L. Johnson (eds.), Fisheries Techniques. American Fisheries Society, Bethesda, MD. 468p.
- Rieman, B.E. 1977. Coeur d'Alene Lake limnology. IDFG Lake and Reservoir Investigations Job Performance Report F-72-R-2 pg. 27-68.
- Rosgen, D.L. 1994. A classification of natural rivers. Catena 22:169-199.
- Scholz, A.T., D.R. Geist, and J.K. Uehara. 1985. Feasibility report on restoration of Coeur d'Alene Tribal Fisheries. Upper Columbia United Tribes Fisheries Center. Cheney, WA. 85 pp.
- Seber, G.A.F., and E.D. LeCren. 1967. Estimating population parameters from catches large relative to the population. Journal Animal Ecology 36:631-643.
- U.S. Department of Agriculture. 1984. General soil map and landform provinces of Idaho: Boise, Idaho. Soil Conservation Service. 1p.
- Varley, J.D. and R.E. Gresswell. 1988. Status, ecology, and management of the Yellowstone cutthroat trout. American Fisheries Society Symposium 4:13-24.

- Weaver, T.M. and J.J. Fraley. 1993. A method to measure emergence success of westslope cutthroat trout fry from varying substrate compositions in a natural stream channel. *North American Journal of Fisheries Management* 13:817-822.
- Wolman, G.M. 1954. A method of sampling coarse river-bed material. *Transactions, American Geophysical union* 35:951-956.
- Woods, P.F. 1989. Hypolimnetic concentrations of dissolved oxygen, nutrients, and trace elements. USGS- Water Resources Investigations Report 89-4032. 56p.
- \_\_\_\_\_. 1994. Shift in trophic state of Coeur d'Alene Lake, Idaho, 1975-92. State of Washington Water Research Center Report 89. pgs. 247-250.
- Woods, P. F. and C. Barenbrock. 1994. Bathymetric map of Coeur d'Alene Lake, Idaho. U.S. Geological Survey Water Resources Investigations Report 94-xxx. 1p.
- Woods, P.F. and M.A. Beckwith. 1995. Nutrient and trace element enrichment of Coeur d'Alene Lake, Idaho. USGS Open File Report 95-740.
- Zippen, C. 1958. The removal method of population estimation. *Journal of Wildlife Management* 22:82-90.

## **Appendix A**

Vertical hydrolab profiles of thirteen stations on Coeur d'Alene Lake, 1997.

| Location     | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|--------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 11       | 0.4       | 12.58                   | 5.64            | 7.34 | 42.4                 | 0.03 | *****              | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 10       | 1.3       | 12.51                   | 5.42            | 7.33 | 42.2                 | 0.03 | 99                 | -       | 373   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 9        | 2.6       | 12.54                   | 5.24            | 7.34 | 41.9                 | 0.03 | 98.8               | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 8        | 4         | 12.48                   | 4.98            | 7.32 | 41.6                 | 0.03 | 97.6               | -       | 373   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 7        | 5.4       | 12.5                    | 4.94            | 7.31 | 41.8                 | 0.03 | 97.7               | -       | 373   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 6        | 6.7       | 12.47                   | 4.87            | 7.31 | 42.5                 | 0.03 | 97.3               | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 5        | 7.8       | 12.47                   | 4.81            | 7.31 | 42.7                 | 0.03 | 97.2               | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 4        | 9         | 12.48                   | 4.74            | 7.3  | 43.2                 | 0.03 | 97                 | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 3        | 10.3      | 12.47                   | 4.69            | 7.29 | 43.2                 | 0.03 | 96.9               | -       | 372   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 2        | 11.8      | 12.79                   | 4.65            | 7.29 | 43.4                 | 0.03 | 99.2               | -       | 370   |
| Rockford Bay | 1597  | ASSS    | 4/18/97 | -    | 1.7    | 1        | 13        | 12.19                   | 4.61            | 7.28 | 43.2                 | 0.03 | 94.4               | -       | 369   |
|              |       |         |         |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 11       | 0.2       | 11.36                   | 13.57           | 7.09 | 36.1                 | 0.02 | 108.4              | -       | 409   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 10       | 2         | 11.31                   | 10.51           | 7.09 | 34.2                 | 0.02 | 100.6              | -       | 412   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 9        | 3.4       | 11.39                   | 10.46           | 7.08 | 34.7                 | 0.02 | 101.3              | -       | 412   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 8        | 5         | 11.54                   | 10.41           | 7.06 | 37.4                 | 0.02 | 102.4              | -       | 413   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 7        | 6.5       | 11.53                   | 10.22           | 7.05 | 37.5                 | 0.02 | 101.9              | -       | 413   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 6        | 8         | 11.53                   | 10              | 7.05 | 38                   | 0.02 | 101.3              | -       | 414   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 5        | 9.5       | 11.49                   | 9.74            | 7.05 | 37.6                 | 0.02 | 100.3              | -       | 413   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 4        | 11        | 11.43                   | 9.55            | 7.04 | 37.6                 | 0.02 | 99.4               | -       | 413   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 3        | 12.5      | 11.37                   | 9.32            | 7.02 | 38.2                 | 0.02 | 98.3               | -       | 414   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 2        | 14        | 11.27                   | 8.83            | 7.01 | 38.7                 | 0.02 | 96.3               | -       | 414   |
| Rockford Bay | 1997  | DBSS    | 5/16/97 | -    | 1.3    | 1        | 15.4      | 11.21                   | 7.66            | 7.01 | 39.9                 | 0.03 | 93.1               | -       | 414   |
|              |       |         |         |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Rockford Bay | 2197  | DBSS    | 5/29/97 | -    | 1.3    | 15       | 0.6       | 10.57                   | 16.8            | 6.91 | 32.1                 | 0.02 | 108.4              | -       | 387   |
| Rockford Bay | 2197  | DBSS    | 5/29/97 | -    | 1.3    | 14       | 1.6       | 10.6                    | 16.43           | 6.88 | 32                   | 0.02 | 107.9              | -       | 388   |
| Rockford Bay | 2197  | DBSS    | 5/29/97 | -    | 1.3    | 13       | 2.5       | 10.42                   | 12.02           | 6.82 | 32.5                 | 0.02 | 96.2               | -       | 393   |
| Rockford Bay | 2197  | DBSS    | 5/29/97 | -    | 1.3    | 12       | 3.5       | 10.75                   | 11.31           | 6.76 | 31                   | 0.02 | 97.7               | -       | 395   |

|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 11 | 4.6  | 10.54 | 10.92 | 6.73 | 30.8 | 0.02 | 94.8  | - | 396 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 10 | 5.5  | 10.49 | 10.36 | 6.72 | 30.4 | 0.02 | 93.2  | - | 396 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 9  | 6.6  | 10.49 | 10.18 | 6.7  | 30.4 | 0.02 | 92.8  | - | 397 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 8  | 7.6  | 10.62 | 9.96  | 6.69 | 29   | 0.02 | 93.4  | - | 397 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 7  | 8.7  | 10.61 | 9.82  | 6.66 | 30.2 | 0.02 | 93.1  | - | 397 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 6  | 9.7  | 10.63 | 9.71  | 6.64 | 30.2 | 0.02 | 93    | - | 398 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 5  | 10.6 | 10.62 | 9.69  | 6.62 | 30.4 | 0.02 | 92.8  | - | 398 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 4  | 11.6 | 10.51 | 9.64  | 6.6  | 31.3 | 0.02 | 91.9  | - | 400 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 3  | 12.6 | 10.51 | 9.36  | 6.57 | 32   | 0.02 | 91.1  | - | 401 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 2  | 13.6 | 10.5  | 9.12  | 6.53 | 32.8 | 0.02 | 90.5  | - | 401 |
| Rockford Bay | 2197 | DBSS | 5/29/97 | - | 1.3 | 1  | 14.8 | 10.53 | 8.63  | 6.48 | 34   | 0.02 | 89.7  | - | 404 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 14 | 0.6  | 10.87 | 14.41 | 7.14 | 34.4 | 0.02 | 106.3 | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 13 | 1.5  | 11.11 | 12.51 | 7.13 | 33.1 | 0.02 | 104.1 | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 12 | 2.5  | 11.13 | 12.08 | 7.13 | 32.4 | 0.02 | 103.1 | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 11 | 3.4  | 11.09 | 11.62 | 7.13 | 32   | 0.02 | 101.8 | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 10 | 4.6  | 10.9  | 10.8  | 7.13 | 30.9 | 0.02 | 98.1  | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 9  | 5.5  | 10.81 | 10.41 | 7.14 | 30   | 0.02 | 96.5  | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 8  | 6.4  | 10.72 | 10.33 | 7.14 | 29.6 | 0.02 | 95.5  | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 7  | 7.6  | 10.65 | 10.13 | 7.14 | 29.9 | 0.02 | 94.4  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 6  | 8.5  | 10.69 | 10.08 | 7.15 | 30.1 | 0.02 | 94.6  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 5  | 9.7  | 10.68 | 9.97  | 7.15 | 30.1 | 0.02 | 94.3  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 4  | 10.6 | 10.66 | 9.92  | 7.16 | 30   | 0.02 | 94.1  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 3  | 11.7 | 10.64 | 9.92  | 7.18 | 30   | 0.02 | 93.8  | - | 413 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 2  | 12.6 | 10.63 | 9.76  | 7.19 | 30.5 | 0.02 | 93.4  | - | 412 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 1  | 13.6 | 10.2  | 10.12 | 7.24 | 30.5 | 0.02 | 90.3  | - | 411 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 14 | 0.2  | 10.64 | 16.29 | 7.11 | 33.8 | 0.02 | 108.2 | - | 442 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 13 | 0.9  | 10.66 | 16.23 | 7.1  | 33.9 | 0.02 | 108.3 | - | 443 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 12 | 1.9  | 10.66 | 16.04 | 7.08 | 33.8 | 0.02 | 107.8 | - | 444 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 11 | 2.9  | 10.63 | 15.71 | 7.07 | 33.6 | 0.02 | 106.7 | - | 444 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 10 | 3.9  | 10.7  | 15.04 | 7.05 | 33.5 | 0.02 | 105.9 | - | 446 |

|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 9  | 4.9  | 10.78 | 14.13 | 7.02 | 33.1 | 0.02 | 104.6 | - | 447 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 8  | 6    | 10.78 | 13.55 | 7    | 33.1 | 0.02 | 103.3 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 7  | 7    | 10.77 | 13.29 | 7    | 32.9 | 0.02 | 102.6 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 6  | 8    | 10.73 | 13.09 | 6.98 | 32.8 | 0.02 | 101.7 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 5  | 8.9  | 10.78 | 12.96 | 6.96 | 32.8 | 0.02 | 101.9 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 4  | 10   | 10.76 | 12.86 | 6.94 | 32.8 | 0.02 | 101.5 | - | 449 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 3  | 11   | 10.69 | 12.77 | 6.91 | 32.8 | 0.02 | 100.7 | - | 450 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 2  | 12   | 10.55 | 12.56 | 6.87 | 32.9 | 0.02 | 98.9  | - | 451 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 1  | 12.9 | 9.92  | 11.23 | 6.81 | 34   | 0.02 | 90.1  | - | 454 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 14 | 0.4  | 10.09 | 17.81 | 7.28 | 36.2 | 0.02 | ***** | - | 381 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 13 | 1.2  | 10.26 | 16.91 | 7.27 | 36.1 | 0.02 | ***** | - | 381 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 12 | 2.2  | 10.33 | 16.58 | 7.21 | 36.2 | 0.02 | ***** | - | 382 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 11 | 3.2  | 10.35 | 16.38 | 7.17 | 36.1 | 0.02 | ***** | - | 383 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 10 | 4.2  | 10.37 | 15.79 | 7.15 | 36.1 | 0.02 | ***** | - | 383 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 9  | 5.2  | 10.36 | 15.4  | 7.09 | 35.8 | 0.02 | ***** | - | 384 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 8  | 6.2  | 10.3  | 14.94 | 7.05 | 35.7 | 0.02 | ***** | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 7  | 7.2  | 10.01 | 13.61 | 7.02 | 34.8 | 0.02 | 96.9  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 6  | 8.2  | 9.88  | 13.17 | 7    | 34.9 | 0.02 | 94.7  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 5  | 9.2  | 9.74  | 12.26 | 6.99 | 34.6 | 0.02 | 91.5  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 4  | 10.2 | 9.68  | 11.16 | 7    | 35.2 | 0.02 | 88.6  | - | 384 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 3  | 11.1 | 9.79  | 10.66 | 7.03 | 35.5 | 0.02 | 88.5  | - | 382 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 2  | 12.3 | 9.81  | 10.31 | 7.04 | 35.9 | 0.02 | 88    | - | 380 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 1  | 13.2 | 10    | 10.02 | 7.09 | 36.4 | 0.02 | 89.1  | - | 377 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 14 | 0.2  | 10.78 | 19.13 | 7.56 | 37.2 | 0.02 | 109.3 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 13 | 0.9  | 10.82 | 18.87 | 7.56 | 37.1 | 0.02 | 109.2 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 12 | 1.9  | 10.96 | 18.2  | 7.56 | 36.8 | 0.02 | 109.1 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 11 | 2.8  | 11.01 | 18.05 | 7.57 | 36.8 | 0.02 | 109.3 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 10 | 3.9  | 11.02 | 18.03 | 7.53 | 36.7 | 0.02 | 109.3 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 9  | 5    | 11    | 18.01 | 7.48 | 36.7 | 0.02 | 109.1 | - | 395 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 8  | 5.9  | 11.01 | 17.9  | 7.44 | 36.8 | 0.02 | 109   | - | 395 |



|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 7  | 6.9  | 10.99 | 17.74 | 7.38 | 36.8 | 0.02 | 108.5 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 6  | 7.9  | 10.87 | 17.49 | 7.32 | 36.8 | 0.02 | 106.7 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 5  | 8.9  | 11    | 16.68 | 7.24 | 36.6 | 0.02 | 106.1 | - | 395 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 4  | 9.9  | 10.77 | 16.38 | 7.16 | 36.5 | 0.02 | 103.2 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 3  | 10.9 | 10.68 | 15.74 | 7.08 | 36.3 | 0.02 | 101   | - | 394 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 2  | 11.9 | 10.59 | 15.24 | 7    | 36.2 | 0.02 | 99.1  | - | 389 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 1  | 12.9 | 10.21 | 14.03 | 6.91 | 35.8 | 0.02 | 93    | - | 385 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 14 | 0.5  | 9.03  | 23.66 | 7.42 | 41.2 | 0.03 | 105.1 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 13 | 1.1  | 9.11  | 23.37 | 7.42 | 41   | 0.03 | 105.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 12 | 2.1  | 9.24  | 23.08 | 7.39 | 40.8 | 0.03 | 106.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 11 | 3.1  | 9.29  | 22.99 | 7.36 | 40.6 | 0.03 | 106.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 10 | 4.1  | 9.3   | 22.86 | 7.34 | 40.4 | 0.03 | 106.7 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 9  | 5.1  | 9.34  | 22.66 | 7.29 | 40.3 | 0.03 | 106.8 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 8  | 6.1  | 9.53  | 21.58 | 7.27 | 39.4 | 0.03 | 106.7 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 7  | 7.1  | 9.52  | 21.44 | 7.14 | 39.5 | 0.03 | 106.3 | - | 369 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 6  | 8.1  | 9.85  | 19.73 | 7.05 | 37.9 | 0.02 | 106.4 | - | 372 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 5  | 9.1  | 9.99  | 15.56 | 6.85 | 34.6 | 0.02 | 99    | - | 379 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 4  | 10.1 | 9.29  | 12.25 | 6.69 | 34   | 0.02 | 85.5  | - | 382 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 3  | 11.1 | 9.15  | 10.74 | 6.63 | 34.4 | 0.02 | 81.3  | - | 383 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 2  | 12.1 | 9.18  | 10.3  | 6.61 | 34.8 | 0.02 | 80.8  | - | 381 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 1  | 13.1 | 9.08  | 10.02 | 6.57 | 35.4 | 0.02 | 79.4  | - | 385 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 14 | 0.3  | 8.54  | 22.49 | 7.43 | 42.7 | 0.03 | 97.9  | - | 378 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 13 | 1.7  | 8.55  | 22.45 | 7.37 | 42.7 | 0.03 | 97.9  | - | 379 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 12 | 2.7  | 8.65  | 22.25 | 7.35 | 41.9 | 0.03 | 98.8  | - | 380 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 11 | 3.7  | 8.68  | 22.06 | 7.33 | 41.1 | 0.03 | 98.7  | - | 379 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 10 | 4.7  | 8.67  | 21.88 | 7.28 | 41.2 | 0.03 | 98.2  | - | 381 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 9  | 5.7  | 8.62  | 21.69 | 7.19 | 41.3 | 0.03 | 97.3  | - | 384 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 8  | 6.7  | 8.64  | 21.35 | 7.13 | 41.3 | 0.03 | 97    | - | 385 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 7  | 7.7  | 8.72  | 20.58 | 7.04 | 39.5 | 0.03 | 96.3  | - | 388 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 6  | 8.7  | 9.23  | 17.2  | 6.92 | 36.6 | 0.02 | 95.2  | - | 395 |

|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 5  | 9.7  | 9.25 | 16.24 | 6.87 | 35.9 | 0.02 | 93.6  | - | 397 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 4  | 10.7 | 9.12 | 14.81 | 6.84 | 35.3 | 0.02 | 89.4  | - | 398 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 3  | 11.7 | 9.02 | 13.12 | 6.82 | 34.9 | 0.02 | 85.2  | - | 399 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 2  | 12.8 | 8.61 | 12.44 | 6.79 | 35.3 | 0.02 | 80.2  | - | 401 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 1  | 13.7 | 8.22 | 10.67 | 6.81 | 36.1 | 0.02 | 73.5  | - | 401 |
|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 14 | 0.3  | 9.16 | 21.23 | 7.43 | 47.7 | 0.03 | 102.6 | - | 372 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 13 | 1.2  | 9.17 | 21.23 | 7.39 | 47.6 | 0.03 | 102.7 | - | 373 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 12 | 2.2  | 9.19 | 21.16 | 7.37 | 47.8 | 0.03 | 102.9 | - | 374 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 11 | 3.1  | 9.23 | 21.11 | 7.35 | 47.8 | 0.03 | 103.1 | - | 374 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 10 | 4.1  | 9.24 | 21.06 | 7.31 | 47.9 | 0.03 | 103.4 | - | 375 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 9  | 5.2  | 9.21 | 21    | 7.26 | 47.7 | 0.03 | 102.8 | - | 377 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 8  | 6.1  | 9.19 | 20.91 | 7.18 | 47.8 | 0.03 | 102.3 | - | 380 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 7  | 7.2  | 9.17 | 20.76 | 7.09 | 47.7 | 0.03 | 101.8 | - | 383 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 6  | 8.2  | 9.26 | 18.17 | 6.96 | 42.3 | 0.03 | 97.6  | - | 389 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 5  | 9.2  | 9.23 | 16.83 | 6.9  | 40.5 | 0.03 | 94.8  | - | 392 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 4  | 10.2 | 9.09 | 15.66 | 6.82 | 39.7 | 0.03 | 90.9  | - | 396 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 3  | 11.3 | 8.94 | 14.06 | 6.79 | 39.1 | 0.03 | 86.3  | - | 398 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 2  | 12.3 | 8.74 | 11.82 | 6.76 | 39   | 0.03 | 80.4  | - | 400 |
| Rockford Bay | 3497 | DBSS | 8/27/97 | - | 8.5 | 1  | 13.2 | 8.8  | 11.23 | 6.78 | 39.2 | 0.03 | 79.8  | - | 400 |
|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 13 | 0.2  | 9.33 | 16.91 | 7.31 | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 12 | 1.8  | 9.34 | 16.9  | 7.31 | 48.5 | 0.03 | 97.5  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 11 | 2.8  | 9.33 | 16.89 | 7.31 | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 10 | 3.8  | 9.33 | 16.86 | 7.3  | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 9  | 4.8  | 9.35 | 16.85 | 7.31 | 48.4 | 0.03 | 97.5  | - | 391 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 8  | 5.8  | 9.33 | 16.83 | 7.29 | 48.5 | 0.03 | 97.2  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 7  | 6.8  | 9.34 | 16.82 | 7.29 | 48.3 | 0.03 | 97.4  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 6  | 7.8  | 9.35 | 16.8  | 7.28 | 48.5 | 0.03 | 97.6  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 5  | 8.8  | 9.34 | 16.63 | 7.27 | 47.7 | 0.03 | 96.9  | - | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 4  | 9.8  | 9.31 | 16.6  | 7.25 | 47.6 | 0.03 | 96.5  | - | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97 | - | 8   | 3  | 10.8 | 9.37 | 16.55 | 7.25 | 47.4 | 0.03 | 97.1  | - | 394 |

|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|--------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 2  | 11.8 | 9.34  | 16.36 | 7.26 | 46.8 | 0.03 | 96.5 | -   | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 1  | 12.8 | 9.38  | 16.33 | 7.25 | 46.6 | 0.03 | 96.8 | -   | 393 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 10 | 0.2  | 9.72  | 15.81 | 7.19 | 43.5 | 0.03 | 99.1 | -   | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 9  | 1    | 9.68  | 15.72 | 7.18 | 43.6 | 0.03 | 98.5 | -   | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 8  | 2.5  | 9.66  | 15.67 | 7.15 | 43.6 | 0.03 | 98.2 | -   | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 7  | 4    | 9.65  | 15.61 | 7.12 | 43.6 | 0.03 | 98   | -   | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 6  | 5.5  | 9.58  | 15.54 | 7.1  | 43.5 | 0.03 | 97.1 | -   | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 5  | 7    | 9.52  | 15.36 | 7.06 | 43.6 | 0.03 | 96   | -   | 448 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 4  | 8.5  | 9.48  | 15.19 | 7.04 | 43.6 | 0.03 | 95.4 | -   | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 3  | 10   | 9.44  | 15.13 | 6.98 | 43.6 | 0.03 | 94.9 | -   | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 2  | 11.3 | 9.51  | 14.94 | 6.95 | 43.6 | 0.03 | 95.1 | -   | 444 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 1  | 12.8 | 9.42  | 14.71 | 6.89 | 43.5 | 0.03 | 93.8 | -   | 441 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 12 | 0.4  | 10.03 | 12.41 | 7.16 | 48.2 | 0.03 | 92.9 | -   | 396 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 11 | 1.8  | 10.02 | 12.41 | 7.13 | 48.2 | 0.03 | 92.9 | -   | 398 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 10 | 2.8  | 10.02 | 12.41 | 7.13 | 48.2 | 0.03 | 92.9 | -   | 397 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 9  | 3.8  | 10.02 | 12.41 | 7.12 | 48.2 | 0.03 | 92.9 | -   | 398 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 8  | 4.8  | 10.02 | 12.41 | 7.09 | 48.3 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 7  | 5.9  | 10.02 | 12.41 | 7.09 | 48.2 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 6  | 6.8  | 10.02 | 12.41 | 7.07 | 48.3 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 5  | 7.8  | 10.04 | 12.41 | 7.04 | 48.3 | 0.03 | 93   | -   | 400 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 4  | 8.8  | 10.04 | 12.41 | 7.02 | 48.3 | 0.03 | 92.9 | -   | 400 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 3  | 9.8  | 10.02 | 12.41 | 7    | 48.2 | 0.03 | 92.9 | -   | 401 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 2  | 11.8 | 10.03 | 12.4  | 6.93 | 48.2 | 0.03 | 92.9 | -   | 402 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 1  | 12.8 | 10.05 | 12.38 | 6.88 | 48.3 | 0.03 | 93   | -   | 404 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 13 | 0.3  | 10.68 | 10.28 | 7.1  | 46.8 | 0.03 | 94.4 | 39  | 420 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 12 | 1.7  | 10.69 | 10.15 | 7.07 | 46.9 | 0.03 | 94.1 | 41  | 423 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 11 | 2.7  | 10.66 | 10.14 | 7.08 | 46.9 | 0.03 | 93.9 | 46  | 422 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 10 | 3.8  | 10.67 | 10.12 | 7.07 | 47   | 0.03 | 93.9 | 122 | 423 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 9  | 4.7  | 10.73 | 10.09 | 7.04 | 46.8 | 0.03 | 94.4 | 50  | 425 |

|              |      |      |         |   |     |   |      |       |       |      |      |      |      |     |     |
|--------------|------|------|---------|---|-----|---|------|-------|-------|------|------|------|------|-----|-----|
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 8 | 5.8  | 10.66 | 10.09 | 7.05 | 46.9 | 0.03 | 93.8 | 49  | 424 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 7 | 6.7  | 10.66 | 10.07 | 7    | 46.9 | 0.03 | 93.7 | 101 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 6 | 7.7  | 10.66 | 10.07 | 7    | 46.9 | 0.03 | 93.7 | 108 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 5 | 8.7  | 10.66 | 10.05 | 7.01 | 47   | 0.03 | 93.7 | 124 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 4 | 9.7  | 10.72 | 10    | 6.99 | 47   | 0.03 | 94   | 101 | 428 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 3 | 10.7 | 10.75 | 10    | 6.95 | 47.2 | 0.03 | 94.3 | 42  | 431 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 2 | 11.7 | 10.71 | 9.99  | 6.96 | 47.2 | 0.03 | 93.9 | 53  | 430 |
| Rockford Bay | 4497 | DBAS | 11/4/97 | - | 7.2 | 1 | 12.6 | 10.66 | 9.99  | 6.88 | 47.2 | 0.03 | 93.5 | 730 | 434 |

| Location          | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|-------------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 10       | 0.2       | 12.97                   | 5.69            | 7.39 | 42.6                 | 0.03 | *****              | -       | 388   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 9        | 1.4       | 12.96                   | 5.67            | 7.39 | 42.7                 | 0.03 | *****              | -       | 388   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 8        | 2.5       | 12.99                   | 5.62            | 7.38 | 42.7                 | 0.03 | *****              | -       | 389   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 7        | 3.7       | 12.97                   | 5.6             | 7.37 | 42.9                 | 0.03 | *****              | -       | 389   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 6        | 5.2       | 12.93                   | 5.52            | 7.37 | 43.1                 | 0.03 | *****              | -       | 389   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 5        | 6.5       | 12.95                   | 5.52            | 7.36 | 43.2                 | 0.03 | *****              | -       | 390   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 4        | 7.8       | 12.95                   | 5.47            | 7.36 | 42.8                 | 0.03 | *****              | -       | 390   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 3        | 9.1       | 12.91                   | 5.46            | 7.36 | 43.1                 | 0.03 | *****              | -       | 390   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 2        | 10.5      | 12.89                   | 5.42            | 7.35 | 43.3                 | 0.03 | *****              | -       | 391   |
| Windy Bay Shallow | 1597  | ASSS    | 4/18/97 | -    | 1.4    | 1        | 11.9      | 12.85                   | 5.22            | 7.36 | 43.7                 | 0.03 | *****              | -       | 392   |
|                   |       |         |         |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 13       | 0.2       | 11.48                   | 12.43           | 7.03 | 21                   | 0.01 | 106.8              | -       | 420   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 12       | 1.1       | 11.6                    | 11.46           | 7    | 36.7                 | 0.02 | 105.4              | -       | 427   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 11       | 2.6       | 11.6                    | 10.94           | 6.95 | 36                   | 0.02 | 104.2              | -       | 431   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 10       | 4.1       | 11.39                   | 10.15           | 6.9  | 34.8                 | 0.02 | 100.4              | -       | 433   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 9        | 5.6       | 11.34                   | 9.74            | 6.89 | 35.3                 | 0.02 | 99                 | -       | 433   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 8        | 7.1       | 11.39                   | 9.53            | 6.86 | 36.5                 | 0.02 | 99                 | -       | 434   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 7        | 8.6       | 11.4                    | 9.15            | 6.84 | 38.2                 | 0.02 | 98.2               | -       | 434   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 6        | 10.1      | 11.38                   | 8.69            | 6.81 | 38.7                 | 0.02 | 97.1               | -       | 435   |
| Windy Bay Shallow | 1997  | DBSS    | 5/16/97 | -    | 1.75   | 5        | 11.6      | 11.36                   | 8.17            | 6.8  | 40.8                 | 0.03 | 95.5               | -       | 436   |

|                   |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 4  | 13.1 | 11.34 | 8.04  | 6.77 | 41.4 | 0.03 | 95    | - | 436 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 3  | 14.6 | 11.45 | 7.47  | 6.76 | 43.4 | 0.03 | 94.6  | - | 437 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 2  | 16.1 | 11.4  | 7.19  | 6.73 | 43.4 | 0.03 | 93.5  | - | 437 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 1  | 17.6 | 11.5  | 7.02  | 6.69 | 43.6 | 0.03 | 94    | - | 436 |
|                   |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 16 | 0.5  | 10.7  | 16.3  | 7.03 | 31.5 | 0.02 | 108.5 | - | 382 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 15 | 1.5  | 10.69 | 16.23 | 7.01 | 31.5 | 0.02 | 108.3 | - | 382 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 14 | 2.4  | 10.74 | 15.76 | 6.99 | 31.5 | 0.02 | 107.7 | - | 383 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 13 | 3.5  | 10.76 | 14.11 | 6.93 | 30.9 | 0.02 | 104.1 | - | 386 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 12 | 4.5  | 9.97  | 12.69 | 6.86 | 31.2 | 0.02 | 93.5  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 11 | 5.6  | 10.42 | 11.36 | 6.88 | 29.9 | 0.02 | 94.7  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 10 | 6.5  | 10.43 | 10.68 | 6.87 | 29.5 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 9  | 7.4  | 10.48 | 10.5  | 6.87 | 29.4 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 8  | 8.5  | 10.52 | 10.33 | 6.87 | 29.6 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 7  | 9.5  | 10.5  | 9.97  | 6.85 | 29.4 | 0.02 | 92.4  | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 6  | 10.6 | 10.45 | 9.09  | 6.85 | 30.4 | 0.02 | 90    | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 5  | 11.5 | 10.37 | 8.97  | 6.84 | 30.6 | 0.02 | 89.1  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 4  | 12.6 | 10.26 | 8.83  | 6.84 | 31.1 | 0.02 | 87.8  | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 3  | 13.6 | 10.34 | 8.61  | 6.84 | 31.7 | 0.02 | 88.1  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 2  | 14.6 | 10.31 | 8.42  | 6.84 | 32.7 | 0.02 | 87.4  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 1  | 15.5 | 10.34 | 8.27  | 6.84 | 33.3 | 0.02 | 87.3  | - | 391 |
|                   |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 16 | 0.5  | 10.8  | 13.95 | 7.02 | 34.9 | 0.02 | 104.5 | - | 423 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 15 | 1.4  | 10.67 | 12.79 | 7.01 | 34.6 | 0.02 | 100.6 | - | 425 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 14 | 2.3  | 10.89 | 11.64 | 6.98 | 32.6 | 0.02 | 100   | - | 426 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 13 | 3.3  | 10.84 | 11.1  | 6.98 | 32   | 0.02 | 98.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 12 | 4.4  | 10.74 | 10.53 | 6.98 | 32.5 | 0.02 | 96.1  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 11 | 5.3  | 10.78 | 10.22 | 6.97 | 30.8 | 0.02 | 95.8  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 10 | 6.3  | 10.69 | 10.04 | 6.97 | 30.5 | 0.02 | 94.7  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 9  | 7.5  | 10.66 | 9.96  | 6.96 | 30.6 | 0.02 | 94.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 8  | 8.3  | 10.58 | 9.89  | 6.96 | 31   | 0.02 | 93.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 7  | 9.5  | 10.63 | 9.66  | 6.96 | 30.9 | 0.02 | 93.2  | - | 427 |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 10.4 | 10.6  | 9.63  | 6.96 | 31.4 | 0.02 | 92.8  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 11.6 | 10.61 | 9.5   | 6.96 | 31.5 | 0.02 | 92.7  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 12.5 | 10.61 | 9.32  | 6.95 | 32   | 0.02 | 92.2  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 13.6 | 10.56 | 9.1   | 6.95 | 33   | 0.02 | 91.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 14.7 | 10.65 | 8.68  | 6.95 | 34.5 | 0.02 | 91.1  | - | 428 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 15.7 | 10.64 | 8.66  | 6.96 | 34.7 | 0.02 | 91    | - | 427 |
|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 16 | 0.4  | 10.71 | 14.84 | 7.13 | 35.5 | 0.02 | 105.6 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 15 | 0.8  | 10.71 | 14.83 | 7.14 | 35.4 | 0.02 | 105.5 | - | 438 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 14 | 1.7  | 10.72 | 14.74 | 7.11 | 35.4 | 0.02 | 105.5 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 13 | 2.7  | 10.74 | 14.21 | 7.11 | 35.4 | 0.02 | 104.4 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 12 | 3.7  | 10.81 | 13.65 | 7.09 | 34.9 | 0.02 | 103.8 | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 11 | 4.7  | 10.83 | 13.22 | 7.08 | 34.3 | 0.02 | 103   | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 10 | 5.6  | 10.82 | 12.61 | 7.08 | 34   | 0.02 | 101.4 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 9  | 6.5  | 10.8  | 12.73 | 7.07 | 34   | 0.02 | 101.4 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 8  | 7.9  | 10.81 | 12.12 | 7.05 | 33.5 | 0.02 | 100.2 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 7  | 8.5  | 10.77 | 12.05 | 7.05 | 33.6 | 0.02 | 99.8  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 6  | 9.7  | 10.81 | 11.61 | 7.04 | 33.5 | 0.02 | 99.1  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 5  | 10.6 | 10.82 | 11.56 | 7.04 | 33.4 | 0.02 | 99.1  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 4  | 11.5 | 10.79 | 11.41 | 7.03 | 33.7 | 0.02 | 98.4  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 3  | 12.5 | 10.78 | 11.2  | 7.03 | 33.7 | 0.02 | 97.9  | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 2  | 13.6 | 10.81 | 11.12 | 7.03 | 33.7 | 0.02 | 98    | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 1  | 14.8 | 10.78 | 11.02 | 7.05 | 33.5 | 0.02 | 97.5  | - | 440 |
|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 16 | 0.4  | 10.28 | 16.6  | 7.25 | 37.3 | 0.02 | ****  | - | 411 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 15 | 1.1  | 10.28 | 16.59 | 7.24 | 37.3 | 0.02 | ****  | - | 412 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 14 | 2    | 10.29 | 16.55 | 7.2  | 37.3 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 13 | 2.9  | 10.28 | 16.53 | 7.18 | 37.3 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 12 | 3.9  | 10.28 | 16.32 | 7.15 | 37.4 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 11 | 5    | 10.39 | 15.56 | 7.12 | 36.7 | 0.02 | ****  | - | 416 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 10 | 6    | 10.4  | 15.31 | 7.1  | 36.5 | 0.02 | ****  | - | 416 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 9  | 7    | 10.35 | 15.03 | 7.06 | 36.8 | 0.02 | ****  | - | 417 |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 8  | 8    | 10.35 | 14.51 | 7.06 | 36   | 0.02 | ***** | - | 418 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 7  | 9    | 10.3  | 14.21 | 7.03 | 36.3 | 0.02 | ***** | - | 418 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 6  | 10.1 | 10.19 | 13.65 | 6.99 | 35.2 | 0.02 | 98.7  | - | 419 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 5  | 11   | 10.04 | 13.32 | 6.94 | 35.3 | 0.02 | 96.5  | - | 420 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 4  | 12   | 9.95  | 12.79 | 6.93 | 34.7 | 0.02 | 94.5  | - | 421 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 3  | 13   | 9.66  | 11.92 | 6.9  | 34.9 | 0.02 | 90    | - | 423 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 2  | 14   | 9.5   | 10.12 | 6.9  | 36.4 | 0.02 | 84.8  | - | 424 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 1  | 15   | 9.57  | 9.74  | 6.94 | 37.4 | 0.02 | 84.7  | - | 422 |
|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 10 | 1.6  | 10.61 | 19.64 | 7.58 | 39.1 | 0.03 | 108.7 | - | 437 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 9  | 3    | 10.74 | 19.09 | 7.58 | 38.8 | 0.02 | 108.9 | - | 437 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 8  | 4.7  | 11.09 | 18.2  | 7.53 | 37.9 | 0.02 | 110.4 | - | 439 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 7  | 6.1  | 11.2  | 17.22 | 7.43 | 37.5 | 0.02 | 109.3 | - | 442 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 6  | 7.6  | 11.19 | 16.86 | 7.34 | 37.3 | 0.02 | 108.3 | - | 444 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 5  | 9    | 11.06 | 16.59 | 7.26 | 37.2 | 0.02 | 106.5 | - | 445 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 4  | 10.5 | 10.86 | 15.92 | 7.17 | 36.6 | 0.02 | 103.1 | - | 447 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 3  | 12   | 10.67 | 15.4  | 7.05 | 36.6 | 0.02 | 100   | - | 449 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 2  | 13.5 | 9.73  | 14.29 | 6.93 | 36.9 | 0.02 | 89.2  | - | 453 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 1  | 15.5 | 9.38  | 11.93 | 6.87 | 36.7 | 0.02 | 81.7  | - | 455 |
|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 12 | 0.4  | 8.97  | 24.09 | 7.47 | 42.1 | 0.03 | 105.3 | - | 369 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 11 | 1    | 8.97  | 24.06 | 7.47 | 42   | 0.03 | 105.3 | - | 370 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 10 | 2.5  | 9.06  | 23.78 | 7.42 | 42.1 | 0.03 | 105.8 | - | 372 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 9  | 4    | 9.06  | 23.38 | 7.4  | 42   | 0.03 | 105   | - | 373 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 8  | 5.5  | 9.39  | 22.43 | 7.35 | 41.3 | 0.03 | 106.8 | - | 374 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 7  | 7    | 9.93  | 20.42 | 7.23 | 39.5 | 0.03 | 108.6 | - | 378 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 6  | 8.5  | 10    | 17.76 | 7.11 | 36.1 | 0.02 | 103.5 | - | 383 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 5  | 10   | 9.48  | 14.74 | 6.98 | 34.9 | 0.02 | 92.2  | - | 389 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 4  | 11.5 | 9.06  | 12.82 | 6.95 | 34.6 | 0.02 | 84.5  | - | 392 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 3  | 13   | 8.88  | 10.21 | 6.94 | 35.4 | 0.02 | 78    | - | 394 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 2  | 14.5 | 8.99  | 9.33  | 6.97 | 36.2 | 0.02 | 77.3  | - | 394 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 1  | 16   | 8.8   | 8.9   | 7.01 | 37.6 | 0.02 | 75.1  | - | 394 |

|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 14 | 0.2  | 8.65 | 22.04 | 7.52 | 41.5 | 0.03 | 98.3  | - | 378 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 13 | 1.6  | 8.65 | 22.01 | 7.41 | 41.5 | 0.03 | 98.3  | - | 381 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 12 | 2.6  | 8.64 | 22.02 | 7.4  | 41.6 | 0.03 | 98.2  | - | 381 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 11 | 3.6  | 8.64 | 21.99 | 7.32 | 41.6 | 0.03 | 98.2  | - | 384 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 10 | 4.6  | 8.64 | 21.92 | 7.26 | 41.6 | 0.03 | 98    | - | 386 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 9  | 5.5  | 8.64 | 21.78 | 7.19 | 41.5 | 0.03 | 97.7  | - | 388 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 8  | 6.6  | 8.6  | 21.32 | 7.05 | 41.6 | 0.03 | 96.4  | - | 394 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 7  | 7.6  | 8.99 | 16.09 | 7.02 | 35.7 | 0.02 | 90.6  | - | 398 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 6  | 8.5  | 8.96 | 15.22 | 7.01 | 35.6 | 0.02 | 88.8  | - | 399 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 5  | 9.6  | 8.77 | 14.04 | 7.01 | 35.3 | 0.02 | 84.7  | - | 400 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 4  | 10.5 | 8.59 | 12.81 | 7.03 | 35   | 0.02 | 80.6  | - | 400 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 3  | 11.5 | 8.43 | 12.08 | 7.07 | 35.2 | 0.02 | 77.9  | - | 399 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 2  | 12.6 | 8.19 | 11.67 | 7.14 | 35.7 | 0.02 | 74.9  | - | 396 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 1  | 13.7 | 8.03 | 10.82 | 7.13 | 36.3 | 0.02 | 72    | - | 398 |
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 12 | 0.4  | 9.26 | 20.74 | 7.51 | 46.5 | 0.03 | 102.8 | - | 368 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 11 | 1    | 9.27 | 20.76 | 7.5  | 46.4 | 0.03 | 102.9 | - | 368 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 10 | 2.5  | 9.26 | 20.74 | 7.48 | 46.5 | 0.03 | 102.8 | - | 369 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 9  | 3.9  | 9.27 | 20.74 | 7.45 | 46.4 | 0.03 | 102.8 | - | 370 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 8  | 5.5  | 9.23 | 20.7  | 7.38 | 46.6 | 0.03 | 102.4 | - | 372 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 7  | 7.1  | 9.21 | 20.65 | 7.27 | 46.8 | 0.03 | 102   | - | 377 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 6  | 8.6  | 9.02 | 17.29 | 6.94 | 39.9 | 0.03 | 93.4  | - | 389 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 5  | 10   | 9.19 | 14.11 | 6.9  | 38.6 | 0.02 | 88.9  | - | 392 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 4  | 11.5 | 8.86 | 12.31 | 6.86 | 38.7 | 0.02 | 82.3  | - | 395 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 3  | 13.1 | 8.52 | 10.95 | 6.86 | 39.6 | 0.03 | 76.7  | - | 396 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 2  | 14.4 | 8.44 | 9.36  | 6.9  | 41.3 | 0.03 | 73.2  | - | 396 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 1  | 16.1 | 8.59 | 8.64  | 7.06 | 42.5 | 0.03 | 73.2  | - | 392 |
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97 | - | 8.1 | 16 | 0.2  | 9.32 | 17.01 | 7.27 | 47.5 | 0.03 | 97.5  | - | 384 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97 | - | 8.1 | 15 | 1.7  | 9.32 | 17.01 | 7.26 | 47.6 | 0.03 | 97.5  | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97 | - | 8.1 | 14 | 2.7  | 9.32 | 17    | 7.24 | 47.6 | 0.03 | 97.5  | - | 385 |



|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |   |     |
|-------------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|---|-----|
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 13 | 3.7  | 9.3   | 17.01 | 7.24 | 47.5 | 0.03 | 97.3 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 12 | 4.7  | 9.3   | 16.98 | 7.23 | 47.6 | 0.03 | 97.2 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 11 | 5.7  | 9.31  | 16.96 | 7.22 | 47.6 | 0.03 | 97.2 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 10 | 6.7  | 9.29  | 16.91 | 7.19 | 47.3 | 0.03 | 97   | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 9  | 7.7  | 9.28  | 16.81 | 7.18 | 47.2 | 0.03 | 96.8 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 8  | 8.7  | 9.29  | 16.75 | 7.15 | 47.1 | 0.03 | 96.8 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 7  | 9.7  | 9.31  | 16.58 | 7.13 | 46.5 | 0.03 | 96.5 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 6  | 10.7 | 9.33  | 16.52 | 7.1  | 46.3 | 0.03 | 96.6 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 5  | 11.7 | 9.31  | 16.48 | 7.07 | 46.1 | 0.03 | 96.3 | - | 386 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 4  | 12.7 | 9.3   | 16.38 | 7.05 | 45.9 | 0.03 | 96   | - | 386 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 3  | 13.7 | 9.24  | 16.04 | 7    | 45.6 | 0.03 | 94.7 | - | 387 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 2  | 14.7 | 9.23  | 15.25 | 6.96 | 43.7 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 1  | 15.7 | 9.13  | 15.04 | 6.94 | 43.9 | 0.03 | 91.8 | - | 388 |
|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |   |     |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 11 | 0.3  | 9.65  | 14.71 | 7.17 | 46.5 | 0.03 | 96.1 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 10 | 2    | 9.62  | 14.67 | 7.18 | 46.6 | 0.03 | 95.7 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 9  | 3.5  | 9.61  | 14.64 | 7.17 | 46.6 | 0.03 | 95.5 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 8  | 5    | 9.59  | 14.61 | 7.17 | 46.6 | 0.03 | 95.3 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 7  | 6.5  | 9.55  | 14.54 | 7.16 | 46.9 | 0.03 | 94.8 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 6  | 8    | 9.54  | 14.52 | 7.15 | 46.9 | 0.03 | 94.7 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 5  | 9.5  | 9.5   | 14.44 | 7.15 | 46.5 | 0.03 | 94   | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 4  | 11   | 9.52  | 14.41 | 7.15 | 46.9 | 0.03 | 94.2 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 3  | 12.5 | 9.47  | 14.27 | 7.15 | 46.6 | 0.03 | 93.7 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 2  | 14   | 9.52  | 14.24 | 7.15 | 46.9 | 0.03 | 93.8 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 1  | 15.2 | 9.48  | 14.16 | 7.15 | 46.8 | 0.03 | 93.2 | - | 447 |
|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |   |     |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 16 | 0.3  | 10.06 | 12.32 | 7.24 | 49.9 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 15 | 1    | 10.07 | 12.33 | 7.25 | 49.8 | 0.03 | 93.1 | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 14 | 2    | 10.05 | 12.33 | 7.25 | 49.8 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 13 | 3    | 10.05 | 12.33 | 7.24 | 49.9 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 12 | 4    | 10.06 | 12.32 | 7.24 | 49.9 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 11 | 5    | 10.05 | 12.33 | 7.23 | 50   | 0.03 | 93   | - | 387 |

|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|-------------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 10 | 6    | 10.05 | 12.33 | 7.21 | 49.9 | 0.03 | 93   | -   | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 9  | 7    | 10.05 | 12.33 | 7.2  | 49.9 | 0.03 | 93   | -   | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 8  | 8    | 10.04 | 12.33 | 7.19 | 49.8 | 0.03 | 93   | -   | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 7  | 9    | 10.06 | 12.32 | 7.18 | 49.9 | 0.03 | 93   | -   | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 6  | 10   | 10.07 | 12.32 | 7.15 | 49.9 | 0.03 | 93.1 | -   | 390 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 5  | 11   | 10.07 | 12.33 | 7.16 | 49.9 | 0.03 | 93.1 | -   | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 4  | 12   | 10.08 | 12.32 | 7.13 | 49.9 | 0.03 | 93.2 | -   | 390 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 3  | 13   | 10.09 | 12.33 | 7.12 | 49.9 | 0.03 | 93.3 | -   | 391 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 2  | 14   | 10.1  | 12.32 | 7.1  | 49.9 | 0.03 | 93.3 | -   | 391 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 1  | 15   | 10.11 | 12.3  | 7.06 | 50   | 0.03 | 93.5 | -   | 392 |
|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 11 | 0.4  | 10.55 | 10.17 | 7.05 | 47.6 | 0.03 | 93   | 57  | 409 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 10 | 1.7  | 10.53 | 9.89  | 7.14 | 47.7 | 0.03 | 92.2 | 105 | 406 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 9  | 3.2  | 10.53 | 9.78  | 7.1  | 48.2 | 0.03 | 91.9 | 144 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 8  | 4.7  | 10.53 | 9.74  | 7.12 | 48.6 | 0.03 | 91.8 | 117 | 407 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 7  | 6.2  | 10.51 | 9.66  | 7.14 | 49.1 | 0.03 | 91.5 | 141 | 406 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 6  | 7.7  | 10.52 | 9.55  | 7.1  | 49.7 | 0.03 | 91.4 | 134 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 5  | 9.2  | 10.55 | 9.51  | 7.12 | 49.8 | 0.03 | 91.5 | 100 | 407 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 4  | 10.7 | 10.56 | 9.46  | 7.09 | 50   | 0.03 | 91.5 | 138 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 3  | 12.2 | 10.49 | 9.42  | 7.09 | 50.2 | 0.03 | 90.8 | 152 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 2  | 13.7 | 10.53 | 9.33  | 7.08 | 50.2 | 0.03 | 90.9 | 115 | 409 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97  | - | 5.8 | 1  | 15.2 | 10.46 | 9.33  | 7.09 | 50.3 | 0.03 | 90.3 | 601 | 409 |

| Location       | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|----------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 27       | 0.4       | 12.89                   | 5.6             | 7.38 | 43.4                 | 0.03 | *****              | -       | 385   |
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 26       | 1         | 12.89                   | 5.6             | 7.36 | 43.5                 | 0.03 | *****              | -       | 387   |
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 25       | 2         | 12.95                   | 5.57            | 7.37 | 43.4                 | 0.03 | *****              | -       | 387   |
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 24       | 3.2       | 12.96                   | 5.54            | 7.37 | 43.5                 | 0.03 | *****              | -       | 387   |
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 23       | 4.2       | 12.77                   | 5.54            | 7.36 | 43.5                 | 0.03 | *****              | -       | 386   |
| Windy Bay Deep | 1597  | ASSS    | 4/18/97 | -    | 1.6    | 22       | 5.5       | 12.88                   | 5.51            | 7.35 | 43.6                 | 0.03 | *****              | -       | 387   |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 21 | 6.7  | 12.83 | 5.5   | 7.35 | 43.5 | 0.03 | ***** |   | 387 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 20 | 7.8  | 12.98 | 5.39  | 7.34 | 43.6 | 0.03 | ***** | - | 387 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 19 | 9    | 12.7  | 5.34  | 7.33 | 43.5 | 0.03 | ***** | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 18 | 10.1 | 12.87 | 5.32  | 7.33 | 43.5 | 0.03 | ***** | - | 387 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 17 | 11.4 | 12.92 | 5.3   | 7.33 | 43.5 | 0.03 | ***** | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 16 | 12.8 | 12.72 | 5.24  | 7.33 | 43.5 | 0.03 | ***** | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 15 | 14   | 12.75 | 5.22  | 7.32 | 43.4 | 0.03 | ***** | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 14 | 15.3 | 12.86 | 5.17  | 7.32 | 43.3 | 0.03 | ***** | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 13 | 16.5 | 12.62 | 5.12  | 7.32 | 43.2 | 0.03 | 99    | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 12 | 17.8 | 12.61 | 5.09  | 7.31 | 43.2 | 0.03 | 99    | - | 389 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 11 | 19   | 12.59 | 5.04  | 7.31 | 43.2 | 0.03 | 98.7  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 10 | 20.2 | 12.63 | 5.06  | 7.31 | 43.2 | 0.03 | 99    | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 9  | 21.5 | 12.77 | 4.99  | 7.31 | 43.2 | 0.03 | 99.9  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 8  | 22.7 | 12.77 | 4.96  | 7.3  | 43.3 | 0.03 | 99.9  | - | 389 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 7  | 24   | 12.84 | 4.74  | 7.31 | 43.9 | 0.03 | 99.9  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 6  | 25.3 | 12.7  | 4.73  | 7.3  | 43.9 | 0.03 | 98.8  | - | 389 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 5  | 26.5 | 12.79 | 4.74  | 7.3  | 43.9 | 0.03 | 99.5  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 4  | 27.9 | 12.85 | 4.74  | 7.31 | 43.9 | 0.03 | 99.7  | - | 387 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 3  | 29.1 | 12.85 | 4.71  | 7.3  | 43.9 | 0.03 | 99.8  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 2  | 30.5 | 12.71 | 4.71  | 7.3  | 44   | 0.03 | 98.8  | - | 388 |
| Windy Bay Deep | 1597 | ASSS | 4/18/97 | - | 1.6 | 1  | 31.8 | 12.93 | 4.74  | 7.31 | 43.9 | 0.03 | ***** | - | 387 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 19 | 0.2  | 11.55 | 12.53 | 7.06 | 36.7 | 0.02 | 107.6 | - | 413 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 18 | 1.8  | 11.42 | 10.91 | 7.05 | 34.6 | 0.02 | 102.4 | - | 417 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 17 | 3.8  | 11.3  | 9.97  | 7.01 | 34.4 | 0.02 | 99.2  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 16 | 5.8  | 11.31 | 9.48  | 7    | 35   | 0.02 | 98.2  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 15 | 7.8  | 11.3  | 9.09  | 6.98 | 35.3 | 0.02 | 97.2  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 14 | 9.8  | 11.44 | 8.3   | 6.98 | 37.5 | 0.02 | 96.5  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 13 | 11.8 | 11.49 | 7.66  | 6.97 | 39.6 | 0.03 | 95.4  |   | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 12 | 13.8 | 11.53 | 7.27  | 6.96 | 40.2 | 0.03 | 94.8  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 11 | 15.8 | 11.54 | 7.12  | 6.96 | 40.6 | 0.03 | 94.6  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 10 | 17.8 | 11.55 | 7.07  | 6.96 | 40.9 | 0.03 | 94.5  | - | 420 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 9  | 19.8 | 11.63 | 6.79  | 6.95 | 42.7 | 0.03 | 94.5  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 8  | 21.8 | 11.65 | 6.66  | 6.95 | 43.3 | 0.03 | 94.4  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 7  | 23.8 | 11.74 | 6.33  | 6.94 | 45.1 | 0.03 | 94.3  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 6  | 25.8 | 11.71 | 6.33  | 6.94 | 45.3 | 0.03 | 94.1  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 5  | 27.8 | 11.72 | 6.25  | 6.93 | 45.8 | 0.03 | 93.8  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 4  | 29.8 | 11.63 | 6.17  | 6.92 | 46   | 0.03 | 93    | - | 418 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 3  | 31.8 | 11.6  | 5.89  | 6.9  | 47   | 0.03 | 92.1  | - | 418 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 2  | 33.8 | 11.57 | 5.75  | 6.91 | 47.2 | 0.03 | 91.6  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 1  | 35.8 | 11.52 | 5.75  | 6.91 | 47.5 | 0.03 | 91.1  | - | 425 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 19 | 0.2  | 10.79 | 15.64 | 7.04 | 31.2 | 0.02 | 107.9 | - | 385 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 18 | 1    | 10.79 | 15.59 | 7.01 | 31.3 | 0.02 | 107.8 | - | 386 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 17 | 3    | 10.89 | 14.01 | 6.95 | 31.2 | 0.02 | 105.1 | - | 390 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 16 | 5    | 10.49 | 11.31 | 6.9  | 30.6 | 0.02 | 95.3  | - | 394 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 15 | 7    | 10.51 | 10.33 | 6.92 | 29.1 | 0.02 | 93.3  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 14 | 9    | 10.6  | 9.99  | 6.92 | 28.2 | 0.02 | 93.3  | - | 392 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 13 | 11.1 | 10.72 | 9.42  | 6.92 | 29.4 | 0.02 | 93.1  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 12 | 12.9 | 10.8  | 8.99  | 6.93 | 29.1 | 0.02 | 92.7  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 11 | 14.9 | 10.88 | 8.48  | 6.92 | 31.1 | 0.02 | 92.3  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 10 | 17   | 11.03 | 7.91  | 6.91 | 34.3 | 0.02 | 92.3  | - | 394 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 9  | 19   | 11.14 | 7.33  | 6.89 | 37.4 | 0.02 | 91.9  | - | 395 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 8  | 21   | 11.05 | 7.1   | 6.88 | 38.4 | 0.02 | 90.7  | - | 395 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 7  | 23   | 11.13 | 6.61  | 6.87 | 40.4 | 0.03 | 90.2  | - | 396 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 6  | 24.9 | 11.12 | 6.33  | 6.86 | 41.5 | 0.03 | 89.5  | - | 396 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 5  | 27.1 | 11.03 | 6.08  | 6.85 | 42.5 | 0.03 | 88.2  | - | 397 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 4  | 28.9 | 11.03 | 5.97  | 6.85 | 42.8 | 0.03 | 87.9  | - | 397 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 3  | 30.9 | 10.89 | 5.87  | 6.84 | 43.3 | 0.03 | 86.6  | - | 398 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 2  | 33.1 | 10.89 | 5.84  | 6.85 | 43.4 | 0.03 | 86.4  | - | 398 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 1  | 35   | 10.81 | 5.9   | 6.85 | 43.4 | 0.03 | 86.1  | - | 398 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 17 | 0.8  | 10.58 | 14.84 | 6.97 | 37   | 0.02 | 104.4 | - | 422 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 16 | 2.8  | 10.68 | 13.55 | 6.94 | 35.7 | 0.02 | 102.6 | - | 424 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 15 | 4.8  | 10.67 | 10.86 | 6.95 | 31.8 | 0.02 | 96.2  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 14 | 6.9  | 10.7  | 10.12 | 6.95 | 30.1 | 0.02 | 94.8  | - | 424 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 13 | 8.8  | 10.61 | 9.91  | 6.94 | 30.1 | 0.02 | 93.5  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 10.8 | 10.67 | 9.2   | 6.94 | 31.8 | 0.02 | 92.5  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 12.8 | 10.66 | 9.02  | 6.93 | 33.2 | 0.02 | 92    | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 14.8 | 10.67 | 8.84  | 6.93 | 33.9 | 0.02 | 91.7  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 16.7 | 10.67 | 8.58  | 6.92 | 35.2 | 0.02 | 91.2  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 18.9 | 10.75 | 7.86  | 6.91 | 38.3 | 0.02 | 90.3  | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 20.7 | 10.81 | 6.91  | 6.89 | 42.8 | 0.03 | 88.6  | - | 427 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 22.6 | 10.81 | 6.76  | 6.89 | 43.7 | 0.03 | 88.2  | - | 427 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 24.8 | 10.82 | 6.61  | 6.89 | 44.3 | 0.03 | 88    | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 26.8 | 10.85 | 6.51  | 6.88 | 44.7 | 0.03 | 88    | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 28.7 | 10.92 | 6.33  | 6.88 | 45.4 | 0.03 | 88.2  | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 30.5 | 10.76 | 6.18  | 6.88 | 46.2 | 0.03 | 86.4  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 32.4 | 10.85 | 5.98  | 6.87 | 47.1 | 0.03 | 86.7  | - | 426 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 16 | 0.3  | 10.75 | 13.52 | 7.13 | 35   | 0.02 | 102.9 | - | 435 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 15 | 1.2  | 10.74 | 13.52 | 7.12 | 35   | 0.02 | 102.8 | - | 436 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 14 | 3.3  | 10.76 | 13.44 | 7.08 | 34.9 | 0.02 | 102.8 | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 13 | 5.3  | 10.79 | 12.84 | 7.05 | 34.5 | 0.02 | 101.6 | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 12 | 7.4  | 10.78 | 12.14 | 7.05 | 33.5 | 0.02 | 100   | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 11 | 9.4  | 10.7  | 11.99 | 7.02 | 33.2 | 0.02 | 99    | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 10 | 11.4 | 10.74 | 11.92 | 7    | 33.1 | 0.02 | 99.2  | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 9  | 13.4 | 10.77 | 11.27 | 6.98 | 33.1 | 0.02 | 98    | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 8  | 15.3 | 10.75 | 10.79 | 6.96 | 33   | 0.02 | 96.7  | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 7  | 17.4 | 10.64 | 9.37  | 6.92 | 34.9 | 0.02 | 92.5  | - | 439 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 6  | 19.4 | 10.66 | 7.47  | 6.9  | 39.3 | 0.03 | 88.5  | - | 441 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 5  | 21.2 | 10.65 | 7.27  | 6.9  | 40.2 | 0.03 | 87.9  | - | 440 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 4  | 23.3 | 10.67 | 6.73  | 6.9  | 42.3 | 0.03 | 87    | - | 440 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 3  | 25.3 | 10.71 | 6.56  | 6.9  | 43   | 0.03 | 86.8  | - | 439 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 2  | 27.3 | 10.62 | 6.23  | 6.91 | 44.4 | 0.03 | 85.5  | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 1  | 29.2 | 10.67 | 6.25  | 6.93 | 44.4 | 0.03 | 85.9  | - | 437 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 16 | 0.3  | 10.09 | 17.22 | 7.16 | 38.2 | 0.02 | ***** | - | 421 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 15 | 1    | 10.11 | 17.17 | 7.15 | 38.1 | 0.02 | ***** | - | 421 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 14 | 3    | 10.19 | 16.89 | 7.06 | 37.9 | 0.02 | ***** | - | 425 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 13 | 5    | 10.35 | 15.28 | 7.04 | 36.2 | 0.02 | ***** | - | 426 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 12 | 7    | 10.26 | 14.31 | 6.98 | 35.7 | 0.02 | ***** | - | 428 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 11 | 9    | 10.18 | 13.32 | 6.93 | 35.2 | 0.02 | 97.9  | - | 429 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 10 | 11   | 9.93  | 12.73 | 6.86 | 34.4 | 0.02 | 94.3  | - | 431 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 9  | 13   | 9.88  | 11.56 | 6.83 | 34.5 | 0.02 | 91.5  | - | 433 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 8  | 15   | 9.83  | 9.32  | 6.78 | 37.1 | 0.02 | 86.2  | - | 435 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 7  | 17   | 10.07 | 7.65  | 6.79 | 40.5 | 0.03 | 84.7  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 6  | 18.9 | 10.12 | 7.43  | 6.78 | 41   | 0.03 | 84.7  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 5  | 20.9 | 10.3  | 7.1   | 6.79 | 41.4 | 0.03 | 85.5  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 4  | 23   | 10.31 | 6.79  | 6.79 | 42.7 | 0.03 | 84.8  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 3  | 25   | 10.32 | 6.69  | 6.79 | 43.1 | 0.03 | 84.8  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 2  | 27   | 10.34 | 6.4   | 6.78 | 44.4 | 0.03 | 84.3  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 1  | 29.1 | 10.13 | 6.13  | 6.78 | 45.6 | 0.03 | 82.1  | - | 436 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 15 | 0.4  | 9.68  | 20.43 | 7.4  | 38.9 | 0.02 | ***** | - | 417 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 14 | 1.5  | 9.88  | 19.74 | 7.38 | 38   | 0.02 | ***** | - | 418 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 13 | 5.6  | 10.3  | 17.73 | 7.24 | 35.3 | 0.02 | ***** | - | 423 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 12 | 7.6  | 10.39 | 16.87 | 7.14 | 35.4 | 0.02 | ***** | - | 426 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 11 | 9.6  | 10.36 | 16.08 | 6.99 | 35.3 | 0.02 | ***** | - | 431 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 10 | 11.6 | 9.99  | 14.64 | 6.85 | 34.2 | 0.02 | 97.7  | - | 437 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 9  | 13.6 | 9.49  | 13.24 | 6.75 | 33.4 | 0.02 | 90    | - | 441 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 8  | 15.6 | 8.99  | 10.54 | 6.69 | 34   | 0.02 | 80.2  | - | 444 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 7  | 17.6 | 9.71  | 8.29  | 6.67 | 37   | 0.02 | 82    | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 6  | 19.6 | 9.92  | 7.28  | 6.67 | 39.2 | 0.03 | 81.7  | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 5  | 21.6 | 9.99  | 6.87  | 6.66 | 40.4 | 0.03 | 81.4  | - | 447 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 4  | 23.6 | 10.1  | 6.66  | 6.66 | 41.1 | 0.03 | 81.8  | - | 447 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 3  | 25.5 | 10.09 | 6.51  | 6.66 | 41.5 | 0.03 | 81.5  | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5 | 2  | 27.7 | 10.07 | 6.31  | 6.66 | 42.2 | 0.03 | 81    | - | 446 |

|                |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 1  | 29.6 | 9.6   | 6.33  | 6.65 | 42.7 | 0.03 | 77.2  | - | 447 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 18 | 0.3  | 8.91  | 24.3  | 7.42 | 42.5 | 0.03 | 105   | - | 371 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 17 | 2.3  | 9.05  | 23.52 | 7.4  | 43.6 | 0.03 | 105.1 | - | 372 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 16 | 4.2  | 9.24  | 22.92 | 7.31 | 43   | 0.03 | 106.1 | - | 375 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 15 | 6.3  | 9.79  | 20.85 | 7.2  | 40   | 0.03 | 108.1 | - | 378 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 14 | 8.3  | 10.19 | 16.78 | 7.03 | 35.7 | 0.02 | 103.5 | - | 385 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 13 | 10.4 | 9.31  | 13.48 | 6.87 | 34.2 | 0.02 | 88    | - | 392 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 12 | 12.3 | 8.86  | 11.15 | 6.84 | 34.7 | 0.02 | 79.6  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 11 | 14.2 | 8.98  | 9.66  | 6.85 | 35.4 | 0.02 | 77.8  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 10 | 16.2 | 9.21  | 8.37  | 6.85 | 37.6 | 0.02 | 77.4  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 9  | 18.3 | 9.66  | 7.5   | 6.87 | 38.7 | 0.02 | 79.5  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 8  | 20.3 | 9.78  | 7.05  | 6.87 | 40.3 | 0.03 | 79.5  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 7  | 22.3 | 9.96  | 6.79  | 6.89 | 40.8 | 0.03 | 80.4  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 6  | 24.2 | 10.08 | 6.57  | 6.89 | 41.3 | 0.03 | 81    | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 5  | 26.3 | 10.1  | 6.38  | 6.91 | 42   | 0.03 | 80.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 4  | 28.3 | 10    | 6.23  | 6.92 | 42.7 | 0.03 | 79.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 3  | 30.3 | 9.74  | 6.26  | 6.93 | 43.2 | 0.03 | 77.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 2  | 32.3 | 9.57  | 6.21  | 6.94 | 43.2 | 0.03 | 76.2  | - | 393 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 1  | 34.3 | 9.56  | 6.18  | 6.98 | 44.8 | 0.03 | 76.1  | - | 392 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 16 | 0.3  | 8.72  | 22.94 | 7.3  | 41.9 | 0.03 | 100.8 | - | 382 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 15 | 1.3  | 8.77  | 22.33 | 7.2  | 42.6 | 0.03 | 100.2 | - | 388 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 14 | 3.3  | 8.8   | 22.03 | 7.13 | 43.1 | 0.03 | 100.1 | - | 391 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 13 | 5.3  | 8.9   | 21.82 | 7.43 | 42   | 0.03 | 100.7 | - | 384 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 12 | 7.3  | 8.87  | 21.61 | 7.31 | 41.5 | 0.03 | 100   | - | 388 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 11 | 9.3  | 8.96  | 20.73 | 7.4  | 41.3 | 0.03 | 99.4  | - | 383 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 10 | 11.3 | 8.86  | 13.82 | 6.78 | 35.3 | 0.02 | 85.6  | - | 408 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 9  | 13.5 | 8.82  | 11.59 | 7.2  | 35.9 | 0.02 | 80.5  | - | 399 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 8  | 15.4 | 8.47  | 10.68 | 6.75 | 35.9 | 0.02 | 75.7  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 7  | 17.4 | 8.57  | 9.04  | 6.76 | 37.3 | 0.02 | 73.7  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 6  | 19.3 | 8.84  | 8.18  | 6.77 | 39   | 0.03 | 74.4  | - | 411 |

|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|------|----|------|------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 5  | 21.4 | 9.11 | 7.27  | 6.78 | 40.8 | 0.03 | 74.9  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 4  | 23.4 | 9.37 | 6.92  | 6.8  | 41.5 | 0.03 | 76.6  | - | 410 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 3  | 25.4 | 9.5  | 6.72  | 6.81 | 42.1 | 0.03 | 77.1  | - | 409 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 2  | 27.4 | 9.52 | 6.56  | 6.82 | 42.5 | 0.03 | 76.9  | - | 408 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 1  | 29.4 | 9.12 | 6.33  | 6.83 | 43.8 | 0.03 | 73.3  | - | 408 |
|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 16 | 0.3  | 9.22 | 20.87 | 7.41 | 47   | 0.03 | 102.6 | - | 372 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 15 | 1.5  | 9.23 | 20.83 | 7.39 | 47.1 | 0.03 | 102.7 | - | 372 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 14 | 3.5  | 9.23 | 20.78 | 7.33 | 46.9 | 0.03 | 102.5 | - | 374 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 13 | 5.4  | 9.24 | 20.67 | 7.26 | 46.8 | 0.03 | 102.4 | - | 375 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 12 | 7.5  | 9.15 | 20.69 | 7.08 | 46.7 | 0.03 | 101.5 | - | 382 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 11 | 9.5  | 9.28 | 15.93 | 6.87 | 38.8 | 0.02 | 93.5  | - | 394 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 10 | 11.5 | 8.91 | 12.38 | 6.82 | 38.3 | 0.02 | 82.9  | - | 397 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 9  | 13.6 | 8.48 | 9.58  | 6.8  | 40.6 | 0.03 | 73.9  | - | 399 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 8  | 15.5 | 9.07 | 7.68  | 6.8  | 43.2 | 0.03 | 75.5  | - | 401 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 7  | 17.2 | 9.43 | 7.07  | 6.83 | 44.3 | 0.03 | 77.4  | - | 400 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 6  | 21.6 | 9.72 | 6.64  | 6.88 | 45.3 | 0.03 | 78.8  | - | 399 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 5  | 23.5 | 9.72 | 6.5   | 6.89 | 45.6 | 0.03 | 78.6  | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 4  | 25.6 | 9.71 | 6.38  | 6.9  | 46.1 | 0.03 | 78.2  | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 3  | 27.5 | 9.58 | 6.3   | 6.92 | 46.4 | 0.03 | 77    | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 2  | 29.5 | 9.47 | 6.28  | 6.96 | 46.5 | 0.03 | 76.2  | - | 397 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 1  | 31.7 | 9.48 | 6.28  | 7.05 | 46.7 | 0.03 | 76.2  | - | 395 |
|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 17 | 0.4  | 9.28 | 17.53 | 7.3  | 52   | 0.03 | 98.1  | - | 389 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 16 | 2.7  | 9.34 | 17.56 | 7.27 | 52   | 0.03 | 98.6  | - | 390 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 15 | 4.7  | 9.22 | 17.51 | 7.22 | 52   | 0.03 | 97.5  | - | 392 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 14 | 6.7  | 9.22 | 17.24 | 7.18 | 50.2 | 0.03 | 96.9  | - | 393 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 13 | 8.7  | 9.25 | 17.17 | 7.15 | 49.7 | 0.03 | 97.1  | - | 393 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 12 | 10.7 | 9.28 | 17.12 | 7.08 | 48.7 | 0.03 | 97.3  | - | 395 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 11 | 12.7 | 9.26 | 16.8  | 6.98 | 47.3 | 0.03 | 96.3  | - | 398 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 10 | 14.7 | 9.07 | 15.91 | 6.86 | 45   | 0.03 | 92.6  | - | 403 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 9  | 16.7 | 8.66 | 12.17 | 6.73 | 40.1 | 0.03 | 81.4  | - | 411 |



|                |      |      |          |   |     |    |      |      |       |      |      |      |      |   |     |
|----------------|------|------|----------|---|-----|----|------|------|-------|------|------|------|------|---|-----|
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 8  | 18.7 | 8.43 | 10.51 | 6.7  | 41.1 | 0.03 | 76.3 | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 7  | 20.7 | 8.47 | 9.46  | 6.71 | 41.9 | 0.03 | 74.8 | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 6  | 22.7 | 8.57 | 8.43  | 6.71 | 42.7 | 0.03 | 73.8 | - | 413 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 5  | 24.7 | 8.89 | 7.38  | 6.72 | 43.9 | 0.03 | 74.6 | - | 413 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 4  | 26.7 | 9.04 | 7.07  | 6.74 | 44.5 | 0.03 | 75.3 | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 3  | 28.7 | 9.16 | 6.73  | 6.76 | 45.2 | 0.03 | 75.6 | - | 411 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 2  | 30.7 | 9.21 | 6.64  | 6.8  | 45.2 | 0.03 | 75.9 | - | 410 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 1  | 32.7 | 9.23 | 6.64  | 6.84 | 45.2 | 0.03 | 76   | - | 408 |
|                |      |      |          |   |     |    |      |      |       |      |      |      |      |   |     |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 17 | 0.2  | 9.77 | 15.09 | 7.16 | 46.3 | 0.03 | 98.1 | - | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 16 | 2    | 9.65 | 15.03 | 7.16 | 46.4 | 0.03 | 96.8 | - | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 15 | 4    | 9.6  | 14.96 | 7.15 | 46.2 | 0.03 | 96.2 | - | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 14 | 6    | 9.59 | 14.78 | 7.11 | 46.6 | 0.03 | 95.7 | - | 445 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 13 | 8    | 9.57 | 14.66 | 7.09 | 46.5 | 0.03 | 95.2 | - | 446 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 12 | 10   | 9.57 | 14.66 | 7.07 | 46.6 | 0.03 | 95.2 | - | 446 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 11 | 12   | 9.5  | 14.58 | 7.02 | 46.5 | 0.03 | 94.3 | - | 447 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 10 | 14   | 9.39 | 14.35 | 6.97 | 45.9 | 0.03 | 92.8 | - | 448 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 9  | 16   | 9.34 | 13.92 | 6.94 | 45.4 | 0.03 | 91.4 | - | 449 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 8  | 18   | 9.28 | 13.6  | 6.91 | 45   | 0.03 | 90.2 | - | 450 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 7  | 20   | 9.23 | 13.3  | 6.85 | 44.7 | 0.03 | 89.1 | - | 451 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 6  | 22   | 8.88 | 11.2  | 6.78 | 42.2 | 0.03 | 81.7 | - | 455 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 5  | 23.9 | 8.9  | 8.48  | 6.75 | 41.4 | 0.03 | 76.8 | - | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 4  | 26   | 8.91 | 7.97  | 6.74 | 41.8 | 0.03 | 75.9 | - | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 3  | 28   | 8.94 | 7.6   | 6.76 | 42   | 0.03 | 75.5 | - | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 2  | 30   | 8.93 | 7.28  | 6.8  | 42.2 | 0.03 | 74.8 | - | 455 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 1  | 31.8 | 9.05 | 7     | 6.83 | 42.6 | 0.03 | 75.3 | - | 454 |
|                |      |      |          |   |     |    |      |      |       |      |      |      |      |   |     |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 16 | 0.3  | 9.92 | 12.28 | 7.14 | 49.7 | 0.03 | 91.7 | - | 386 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 15 | 1    | 9.94 | 12.28 | 7.14 | 49.7 | 0.03 | 91.8 | - | 387 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 14 | 3    | 9.92 | 12.3  | 7.14 | 49.7 | 0.03 | 91.6 | - | 386 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 13 | 5    | 9.91 | 12.3  | 7.12 | 49.7 | 0.03 | 91.5 | - | 387 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 12 | 7    | 9.91 | 12.28 | 7.1  | 49.9 | 0.03 | 91.6 | - | 387 |

|                |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|----------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 11 | 9    | 9.92  | 12.27 | 7.07 | 49.8 | 0.03 | 91.6 | -   | 388 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 10 | 11   | 9.91  | 12.27 | 7.03 | 49.7 | 0.03 | 91.5 | -   | 388 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 9  | 13   | 9.88  | 12.27 | 6.99 | 49.7 | 0.03 | 91.2 | -   | 389 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 8  | 15   | 9.82  | 12.27 | 6.91 | 49.8 | 0.03 | 90.7 | -   | 392 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 7  | 17   | 9.74  | 12.18 | 6.8  | 49.8 | 0.03 | 89.7 | -   | 395 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 6  | 19   | 8.21  | 10.1  | 6.64 | 48.8 | 0.03 | 72.1 | -   | 401 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 5  | 21   | 8.3   | 7.99  | 6.64 | 47.5 | 0.03 | 69.2 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 4  | 23   | 8.4   | 7.35  | 6.64 | 47.5 | 0.03 | 69   | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 3  | 25   | 8.27  | 6.96  | 6.65 | 47.6 | 0.03 | 67.3 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 2  | 27   | 8.28  | 6.91  | 6.65 | 47.9 | 0.03 | 67.2 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 1  | 29.1 | 8.27  | 6.84  | 6.71 | 47.9 | 0.03 | 67   | -   | 401 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 16 | 0.6  | 10.6  | 10.33 | 7.11 | 46.8 | 0.03 | 93.7 | 57  | 411 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 15 | 2.8  | 10.56 | 10.23 | 7.1  | 46.9 | 0.03 | 93.1 | 142 | 413 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 14 | 4.8  | 10.5  | 10.22 | 7.11 | 46.9 | 0.03 | 92.6 | 207 | 413 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 13 | 6.8  | 10.43 | 10.07 | 7.09 | 47.1 | 0.03 | 91.6 | 56  | 414 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 12 | 8.8  | 10.46 | 10.07 | 7.08 | 47.1 | 0.03 | 91.9 | 46  | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 11 | 10.8 | 10.45 | 10.09 | 7.09 | 47.2 | 0.03 | 91.9 | 40  | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 10 | 12.8 | 10.46 | 10.07 | 7.09 | 47.2 | 0.03 | 91.9 | 117 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 9  | 14.8 | 10.47 | 10.05 | 7.09 | 47.2 | 0.03 | 92   | 159 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 8  | 16.8 | 10.42 | 10.04 | 7.08 | 47.4 | 0.03 | 91.5 | 132 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 7  | 18.8 | 10.44 | 10.04 | 7.08 | 47.3 | 0.03 | 91.7 | 102 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 6  | 20.8 | 10.45 | 10.02 | 7.07 | 47.3 | 0.03 | 91.7 | 112 | 416 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 5  | 22.8 | 10.43 | 9.97  | 7.07 | 47.5 | 0.03 | 91.5 | 52  | 416 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 4  | 24.8 | 10.39 | 9.91  | 7.04 | 47.7 | 0.03 | 91   | 55  | 417 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 3  | 26.8 | 10.34 | 9.71  | 7.03 | 48.3 | 0.03 | 90.1 | 119 | 418 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 2  | 28.8 | 10.24 | 9.33  | 7.02 | 49.5 | 0.03 | 88.5 | 111 | 419 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 1  | 30.8 | 10.05 | 9.02  | 7.01 | 50   | 0.03 | 86.1 | 631 | 421 |

| Location | Phase | Sampler | Date | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH | Conductivity (µs/cm) | TDS | Percent Saturation | Statime | Redox |
|----------|-------|---------|------|------|--------|----------|-----------|-------------------------|-----------------|----|----------------------|-----|--------------------|---------|-------|
|----------|-------|---------|------|------|--------|----------|-----------|-------------------------|-----------------|----|----------------------|-----|--------------------|---------|-------|

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 17 | 0.4  | 12.48 | 6.49  | 7.48 | 37.4 | 0.02 | ***** | - | 372 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 16 | 1.4  | 12.52 | 6.33  | 7.46 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 15 | 2.4  | 12.46 | 6.3   | 7.46 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 14 | 3.5  | 12.47 | 6.28  | 7.45 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 13 | 4.6  | 12.42 | 6.28  | 7.44 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 12 | 5.5  | 12.4  | 6.28  | 7.44 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 11 | 6.6  | 12.41 | 6.26  | 7.43 | 37.5 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 10 | 7.8  | 12.39 | 6.25  | 7.43 | 37.4 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 9  | 8.8  | 12.45 | 6.23  | 7.42 | 37.5 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 8  | 10   | 12.42 | 6.2   | 7.42 | 37.5 | 0.02 | ***** | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 7  | 11.1 | 12.38 | 6.17  | 7.42 | 37.6 | 0.02 | 99.9  | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 6  | 12.4 | 12.34 | 6.08  | 7.41 | 38   | 0.02 | 99.3  | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 5  | 13.5 | 12.32 | 5.88  | 7.41 | 38.4 | 0.02 | 98.6  | - | 374 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 4  | 14.8 | 12.27 | 5.77  | 7.4  | 38.5 | 0.02 | 98    | - | 375 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 3  | 16   | 12.23 | 5.67  | 7.39 | 38.4 | 0.02 | 97.4  | - | 376 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 2  | 17.2 | 12.19 | 5.59  | 7.39 | 38.3 | 0.02 | 96.9  | - | 380 |
| CDA River | 1597 | ASSS | 4/18/97 | - | 1.8 | 1  | 18.6 | 12.59 | 5.44  | 7.38 | 38.4 | 0.02 | 99.9  | - | 382 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 15 | 0.2  | 10.61 | 14.93 | 6.94 | 32   | 0.02 | 104.3 | - | 390 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 14 | 1.6  | 10.56 | 12.51 | 6.86 | 32   | 0.02 | 98.3  | - | 399 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 13 | 2.7  | 10.43 | 12.37 | 6.84 | 32.1 | 0.02 | 96.8  | - | 401 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 12 | 3.7  | 10.41 | 12.27 | 6.83 | 32.3 | 0.02 | 96.4  | - | 402 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 11 | 4.7  | 10.33 | 12.18 | 6.82 | 32.2 | 0.02 | 95.5  | - | 402 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 10 | 5.7  | 10.25 | 11.91 | 6.82 | 32.4 | 0.02 | 94.1  | - | 404 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 9  | 6.7  | 10.25 | 11.76 | 6.82 | 32.3 | 0.02 | 93.8  | - | 404 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 8  | 7.7  | 10.24 | 11.64 | 6.85 | 32.3 | 0.02 | 93.5  | - | 403 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 7  | 8.7  | 10.33 | 11.31 | 6.93 | 32   | 0.02 | 93.7  | - | 400 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 6  | 9.7  | 10.67 | 10.4  | 6.99 | 31   | 0.02 | 94.6  | - | 396 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 5  | 10.7 | 11.02 | 9.42  | 7.01 | 29.6 | 0.02 | 95.5  | - | 394 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 4  | 11.7 | 11    | 9.06  | 7.01 | 31.9 | 0.02 | 94.5  | - | 394 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 3  | 12.7 | 10.99 | 8.88  | 7.01 | 32.5 | 0.02 | 94    | - | 393 |
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 2  | 13.7 | 11.02 | 8.77  | 7.01 | 32.9 | 0.02 | 94    | - | 392 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 1997 | DBSS | 5/16/97 | - | 0.6 | 1  | 14.7 | 10.97 | 8.73  | 7.01 | 33   | 0.02 | 93.5  | - | 395 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 12 | 0.5  | 10.55 | 14.36 | 6.93 | 35.9 | 0.02 | 102.5 | - | 389 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 11 | 1.4  | 10.7  | 13.83 | 6.95 | 35.4 | 0.02 | 102.8 | - | 389 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 10 | 2.3  | 10.71 | 13.67 | 6.94 | 34.6 | 0.02 | 102.5 | - | 390 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 9  | 3.5  | 10.8  | 12.2  | 6.92 | 35.6 | 0.02 | 100.1 | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 8  | 4.3  | 10.8  | 12.05 | 6.91 | 35.7 | 0.02 | 99.8  | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 7  | 5.4  | 10.8  | 12    | 6.91 | 35.8 | 0.02 | 99.6  | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 6  | 6.3  | 10.78 | 11.94 | 6.91 | 35.8 | 0.02 | 99.3  | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 5  | 7.3  | 10.77 | 11.92 | 6.91 | 35.8 | 0.02 | 99.2  | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 4  | 8.3  | 10.65 | 11.2  | 6.94 | 33.3 | 0.02 | 96.4  | - | 392 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 3  | 9.3  | 10.87 | 9.58  | 7.01 | 27.5 | 0.02 | 94.7  | - | 389 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 2  | 10.3 | 10.87 | 9.33  | 7.02 | 26.8 | 0.02 | 94.2  | - | 390 |
| CDA River | 2197 | DBSS | 5/29/97 | - | 1.7 | 1  | 11.4 | 10.86 | 9.33  | 7.01 | 27.1 | 0.02 | 94.1  | - | 392 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 13 | 0.4  | 10.07 | 16.99 | 6.99 | 43.2 | 0.03 | 104   | - | 420 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 12 | 1    | 10.06 | 15.56 | 7    | 44.3 | 0.03 | 100.8 | - | 420 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 11 | 2    | 10.06 | 15.31 | 6.98 | 44.8 | 0.03 | 100.3 | - | 421 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 10 | 2.9  | 10    | 14.53 | 6.97 | 45   | 0.03 | 98    | - | 422 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 9  | 4    | 10.02 | 14.59 | 6.98 | 45.1 | 0.03 | 98.3  | - | 421 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 8  | 5    | 10.06 | 14.36 | 6.99 | 44.7 | 0.03 | 98.2  | - | 421 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 7  | 5.9  | 10.05 | 14.28 | 6.99 | 44   | 0.03 | 97.9  | - | 421 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 6  | 6.9  | 10.52 | 12.6  | 7.03 | 36.9 | 0.02 | 98.7  | - | 419 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 5  | 7.8  | 10.59 | 12.22 | 7.05 | 34.8 | 0.02 | 98.6  | - | 417 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 4  | 8.9  | 10.94 | 11.51 | 7.08 | 30.8 | 0.02 | 100.1 | - | 415 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 3  | 9.9  | 11.04 | 11.3  | 7.08 | 30.1 | 0.02 | 100.5 | - | 414 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 2  | 10.9 | 11.08 | 11.25 | 7.07 | 29.2 | 0.02 | 100.8 | - | 414 |
| CDA River | 2397 | DBAS | 6/11/97 | - | 1.9 | 1  | 11.9 | 10.86 | 10.87 | 7.07 | 29.1 | 0.02 | 98.1  | - | 414 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 13 | 0.3  | 10.46 | 14.23 | 7.08 | 45.3 | 0.03 | 101.7 | - | 435 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 12 | 1.2  | 10.46 | 14.25 | 7.07 | 45.3 | 0.03 | 101.7 | - | 437 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 11 | 2.3  | 10.45 | 14.16 | 7.05 | 45   | 0.03 | 101.5 | - | 438 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 10 | 3.3  | 10.49 | 13.73 | 7.01 | 44.6 | 0.03 | 100.7 | - | 439 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 9  | 4.2  | 10.47 | 13.6  | 7.03 | 42.4 | 0.03 | 100.4 | - | 437 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 8  | 5.2  | 10.52 | 13.4  | 7.02 | 37.8 | 0.02 | 100.5 | - | 436 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 7  | 6.2  | 10.56 | 12.5  | 7    | 34.2 | 0.02 | 98.8  | - | 436 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 6  | 7.2  | 10.55 | 12.18 | 7    | 32.4 | 0.02 | 98    | - | 435 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 5  | 8.3  | 10.45 | 11.41 | 7    | 30.5 | 0.02 | 95.3  | - | 435 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 4  | 9    | 10.39 | 11.36 | 7    | 30.4 | 0.02 | 94.7  | - | 435 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 3  | 10.2 | 10.39 | 11.33 | 6.99 | 30.4 | 0.02 | 94.7  | - | 434 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 2  | 11.5 | 10.22 | 11.12 | 6.98 | 30.4 | 0.02 | 92.6  | - | 434 |
| CDA River | 2597 | DBSS | 6/26/97 | - | 2.2 | 1  | 12.6 | 10.07 | 10.82 | 7    | 30.6 | 0.02 | 90.6  | - | 433 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 11 | 0.4  | 9.72  | 18.54 | 7.11 | 49.9 | 0.03 | ***** | - | 421 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 10 | 1.1  | 9.72  | 18.53 | 7.09 | 50   | 0.03 | ***** | - | 421 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 9  | 2.1  | 9.73  | 18.5  | 7.06 | 52.1 | 0.03 | ***** | - | 422 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 8  | 3    | 9.78  | 18.37 | 7    | 54.8 | 0.04 | ***** | - | 424 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 7  | 4.1  | 9.6   | 18.03 | 6.91 | 69.8 | 0.04 | ***** | - | 428 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 6  | 5    | 9.65  | 17.73 | 6.86 | 75.4 | 0.05 | ***** | - | 430 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 5  | 6    | 9.96  | 14.99 | -    | 43   | 0.03 | 99.4  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 4  | 7    | 10    | 13.1  | 6.94 | 34.1 | 0.02 | 95.7  | - | 425 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 3  | 9    | 9.76  | 11.94 | 6.91 | 33.4 | 0.02 | 90.9  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 2  | 10.1 | 9.75  | 11.82 | 6.92 | 33.4 | 0.02 | 90.6  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 1  | 11   | 9.71  | 11.82 | 6.94 | 33.4 | 0.02 | 90.3  | - | 430 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 12 | 0.5  | 9.38  | 21.69 | 7.49 | 39.8 | 0.03 | ***** | - | 395 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 11 | 1.2  | 9.4   | 21.37 | 7.46 | 42.7 | 0.03 | ***** | - | 396 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 10 | 2.3  | 9.4   | 21.33 | 7.4  | 48.4 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 9  | 3.2  | 9.4   | 21.33 | 7.37 | 50.1 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 8  | 4.2  | 9.4   | 21.3  | 7.33 | 51.3 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 7  | 5.3  | 9.42  | 21.11 | 7.26 | 49.9 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 6  | 6.2  | 9.33  | 20.67 | 7.11 | 58.2 | 0.04 | ***** | - | 402 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 5  | 7.2  | 10.2  | 16.19 | 7.12 | 35.9 | 0.02 | ***** | - | 401 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 4  | 8.2  | 9.82  | 15.19 | 7.02 | 34.9 | 0.02 | 97.2  | - | 403 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 3  | 9.2  | 9.45  | 14.18 | 6.98 | 34   | 0.02 | 91.5  | - | 402 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 2  | 10.2 | 9.28  | 13.84 | 0    | 34   | 0.02 | 89.2  | - | 400 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 1  | 11.2 | 8.84  | 13.68 | 7.01 | 34.2 | 0.02 | 84.7  | - | 397 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 11 | 0.4  | 9.38  | 25.18 | 7.65 | 61.4 | 0.04 | 112.4 | - | 359 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 10 | 1.1  | 9.29  | 24.06 | 7.58 | 53.3 | 0.03 | 109   | - | 361 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 9  | 2.1  | 9.23  | 23.67 | 7.54 | 50.8 | 0.03 | 107.6 | - | 362 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 8  | 3.2  | 9.1   | 23.53 | 7.5  | 46.6 | 0.03 | 105.8 | - | 363 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 7  | 4.2  | 9.14  | 23.26 | 7.46 | 46.6 | 0.03 | 105.8 | - | 365 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 6  | 5.2  | 9.22  | 23.15 | 7.39 | 49.1 | 0.03 | 106.3 | - | 366 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 5  | 6    | 10.06 | 19.97 | 7.38 | 41.6 | 0.03 | 109.2 | - | 367 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 4  | 7.1  | 9.71  | 19.4  | 7.31 | 42.1 | 0.03 | 0     | - | 369 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 3  | 8    | 9.81  | 18.17 | 7.33 | 38.3 | 0.02 | 102.7 | - | 368 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 2  | 8.9  | 10.49 | 17.23 | 7.35 | 36.1 | 0.02 | 107.7 | - | 366 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 1  | 10   | 9.81  | 16.75 | 7.32 | 36.6 | 0.02 | 99.7  | - | 368 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 10 | 0.2  | 8.82  | 23.38 | 7.74 | 54.8 | 0.04 | 102.8 | - | 363 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 9  | 0.7  | 8.77  | 23.13 | 7.65 | 53.9 | 0.03 | 101.8 | - | 369 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 8  | 1.7  | 8.74  | 22.74 | 7.66 | 47.7 | 0.03 | 100.7 | - | 368 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 7  | 2.7  | 8.8   | 22.65 | 7.68 | 50.3 | 0.03 | 101.3 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 6  | 3.7  | 8.88  | 22.63 | 7.66 | 52.9 | 0.03 | 102.2 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 5  | 4.7  | 8.91  | 22.33 | 7.64 | 56   | 0.04 | 101.9 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 4  | 5.6  | 8.95  | 22.24 | 7.58 | 61.9 | 0.04 | 102.2 | - | 368 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 3  | 6.7  | 8.95  | 22.08 | 7.3  | 67.3 | 0.04 | 101.8 | - | 375 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 2  | 7.7  | 8.67  | 19.88 | 7.15 | 46.5 | 0.03 | 94.5  | - | 381 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 1  | 8.7  | 8.73  | 17.78 | 7.14 | 39.5 | 0.03 | 91.2  | - | 382 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 13 | 0.3  | 9.13  | 20.86 | 7.53 | 52.7 | 0.03 | 101.5 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 12 | 0.9  | 9.14  | 20.84 | 7.54 | 52.6 | 0.03 | 101.7 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 11 | 2    | 9.14  | 20.84 | 7.54 | 52.5 | 0.03 | 101.7 | - | 344 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 10 | 2.9  | 9.14  | 20.83 | 7.54 | 53.3 | 0.03 | 101.7 | - | 344 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 9  | 4    | 9.14  | 20.81 | 7.52 | 55.2 | 0.04 | 101.6 | - | 345 |

|           |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 8  | 5.1  | 9.12 | 20.81 | 7.5  | 55.4 | 0.04 | 101.4 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 7  | 6.1  | 9.11 | 20.77 | 7.48 | 58   | 0.04 | 101.2 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 6  | 6.9  | 9.08 | 20.76 | 7.47 | 58.3 | 0.04 | 100.9 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 5  | 7.9  | 9.09 | 20.72 | 7.44 | 60   | 0.04 | 100.9 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 4  | 9    | 9.06 | 20.69 | 7.41 | 59.8 | 0.04 | 100.4 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 3  | 10   | 9.03 | 20.67 | 7.39 | 59.1 | 0.04 | 100.2 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 2  | 10.9 | 8.92 | 20.56 | 7.37 | 59   | 0.04 | 98.7  | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97 | - | 6.1 | 1  | 12   | 8.08 | 16.69 | 7.11 | 42.1 | 0.03 | 82.6  | - | 359 |
|           |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 11 | 0.1  | 9.38 | 17.47 | 7.44 | 58.9 | 0.04 | 99.1  | - | 370 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 10 | 1    | 9.32 | 17.49 | 7.43 | 58.9 | 0.04 | 98.4  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 9  | 2    | 9.24 | 17.48 | 7.45 | 59.1 | 0.04 | 97.6  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 8  | 3    | 9.35 | 17.49 | 7.42 | 59   | 0.04 | 98.8  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 7  | 4    | 9.37 | 17.48 | 7.4  | 59.6 | 0.04 | 99    | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 6  | 5    | 9.35 | 17.49 | 7.39 | 59.3 | 0.04 | 98.8  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 5  | 6    | 9.29 | 17.48 | 7.37 | 59.8 | 0.04 | 97.9  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 4  | 7    | 9.28 | 17.43 | 7.34 | 61   | 0.04 | 97.8  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 3  | 8    | 9.19 | 17.44 | 7.22 | 61   | 0.04 | 97    | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 2  | 9    | 8.59 | 9.82  | 7.04 | 50.3 | 0.03 | 76.9  | - | 380 |
| CDA River | 3797 | ASRP | 9/17/97 | - | 5.5 | 1  | 10.1 | 8.38 | 9.79  | 6.96 | 52.8 | 0.03 | 74.6  | - | 385 |
|           |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 11 | 0.2  | 9.72 | 16.01 | 7.37 | 52.6 | 0.03 | 99.4  | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 10 | 1.8  | 9.66 | 15.76 | 7.36 | 52.2 | 0.03 | 98.4  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 9  | 2.8  | 9.66 | 15.71 | 7.33 | 53.3 | 0.03 | 98.3  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 8  | 3.8  | 9.62 | 15.66 | 7.31 | 53.9 | 0.03 | 97.8  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 7  | 4.8  | 9.59 | 15.61 | 7.29 | 53.4 | 0.03 | 97.4  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 6  | 5.9  | 9.56 | 15.54 | 7.26 | 53.8 | 0.03 | 97    | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 5  | 6.8  | 9.62 | 15.52 | 7.23 | 56.7 | 0.04 | 97.5  | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 4  | 7.8  | 9.74 | 15.12 | 7.15 | 64.5 | 0.04 | 97.9  | - | 416 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 3  | 8.8  | 9.62 | 14.99 | 7.03 | 68   | 0.04 | 96.5  | - | 419 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 2  | 9.8  | 7.74 | 12.13 | 6.9  | 48.6 | 0.03 | 72.8  | - | 425 |
| CDA River | 3997 | ASRA | 9/29/97 | - | 6   | 1  | 10.8 | 7.85 | 10.26 | 6.97 | 44.5 | 0.03 | 70.6  | - | 424 |

|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|-----------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 11 | 0.3  | 10.11 | 11.79 | 7.24 | 54.6 | 0.03 | 92.3 | -   | 371 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 10 | 1    | 10.1  | 11.81 | 7.27 | 54.6 | 0.03 | 92.2 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 9  | 2    | 10.11 | 11.81 | 7.23 | 54.8 | 0.04 | 92.4 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 8  | 3    | 10.08 | 11.81 | 7.24 | 54.7 | 0.04 | 92.1 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 7  | 4.1  | 10.12 | 11.77 | 7.22 | 55.3 | 0.04 | 92.4 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 6  | 5    | 10.13 | 11.74 | 7.19 | 55.7 | 0.04 | 92.4 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 5  | 6    | 10.16 | 11.68 | 7.18 | 56.7 | 0.04 | 92.5 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 4  | 7    | 10.17 | 11.59 | 7.17 | 57.9 | 0.04 | 92.5 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 3  | 8    | 10.17 | 11.59 | 7.13 | 57.9 | 0.04 | 92.5 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 2  | 9    | 10.22 | 11.41 | 7.1  | 63.2 | 0.04 | 92.5 | -   | 371 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 1  | 10   | 10.23 | 10.94 | 7.01 | 81.3 | 0.05 | 91.6 | -   | 379 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 11 | 0.4  | 10.87 | 10.02 | 7.35 | 53.9 | 0.03 | 95.5 | 54  | 392 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 10 | 1.5  | 10.94 | 9.79  | 7.34 | 53.7 | 0.03 | 95.5 | 50  | 393 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 9  | 2.5  | 10.87 | 9.79  | 7.37 | 53.7 | 0.03 | 94.9 | 146 | 392 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 8  | 3.5  | 10.92 | 9.74  | 7.33 | 53.6 | 0.03 | 95.3 | 51  | 393 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 7  | 4.5  | 10.88 | 9.68  | 7.28 | 53.7 | 0.03 | 94.8 | 102 | 395 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 6  | 5.5  | 10.9  | 9.4   | 7.21 | 54.5 | 0.03 | 94.3 | 59  | 398 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 5  | 6.5  | 10.91 | 9.38  | 7.26 | 54.6 | 0.03 | 94.3 | 121 | 395 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 4  | 7.5  | 11    | 9.25  | 7.19 | 54.9 | 0.04 | 94.8 | 49  | 398 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 3  | 8.5  | 11.23 | 8.55  | 7.14 | 57.3 | 0.04 | 95.1 | 42  | 401 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 2  | 9.5  | 11.29 | 8.25  | 7.12 | 58.2 | 0.04 | 95   | 59  | 402 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 1  | 10.5 | 11.26 | 8.17  | 7.1  | 58.6 | 0.04 | 94.5 | 536 | 404 |

| Location | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|----------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Mid Lake | 1597  | ASSS    | 4/18/97 | -    | 2      | 10       | 0.4       | 12.33                   | 6.36            | 7.36 | 44.6                 | 0.03 | 99.9               | -       | 377   |



|          |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|----------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 9  | 1.4  | 12.36 | 6.3   | 7.35 | 45.2 | 0.03 | ***** | - | 378 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 8  | 2.6  | 12.27 | 6.33  | 7.35 | 44.4 | 0.03 | 99.4  | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 7  | 3.8  | 12.31 | 6.28  | 7.34 | 45.3 | 0.03 | 99.6  | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 6  | 5.1  | 12.32 | 6.3   | 7.34 | 45.4 | 0.03 | 99.7  | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 5  | 6.4  | 12.35 | 6.31  | 7.34 | 44.9 | 0.03 | ***** | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 4  | 7.5  | 12.43 | 6.31  | 7.33 | 44.6 | 0.03 | ***** | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 3  | 8.7  | 12.4  | 6.31  | 7.33 | 44.7 | 0.03 | ***** | - | 378 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 2  | 10.1 | 12.44 | 6.26  | 7.32 | 44.8 | 0.03 | ***** | - | 379 |
| Mid Lake | 1597 | ASSS | 4/18/97 | - | 2    | 1  | 11.2 | 12.57 | 6.28  | 7.32 | 44.8 | 0.03 | ***** | - | 378 |
|          |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 15 | 0.2  | 12.04 | 13.83 | 6.33 | 0.7  | 0    | 115.5 | - | 422 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 14 | 1.5  | 11.12 | 10.32 | 7    | 28.2 | 0.02 | 98.5  | - | 392 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 13 | 3    | 11.1  | 9.86  | 7    | 28.8 | 0.02 | 97.3  | - | 392 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 12 | 4.5  | 11.16 | 9.79  | 6.99 | 28.3 | 0.02 | 97.5  | - | 392 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 11 | 6    | 11.1  | 9.51  | 6.98 | 29   | 0.02 | 96.4  | - | 393 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 10 | 7.5  | 11.04 | 9.25  | 6.98 | 29.4 | 0.02 | 95.3  | - | 392 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 9  | 9    | 11.04 | 9.17  | 6.99 | 29.6 | 0.02 | 95    | - | 391 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 8  | 10.5 | 11.08 | 8.92  | 7.02 | 30.7 | 0.02 | 94.8  | - | 389 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 7  | 12   | 11.29 | 8.45  | 7.01 | 33.4 | 0.02 | 95.6  | - | 389 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 6  | 13.5 | 11.32 | 7.81  | 6.99 | 34.7 | 0.02 | 94.3  | - | 390 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 5  | 15   | 11.28 | 7.47  | 6.98 | 34.9 | 0.02 | 93.2  | - | 389 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 4  | 16.5 | 11.25 | 6.66  | 6.96 | 37.2 | 0.02 | 91.1  | - | 390 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 3  | 18   | 11.12 | 6.08  | 6.94 | 40.1 | 0.03 | 88.7  | - | 391 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 2  | 19.5 | 11.15 | 5.69  | 6.93 | 45.9 | 0.03 | 88.1  | - | 391 |
| Mid Lake | 1997 | DBAC | 5/16/97 | - | 1.25 | 1  | 20.9 | 10.91 | 5.6   | 6.92 | 47.2 | 0.03 | 85.9  | - | 390 |
|          |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 13 | 0.2  | 11.32 | 12.63 | 7.06 | 27.6 | 0.02 | 105.9 | - | 351 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 12 | 0.9  | 11.33 | 12.51 | 7.05 | 27.6 | 0.02 | 105.8 | - | 352 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 11 | 2.5  | 11.19 | 10.72 | 7    | 26.8 | 0.02 | 100.2 | - | 353 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 10 | 4    | 10.83 | 10.3  | 6.97 | 26.6 | 0.02 | 96    | - | 354 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 9  | 5.5  | 10.7  | 10.07 | 6.96 | 26.4 | 0.02 | 94.4  | - | 353 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7  | 8  | 7    | 10.68 | 9.64  | 6.95 | 26.3 | 0.02 | 93.2  | - | 352 |

|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 7  | 8.5  | 10.71 | 9.45  | 6.95 | 26.5 | 0.02 | 93    | - | 349 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 6  | 9.9  | 10.73 | 9.25  | 6.96 | 26.6 | 0.02 | 92.8  | - | 348 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 5  | 12.5 | 10.75 | 8.84  | 6.95 | 27   | 0.02 | 92    | - | 346 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 4  | 14.4 | 10.82 | 8.38  | 6.94 | 28.3 | 0.02 | 91.6  | - | 343 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 3  | 16.1 | 10.73 | 7.96  | 6.92 | 30.2 | 0.02 | 89.9  | - | 342 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 2  | 18.7 | 10.48 | 6.49  | 6.87 | 40.5 | 0.03 | 84.7  | - | 339 |
| Mid Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 1  | 20.2 | 10.17 | 6.71  | 6.87 | 40.7 | 0.03 | 82.6  | - | 332 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 14 | 0.4  | 10.81 | 15.76 | 7.1  | 30.6 | 0.02 | 108.8 | - | 411 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 13 | 1.3  | 10.86 | 15.58 | 7.11 | 30.7 | 0.02 | 108.9 | - | 411 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 2.9  | 10.9  | 13.8  | 7.1  | 29.8 | 0.02 | 105.2 | - | 411 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 4.3  | 10.86 | 13.09 | 7.05 | 29.4 | 0.02 | 103   | - | 414 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 5.9  | 11.2  | 11.54 | 7.04 | 28.9 | 0.02 | 102.6 | - | 415 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 7.4  | 11    | 10.45 | 7.02 | 29   | 0.02 | 98.2  | - | 416 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 9    | 10.92 | 10.12 | 7    | 28.6 | 0.02 | 96.8  | - | 417 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 10.5 | 10.84 | 9.71  | 6.99 | 28.6 | 0.02 | 95.2  | - | 417 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 11.9 | 10.62 | 8.94  | 6.96 | 31.6 | 0.02 | 91.4  | - | 419 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 13.4 | 10.6  | 8.42  | 6.95 | 34.4 | 0.02 | 90.1  | - | 420 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 15   | 10.23 | 7.81  | 6.92 | 37.7 | 0.02 | 85.7  | - | 421 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 16.4 | 10.36 | 7.05  | 6.92 | 41.9 | 0.03 | 85.2  | - | 421 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 17.9 | 10.31 | 6.82  | 6.92 | 43.2 | 0.03 | 84.3  | - | 425 |
| Mid Lake | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 19.5 | 10.02 | 6.99  | 6.93 | 43.2 | 0.03 | 82.3  | - | 424 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 13 | 0.4  | 10.7  | 15.09 | 7.14 | 31.7 | 0.02 | 106.1 | - | 405 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 12 | 1.9  | 10.69 | 14.89 | 7.12 | 31.8 | 0.02 | 105.4 | - | 405 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 11 | 3.5  | 10.74 | 14.36 | 7.1  | 31.9 | 0.02 | 104.8 | - | 406 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 10 | 5    | 10.68 | 13.32 | 6.97 | 30.2 | 0.02 | 101.8 | - | 410 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 9  | 6.6  | 10.73 | 12.33 | 6.96 | 30.5 | 0.02 | 100   | - | 410 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 8  | 8.1  | 10.36 | 11.45 | 6.92 | 30.4 | 0.02 | 94.6  | - | 411 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 7  | 9.6  | 10.26 | 10.59 | 6.89 | 29.6 | 0.02 | 91.8  | - | 412 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 6  | 10.9 | 10.14 | 9.27  | 6.86 | 30.8 | 0.02 | 88    | - | 413 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 5  | 12.5 | 10.11 | 8.25  | 6.83 | 34.6 | 0.02 | 85.6  | - | 414 |

|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 4  | 14   | 9.57  | 7.53  | 6.81 | 38   | 0.02 | 79.6  | - | 414 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 3  | 15.6 | 9.44  | 7.05  | 6.8  | 40.5 | 0.03 | 77.7  | - | 413 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 2  | 17   | 9.03  | 6.82  | 6.8  | 42.4 | 0.03 | 73.8  | - | 411 |
| Mid Lake | 2597 | DBSS | 6/26/97 | - | 3   | 1  | 18.6 | 9.08  | 6.81  | 6.86 | 42.4 | 0.03 | 74.1  | - | 407 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 13 | 0.3  | 9.95  | 19.4  | 7.02 | 35.5 | 0.02 | ***** | - | 444 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 12 | 1.7  | 10.11 | 17.84 | 7    | 34.8 | 0.02 | ***** | - | 445 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 11 | 3.3  | 10.29 | 17.17 | 6.92 | 35   | 0.02 | ***** | - | 447 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 10 | 4.8  | 10.55 | 14.84 | 6.81 | 38.2 | 0.02 | ***** | - | 451 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 9  | 6.2  | 10.41 | 13.81 | 6.77 | 35.6 | 0.02 | 99.9  | - | 451 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 8  | 7.8  | 10.26 | 13.02 | 6.7  | 34.6 | 0.02 | 96.7  | - | 451 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 7  | 9.2  | 10.03 | 12.2  | 6.66 | 33.4 | 0.02 | 92.8  | - | 451 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 6  | 10.8 | 9.78  | 10.8  | 6.6  | 33.3 | 0.02 | 87.6  | - | 450 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 5  | 12.2 | 9.53  | 9.76  | 6.57 | 34.5 | 0.02 | 83.3  | - | 450 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 4  | 13.7 | 9.55  | 8.95  | 6.54 | 36.6 | 0.02 | 82    | - | 448 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 3  | 15.3 | 9.33  | 8.46  | 6.52 | 38.2 | 0.02 | 79.1  | - | 447 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 2  | 16.6 | 9.28  | 7.91  | 6.49 | 39.8 | 0.03 | 77.7  | - | 445 |
| Mid Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 1  | 18.2 | 9.37  | 7.51  | 6.47 | 42   | 0.03 | 77.6  | - | 447 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 13 | 1.3  | 9.57  | 21.01 | 7.48 | 38.8 | 0.02 | ***** | - | 394 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 12 | 2.8  | 9.92  | 20.47 | 7.43 | 37.9 | 0.02 | ***** | - | 393 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 11 | 4.3  | 10.52 | 17.56 | 7.27 | 37.7 | 0.02 | ***** | - | 400 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 10 | 5.8  | 10.49 | 16.73 | 7.18 | 36.2 | 0.02 | ***** | - | 402 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 9  | 7.3  | 10.22 | 15.79 | 7.03 | 35.2 | 0.02 | ***** | - | 406 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 8  | 8.8  | 9.96  | 15.01 | 6.95 | 34.4 | 0.02 | 98.2  | - | 408 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 7  | 10.3 | 9.14  | 13.61 | 6.84 | 33.2 | 0.02 | 87.5  | - | 411 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 6  | 11.8 | 9.33  | 11.29 | 6.79 | 33.2 | 0.02 | 84.6  | - | 414 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 5  | 13.3 | 8.89  | 9.48  | 6.78 | 35.2 | 0.02 | 77.3  | - | 414 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 4  | 14.3 | 9.1   | 8.22  | 6.8  | 37.7 | 0.02 | 76.7  | - | 413 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 3  | 15.8 | 9.07  | 8.17  | 6.83 | 37.8 | 0.02 | 76.4  | - | 411 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 2  | 17.3 | 9.15  | 7.95  | 6.87 | 38.3 | 0.02 | 76.6  | - | 414 |
| Mid Lake | 2997 | SSJD | 7/24/97 | - | 4.5 | 1  | 18.8 | 9.24  | 7.46  | 6.93 | 39.7 | 0.03 | 76.2  | - | 422 |

|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 14 | 0.4  | 9.23  | 24.49 | 7.8  | 41.3 | 0.03 | 109.2 | - | 345 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 13 | 1    | 9.29  | 23.73 | 7.7  | 41.5 | 0.03 | 108.4 | - | 348 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 12 | 2.6  | 9.34  | 22.74 | 7.49 | 43.4 | 0.03 | 106.9 | - | 354 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 11 | 4    | 9.57  | 21.92 | 7.41 | 49.1 | 0.03 | 107.8 | - | 357 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 10 | 5.5  | 9.89  | 20.76 | 7.43 | 41.8 | 0.03 | 108.9 | - | 354 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 9  | 7    | 10.15 | 19.78 | 7.29 | 39.8 | 0.03 | 109.7 | - | 356 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 8  | 8.6  | 10.13 | 17.86 | 7.05 | 36.9 | 0.02 | 105.4 | - | 361 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 7  | 10   | 9.33  | 15.01 | 6.96 | 34.5 | 0.02 | 91.3  | - | 364 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 6  | 11.5 | 8.92  | 12.66 | 6.9  | 34.4 | 0.02 | 82.9  | - | 364 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 5  | 13.1 | 8.74  | 10.18 | 6.88 | 35.9 | 0.02 | 76.7  | - | 361 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 4  | 14.5 | 8.58  | 9.05  | 6.91 | 37.2 | 0.02 | 73.3  | - | 357 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 3  | 16.1 | 8.76  | 8.04  | 6.94 | 38.9 | 0.02 | 73    | - | 354 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 2  | 17.5 | 8.83  | 7.94  | 6.99 | 39   | 0.03 | 73.5  | - | 347 |
| Mid Lake | 3197 | DBSS | 8/5/97  | - | 5.6 | 1  | 18.9 | 8.88  | 8.18  | 7.08 | 39   | 0.03 | 74    | - | 374 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 13 | 0.5  | 8.7   | 22.93 | 7.65 | 45.4 | 0.03 | 100.6 | - | 361 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 12 | 2    | 8.72  | 22.56 | 7.6  | 45.2 | 0.03 | 100.1 | - | 362 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 11 | 3.5  | 8.77  | 22.29 | 7.56 | 45   | 0.03 | 100.2 | - | 362 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 10 | 5    | 8.81  | 22.12 | 7.44 | 45   | 0.03 | 100.4 | - | 365 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 9  | 6.5  | 8.87  | 21.44 | 7.28 | 49.4 | 0.03 | 99.7  | - | 370 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 8  | 8    | 9.05  | 20.09 | 7.19 | 42.6 | 0.03 | 99.1  | - | 373 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 7  | 9.5  | 9.2   | 18.61 | 7.05 | 38.7 | 0.02 | 97.7  | - | 374 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 6  | 11   | 8.12  | 14.03 | 6.91 | 35.4 | 0.02 | 78.3  | - | 379 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 5  | 12.5 | 7.96  | 10.62 | 6.9  | 36.8 | 0.02 | 71.1  | - | 380 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 4  | 14   | 7.79  | 9.22  | 6.93 | 38.5 | 0.02 | 67.3  | - | 378 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 3  | 15.5 | 7.71  | 8.49  | 6.97 | 39.3 | 0.03 | 65.4  | - | 376 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 2  | 17   | 7.29  | 8.15  | 7.08 | 40.1 | 0.03 | 61.3  | - | 370 |
| Mid Lake | 3297 | DBSS | 8/13/97 | - | 6.5 | 1  | 18.7 | 8.21  | 7.69  | 6.92 | 41.3 | 0.03 | 68.3  | - | 393 |
|          |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 13 | 0.4  | 9.08  | 21.49 | 7.54 | 50.5 | 0.03 | 102.3 | - | 353 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 12 | 1.1  | 9.08  | 21.48 | 7.54 | 50.4 | 0.03 | 102.2 | - | 353 |

|          |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|----------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 11 | 2.4  | 9.08 | 21.42 | 7.53 | 50.4 | 0.03 | 102.1 | - | 353 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 10 | 4.1  | 9.09 | 21.28 | 7.52 | 50.7 | 0.03 | 102   | - | 353 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 9  | 5.6  | 9.1  | 21.18 | 7.5  | 50.7 | 0.03 | 101.9 | - | 353 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 8  | 6.9  | 9.1  | 21.11 | 7.46 | 50.7 | 0.03 | 101.7 | - | 354 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 7  | 8.5  | 8.96 | 21.05 | 7.33 | 49.9 | 0.03 | 100.1 | - | 358 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 6  | 10   | 8.48 | 20.3  | 7.17 | 48.7 | 0.03 | 93.4  | - | 363 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 5  | 11.5 | 8.19 | 19.32 | 7.01 | 45.8 | 0.03 | 88.4  | - | 370 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 4  | 12.9 | 8.39 | 15.84 | 6.91 | 40.1 | 0.03 | 84.2  | - | 375 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 3  | 14.4 | 6.45 | 12.02 | 6.88 | 40.9 | 0.03 | 59.5  | - | 379 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 2  | 15.9 | 6.81 | 9.48  | 6.95 | 42.1 | 0.03 | 59.2  | - | 377 |
| Mid Lake | 3497 | DBSS | 8/27/97 | - | 8   | 1  | 17.4 | 6.69 | 9.22  | 7.07 | 42.6 | 0.03 | 57.8  | - | 373 |
|          |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 11 | 0.2  | 8.99 | 16.75 | 7.02 | 49.9 | 0.03 | 93.5  | - | 384 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 10 | 1    | 8.9  | 16.75 | 6.94 | 49.9 | 0.03 | 92.6  | - | 386 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 9  | 3    | 8.75 | 16.43 | 6.79 | 49.7 | 0.03 | 90.4  | - | 392 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 8  | 5    | 8.08 | 10.87 | 6.71 | 43.5 | 0.03 | 73.4  | - | 398 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 7  | 7    | 8.27 | 9.37  | 6.69 | 42.7 | 0.03 | 72.8  | - | 399 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 6  | 9    | 8.19 | 8.88  | 6.7  | 43.4 | 0.03 | 71.3  | - | 399 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 5  | 10.9 | 9    | 7.66  | 6.71 | 43.9 | 0.03 | 76.1  | - | 398 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 4  | 13   | 8.96 | 7.58  | 6.74 | 44.1 | 0.03 | 75.6  | - | 396 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 3  | 15   | 9.05 | 7.43  | 6.75 | 44.2 | 0.03 | 76.2  | - | 395 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 2  | 16.9 | 9.45 | 7.12  | 6.78 | 44.6 | 0.03 | 78.8  | - | 393 |
| Mid Lake | 3797 | ASRP | 9/17/97 | - | 5   | 1  | 18.6 | 9.41 | 6.89  | 6.84 | 45.1 | 0.03 | 78    | - | 390 |
|          |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 10 | 0.5  | 9.59 | 16.46 | 7.17 | 50.6 | 0.03 | 99.2  | - | 407 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 9  | 2.7  | 9.45 | 15.86 | 7.07 | 50.1 | 0.03 | 96.5  | - | 411 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 8  | 4.6  | 9.24 | 15.24 | 6.92 | 50.3 | 0.03 | 93.1  | - | 417 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 7  | 6.6  | 8.62 | 13.82 | 6.82 | 49.6 | 0.03 | 84.2  | - | 421 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 6  | 8.6  | 8.64 | 12.23 | 6.78 | 46.7 | 0.03 | 81.4  | - | 423 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 5  | 10.6 | 8.58 | 9.81  | 6.74 | 42.5 | 0.03 | 76.4  | - | 425 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 4  | 12.6 | 8.65 | 8.46  | 6.74 | 42   | 0.03 | 74.5  | - | 424 |
| Mid Lake | 3997 | ASRA | 9/29/97 | - | 6.5 | 3  | 14.6 | 9.04 | 7.79  | 6.74 | 42   | 0.03 | 77.4  | - | 424 |

|           |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
|-----------|-------|---------|----------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Mid Lake  | 3997  | ASRA    | 9/29/97  | -    | 6.5    | 2        | 16.6      | 8.92                    | 7.46            | 6.81 | 42.2                 | 0.03 | 75                 | -       | 420   |
| Mid Lake  | 3997  | ASRA    | 9/29/97  | -    | 6.5    | 1        | 18.6      | 8.91                    | 7.47            | 6.87 | 42.1                 | 0.03 | 75                 | -       | 417   |
|           |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 10       | 0.2       | 10.5                    | 11.58           | 7.09 | 49                   | 0.03 | 96                 | -       | 413   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 9        | 2         | 10.48                   | 11.59           | 7.1  | 48.9                 | 0.03 | 95.8               | -       | 412   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 8        | 4         | 10.47                   | 11.59           | 7.06 | 49.1                 | 0.03 | 95.7               | -       | 413   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 7        | 6         | 10.45                   | 11.58           | 7.02 | 49.1                 | 0.03 | 95.5               | -       | 414   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 6        | 8         | 10.43                   | 11.59           | 7    | 49.2                 | 0.03 | 95.3               | -       | 414   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 5        | 10        | 10.42                   | 11.58           | 6.94 | 49.2                 | 0.03 | 95.2               | -       | 415   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 4        | 12        | 10.4                    | 11.56           | 6.88 | 49.2                 | 0.03 | 95                 | -       | 417   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 3        | 14        | 10.37                   | 11.56           | 6.82 | 49.2                 | 0.03 | 94.8               | -       | 418   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 2        | 16        | 10.21                   | 11.51           | 6.69 | 49.3                 | 0.03 | 93.2               | -       | 421   |
| Mid Lake  | 4297  | DBAS    | 10/22/97 | -    | 4.6    | 1        | 18        | 4.53                    | 8.53            | 6.32 | 45.1                 | 0.03 | 38.5               | -       | 434   |
|           |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 13       | 0.5       | 10.64                   | 10.18           | 7.08 | 53.4                 | 0.03 | 93.8               | 104     | 396   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 12       | 2         | 10.49                   | 9.51            | 7.02 | 53.4                 | 0.03 | 91                 | 56      | 398   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 11       | 3.5       | 10.25                   | 9.4             | 6.97 | 53.7                 | 0.03 | 88.7               | 55      | 401   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 10       | 5         | 10.05                   | 9.32            | 6.94 | 53.7                 | 0.03 | 86.7               | 49      | 401   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 9        | 6.5       | 9.89                    | 9.24            | 6.9  | 53.4                 | 0.03 | 85.2               | 120     | 402   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 8        | 8         | 8.86                    | 9.01            | 6.85 | 52.9                 | 0.03 | 75.9               | 119     | 403   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 7        | 9.5       | 9.57                    | 8.71            | 6.85 | 54.6                 | 0.03 | 81.4               | 102     | 403   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 6        | 11        | 9.28                    | 8.6             | 6.84 | 53.2                 | 0.03 | 78.8               | 58      | 403   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 5        | 12.5      | 9.13                    | 8.5             | 6.79 | 52.6                 | 0.03 | 77.3               | 102     | 405   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 4        | 14        | 9.01                    | 8.43            | 6.75 | 52.1                 | 0.03 | 76.2               | 227     | 406   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 3        | 15.5      | 8.76                    | 7.68            | 6.72 | 47.9                 | 0.03 | 72.6               | 51      | 407   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 2        | 17        | 8.81                    | 7.38            | 6.72 | 46.3                 | 0.03 | 72.6               | 58      | 408   |
| Mid Lake  | 4497  | DBAS    | 11/4/97  | -    | 4      | 1        | 18.4      | 8.78                    | 7.27            | 6.75 | 46.2                 | 0.03 | 72.1               | 713     | 407   |
| Location  | Phase | Sampler | Date     | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
| Carey Bay | 1597  | ASSS    | 4/18/97  | -    | 1.9    | 10       | 0.4       | 12.31                   | 6.22            | 7.43 | 38.4                 | 0.02 | 99.4               | -       | 374   |
| Carey Bay | 1597  | ASSS    | 4/18/97  | -    | 1.9    | 9        | 1         | 12.3                    | 6.07            | 7.41 | 38.5                 | 0.02 | 99                 | -       | 375   |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 8  | 2    | 12.29 | 5.77  | 7.41 | 38.2 | 0.02 | 98.1  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 7  | 3.1  | 12.21 | 5.64  | 7.41 | 38.4 | 0.02 | 97.2  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 6  | 4.3  | 12.2  | 5.6   | 7.4  | 38.3 | 0.02 | 97    | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 5  | 5.5  | 12.14 | 5.57  | 7.39 | 38.2 | 0.02 | 96.6  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 4  | 6.8  | 12.15 | 5.56  | 7.39 | 38.3 | 0.02 | 96.5  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 3  | 8.1  | 12.12 | 5.54  | 7.38 | 38.4 | 0.02 | 96.2  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 2  | 9.3  | 12.12 | 5.49  | 7.38 | 38.4 | 0.02 | 96.1  | - | 374 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 1  | 10.5 | 12.09 | 5.47  | 7.37 | 38.5 | 0.02 | 95.8  | - | 374 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 12 | 0.2  | 11.52 | 15.54 | 7.11 | 35.1 | 0.02 | 114.7 | - | 389 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 11 | 1    | 11.94 | 11.22 | 7.15 | 33.5 | 0.02 | 107.9 | - | 390 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 10 | 2    | 12.11 | 10.1  | 7.1  | 33.6 | 0.02 | 106.7 | - | 392 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 9  | 3.5  | 11.57 | 9.45  | 7.05 | 32.8 | 0.02 | 100.3 | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 8  | 5    | 11.39 | 9.02  | 7.02 | 33.9 | 0.02 | 97.7  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 7  | 6.5  | 11.37 | 8.88  | 7.01 | 34.2 | 0.02 | 97.2  | - | 396 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 6  | 8    | 11.34 | 8.69  | 7.01 | 34.6 | 0.02 | 96.6  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 5  | 9.5  | 11.36 | 8.33  | 7    | 34.4 | 0.02 | 95.8  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 4  | 11   | 11.32 | 8.07  | 6.99 | 34.8 | 0.02 | 95    | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 3  | 12.5 | 11.24 | 7.83  | 6.98 | 34.8 | 0.02 | 93.7  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 2  | 14   | 10.99 | 7.61  | 6.96 | 35.2 | 0.02 | 91    | - | 396 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 1  | 15.3 | 10.38 | 6.86  | 6.95 | 37.3 | 0.02 | 84.5  | - | 396 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 14 | 0.3  | 10.67 | 14.15 | 6.97 | 28.5 | 0.02 | 103.3 | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 13 | 1    | 11.21 | 10.91 | 6.99 | 26.2 | 0.02 | 100.9 | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 12 | 2.1  | 10.88 | 10.38 | 6.97 | 26.3 | 0.02 | 96.7  | - | 367 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 11 | 3.1  | 10.63 | 10.2  | 6.94 | 26.3 | 0.02 | 94.1  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 10 | 4.1  | 10.64 | 9.86  | 6.91 | 26.2 | 0.02 | 93.3  | - | 369 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 9  | 5.1  | 10.43 | 9.79  | 6.92 | 26.4 | 0.02 | 91.3  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 8  | 6.2  | 10.43 | 9.74  | 6.93 | 26.5 | 0.02 | 91.3  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 7  | 7.2  | 10.61 | 9.64  | 6.92 | 26.3 | 0.02 | 92.6  | - | 367 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 6  | 8.2  | 10.61 | 9.53  | 6.94 | 26.4 | 0.02 | 92.4  | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 5  | 9.2  | 10.64 | 9.43  | 6.93 | 26.4 | 0.02 | 92.4  | - | 366 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 4  | 10.2 | 10.64 | 9.3   | 6.93 | 26.5 | 0.02 | 92.1  | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 3  | 11.2 | 10.61 | 9.12  | 6.94 | 26.7 | 0.02 | 91.5  | - | 365 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 2  | 12.3 | 10.56 | 8.86  | 6.93 | 27.5 | 0.02 | 90.3  | - | 365 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 1  | 13.2 | 10.64 | 8.66  | 6.94 | 27.8 | 0.02 | 90.7  | - | 364 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 14 | 0.7  | 10.87 | 15.48 | 7.17 | 31.2 | 0.02 | 108.7 | - | 411 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 13 | 1.6  | 11.2  | 14.35 | 7.1  | 30.9 | 0.02 | 109.3 | - | 414 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 2.5  | 11.22 | 11.79 | 7.09 | 29.1 | 0.02 | 103.4 | - | 415 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 3.5  | 11.32 | 11.3  | 7.07 | 28.9 | 0.02 | 103.2 | - | 416 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 4.5  | 11.27 | 11.22 | 7.06 | 28.8 | 0.02 | 102.5 | - | 417 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 5.5  | 10.97 | 10.99 | 7.03 | 28.8 | 0.02 | 99.2  | - | 418 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 6.5  | 10.9  | 10.38 | 7.01 | 28.8 | 0.02 | 97.2  | - | 419 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 7.5  | 10.84 | 10.17 | 7.01 | 28.7 | 0.02 | 96.2  | - | 419 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 8.5  | 10.69 | 9.91  | 6.98 | 28.9 | 0.02 | 94.2  | - | 420 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 9.5  | 10.63 | 9.45  | 6.98 | 29.4 | 0.02 | 92.7  | - | 420 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 10.5 | 10.61 | 9.22  | 6.97 | 30.2 | 0.02 | 92    | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 11.4 | 10.58 | 8.99  | 6.97 | 31.1 | 0.02 | 91.2  | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 12.5 | 10.56 | 8.84  | 6.96 | 31.9 | 0.02 | 90.6  | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 13.4 | 10.14 | 8.11  | 6.95 | 36.1 | 0.02 | 85.6  | - | 423 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 12 | 0.3  | 10.81 | 14.98 | 7.2  | 30.4 | 0.02 | 106.8 | - | 427 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 0.9  | 10.81 | 14.93 | 7.17 | 30.4 | 0.02 | 106.8 | - | 429 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 1.8  | 10.85 | 14.49 | 7.11 | 30.2 | 0.02 | 106.1 | - | 432 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 2.8  | 10.91 | 13.73 | 7.12 | 29.9 | 0.02 | 104.9 | - | 431 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 3.8  | 10.94 | 13.11 | 7.1  | 29.8 | 0.02 | 103.8 | - | 432 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 4.8  | 10.91 | 12.73 | 7.06 | 29.6 | 0.02 | 102.6 | - | 434 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 5.7  | 10.84 | 12.33 | 6.99 | 29.8 | 0.02 | 101.1 | - | 436 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 6.8  | 10.61 | 11.74 | 6.94 | 29.5 | 0.02 | 97.5  | - | 438 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 7.7  | 10.42 | 10.66 | 6.92 | 29.3 | 0.02 | 93.4  | - | 440 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 3  | 8.7  | 10.01 | 10.1  | 6.89 | 29.9 | 0.02 | 88.6  | - | 441 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 2  | 9.7  | 9.88  | 9.97  | 6.91 | 30.4 | 0.02 | 87.1  | - | 440 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 1  | 10.6 | 9.78  | 9.66  | 6.92 | 31.3 | 0.02 | 85.6  | - | 440 |



|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 13 | 0.3  | 10.09 | 18.7  | 7.18 | 36.8 | 0.02 | ***** | - | 444 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 12 | 1.1  | 10.09 | 18.66 | 7.17 | 36.9 | 0.02 | ***** | - | 444 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 11 | 1.9  | 10.24 | 17.47 | 7.13 | 36.3 | 0.02 | ***** | - | 446 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 10 | 2.9  | 10.39 | 16.41 | 7.09 | 37.1 | 0.02 | ***** | - | 449 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 9  | 3.9  | 10.25 | 16.19 | 7.04 | 37.4 | 0.02 | ***** | - | 450 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 8  | 5    | 10.47 | 15.01 | 7.01 | 36.9 | 0.02 | ***** | - | 452 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 7  | 5.9  | 10.37 | 14.42 | 6.97 | 35.9 | 0.02 | ***** | - | 454 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 6  | 7.1  | 10.16 | 13.17 | 6.94 | 34.3 | 0.02 | 96.2  | - | 455 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 5  | 8    | 10    | 12.81 | 6.93 | 34.2 | 0.02 | 93.9  | - | 456 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 4  | 9    | 9.91  | 12.15 | 6.91 | 33.4 | 0.02 | 91.7  | - | 457 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 3  | 10   | 9.55  | 11.25 | 6.9  | 33   | 0.02 | 86.5  | - | 458 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 2  | 11   | 9.03  | 9.98  | 6.91 | 34.7 | 0.02 | 79.4  | - | 460 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 1  | 11.9 | 9.2   | 9.77  | 6.97 | 35.1 | 0.02 | 80.4  | - | 457 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 14 | 0.3  | 9.94  | 20.53 | 7.53 | 40   | 0.03 | ***** | - | 408 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 13 | 0.8  | 9.99  | 20.55 | 7.51 | 40   | 0.03 | ***** | - | 408 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 12 | 1.8  | 10.43 | 19.63 | 7.5  | 39.7 | 0.03 | ***** | - | 410 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 11 | 2.8  | 10.53 | 18.87 | 7.48 | 38.4 | 0.02 | ***** | - | 410 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 10 | 3.7  | 10.45 | 18.25 | 7.34 | 38.9 | 0.02 | ***** | - | 414 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 9  | 4.8  | 10.36 | 17.29 | 7.24 | 37.6 | 0.02 | ***** | - | 418 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 8  | 5.8  | 10.3  | 16.48 | 7.19 | 36.7 | 0.02 | ***** | - | 419 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 7  | 6.7  | 10.29 | 16.07 | 7.16 | 36   | 0.02 | ***** | - | 419 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 6  | 7.9  | 9.77  | 15.61 | 7.07 | 35.2 | 0.02 | 97.7  | - | 422 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 5  | 8.4  | 10.09 | 15.67 | 7.02 | 35.3 | 0.02 | ***** | - | 422 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 4  | 9.4  | 9.66  | 14.52 | 6.98 | 33.7 | 0.02 | 94.3  | - | 423 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 3  | 10.4 | 9.48  | 13.88 | 6.94 | 33.3 | 0.02 | 91.2  | - | 424 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 2  | 11.4 | 8.87  | 12.67 | 6.91 | 33.1 | 0.02 | 83    | - | 426 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 1  | 12.4 | 7.87  | 10.18 | 6.93 | 36.1 | 0.02 | 69.6  | - | 427 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 12 | 0.3  | 9.47  | 23.01 | 7.6  | 43.4 | 0.03 | 109   | - | 364 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 11 | 1    | 9.51  | 22.72 | 7.59 | 43.5 | 0.03 | 108.8 | - | 365 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 10 | 2.1  | 9.61  | 22.15 | 7.53 | 43.8 | 0.03 | 108.7 | - | 366 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 9  | 3    | 9.66  | 21.88 | 7.5  | 42.6 | 0.03 | 108.8 | - | 367 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 8  | 3.9  | 9.81  | 21.53 | 7.47 | 43.8 | 0.03 | 109.7 | - | 368 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 7  | 4.9  | 9.89  | 21.11 | 7.42 | 43   | 0.03 | 109.5 | - | 369 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 6  | 6    | 9.91  | 20.84 | 7.34 | 41.5 | 0.03 | 109.4 | - | 371 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 5  | 6.9  | 9.66  | 19.4  | 7.27 | 39.5 | 0.03 | 103.6 | - | 374 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 4  | 7.9  | 10.34 | 18.36 | 7.26 | 37.2 | 0.02 | 108.6 | - | 373 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 3  | 9    | 9.68  | 17.1  | 7.14 | 36.6 | 0.02 | 99.1  | - | 377 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 2  | 10   | 8.87  | 15.07 | 7.06 | 35.5 | 0.02 | 86.9  | - | 380 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 1  | 11   | 7.89  | 14.19 | 7.1  | 35.6 | 0.02 | 75.9  | - | 381 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 13 | 0.4  | 8.59  | 22.47 | 7.5  | 45.1 | 0.03 | 98.4  | - | 378 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 12 | 1.4  | 8.6   | 22.43 | 7.54 | 45.1 | 0.03 | 98.5  | - | 376 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 11 | 2.4  | 8.65  | 22.33 | 7.47 | 45.1 | 0.03 | 98.9  | - | 377 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 10 | 3.3  | 8.67  | 22.22 | 7.42 | 45.3 | 0.03 | 98.8  | - | 378 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 9  | 4.3  | 8.66  | 21.9  | 7.3  | 45.9 | 0.03 | 98.2  | - | 383 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 8  | 5.3  | 8.61  | 21.55 | 7.25 | 45.7 | 0.03 | 97    | - | 385 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 7  | 6.4  | 8.62  | 21.19 | 7.2  | 45.6 | 0.03 | 96.4  | - | 386 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 6  | 7.4  | 8.39  | 20.53 | 7.15 | 43   | 0.03 | 92.6  | - | 388 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 5  | 8.4  | 8.64  | 19.56 | 7.01 | 40.1 | 0.03 | 93.5  | - | 391 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 4  | 9.4  | 8.27  | 18.94 | 6.95 | 39.7 | 0.03 | 88.4  | - | 393 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 3  | 10.4 | 8.26  | 18.41 | 6.88 | 38.9 | 0.02 | 87.5  | - | 395 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 2  | 11.5 | 7.89  | 15.96 | 6.83 | 36.5 | 0.02 | 79.3  | - | 398 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 1  | 12.4 | 6.74  | 13.07 | 6.82 | 37.1 | 0.02 | 62.7  | - | 401 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 14 | 0.3  | 8.97  | 21.76 | 7.5  | 49.7 | 0.03 | 101.6 | - | 356 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 13 | 0.9  | 8.98  | 21.74 | 7.52 | 49.6 | 0.03 | 101.6 | - | 355 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 12 | 1.9  | 8.99  | 21.67 | 7.5  | 49.6 | 0.03 | 101.6 | - | 356 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 11 | 2.9  | 9     | 21.53 | 7.45 | 49.8 | 0.03 | 101.5 | - | 358 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 10 | 3.9  | 8.93  | 21.37 | 7.43 | 49.3 | 0.03 | 100.4 | - | 358 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 9  | 4.9  | 8.91  | 21.26 | 7.39 | 49.3 | 0.03 | 99.9  | - | 359 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 8  | 6.1  | 8.89  | 21.09 | 7.36 | 49.3 | 0.03 | 99.4  | - | 360 |

|           |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
|-----------|------|------|---------|---|-----|----|------|------|-------|------|------|------|------|---|-----|
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 7  | 7    | 8.95 | 20.91 | 7.34 | 49.6 | 0.03 | 99.7 | - | 360 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 6  | 8    | 8.92 | 20.84 | 7.28 | 49.5 | 0.03 | 99.2 | - | 362 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 5  | 9    | 8.84 | 20.7  | 7.18 | 49.1 | 0.03 | 98.1 | - | 365 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 4  | 10   | 8.6  | 20.42 | 7.08 | 48.3 | 0.03 | 95   | - | 370 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 3  | 11   | 8.16 | 16.21 | 6.99 | 40.9 | 0.03 | 82.6 | - | 376 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 2  | 11.9 | 8.02 | 15.24 | 7.04 | 40.2 | 0.03 | 79.5 | - | 374 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 1  | 12.9 | 8.03 | 14.64 | 7.15 | 40.2 | 0.03 | 78.6 | - | 372 |
|           |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 13 | 0.5  | 8.05 | 14.69 | 6.76 | 44.3 | 0.03 | 80.1 | - | 408 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 12 | 1.5  | 7.89 | 14.03 | 6.72 | 43.7 | 0.03 | 77.4 | - | 411 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 11 | 2.5  | 7.64 | 11.22 | 6.71 | 42.5 | 0.03 | 70.3 | - | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 10 | 3.5  | 7.84 | 9.99  | 6.71 | 42.4 | 0.03 | 70.1 | - | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 9  | 4.6  | 8.21 | 9.33  | 6.7  | 42.9 | 0.03 | 72.1 | - | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 8  | 5.5  | 8.13 | 9.17  | 6.71 | 42.9 | 0.03 | 71.1 | - | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 7  | 6.5  | 8.2  | 9.1   | 6.72 | 43.1 | 0.03 | 71.8 | - | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 6  | 7.5  | 8.27 | 8.99  | 6.73 | 43.2 | 0.03 | 72.2 | - | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 5  | 8.5  | 8.31 | 8.89  | 6.75 | 43.3 | 0.03 | 72.3 | - | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 4  | 9.3  | 8.34 | 8.78  | 6.75 | 43.3 | 0.03 | 72.3 | - | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 3  | 10.7 | 8.67 | 8.6   | 6.77 | 43.5 | 0.03 | 75.2 | - | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 2  | 11.5 | 8.4  | 8.6   | 6.78 | 43.3 | 0.03 | 72.7 | - | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97 | - | 3.5 | 1  | 12.5 | 8.46 | 8.48  | 6.82 | 43.4 | 0.03 | 73.3 | - | 412 |
|           |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 13 | 0.2  | 9.16 | 16.28 | 7.09 | 50   | 0.03 | 94.3 | - | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 12 | 1    | 9.09 | 15.89 | 7.08 | 49.9 | 0.03 | 92.9 | - | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 11 | 2    | 9.02 | 15.58 | 7.07 | 49.9 | 0.03 | 91.6 | - | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 10 | 3    | 8.95 | 15.44 | 7.04 | 50   | 0.03 | 90.6 | - | 412 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 9  | 4    | 8.83 | 15.34 | 6.99 | 49.9 | 0.03 | 89.2 | - | 413 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 8  | 5    | 8.59 | 15.11 | 6.92 | 49.4 | 0.03 | 86.3 | - | 415 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 7  | 6    | 7.93 | 14.51 | 6.8  | 47.5 | 0.03 | 78.7 | - | 419 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 6  | 7    | 7.68 | 12.41 | 6.77 | 44.2 | 0.03 | 72.7 | - | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 5  | 8    | 7.63 | 12.32 | 6.77 | 44.2 | 0.03 | 72.1 | - | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97 | - | 6.3 | 4  | 9    | 7.56 | 12.14 | 6.77 | 43.9 | 0.03 | 71.1 | - | 420 |

|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|-----------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 3  | 10   | 7.3   | 10.38 | 6.75 | 42.4 | 0.03 | 65.9 | -   | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 2  | 11   | 7.79  | 8.86  | 6.84 | 42.2 | 0.03 | 67.8 | -   | 419 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 1  | 12   | 8.33  | 8.37  | 6.93 | 42.3 | 0.03 | 71.6 | -   | 416 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 10 | 0.3  | 10.25 | 11.54 | 7.25 | 49   | 0.03 | 93.6 | -   | 401 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 9  | 1.7  | 10.24 | 11.56 | 7.23 | 49.1 | 0.03 | 93.6 | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 8  | 2.7  | 10.24 | 11.54 | 7.24 | 49.2 | 0.03 | 93.5 | -   | 401 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 7  | 3.7  | 10.21 | 11.54 | 7.21 | 49.2 | 0.03 | 93.2 | -   | 403 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 6  | 4.7  | 10.18 | 11.54 | 7.23 | 49.1 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 5  | 5.7  | 10.19 | 11.53 | 7.22 | 49.2 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 4  | 6.7  | 10.19 | 11.53 | 7.22 | 49.3 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 3  | 7.7  | 10.19 | 11.53 | 7.22 | 49.3 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 2  | 8.7  | 10.1  | 11.53 | 7.21 | 49.3 | 0.03 | 92.2 | -   | 403 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 1  | 9.7  | 10.05 | 11.51 | 7.19 | 49.2 | 0.03 | 91.7 | -   | 404 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 13 | 0.4  | 10.06 | 9.87  | 7.07 | 53.3 | 0.03 | 88   | 56  | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 12 | 1.5  | 10.34 | 9.55  | 7.09 | 53.1 | 0.03 | 89.8 | 215 | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 11 | 2.5  | 10.46 | 9.5   | 7.12 | 52.9 | 0.03 | 90.7 | 54  | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 10 | 3.5  | 10.5  | 9.53  | 7.11 | 53   | 0.03 | 91.1 | 41  | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 9  | 4.5  | 10.5  | 9.45  | 7.12 | 52.9 | 0.03 | 90.9 | 56  | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 8  | 5.5  | 10.45 | 9.4   | 7.11 | 53   | 0.03 | 90.4 | 113 | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 7  | 6.5  | 10.28 | 9.33  | 7.06 | 53   | 0.03 | 88.8 | 103 | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 6  | 7.5  | 10.14 | 9.22  | 7.01 | 52.9 | 0.03 | 87.3 | 210 | 392 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 5  | 8.5  | 9.65  | 8.89  | 6.93 | 53.1 | 0.03 | 82.5 | 101 | 393 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 4  | 9.5  | 9.51  | 8.81  | 6.89 | 52.9 | 0.03 | 81.1 | 54  | 394 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 3  | 10.6 | 9.08  | 8.58  | 6.86 | 52.4 | 0.03 | 77   | 37  | 395 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 2  | 11.5 | 8.98  | 8.47  | 6.87 | 52.1 | 0.03 | 76   | 111 | 394 |

| Location       | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|----------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Conkling Point | 1597  | ASSS    | 4/18/97 | -    | 1.1    | 13       | 0.4       | 12.35                   | 6.31            | 7.45 | 36.4                 | 0.02 | 99.9               | -       | 372   |
| Conkling Point | 1597  | ASSS    | 4/18/97 | -    | 1.1    | 12       | 1.9       | 12.27                   | 6.23            | 7.43 | 36.4                 | 0.02 | 99.2               | -       | 373   |

|                |      |      |         |   |     |    |      |       |       |      |      |      |      |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|------|---|-----|
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 11 | 3.2  | 12.29 | 6.2   | 7.43 | 36.6 | 0.02 | 99.2 | - | 373 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 10 | 4.5  | 12.27 | 6.13  | 7.43 | 36.7 | 0.02 | 98.7 | - | 373 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 9  | 5.7  | 12.23 | 6.05  | 7.42 | 37   | 0.02 | 98.4 | - | 373 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 8  | 7    | 12.26 | 5.97  | 7.41 | 37.4 | 0.02 | 98.4 | - | 374 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 7  | 8.2  | 12.24 | 5.97  | 7.4  | 37.3 | 0.02 | 98.2 | - | 373 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 6  | 9.4  | 12.26 | 5.9   | 7.4  | 37.6 | 0.02 | 98.2 | - | 373 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 5  | 10.4 | 12.24 | 5.82  | 7.39 | 37.8 | 0.02 | 97.9 | - | 374 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 4  | 11.8 | 12.21 | 5.62  | 7.38 | 38.2 | 0.02 | 97.1 | - | 374 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 3  | 13.1 | 12.2  | 5.57  | 7.37 | 38.3 | 0.02 | 96.9 | - | 374 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 2  | 14.4 | 12.14 | 5.51  | 7.36 | 38.2 | 0.02 | 96.3 | - | 374 |
| Conkling Point | 1597 | ASSS | 4/18/97 | - | 1.1 | 1  | 15.9 | 12.16 | 5.29  | 7.36 | 38.5 | 0.02 | 95.9 | - | 374 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |      |   |     |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 12 | 0.2  | 10.99 | 11.35 | 6.99 | 28   | 0.02 | 99.6 | - | 367 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 11 | 2.1  | 11.1  | 9.61  | 6.98 | 27.8 | 0.02 | 96.7 | - | 369 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 10 | 3.9  | 11.09 | 9.46  | 7    | 27.8 | 0.02 | 96.2 | - | 368 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 9  | 5.5  | 11.08 | 9.42  | 6.98 | 27.7 | 0.02 | 96   | - | 369 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 8  | 7    | 11.05 | 9.35  | 6.97 | 27.9 | 0.02 | 95.6 | - | 369 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 7  | 8.6  | 11.04 | 9.28  | 6.98 | 28.2 | 0.02 | 95.3 | - | 368 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 6  | 10.1 | 11.03 | 9.28  | 6.97 | 28.2 | 0.02 | 95.3 | - | 368 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 5  | 11.6 | 11.01 | 9.15  | 6.97 | 28.6 | 0.02 | 94.7 | - | 368 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 4  | 13   | 11.03 | 8.35  | 6.95 | 31.4 | 0.02 | 93.1 | - | 369 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 3  | 14.5 | 11.15 | 7.45  | 6.93 | 34.8 | 0.02 | 92.1 | - | 369 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 2  | 16   | 10.71 | 6.08  | 6.89 | 41   | 0.03 | 85.6 | - | 372 |
| Conkling Point | 1997 | DBAC | 5/16/97 | - | 0.8 | 1  | 17.5 | 10.51 | 5.9   | 6.89 | 44.8 | 0.03 | 83.4 | - | 370 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |      |   |     |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 12 | 0.2  | 11.09 | 11.41 | 7.01 | 26.9 | 0.02 | 101  | - | 362 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 11 | 1.4  | 11.11 | 10.79 | 7.02 | 26.9 | 0.02 | 99.7 | - | 363 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 10 | 2.9  | 11.16 | 10.36 | 7    | 26.8 | 0.02 | 99   | - | 364 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 9  | 4.5  | 11.12 | 10.22 | 6.98 | 26.8 | 0.02 | 98.4 | - | 365 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 8  | 6    | 11.12 | 10.05 | 6.96 | 26.9 | 0.02 | 98   | - | 365 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 7  | 7.4  | 11.05 | 9.74  | 6.94 | 26.8 | 0.02 | 96.6 | - | 364 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 6  | 8.9  | 11.01 | 9.61  | 6.93 | 26.9 | 0.02 | 96   | - | 364 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 5  | 10.4 | 10.97 | 9.43  | 6.92 | 26.9 | 0.02 | 95.3  | - | 363 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 4  | 12.1 | 10.89 | 8.96  | 6.91 | 27.2 | 0.02 | 93.5  | - | 362 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 3  | 13.4 | 10.75 | 8.71  | 6.9  | 27.8 | 0.02 | 91.8  | - | 362 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 2  | 15   | 10.75 | 7.97  | 6.88 | 31.3 | 0.02 | 90.1  | - | 361 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 1  | 16.8 | 10.29 | 7.01  | 6.86 | 36   | 0.02 | 84.3  | - | 360 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 0.2  | 10.9  | 15.61 | 7.09 | 30.8 | 0.02 | 109.4 | - | 395 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 1    | 10.7  | 14.08 | 7.08 | 30.1 | 0.02 | 103.8 | - | 396 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 2.5  | 10.95 | 12.4  | 7.06 | 29.4 | 0.02 | 102.2 | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 4    | 10.98 | 11.91 | 7.04 | 29.1 | 0.02 | 101.5 | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 5.3  | 10.87 | 11.77 | 6.99 | 29.2 | 0.02 | 100.1 | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 6.9  | 10.85 | 10.68 | 6.99 | 29.6 | 0.02 | 97.3  | - | 400 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 8.3  | 10.82 | 9.94  | 6.98 | 29.1 | 0.02 | 95.5  | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 9.9  | 10.78 | 9.78  | 6.97 | 28.9 | 0.02 | 94.7  | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 11.4 | 10.73 | 9.33  | 6.96 | 29.8 | 0.02 | 93.3  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 13   | 10.53 | 8.72  | 6.94 | 32.2 | 0.02 | 90.2  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 14.5 | 10.48 | 8.47  | 6.93 | 33.3 | 0.02 | 89.3  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 16.2 | 10.17 | 7.86  | 6.93 | 37.1 | 0.02 | 85.3  | - | 395 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 16 | 0.4  | 10.79 | 15.18 | 7.23 | 30.6 | 0.02 | 107.1 | - | 421 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 15 | 1    | 10.79 | 15.14 | 7.23 | 30.6 | 0.02 | 107.1 | - | 422 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 14 | 2.2  | 10.8  | 14.94 | 7.21 | 30.7 | 0.02 | 106.7 | - | 424 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 13 | 3.3  | 10.82 | 14.66 | 7.15 | 31   | 0.02 | 106.2 | - | 426 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 12 | 4.2  | 10.75 | 14.46 | 7.11 | 30.9 | 0.02 | 105   | - | 427 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 5.2  | 10.87 | 13.07 | 7.06 | 31.2 | 0.02 | 103   | - | 430 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 6.2  | 10.75 | 12.66 | 7.02 | 30.8 | 0.02 | 101   | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 7.2  | 10.78 | 12.17 | 7    | 30.6 | 0.02 | 100.1 | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 8.2  | 10.73 | 12    | 6.95 | 30.8 | 0.02 | 99.3  | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 9.3  | 10.02 | 10.77 | 6.9  | 29.5 | 0.02 | 90    | - | 434 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 10.2 | 10.21 | 9.2   | 6.83 | 31.2 | 0.02 | 88.4  | - | 435 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 11.3 | 9.97  | 8.66  | 6.8  | 32.7 | 0.02 | 85.2  | - | 435 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 12.4 | 9.36  | 8.27  | 6.78 | 35.8 | 0.02 | 79.3  | - | 436 |



|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 15 | 0.3  | 9.57  | 24    | 8.18 | 41.5 | 0.03 | 112.2 | - | 329 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 14 | 1.8  | 9.56  | 22.25 | 7.5  | 43.2 | 0.03 | 108.4 | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 13 | 2.7  | 9.49  | 22.04 | 7.46 | 42.3 | 0.03 | 107.2 | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 12 | 3.6  | 9.48  | 21.7  | 7.39 | 41.4 | 0.03 | 106.4 | - | 352 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 11 | 4.7  | 9.88  | 21.04 | 7.33 | 43.5 | 0.03 | 109.5 | - | 352 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 10 | 5.6  | 9.4   | 20.23 | 7.19 | 40.7 | 0.03 | 102.5 | - | 356 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 9  | 6.7  | 9.32  | 19.21 | 7.13 | 39.5 | 0.03 | 99.6  | - | 356 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 8  | 7.7  | 9.69  | 18.15 | 7.13 | 37.4 | 0.02 | 101.4 | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 7  | 8.7  | 10.04 | 17.18 | 7.09 | 36.3 | 0.02 | 103   | - | 353 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 6  | 9.7  | 9.28  | 15.82 | 6.99 | 35.4 | 0.02 | 92.4  | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 5  | 10.7 | 8.73  | 14.39 | 6.93 | 34.7 | 0.02 | 84.3  | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 4  | 11.8 | 8.22  | 12.28 | 6.91 | 33.7 | 0.02 | 75.9  | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 3  | 12.6 | 8.21  | 11.23 | 6.93 | 34.5 | 0.02 | 73.8  | - | 346 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 2  | 13.7 | 8     | 10.16 | 6.95 | 35.9 | 0.02 | 70.2  | - | 342 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 1  | 14.8 | 7.22  | 9.44  | 7.02 | 38.4 | 0.02 | 62.2  | - | 388 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 11 | 0.3  | 8.64  | 22.75 | 7.98 | 44.7 | 0.03 | 99.5  | - | 368 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 10 | 2.1  | 8.73  | 22.56 | 8.11 | 44.7 | 0.03 | 100.3 | - | 361 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 9  | 3.6  | 8.81  | 22.33 | 7.87 | 44.3 | 0.03 | 100.7 | - | 363 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 8  | 5    | 8.57  | 21.86 | 7.31 | 45.2 | 0.03 | 97.1  | - | 384 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 7  | 6.5  | 8.56  | 21.26 | 7.14 | 46.4 | 0.03 | 95.9  | - | 389 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 6  | 8    | 8.56  | 20.65 | 7.06 | 43.6 | 0.03 | 94.7  | - | 391 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 5  | 9.6  | 8.58  | 18.75 | 6.94 | 39   | 0.03 | 91.4  | - | 396 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 4  | 11   | 7.69  | 15.02 | 6.9  | 35.7 | 0.02 | 75.8  | - | 399 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 3  | 12.6 | 6.19  | 15.07 | 6.87 | 36.5 | 0.02 | 77    | - | 399 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 2  | 14   | 6.57  | 9.2   | 6.95 | 39.1 | 0.03 | 56.7  | - | 401 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 1  | 15.6 | 6.79  | 8.9   | 7.08 | 39.3 | 0.03 | 58.2  | - | 398 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 11 | 0.4  | 9.02  | 21.92 | 7.53 | 50   | 0.03 | 102.3 | - | 346 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 10 | 2    | 9.03  | 21.7  | 7.51 | 50   | 0.03 | 102.1 | - | 346 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 9  | 3.6  | 9     | 21.56 | 7.45 | 50   | 0.03 | 101.6 | - | 347 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 8  | 5    | 9.01  | 21.44 | 7.4  | 50.3 | 0.03 | 101.4 | - | 349 |



|                |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
|----------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|------|---|-----|
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 7  | 6.6  | 8.84 | 21.14 | 7.33 | 49.2 | 0.03 | 98.9 | - | 350 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 6  | 8    | 8.73 | 21    | 7.17 | 49.5 | 0.03 | 97.4 | - | 355 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 5  | 9.5  | 8.19 | 20.49 | 6.96 | 49.5 | 0.03 | 90.5 | - | 363 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 4  | 11   | 7.84 | 17.3  | 6.89 | 42.7 | 0.03 | 81.2 | - | 368 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 3  | 12.6 | 8.37 | 15.61 | 6.88 | 40.2 | 0.03 | 83.6 | - | 368 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 2  | 14.1 | 7.65 | 12.49 | 6.86 | 39.9 | 0.03 | 71.3 | - | 369 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 1  | 15.8 | 6.43 | 9.48  | 6.87 | 42.8 | 0.03 | 56   | - | 369 |
|                |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 16 | 0.2  | 8.46 | 14.83 | 6.87 | 47.6 | 0.03 | 84.4 | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 15 | 1    | 8.42 | 14.86 | 6.86 | 47.6 | 0.03 | 84.1 | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 14 | 2    | 8.3  | 14.58 | 6.79 | 46.9 | 0.03 | 82.4 | - | 395 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 13 | 3    | 7.76 | 13.75 | 6.68 | 45.7 | 0.03 | 76.9 | - | 399 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 12 | 4    | 7.6  | 10.69 | 6.62 | 42.6 | 0.03 | 69.1 | - | 404 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 11 | 5    | 7.76 | 9.81  | 6.62 | 42.7 | 0.03 | 69.1 | - | 404 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 10 | 6    | 7.91 | 9.5   | 6.63 | 42.7 | 0.03 | 69.9 | - | 403 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 9  | 7    | 7.78 | 9.43  | 6.63 | 43.1 | 0.03 | 68.6 | - | 402 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 8  | 8    | 7.61 | 9.37  | 6.63 | 42.9 | 0.03 | 67.2 | - | 402 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 7  | 9    | 7.5  | 9.06  | 6.62 | 43   | 0.03 | 65.7 | - | 401 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 6  | 10   | 7.18 | 8.74  | 6.63 | 43.3 | 0.03 | 62.5 | - | 400 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 5  | 11   | 7.35 | 8.68  | 6.66 | 43.4 | 0.03 | 63.7 | - | 398 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 4  | 12   | 8.42 | 8.29  | 6.71 | 43.8 | 0.03 | 72.1 | - | 395 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 3  | 13   | 8.59 | 8.2   | 6.74 | 43.8 | 0.03 | 74.2 | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 2  | 14.1 | 8.63 | 8.19  | 6.76 | 43.8 | 0.03 | 73.9 | - | 391 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 1  | 15.2 | 8.7  | 8.19  | 6.79 | 43.8 | 0.03 | 74.5 | - | 389 |
|                |      |      |         |   |     |    |      |      |       |      |      |      |      |   |     |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 10 | 0.2  | 9.53 | 16.26 | 7.21 | 50.8 | 0.03 | 98.2 | - | 388 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 9  | 1.5  | 9.3  | 15.79 | 7.13 | 50.3 | 0.03 | 94.8 | - | 391 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 8  | 3    | 9.24 | 15.61 | 7.06 | 50.2 | 0.03 | 93.9 | - | 392 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 7  | 4.5  | 8.78 | 14.69 | 6.94 | 48.9 | 0.03 | 87.4 | - | 395 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 6  | 6    | 8.32 | 13.67 | 6.83 | 47.1 | 0.03 | 81   | - | 397 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 5  | 7.5  | 7.78 | 12.38 | 6.77 | 44.3 | 0.03 | 73.6 | - | 399 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 4  | 9    | 7.84 | 10.9  | 6.76 | 44   | 0.03 | 71.7 | - | 398 |

|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
|----------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-------|-----|
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 3  | 10.5 | 7.38  | 9.64  | 6.77 | 42.2 | 0.03 | 65.1 | -     | 397 |
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 2  | 12   | 7.86  | 8.4   | 6.82 | 42.3 | 0.03 | 67.6 | -     | 395 |
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 1  | 13.4 | 8.18  | 8.22  | 6.92 | 42.4 | 0.03 | 70.1 | -     | 389 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 16 | 0.2  | 10.3  | 11.54 | 7.28 | 48   | 0.03 | 94   | -     | 378 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 15 | 1    | 10.28 | 11.56 | 7.31 | 49   | 0.03 | 93.9 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 14 | 2    | 10.27 | 11.57 | 7.29 | 49.1 | 0.03 | 93.8 | -     | 377 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 13 | 3    | 10.26 | 11.54 | 7.31 | 49.2 | 0.03 | 93.7 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 12 | 4    | 10.28 | 11.54 | 7.29 | 49.3 | 0.03 | 93.8 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 11 | 5    | 10.25 | 11.54 | 7.29 | 49.1 | 0.03 | 93.6 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 10 | 6    | 10.28 | 11.54 | 7.29 | 49.2 | 0.03 | 93.8 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 9  | 7    | 10.26 | 11.54 | 7.28 | 49.3 | 0.03 | 93.7 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 8  | 8    | 10.25 | 11.54 | 7.29 | 49.3 | 0.03 | 93.6 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 7  | 9    | 10.23 | 11.53 | 7.27 | 49.1 | 0.03 | 93.4 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 6  | 10   | 10.24 | 11.53 | 7.27 | 49.1 | 0.03 | 93.5 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 5  | 11   | 10.25 | 11.51 | 7.31 | 49.3 | 0.03 | 93.5 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 4  | 12   | 10.26 | 11.5  | 7.3  | 49.3 | 0.03 | 93.4 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 3  | 13   | 10.25 | 11.43 | 7.3  | 49.3 | 0.03 | 93.4 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 2  | 14   | 10.63 | 11.15 | 7.31 | 49.5 | 0.03 | 96.2 | -     | 371 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 1  | 14.9 | 10.61 | 10.86 | 7.32 | 50   | 0.03 | 95.3 | -     | 370 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 11 | 0.3  | 10.59 | 7.04  | 9.79 | 53   | 0.03 | 93.2 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 10 | 1.5  | 10.54 | 7.05  | 9.74 | 53   | 0.03 | 92.5 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 9  | 3    | 10.52 | 7.02  | 9.6  | 53.1 | 0.03 | 92   | ***** | 442 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 8  | 4.5  | 10.75 | 6.99  | 9.24 | 52.9 | 0.03 | 93.3 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 7  | 6    | 10.97 | 6.93  | 8.74 | 53.7 | 0.03 | 94   | ***** | 440 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 6  | 7.5  | 9.25  | 6.79  | 8.48 | 52.9 | 0.03 | 78.8 | ***** | 442 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 5  | 9    | 9.27  | 6.77  | 8.48 | 52.8 | 0.03 | 78.9 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 4  | 10.5 | 9.32  | 6.75  | 8.43 | 52.5 | 0.03 | 79.3 | ***** | 439 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 3  | 12   | 8.81  | 6.66  | 7.73 | 48.8 | 0.03 | 73.7 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 2  | 13.6 | 8.64  | 6.66  | 7.41 | 47   | 0.03 | 71.6 | ***** | 439 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 1  | 14.9 | 8.65  | 6.71  | 7.37 | 46.7 | 0.03 | 71.7 | ***** | 441 |

| Location    | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|-------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Hidden Lake | 1597  | ASSS    | 4/18/97 | -    | 1.9    | 5        | 0.2       | 12.49                   | 6.73            | 7.46 | 38.6                 | 0.02 | *****              | -       | 374   |
| Hidden Lake | 1597  | ASSS    | 4/18/97 | -    | 1.9    | 4        | 2.3       | 12.51                   | 6.31            | 7.45 | 38.4                 | 0.02 | *****              | -       | 375   |
| Hidden Lake | 1597  | ASSS    | 4/18/97 | -    | 1.9    | 3        | 3.7       | 12.54                   | 6.28            | 7.45 | 38.4                 | 0.02 | *****              | -       | 375   |
| Hidden Lake | 1597  | ASSS    | 4/18/97 | -    | 1.9    | 2        | 5.3       | 12.47                   | 6.26            | 7.43 | 38.5                 | 0.02 | *****              | -       | 375   |
| Hidden Lake | 1597  | ASSS    | 4/18/97 | -    | 1.9    | 1        | 6.9       | 12.38                   | 6.23            | 7.42 | 38.3                 | 0.02 | 99.9               | -       | 375   |
|             |       |         |         |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 11       | 0.2       | 11.1                    | 11.97           | 7    | 29.9                 | 0.02 | 102.2              | -       | 371   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 10       | 1.4       | 11                      | 10.64           | 6.97 | 28.7                 | 0.02 | 98.1               | -       | 376   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 9        | 2.4       | 11.06                   | 9.38            | 6.97 | 28.8                 | 0.02 | 95.8               | -       | 377   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 8        | 3.4       | 11.03                   | 9.3             | 6.97 | 29                   | 0.02 | 95.3               | -       | 377   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 7        | 4.5       | 11.03                   | 9.27            | 6.96 | 29.3                 | 0.02 | 95.3               | -       | 377   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 6        | 5.4       | 11.01                   | 9.25            | 6.98 | 29.4                 | 0.02 | 95                 | -       | 375   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 5        | 6.4       | 10.95                   | 9.17            | 6.96 | 30.2                 | 0.02 | 94.4               | -       | 376   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 4        | 7.5       | 10.89                   | 9.1             | 6.98 | 29.1                 | 0.02 | 93.6               | -       | 375   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 3        | 8.5       | 10.87                   | 9.07            | 6.98 | 28.8                 | 0.02 | 93.4               | -       | 374   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 2        | 9.3       | 10.86                   | 9.04            | 6.99 | 28.9                 | 0.02 | 93.3               | -       | 373   |
| Hidden Lake | 1997  | DBAC    | 5/16/97 | -    | 0.7    | 1        | 10.4      | 10.86                   | 8.97            | 6.99 | 30                   | 0.02 | 93.1               | -       | 373   |
|             |       |         |         |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 10       | 0.7       | 10.91                   | 12.61           | 6.99 | 26.6                 | 0.02 | 101.9              | -       | 361   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 9        | 1.8       | 11.25                   | 9.84            | 7    | 26.9                 | 0.02 | 98.6               | -       | 361   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 8        | 2.8       | 11.18                   | 9.53            | 6.99 | 26.7                 | 0.02 | 97.3               | -       | 361   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 7        | 3.7       | 11.14                   | 9.3             | 6.97 | 26.6                 | 0.02 | 96.5               | -       | 362   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 6        | 4.7       | 11.13                   | 9.19            | 6.95 | 26.6                 | 0.02 | 96.1               | -       | 362   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 5        | 5.7       | 11.03                   | 9.12            | 6.95 | 26.4                 | 0.02 | 95.3               | -       | 361   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 4        | 6.8       | 10.74                   | 8.45            | 6.92 | 26.4                 | 0.02 | 91.1               | -       | 361   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 3        | 7.7       | 10.76                   | 8.38            | 6.92 | 26.3                 | 0.02 | 91.1               | -       | 360   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 2        | 8.8       | 10.62                   | 8.33            | 6.91 | 26.5                 | 0.02 | 89.8               | -       | 359   |
| Hidden Lake | 2197  | DBSS    | 5/29/97 | -    | 1.7    | 1        | 9.8       | 10                      | 8.3             | 6.91 | 27                   | 0.02 | 84.5               | -       | 367   |

|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|-------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 0.5 | 11.03 | 13.17 | 7.09 | 29.3 | 0.02 | 104.9 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 1.6 | 11.24 | 12.28 | 7.11 | 29.1 | 0.02 | 104.7 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 2.6 | 11.15 | 11.98 | 7.1  | 29   | 0.02 | 103.2 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 3.6 | 11.13 | 11.91 | 7.06 | 29   | 0.02 | 102.9 | - | 407 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 4.3 | 11.1  | 10.59 | 7.02 | 28.5 | 0.02 | 99.5  | - | 409 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 5.3 | 10.97 | 10.02 | 7.01 | 28.2 | 0.02 | 97.1  | - | 410 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 6.3 | 10.96 | 9.61  | 6.99 | 27.7 | 0.02 | 96    | - | 410 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 7.2 | 10.72 | 9.35  | 6.98 | 27.7 | 0.02 | 93.3  | - | 411 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 8.1 | 10.16 | 9.13  | 6.98 | 28.1 | 0.02 | 88    | - | 411 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 8.2 | 10.15 | 9.12  | 6.98 | 28.1 | 0.02 | 87.8  | - | 411 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 11 | 0.4 | 10.75 | 13.72 | 7.16 | 30.6 | 0.02 | 103.3 | - | 406 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 10 | 0.3 | 10.75 | 13.77 | 7.15 | 30.6 | 0.02 | 103.5 | - | 406 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 9  | 1.2 | 10.78 | 13.3  | 7.14 | 30.4 | 0.02 | 102.7 | - | 409 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 8  | 2.4 | 10.8  | 12.94 | 7.13 | 30.5 | 0.02 | 102   | - | 409 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 7  | 3.4 | 10.75 | 12.74 | 7.11 | 30.2 | 0.02 | 101   | - | 410 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 6  | 4.5 | 10.52 | 12.44 | 7.06 | 30.3 | 0.02 | 98.3  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 5  | 4.5 | 10.54 | 12.36 | 7.02 | 30.3 | 0.02 | 98.3  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 4  | 5.5 | 10.38 | 11.89 | 6.98 | 30   | 0.02 | 95.8  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 3  | 6.5 | 7.98  | 11.15 | 6.83 | 29.8 | 0.02 | 72.3  | - | 416 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 2  | 7.5 | 6.41  | 10.43 | 6.86 | 30.3 | 0.02 | 57.2  | - | 418 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 1  | 7.4 | 6.45  | 10.38 | 6.88 | 30.3 | 0.02 | 57.4  | - | 419 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 12 | 0.3 | 10.08 | 18.56 | 7.14 | 34.7 | 0.02 | ***** | - | 392 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 11 | 1.4 | 10.29 | 16.49 | 7.1  | 34.2 | 0.02 | ***** | - | 395 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 10 | 2.1 | 10.45 | 16.29 | 7.08 | 34.2 | 0.02 | ***** | - | 393 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 9  | 2.7 | 10.27 | 15.84 | 7.02 | 34.1 | 0.02 | ***** | - | 394 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 8  | 3.4 | 10.49 | 15.25 | 6.98 | 33.8 | 0.02 | ***** | - | 394 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 7  | 4.1 | 10.27 | 14.49 | 6.89 | 33.2 | 0.02 | ***** | - | 396 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 6  | 4.8 | 9.94  | 13.4  | 6.84 | 32.4 | 0.02 | 94.5  | - | 396 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 5  | 5.4 | 9.42  | 13.18 | 6.77 | 32.6 | 0.02 | 89.2  | - | 396 |

|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|-------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 4  | 6.1 | 8.57  | 12.67 | 6.72 | 32.4 | 0.02 | 80.2  | - | 395 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 3  | 6.9 | 7.84  | 12.2  | 6.7  | 32.5 | 0.02 | 72.4  | - | 392 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 2  | 7.6 | 6.94  | 11.93 | 6.7  | 32.6 | 0.02 | 63.7  | - | 388 |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 1  | 8.3 | 6.46  | 11.82 | 6.82 | 32.8 | 0.02 | 59.3  | - | 375 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 10 | 0.5 | 9.66  | 22.27 | 7.88 | 35.2 | 0.02 | ***** | - | 391 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 9  | 1   | 9.74  | 22.11 | 7.9  | 35.3 | 0.02 | ***** | - | 390 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 8  | 1.5 | 9.8   | 21.99 | 7.82 | 35.2 | 0.02 | ***** | - | 391 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 7  | 2.5 | 9.96  | 21.67 | 7.59 | 35.5 | 0.02 | ***** | - | 395 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 6  | 3.5 | 10.45 | 17.59 | 7.02 | 32.7 | 0.02 | ***** | - | 417 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 5  | 4.5 | 10    | 16.07 | 6.88 | 32   | 0.02 | ***** | - | 421 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 4  | 5.4 | 8.64  | 14.36 | 6.67 | 31.6 | 0.02 | 84    | - | 425 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 3  | 6.5 | 6.29  | 13.18 | 6.61 | 33.2 | 0.02 | 59.8  | - | 427 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 2  | 7.3 | 5.5   | 12.9  | 6.65 | 33.7 | 0.02 | 51.8  | - | 424 |
| Hidden Lake | 2997 | SSJD | 7/24/97 | - | 3.5 | 1  | 8   | 4.83  | 12.77 | 6.82 | 34   | 0.02 | 45.4  | - | 432 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 10 | 0.3 | 10.19 | 25.4  | 8.67 | 38.8 | 0.02 | 122.5 | - | 256 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 9  | 1.5 | 10    | 24.66 | 8.44 | 38.2 | 0.02 | 118.5 | - | 258 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 8  | 2.5 | 10    | 24.16 | 8.19 | 38.2 | 0.02 | 117.4 | - | 258 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 7  | 3.5 | 10.08 | 21.44 | 7.21 | 35.1 | 0.02 | 112.5 | - | 289 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 6  | 4.5 | 10.33 | 18.29 | 7.05 | 32.4 | 0.02 | 108.5 | - | 293 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 5  | 5.5 | 8.93  | 16.04 | 6.89 | 32.7 | 0.02 | 89.4  | - | 284 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 4  | 6.6 | 5.91  | 14.49 | 6.83 | 36.5 | 0.02 | 57.1  | - | 269 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 3  | 7.5 | 2.2   | 13.35 | 6.85 | 40.1 | 0.03 | 20.7  | - | 237 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 2  | 8.5 | 0.44  | 12.85 | 7.02 | 52.9 | 0.03 | 4.1   | - | 178 |
| Hidden Lake | 3197 | DBSS | 8/5/97  | - | 5.1 | 1  | 8.5 | 0.48  | 12.82 | 7.09 | 52.2 | 0.03 | 4.5   | - | 174 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 10 | 0.3 | 8.76  | 23.28 | 7.95 | 39.9 | 0.03 | 101.9 | - | 275 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 9  | 1   | 8.77  | 23.22 | 7.89 | 39.8 | 0.03 | 101.9 | - | 271 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 8  | 2   | 8.82  | 22.95 | 7.8  | 39.9 | 0.03 | 102.1 | - | 269 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 7  | 3   | 8.91  | 22.74 | 7.54 | 40.3 | 0.03 | 102.6 | - | 271 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 6  | 4   | 8.64  | 21.72 | 6.98 | 38.8 | 0.02 | 97.6  | - | 290 |

|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|-------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 5  | 5   | 8.67  | 18.32 | 6.76 | 34.5 | 0.02 | 91.6  | - | 296 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 4  | 6   | 6.25  | 16.11 | 6.67 | 37.6 | 0.02 | 63    | - | 288 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 3  | 6.9 | 2.37  | 14.47 | 6.6  | 41.7 | 0.03 | 22.7  | - | 266 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 2  | 8   | 0.46  | 13.43 | 6.59 | 50.7 | 0.03 | 4.4   | - | 242 |
| Hidden Lake | 3297 | DBSS | 8/14/97 | - | 6   | 1  | 8   | 0.81  | 13.86 | 6.92 | 47.1 | 0.03 | 8.1   | - | 225 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 10 | 0.5 | 9.39  | 22.02 | 7.9  | 44.9 | 0.03 | 106.8 | - | 235 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 9  | 0.3 | 9.39  | 22.02 | 7.9  | 44.9 | 0.03 | 106.8 | - | 235 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 8  | 1.6 | 9.41  | 21.92 | 7.88 | 45   | 0.03 | 106.9 | - | 231 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 7  | 2.4 | 9.43  | 21.88 | 7.83 | 45   | 0.03 | 107   | - | 223 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 6  | 3.5 | 9.45  | 21.85 | 7.74 | 45.1 | 0.03 | 107.2 | - | 219 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 5  | 4.6 | 9.33  | 21.58 | 7.44 | 44.9 | 0.03 | 105.3 | - | 218 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 4  | 5.4 | 10.33 | 19.92 | 7.05 | 42.8 | 0.03 | 112.9 | - | 228 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 3  | 6.5 | 3.7   | 17.44 | 6.73 | 47.1 | 0.03 | 38.4  | - | 208 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 2  | 7.5 | 0.3   | 14.92 | 6.67 | 57.7 | 0.04 | 2.9   | - | 147 |
| Hidden Lake | 3497 | DBSS | 8/27/97 | - | 5.7 | 1  | 8.4 | 0.4   | 13.88 | 6.6  | 79.9 | 0.05 | 3.9   | - | 139 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 9  | 0.2 | 8.74  | 18.01 | 7.31 | 47.1 | 0.03 | 92.1  | - | 296 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 8  | 1.2 | 8.75  | 18    | 7.3  | 47.2 | 0.03 | 92.3  | - | 295 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 7  | 2.2 | 8.68  | 17.9  | 7.27 | 47.1 | 0.03 | 91.3  | - | 294 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 6  | 3.2 | 8.57  | 17.84 | 7.27 | 47.1 | 0.03 | 90    | - | 290 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 5  | 4.2 | 8.63  | 17.79 | 7.27 | 47.1 | 0.03 | 90.4  | - | 286 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 4  | 5.2 | 8.75  | 17.73 | 7.24 | 47.1 | 0.03 | 91.7  | - | 281 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 3  | 6.2 | 8     | 17.44 | 7.12 | 49.2 | 0.03 | 83.4  | - | 278 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 2  | 7.2 | 7.77  | 17.12 | 7.07 | 50.9 | 0.03 | 80.4  | - | 264 |
| Hidden Lake | 3797 | DBAS | 9/16/97 | - | 2.5 | 1  | 8.2 | 7.96  | 17.05 | 7.05 | 51   | 0.03 | 82.3  | - | 245 |
|             |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Hidden Lake | 3997 | ASRA | 9/29/97 | - | 3.5 | 10 | 0.2 | 9.65  | 17.34 | 7.45 | 48.2 | 0.03 | 101.6 | - | 350 |
| Hidden Lake | 3997 | ASRA | 9/29/97 | - | 3.5 | 9  | 1.6 | 9.68  | 16.21 | 7.44 | 48.1 | 0.03 | 99.6  | - | 349 |
| Hidden Lake | 3997 | ASRA | 9/29/97 | - | 3.5 | 8  | 2.4 | 9.67  | 15.91 | 7.44 | 48.2 | 0.03 | 98.8  | - | 348 |
| Hidden Lake | 3997 | ASRA | 9/29/97 | - | 3.5 | 7  | 3.2 | 9.61  | 15.86 | 7.42 | 48.2 | 0.03 | 98.1  | - | 347 |
| Hidden Lake | 3997 | ASRA | 9/29/97 | - | 3.5 | 6  | 4   | 9.5   | 15.82 | 7.4  | 48.2 | 0.03 | 97.1  | - | 346 |

| Hidden Lake | 3997  | ASRA    | 9/29/97  | -    | 3.5    | 5        | 4.7       | 9.4                     | 15.77           | 7.37 | 48.3                 | 0.03 | 95.7               | -       | 344   |
|-------------|-------|---------|----------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Hidden Lake | 3997  | ASRA    | 9/29/97  | -    | 3.5    | 4        | 5.6       | 9.36                    | 15.67           | 7.34 | 48.1                 | 0.03 | 95.2               | -       | 340   |
| Hidden Lake | 3997  | ASRA    | 9/29/97  | -    | 3.5    | 3        | 6.4       | 8.83                    | 15.47           | 7.33 | 48.7                 | 0.03 | 89.4               | -       | 333   |
| Hidden Lake | 3997  | ASRA    | 9/29/97  | -    | 3.5    | 2        | 7.2       | 9.16                    | 15.34           | 7.35 | 48.9                 | 0.03 | 92.4               | -       | 328   |
| Hidden Lake | 3997  | ASRA    | 9/29/97  | -    | 3.5    | 1        | 8         | 9.44                    | 15.32           | 7.36 | 48.6                 | 0.03 | 95.2               | -       | 319   |
|             |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 9        | 0.3       | 10.76                   | 11.63           | 7.44 | 54.2                 | 0.03 | 97.8               | -       | 350   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 8        | 1         | 10.76                   | 11.43           | 7.46 | 54                   | 0.03 | 97.5               | -       | 349   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 7        | 2         | 10.76                   | 11.25           | 7.42 | 54.2                 | 0.03 | 97.1               | -       | 351   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 6        | 3         | 10.64                   | 11.12           | 7.39 | 54.3                 | 0.03 | 95.7               | -       | 351   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 5        | 4         | 10.71                   | 11.05           | 7.38 | 54.1                 | 0.03 | 96.2               | -       | 350   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 4        | 5         | 10.72                   | 11.02           | 7.37 | 54.2                 | 0.03 | 96.2               | -       | 349   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 3        | 6         | 10.75                   | 11              | 7.31 | 54.3                 | 0.03 | 96.5               | -       | 351   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 2        | 7         | 10.77                   | 11              | 7.28 | 54.2                 | 0.03 | 96.6               | -       | 351   |
| Hidden Lake | 4297  | DBAS    | 10/20/97 | -    | 2.5    | 1        | 8         | 10.89                   | 10.86           | 7.25 | 54.3                 | 0.03 | 97.4               | -       | 350   |
|             |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 9        | 0.2       | 11.64                   | 7.48            | 8.83 | 53.2                 | 0.03 | 100                | *****   | 425   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 8        | 1         | 11.57                   | 7.44            | 8.73 | 53.2                 | 0.03 | 99.1               | *****   | 427   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 7        | 2         | 11.5                    | 7.44            | 8.65 | 53.1                 | 0.03 | 98.3               | *****   | 426   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 6        | 3         | 11.5                    | 7.41            | 8.61 | 53.3                 | 0.03 | 98.2               | *****   | 425   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 5        | 4         | 11.48                   | 7.43            | 8.61 | 53.2                 | 0.03 | 98.1               | *****   | 422   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 4        | 5         | 11.38                   | 7.37            | 8.6  | 53.2                 | 0.03 | 97.2               | *****   | 423   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 3        | 6         | 11.17                   | 7.36            | 8.45 | 53.5                 | 0.03 | 95                 | *****   | 422   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 2        | 7         | 11.03                   | 7.35            | 8.32 | 53.6                 | 0.03 | 93.6               | *****   | 422   |
| Hidden Lake | 4497  | DBAS    | 11/3/97  | -    | 2.4    | 1        | 7.8       | 11.02                   | 7.36            | 8.32 | 53.7                 | 0.03 | 93.3               | *****   | 433   |
| Location    | Phase | Sampler | Date     | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
| Round Lake  | 1597  | DBAC    | 5/16/97  | -    | 0.4    | 12       | 0.2       | 11.03                   | 10.79           | 7    | 26.7                 | 0.02 | 99                 | -       | 372   |
| Round Lake  | 1597  | DBAC    | 5/16/97  | -    | 0.4    | 11       | 0.2       | 11.11                   | 10.63           | 7.03 | 27                   | 0.02 | 99                 | -       | 372   |
| Round Lake  | 1597  | DBAC    | 5/16/97  | -    | 0.4    | 10       | 0.2       | 11.09                   | 10.63           | 7.05 | 27                   | 0.02 | 98.9               | -       | 371   |
| Round Lake  | 1597  | DBAC    | 5/16/97  | -    | 0.4    | 9        | 1.7       | 11.09                   | 10.33           | 7.04 | 26.9                 | 0.02 | 98.3               | -       | 373   |

|            |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 8  | 1.7 | 11.06 | 10.35 | 7.03 | 26.8 | 0.02 | 98    | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 7  | 1.7 | 11.07 | 10.36 | 7.02 | 26.9 | 0.02 | 98.2  | - | 374 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 6  | 2.8 | 11.15 | 8.61  | 7.06 | 26.5 | 0.02 | 94.7  | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 5  | 2.8 | 11.13 | 8.6   | 7.06 | 26.5 | 0.02 | 94.5  | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 4  | 2.8 | 11.1  | 8.59  | 7.06 | 26.5 | 0.02 | 94.2  | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 3  | 3.9 | 11.17 | 8.33  | 7.06 | 26.4 | 0.02 | 94.3  | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 2  | 3.9 | 11.2  | 8.32  | 7.06 | 26.3 | 0.02 | 94.5  | - | 373 |
| Round Lake | 1597 | DBAC | 5/16/97 | - | 0.4 | 1  | 3.9 | 11.24 | 8.3   | 7.06 | 26.6 | 0.02 | 94.8  | - | 373 |
|            |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 12 | 0.3 | 11.02 | 10.86 | 7.11 | 26.5 | 0.02 | 99    | - | 365 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 11 | 0.3 | 11.03 | 10.92 | 7.1  | 26.4 | 0.02 | 99.3  | - | 366 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 10 | 0.2 | 11.01 | 11    | 7.1  | 26.2 | 0.02 | 99.3  | - | 366 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 9  | 1.2 | 11.39 | 9.71  | 7.11 | 26.6 | 0.02 | 99.7  | - | 366 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 8  | 1.2 | 11.37 | 9.69  | 7.13 | 26.7 | 0.02 | 99.4  | - | 366 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 7  | 1.2 | 11.35 | 9.73  | 7.12 | 26.7 | 0.02 | 99.4  | - | 366 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 6  | 2.4 | 11.44 | 9.5   | 7.09 | 26.7 | 0.02 | 99.5  | - | 368 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 5  | 2.4 | 11.44 | 9.51  | 7.1  | 26.3 | 0.02 | 99.4  | - | 368 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 4  | 2.4 | 11.5  | 9.45  | 7.1  | 26.7 | 0.02 | 99.8  | - | 367 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 3  | 3.3 | 11.53 | 9.2   | 7.09 | 26.7 | 0.02 | 99.6  | - | 368 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 2  | 3.3 | 11.54 | 9.22  | 7.08 | 26.4 | 0.02 | 99.7  | - | 369 |
| Round Lake | 2197 | DBSS | 5/29/97 | - | 1.4 | 1  | 3.3 | 11.54 | 9.23  | 7.09 | 26.4 | 0.02 | 99.8  | - | 368 |
|            |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 10 | 0.3 | 10.67 | 13.47 | 7.14 | 29.4 | 0.02 | 102.1 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 9  | 0.3 | 10.65 | 13.45 | 7.15 | 29.6 | 0.02 | 101.9 | - | 397 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 8  | 0.3 | 10.67 | 13.45 | 7.15 | 29.5 | 0.02 | 102.1 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 7  | 1   | 11.08 | 11.69 | 7.19 | 29   | 0.02 | 101.9 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 6  | 1   | 11.02 | 11.81 | 7.19 | 29.1 | 0.02 | 101.6 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 5  | 1   | 10.99 | 12    | 7.19 | 29   | 0.02 | 101.8 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 4  | 0.9 | 10.85 | 12.21 | 7.19 | 29   | 0.02 | 100.9 | - | 398 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 3  | 1.5 | 11.29 | 11.28 | 7.21 | 28.6 | 0.02 | 102.8 | - | 401 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 2  | 1.5 | 11.25 | 11.38 | 7.22 | 28.7 | 0.02 | 102.7 | - | 400 |
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 1  | 1.6 | 11.27 | 11.43 | 7.21 | 28.6 | 0.02 | 103   | - | 401 |







|                |      |      |         |   |     |    |      |       |       |      |       |       |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|-------|-------|-------|---|-----|
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 14 | 0.2  | 11.31 | 14.46 | 7.12 | 33.7  | 0.02  | 110   | - | 368 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 13 | 0.8  | 11.32 | 13.83 | 7.08 | 33.3  | 0.02  | 108.6 | - | 371 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 12 | 1.8  | 11.46 | 12.2  | 7.03 | 32.6  | 0.02  | 106.3 | - | 373 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 11 | 2.7  | 11.37 | 11.66 | 7.01 | 32    | 0.02  | 103.9 | - | 374 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 10 | 3.7  | 11    | 10.94 | 6.95 | 30.5  | 0.02  | 98.7  | - | 376 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 9  | 4.7  | 10.91 | 9.76  | 6.93 | 28.8  | 0.02  | 95.4  | - | 378 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 8  | 5.6  | 10.84 | 9.3   | 6.9  | 10.84 | ***** | 5.6   | - | 376 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 7  | 6.7  | 10.87 | 9.09  | 6.95 | 27.9  | 0.02  | 93.4  | - | 376 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 6  | 7.7  | 10.88 | 9.02  | 6.95 | 27.9  | 0.02  | 93.4  | - | 376 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 5  | 8.7  | 10.89 | 8.6   | 6.94 | 30.2  | 0.02  | 92.5  | - | 377 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 4  | 9.7  | 10.86 | 8.5   | 6.94 | 30.3  | 0.02  | 92.1  | - | 377 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 3  | 10.6 | 10.75 | 8.32  | 6.94 | 31    | 0.02  | 90.7  | - | 376 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 2  | 11.6 | 10.03 | 7.54  | 6.92 | 35.2  | 0.02  | 83.2  | - | 378 |
| Chatcolet Lake | 1997 | DBSS | 5/16/97 | - | 1.4 | 1  | 12.7 | 10.04 | 7.48  | 6.92 | 35.4  | 0.02  | 83    | - | 377 |
|                |      |      |         |   |     |    |      |       |       |      |       |       |       |   |     |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 11 | 0.3  | 10.62 | 13.44 | 6.98 | 27.4  | 0.02  | 101.2 | - | 360 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 10 | 2.3  | 10.9  | 9.6   | 6.97 | 26.8  | 0.02  | 95    | - | 363 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 9  | 3.4  | 10.8  | 8.6   | 6.98 | 26.7  | 0.02  | 91.9  | - | 362 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 8  | 4.5  | 10.85 | 8.55  | 6.96 | 26.6  | 0.02  | 92.2  | - | 363 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 7  | 6.5  | 10.88 | 8.35  | 6.96 | 26.5  | 0.02  | 92.1  | - | 362 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 6  | 7.4  | 10.88 | 8.29  | 6.95 | 26.4  | 0.02  | 91.9  | - | 362 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 5  | 8.4  | 11.07 | 8.11  | 6.95 | 26.5  | 0.02  | 93.1  | - | 361 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 4  | 9.3  | 11.05 | 7.88  | 6.95 | 26.4  | 0.02  | 92.4  | - | 360 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 3  | 10.4 | 11.03 | 7.86  | 6.93 | 26.4  | 0.02  | 92.2  | - | 360 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 2  | 11.4 | 11    | 7.78  | 6.92 | 26.4  | 0.02  | 91.8  | - | 359 |
| Chatcolet Lake | 2197 | DBSS | 5/29/97 | - | 1.6 | 1  | 12.3 | 10.3  | 7.73  | 6.91 | 27.3  | 0.02  | 85.8  | - | 359 |
|                |      |      |         |   |     |    |      |       |       |      |       |       |       |   |     |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 12 | 0.1  | 11.24 | 12.99 | 7.2  | 29.1  | 0.02  | 106.4 | - | 417 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 11 | 1    | 11.23 | 11.66 | 7.23 | 28.3  | 0.02  | 103.2 | - | 417 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 10 | 1.9  | 11.08 | 10.38 | 7.21 | 27.5  | 0.02  | 98.8  | - | 419 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 9  | 3    | 11.09 | 10    | 7.19 | 27.5  | 0.02  | 98    | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 8  | 3.9  | 11.16 | 9.82  | 7.19 | 27.2  | 0.02  | 98.1  | - | 421 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 7  | 4.9  | 11.2  | 9.72  | 7.19 | 27.1 | 0.02 | 98.2  | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 6  | 6    | 11.07 | 9.67  | 7.2  | 27.1 | 0.02 | 97    | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 5  | 7    | 11.15 | 9.15  | 7.21 | 26.9 | 0.02 | 96.5  | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 4  | 8    | 11.07 | 9.02  | 7.21 | 27   | 0.02 | 95.5  | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 3  | 8.9  | 10.94 | 8.99  | 7.23 | 27.1 | 0.02 | 94.3  | - | 421 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 2  | 9.8  | 10.84 | 9.04  | 7.26 | 27.3 | 0.02 | 93.5  | - | 419 |
| Chatcolet Lake | 2397 | DBAS | 6/13/97 | - | 2.1 | 1  | 10.9 | 10.85 | 8.99  | 7.29 | 27.3 | 0.02 | 93.4  | - | 417 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 0.3  | 10.94 | 14.59 | 7.21 | 31.5 | 0.02 | 107.3 | - | 373 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 1.4  | 10.94 | 14.45 | 7.17 | 31.5 | 0.02 | 106.9 | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 2.5  | 11.03 | 13.45 | 7.17 | 31.5 | 0.02 | 105.5 | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 3.6  | 10.97 | 13.25 | 7.15 | 31.4 | 0.02 | 104.4 | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 4.6  | 10.99 | 12.95 | 7.12 | 31.2 | 0.02 | 103.8 | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 5.6  | 10.93 | 12.74 | 7.09 | 31.5 | 0.02 | 102.8 | - | 374 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 6.6  | 10.97 | 12.1  | 7.04 | 31.6 | 0.02 | 101.8 | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 7.7  | 10.35 | 10.92 | 7    | 30.9 | 0.02 | 93.4  | - | 375 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 3  | 8.7  | 9.68  | 10.68 | 6.98 | 30.2 | 0.02 | 86.8  | - | 374 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 2  | 9.7  | 9.38  | 10.12 | 6.99 | 30.4 | 0.02 | 83    | - | 372 |
| Chatcolet Lake | 2597 | DBSS | 6/26/97 | - | 2.5 | 1  | 10.6 | 9.44  | 10.1  | 7.02 | 30.6 | 0.02 | 83.5  | - | 370 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 12 | 0.3  | 10.25 | 17.45 | 7.19 | 35.2 | 0.02 | ***** | - | 411 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 11 | 0.9  | 10.23 | 17.39 | 7.17 | 35.2 | 0.02 | ***** | - | 411 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 10 | 2.1  | 10.2  | 17.17 | 7.11 | 35.1 | 0.02 | ***** | - | 413 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 9  | 3.1  | 9.99  | 16.11 | 7.05 | 35.1 | 0.02 | ***** | - | 416 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 8  | 4    | 9.88  | 15.35 | 7.02 | 34.9 | 0.02 | 98.1  | - | 417 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 7  | 5    | 9.83  | 14.71 | 6.97 | 33.8 | 0.02 | 96.2  | - | 418 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 6  | 6    | 9.84  | 13.13 | 6.94 | 33.3 | 0.02 | 93    | - | 420 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 5  | 7    | 9.5   | 12.69 | 6.91 | 33.1 | 0.02 | 88.9  | - | 421 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 4  | 8.1  | 8.93  | 12.39 | 6.89 | 33.3 | 0.02 | 83    | - | 421 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 3  | 9    | 9.12  | 12.25 | 6.89 | 33.2 | 0.02 | 84.5  | - | 421 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 2  | 10.1 | 8.4   | 11.77 | 6.9  | 33.3 | 0.02 | 77    | - | 418 |
| Chatcolet Lake | 2797 | DBSS | 7/8/97  | - | 1.8 | 1  | 11   | 8.23  | 11.64 | 7    | 33.7 | 0.02 | 75.2  | - | 423 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 12 | 0.6  | 9.86  | 22.15 | 7.99 | 36.4 | 0.02 | ***** | - | 378 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 11 | 1.8  | 9.85  | 22.15 | 7.96 | 36.4 | 0.02 | ***** | - | 378 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 10 | 2.9  | 9.97  | 21.67 | 7.85 | 36.8 | 0.02 | ***** | - | 379 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 9  | 3.9  | 9.81  | 19.33 | 7.18 | 35.8 | 0.02 | ***** | - | 402 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 8  | 4.9  | 10.24 | 16.32 | 7.08 | 33.2 | 0.02 | ***** | - | 406 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 7  | 5.9  | 9.79  | 14.49 | 6.93 | 32.4 | 0.02 | 95.4  | - | 410 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 6  | 6.9  | 8.29  | 13.58 | 6.84 | 32.1 | 0.02 | 79.2  | - | 413 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 5  | 7.9  | 7.71  | 13.1  | 6.82 | 32.2 | 0.02 | 72.9  | - | 413 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 4  | 8.9  | 6.88  | 12.72 | 6.82 | 32.4 | 0.02 | 64.5  | - | 411 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 3  | 10   | 5.03  | 12.03 | 6.87 | 34.1 | 0.02 | 46.5  | - | 408 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 2  | 10.9 | 5.19  | 12.08 | 6.97 | 34.1 | 0.02 | 47.9  | - | 402 |
| Chatcolet Lake | 2997 | SSJD | 7/24/97 | - | 3.7 | 1  | 0    | 0     | 0     | 0    | 0    | 0    | 0     | - | 0   |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 13 | 0.3  | 11.59 | 24.51 | 9.28 | 42.3 | 0.03 | 137.1 | - | 289 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 12 | 1.7  | 11.55 | 24.37 | 9.22 | 42.2 | 0.03 | 136.3 | - | 290 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 11 | 2.7  | 10.64 | 23.4  | 8.86 | 40.4 | 0.03 | 123.3 | - | 299 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 10 | 3.6  | 10.13 | 22.02 | 7.49 | 39.2 | 0.03 | 114.4 | - | 331 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 9  | 4.6  | 9.8   | 18.9  | 7.05 | 35.8 | 0.02 | 104.2 | - | 349 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 8  | 5.6  | 9.77  | 16.32 | 6.97 | 33   | 0.02 | 98.4  | - | 351 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 7  | 6.5  | 8.09  | 14.44 | 6.84 | 33   | 0.02 | 78.2  | - | 353 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 6  | 7.5  | 6.42  | 13.55 | 6.79 | 33.4 | 0.02 | 60.9  | - | 353 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 5  | 8.4  | 5.49  | 13.15 | 6.79 | 33.6 | 0.02 | 51.6  | - | 351 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 4  | 9.5  | 3.45  | 12.3  | 6.81 | 36.1 | 0.02 | 31.8  | - | 347 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 3  | 10.5 | 3.05  | 12.11 | 6.89 | 36.7 | 0.02 | 28    | - | 342 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 2  | 11.5 | 2.57  | 12.07 | 7.03 | 38.4 | 0.02 | 23.5  | - | 335 |
| Chatcolet Lake | 3197 | DBSS | 8/5/97  | - | 3   | 1  | 12.4 | 2.5   | 12.13 | 7.18 | 39.3 | 0.03 | 23    | - | 343 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Chatcolet Lake | 3297 | DBSS | 8/13/97 | - | 3.5 | 10 | 0.8  | 9.41  | 22.99 | 8.77 | 42.9 | 0.03 | 108.9 | - | 339 |
| Chatcolet Lake | 3297 | DBSS | 8/13/97 | - | 3.5 | 9  | 1.8  | 9.41  | 22.58 | 8.71 | 42.7 | 0.03 | 108.2 | - | 341 |
| Chatcolet Lake | 3297 | DBSS | 8/13/97 | - | 3.5 | 8  | 2.8  | 9.27  | 21.79 | 8.4  | 42   | 0.03 | 104.7 | - | 348 |
| Chatcolet Lake | 3297 | DBSS | 8/13/97 | - | 3.5 | 7  | 3.8  | 8.8   | 21    | 7.58 | 39.8 | 0.03 | 98    | - | 371 |



|                |      |      |          |   |     |    |      |       |       |      |      |      |       |       |     |
|----------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|-------|-------|-----|
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 11 | 0.6  | 9.55  | 16.12 | 7.43 | 48.5 | 0.03 | 98    | -     | 362 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 10 | 1.6  | 9.51  | 15.72 | 7.41 | 48.3 | 0.03 | 96.8  | -     | 362 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 9  | 2.6  | 9.25  | 15.67 | 7.37 | 48.3 | 0.03 | 94.1  | -     | 363 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 8  | 3.6  | 9.15  | 15.54 | 7.35 | 48.4 | 0.03 | 92.8  | -     | 362 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 7  | 4.6  | 8.95  | 15.47 | 7.31 | 48.4 | 0.03 | 90.7  | -     | 362 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 6  | 5.6  | 8.67  | 15.37 | 7.3  | 48.7 | 0.03 | 87.6  | -     | 362 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 5  | 6.6  | 8.63  | 15.32 | 7.3  | 48.7 | 0.03 | 87.1  | -     | 361 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 4  | 7.6  | 8.53  | 15.27 | 7.3  | 48.8 | 0.03 | 86.2  | -     | 359 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 3  | 8.6  | 8.41  | 15.22 | 7.3  | 48.9 | 0.03 | 84.7  | -     | 358 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 2  | 9.6  | 8.06  | 15.06 | 7.29 | 49.4 | 0.03 | 80.8  | -     | 356 |
| Chatcolet Lake | 3997 | ASRA | 9/29/97  | - | 3.3 | 1  | 10.6 | 7.99  | 14.99 | 7.32 | 50.4 | 0.03 | 80    | -     | 353 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |       |       |     |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 11 | 0.3  | 11.17 | 11.63 | 7.54 | 54.9 | 0.04 | 101.7 | -     | 372 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 10 | 1    | 11.1  | 11.56 | 7.5  | 55   | 0.04 | 100.9 | -     | 373 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 9  | 2    | 11.08 | 11.25 | 7.5  | 54.7 | 0.04 | 100   | -     | 373 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 8  | 3    | 11.05 | 11.23 | 7.45 | 54.9 | 0.04 | 99.6  | -     | 375 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 7  | 4    | 10.97 | 11.17 | 7.4  | 54.6 | 0.03 | 98.8  | -     | 376 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 6  | 5    | 10.89 | 11.13 | 7.32 | 54.5 | 0.03 | 98    | -     | 378 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 5  | 6    | 10.81 | 11.07 | 7.24 | 54.5 | 0.03 | 97.1  | -     | 381 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 4  | 7    | 10.67 | 11.04 | 7.15 | 54.6 | 0.03 | 95.8  | -     | 384 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 3  | 8    | 10.53 | 11    | 7.08 | 54.6 | 0.03 | 94.5  | -     | 386 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 2  | 9    | 9.17  | 10.74 | 6.92 | 56.4 | 0.04 | 81.8  | -     | 392 |
| Chatcolet Lake | 4297 | DBAS | 10/20/97 | - | 2.8 | 1  | 10   | 8.97  | 10.64 | 6.88 | 56.8 | 0.04 | 79.8  | -     | 393 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |       |       |     |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 11 | 0.2  | 11.7  | 7.53  | 8.55 | 53.1 | 0.03 | 99.8  | ***** | 418 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 10 | 1    | 11.7  | 7.5   | 8.55 | 53.3 | 0.03 | 99.8  | ***** | 419 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 9  | 2    | 11.69 | 7.45  | 8.5  | 53.2 | 0.03 | 99.5  | ***** | 421 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 8  | 3    | 11.69 | 7.43  | 8.4  | 53.2 | 0.03 | 99.3  | ***** | 419 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 7  | 4    | 11.69 | 7.39  | 8.38 | 53.1 | 0.03 | 99.3  | ***** | 420 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 6  | 5    | 11.7  | 7.38  | 8.37 | 53.1 | 0.03 | 99.4  | ***** | 417 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 5  | 6    | 11.72 | 7.37  | 8.33 | 53.1 | 0.03 | 99.4  | ***** | 416 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97  | - | 2   | 4  | 7    | 11.77 | 7.31  | 8.19 | 52.5 | 0.03 | 99.5  | ***** | 417 |

|                |      |      |         |   |   |   |     |       |      |      |      |      |      |       |     |
|----------------|------|------|---------|---|---|---|-----|-------|------|------|------|------|------|-------|-----|
| Chatcolet Lake | 4497 | DBAS | 11/3/97 | - | 2 | 3 | 8   | 11.73 | 7.29 | 7.97 | 52   | 0.03 | 98.6 | ***** | 416 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97 | - | 2 | 2 | 9   | 11.46 | 7.25 | 7.55 | 52.2 | 0.03 | 95.4 | ***** | 415 |
| Chatcolet Lake | 4497 | DBAS | 11/3/97 | - | 2 | 1 | 9.8 | 11.55 | 7.26 | 6.91 | 47.7 | 0.03 | 94.6 | ***** | 414 |

| Location          | Phase | Sampler | Date    | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
|-------------------|-------|---------|---------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Chatcolet Shallow | 1597  | ASSS    | 4/18/97 | -    | 0.5    | 2        | 0.2       | 12.75                   | 5.46            | 7.43 | 32.4                 | 0.02 | *****              | -       | 325   |
| Chatcolet Shallow | 1597  | ASSS    | 4/18/97 | -    | 0.5    | 1        | 0.7       | 13.38                   | 5.47            | 7.41 | 32.6                 | 0.02 | *****              | -       | 320   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 13       | 0.3       | 11.04                   | 10.48           | 6.99 | 27.2                 | 0.02 | 98.1               | -       | 369   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 12       | 0.3       | 11.05                   | 10.49           | 7    | 27.2                 | 0.02 | 98.2               | -       | 369   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 11       | 0.3       | 11.04                   | 10.44           | 7.01 | 27.3                 | 0.02 | 98                 | -       | 369   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 10       | 0.3       | 11.03                   | 10.46           | 7.01 | 27.2                 | 0.02 | 98                 | -       | 369   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 9        | 1.3       | 11.02                   | 9.97            | 6.97 | 27                   | 0.02 | 95.7               | -       | 374   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 8        | 1.3       | 11                      | 9.89            | 6.96 | 27.1                 | 0.02 | 96.4               | -       | 374   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 7        | 1.3       | 11.01                   | 9.51            | 6.99 | 27                   | 0.02 | 95.6               | -       | 373   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 6        | 1.3       | 10.97                   | 10.15           | 6.95 | 27.4                 | 0.02 | 96.7               | -       | 375   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 5        | 3.1       | 11.21                   | 8.56            | 7    | 26.4                 | 0.02 | 95.1               | -       | 375   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 4        | 3.1       | 11.21                   | 8.58            | 6.99 | 26.3                 | 0.02 | 95.2               | -       | 375   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 3        | 3.1       | 11.22                   | 8.55            | 6.99 | 26.3                 | 0.02 | 95.2               | -       | 376   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 2        | 3.1       | 11.22                   | 8.59            | 7    | 26.3                 | 0.02 | 95.3               | -       | 375   |
| Chatcolet Shallow | 1997  | DBSS    | 5/16/97 | -    | 0.3    | 1        | 3.1       | 11.21                   | 8.56            | 7.01 | 26.4                 | 0.02 | 95.1               | -       | 375   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 10       | 0.2       | 10.8                    | 11.38           | 7.04 | 26.8                 | 0.02 | 98.2               | -       | 368   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 9        | 0.2       | 10.85                   | 11.72           | 7.1  | 18.3                 | 0.01 | 99.5               | -       | 363   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 8        | 0.2       | 10.87                   | 11.59           | 7.1  | 15.3                 | 0.01 | 99.3               | -       | 364   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 7        | 1.4       | 11.32                   | 9.74            | 7.07 | 26.7                 | 0.02 | 99                 | -       | 368   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 6        | 1.4       | 11.32                   | 9.69            | 7.07 | 26.7                 | 0.02 | 99.1               | -       | 368   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 5        | 1.4       | 11.35                   | 9.68            | 7.08 | 26.6                 | 0.02 | 99.4               | -       | 368   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 4        | 2.3       | 11.51                   | 9.32            | 7.08 | 26.5                 | 0.02 | 99.7               | -       | 369   |
| Chatcolet Shallow | 2197  | DBSS    | 5/29/97 | -    | 1.6    | 3        | 2.3       | 11.53                   | 9.32            | 7.08 | 26.5                 | 0.02 | 99.9               | -       | 369   |



|                   |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Chatcolet Shallow | 2197 | DBSS | 5/29/97 | - | 1.6 | 2  | 2.3 | 11.52 | 9.33  | 7.07 | 26.5 | 0.02 | 100   | - | 369 |
| Chatcolet Shallow | 2197 | DBSS | 5/29/97 | - | 1.6 | 1  | 2.4 | 11.57 | 9.35  | 7.08 | 26.6 | 0.02 | 100.3 | - | 368 |
|                   |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 10 | 0.4 | 10.75 | 11.8  | 7.09 | 28.7 | 0.02 | 99.1  | - | 394 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 9  | 0.4 | 10.7  | 11.74 | 7.1  | 28.8 | 0.02 | 98.5  | - | 395 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 8  | 0.3 | 10.71 | 11.86 | 7.11 | 28.7 | 0.02 | 98.8  | - | 394 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 7  | 0.9 | 10.95 | 11.15 | 7.1  | 28.7 | 0.02 | 99.4  | - | 396 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 6  | 0.9 | 10.93 | 11.31 | 7.11 | 28.5 | 0.02 | 99.6  | - | 396 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 5  | 0.9 | 10.94 | 11.18 | 7.12 | 28.3 | 0.02 | 99.4  | - | 395 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 4  | 0.9 | 10.97 | 11.1  | 7.11 | 28.4 | 0.02 | 99.5  | - | 396 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 3  | 1.5 | 11.07 | 10.98 | 7.11 | 28.3 | 0.02 | 100.1 | - | 398 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 2  | 1.5 | 11.12 | 10.89 | 7.11 | 28.4 | 0.02 | 100.4 | - | 398 |
| Chatcolet Shallow | 2397 | DBAS | 6/11/97 | - | 1.6 | 1  | 1.5 | 11.16 | 10.88 | 7.12 | 28.3 | 0.02 | 100.7 | - | 398 |
|                   |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 10 | 0.2 | 11.42 | 11.88 | 7.25 | 31.8 | 0.02 | 105.4 | - | 388 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 9  | 0.2 | 11.38 | 11.97 | 6.9  | 0.7  | 0    | 105.1 | - | 405 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 8  | 0.2 | 11.43 | 11.9  | 7.25 | 31.8 | 0.02 | 105.4 | - | 389 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 7  | 0.2 | 11.44 | 11.88 | 7.26 | 31.8 | 0.02 | 105.6 | - | 389 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 6  | 0.2 | 11.44 | 11.92 | 7.26 | 31.8 | 0.02 | 105.5 | - | 389 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 5  | 0.7 | 11.39 | 11.87 | 7.24 | 31.9 | 0.02 | 105   | - | 392 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 4  | 0.7 | 11.41 | 11.85 | 7.27 | 31.8 | 0.02 | 105.2 | - | 391 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 3  | 0.7 | 11.43 | 11.84 | 7.27 | 31.8 | 0.02 | 105.3 | - | 391 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 2  | 0.7 | 11.4  | 11.87 | 7.27 | 31.8 | 0.02 | 105.2 | - | 391 |
| Chatcolet Shallow | 2597 | DBAC | 6/26/97 | - | 0.7 | 1  | 0.7 | 11.44 | 11.87 | 7.27 | 31.8 | 0.02 | 105.5 | - | 391 |
|                   |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Chatcolet Shallow | 2797 | DBSS | 7/8/97  | - | 0.7 | 2  | 0.2 | 10.12 | 16.22 | 7.28 | 35.2 | 0.02 | ***** | - | 406 |
| Chatcolet Shallow | 2797 | DBSS | 7/8/97  | - | 0.7 | 1  | 0.8 | 10.17 | 16.19 | 7.29 | 35.2 | 0.02 | ***** | - | 409 |
|                   |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Chatcolet Shallow | 2997 | SSJD | 7/24/97 | - | 0.8 | 10 | 0.3 | 11.24 | 22.77 | 9.09 | 38.9 | 0.02 | ***** | - | 356 |
| Chatcolet Shallow | 2997 | SSJD | 7/24/97 | - | 0.8 | 9  | 0.3 | 11.33 | 22.93 | 9.1  | 39   | 0.02 | ***** | - | 357 |
| Chatcolet Shallow | 2997 | SSJD | 7/24/97 | - | 0.8 | 8  | 0.3 | 11.39 | 23.04 | 9.12 | 39   | 0.03 | ***** | - | 356 |
| Chatcolet Shallow | 2997 | SSJD | 7/24/97 | - | 0.8 | 7  | 0.3 | 11.5  | 22.99 | 9.1  | 39   | 0.03 | ***** | - | 357 |

|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
|-------------------|-------|---------|----------|------|--------|----------|-----------|-------------------------|-----------------|------|----------------------|------|--------------------|---------|-------|
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 6        | 0.2       | 11.19                   | 22.95           | 9.03 | 39                   | 0.02 | *****              | -       | 361   |
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 5        | 1         | 11.36                   | 20.13           | 8.43 | 38.2                 | 0.02 | *****              | -       | 382   |
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 4        | 1         | 11.45                   | 20.09           | 8.47 | 38                   | 0.02 | *****              | -       | 379   |
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 3        | 1         | 11.51                   | 20.11           | 8.49 | 38.1                 | 0.02 | *****              | -       | 378   |
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 2        | 0         | 11.47                   | 20.18           | 8.45 | 38.1                 | 0.02 | *****              | -       | 379   |
| Chatcolet Shallow | 2997  | SSJD    | 7/24/97  | -    | 0.8    | 1        | 1         | 11.55                   | 20.09           | 8.47 | 38                   | 0.02 | *****              | -       | 378   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 3197  | DBSS    | 8/5/97   | -    | 1      | 2        | 0.3       | 12.26                   | 25.72           | 9.28 | 46                   | 0.03 | 148.1              | -       | 295   |
| Chatcolet Shallow | 3197  | DBSS    | 8/5/97   | -    | 1      | 1        | 0.9       | 12.94                   | 22.95           | 9.04 | 46.6                 | 0.03 | 148.9              | -       | 304   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 3297  | DBSS    | 8/14/97  | -    | 1.1    | 2        | 0.3       | 11.62                   | 24.37           | 9.54 | 53.3                 | 0.03 | 138                | -       | 279   |
| Chatcolet Shallow | 3297  | DBSS    | 8/14/97  | -    | 1.1    | 1        | 1.1       | 11.73                   | 24.29           | 9.63 | 52.4                 | 0.03 | 139.1              | -       | 276   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| CHATSH            | 3497  | DBSS    | 8/27/97  | -    | 0.8    | 1        | 0.8       | 11.39                   | 21.49           | 8.62 | 56.9                 | 0.04 | 128.4              | -       | 277   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 3797  | DBAS    | 9/16/97  | -    | 1      | 1        | 0.8       | 10.41                   | 15.56           | 7.79 | 54.3                 | 0.03 | 104.2              | -       | 256   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 3997  | ASRA    | 9/29/97  | -    | 0.4    | 1        | 0.4       | 12.26                   | 16.46           | 8.31 | 53.6                 | 0.03 | 126.8              | -       | 362   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 4297  | DBAS    | 10/20/97 | -    | 1      | 2        | 0.3       | 10.98                   | 10.61           | 7.11 | 53.4                 | 0.03 | 97.6               | -       | 372   |
| Chatcolet Shallow | 4297  | DBAS    | 10/20/97 | -    | 1      | 1        | 0.7       | 10.16                   | 9.02            | 7    | 57.2                 | 0.04 | 87                 | -       | 379   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Chatcolet Shallow | 4497  | DBAS    | 11/3/97  | -    | 0.9    | 2        | 0.2       | 11.72                   | 7.28            | 6.97 | 42.6                 | 0.03 | 96.2               | *****   | 418   |
| Chatcolet Shallow | 4497  | DBAS    | 11/3/97  | -    | 0.9    | 1        | 0.8       | 11.86                   | 7.26            | 6.96 | 43                   | 0.03 | 97.3               | *****   | 424   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Location          | Phase | Sampler | Date     | Time | Secchi | Sequence | Depth (m) | Dissolved Oxygen (mg/l) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Percent Saturation | Statime | Redox |
| Benewah Lake      | 1597  | ASSS    | 4/18/97  | -    | 0.8    | 3        | 0.2       | 11.66                   | 9.27            | 7.36 | 28.1                 | 0.02 | *****              | -       | 363   |
| Benewah Lake      | 1597  | ASSS    | 4/18/97  | -    | 0.8    | 2        | 1.6       | 11.46                   | 8.92            | 7.32 | 27.9                 | 0.02 | 99                 | -       | 367   |
| Benewah Lake      | 1597  | ASSS    | 4/18/97  | -    | 0.8    | 1        | 4.3       | 11.42                   | 7.02            | 7.36 | 24.8                 | 0.02 | *****              | -       | 366   |
|                   |       |         |          |      |        |          |           |                         |                 |      |                      |      |                    |         |       |
| Benewah Lake      | 1997  | DBSS    | 5/16/97  | -    | 1.1    | 10       | 0.3       | 10.12                   | 16.39           | 6.97 | 31.6                 | 0.02 | 102.6              | -       | 369   |

|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 9  | 1.1 | 9.81  | 15.1  | 6.95 | 31.2 | 0.02 | 96.8  | - | 371 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 8  | 1.8 | 11.11 | 8.71  | 7    | 26.6 | 0.02 | 94.8  | - | 372 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 7  | 2.5 | 11.1  | 8.53  | 7.02 | 26.6 | 0.02 | 94.1  | - | 372 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 6  | 3.3 | 11.07 | 8.4   | 7.01 | 26.5 | 0.02 | 93.6  | - | 372 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 5  | 3.9 | 11.04 | 8.38  | 7.01 | 26.7 | 0.02 | 93.3  | - | 372 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 4  | 4.7 | 11    | 8.36  | 7.02 | 26.7 | 0.02 | 92.9  | - | 372 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 3  | 5.6 | 10.94 | 8.36  | 7.03 | 27   | 0.02 | 92.5  | - | 371 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 2  | 6.2 | 10.8  | 8.35  | 7.02 | 27.2 | 0.02 | 91    | - | 371 |
| Benewah Lake | 1997 | DBSS | 5/16/97 | - | 1.1 | 1  | 7   | 10.47 | 8.25  | 7.04 | 27.7 | 0.02 | 88.2  | - | 371 |
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 12 | 0.8 | 10.09 | 13.9  | 6.97 | 28   | 0.02 | 97.1  | - | 351 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 11 | 0.8 | 10.05 | 13.88 | 6.97 | 27.6 | 0.02 | 96.7  | - | 350 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 10 | 1.8 | 10.85 | 9.35  | 6.97 | 27   | 0.02 | 94.1  | - | 351 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 9  | 1.8 | 11.02 | 9.17  | 6.96 | 27   | 0.02 | 95.1  | - | 351 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 8  | 2.8 | 11.1  | 8.71  | 6.94 | 26.7 | 0.02 | 94.8  | - | 351 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 7  | 2.8 | 11.08 | 8.65  | 6.96 | 26.7 | 0.02 | 94.4  | - | 350 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 6  | 3.9 | 10.82 | 8.37  | 6.92 | 26.8 | 0.02 | 91.6  | - | 350 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 5  | 3.9 | 10.83 | 8.35  | 6.93 | 26.6 | 0.02 | 91.6  | - | 349 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 4  | 4.9 | 10.54 | 7.89  | 6.92 | 26.5 | 0.02 | 88.2  | - | 346 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 3  | 4.9 | 10.59 | 7.89  | 6.92 | 26.7 | 0.02 | 88.6  | - | 347 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 2  | 5.9 | 10.3  | 7.66  | 6.92 | 26.6 | 0.02 | 85.7  | - | 352 |
| Benewah Lake | 2197 | DBSS | 5/29/97 | - | 2   | 1  | 5.9 | 10.35 | 7.65  | 6.92 | 26.5 | 0.02 | 86.1  | - | 350 |
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 10 | 0.4 | 9.39  | 17.73 | 6.99 | 32.6 | 0.02 | 98.5  | - | 399 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 9  | 0.4 | 9.43  | 17.83 | 7    | 32.5 | 0.02 | 99.1  | - | 399 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 8  | 0.3 | 9.43  | 17.9  | 7    | 32.4 | 0.02 | 99.2  | - | 399 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 7  | 0.9 | 9.82  | 15.89 | 7.01 | 32.7 | 0.02 | 99.1  | - | 399 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 6  | 1.9 | 10.25 | 14.21 | 7    | 30.3 | 0.02 | 99.8  | - | 400 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 5  | 2.8 | 11.19 | 11.04 | 6.95 | 28.6 | 0.02 | 101.4 | - | 402 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 4  | 3.9 | 9.41  | 8.78  | 6.86 | 28.8 | 0.02 | 80.9  | - | 407 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 3  | 4.8 | 8.36  | 8.6   | 6.85 | 30.2 | 0.02 | 71.5  | - | 407 |
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 2  | 4.8 | 8.38  | 8.6   | 6.86 | 29.9 | 0.02 | 71.6  | - | 407 |

|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 1  | 4.8 | 8.41  | 8.56  | 6.87 | 29.7 | 0.02 | 71.8  | - | 406 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 12 | 0.3 | 10.31 | 18.63 | 7.36 | 33.9 | 0.02 | 110   | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 11 | 0.3 | 10.3  | 18.65 | 7.36 | 33.9 | 0.02 | 110   | - | 383 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 10 | 0.9 | 10.32 | 18.57 | 7.34 | 33.9 | 0.02 | 110   | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 9  | 0.9 | 10.31 | 18.59 | 7.32 | 33.9 | 0.02 | 109.9 | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 8  | 1.9 | 10.39 | 17.97 | 7.22 | 34   | 0.02 | 109.4 | - | 386 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 7  | 1.9 | 10.33 | 18.02 | 7.21 | 33.8 | 0.02 | 108.9 | - | 386 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 6  | 2.8 | 11.07 | 13.35 | 7.02 | 31.1 | 0.02 | 105.6 | - | 391 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 5  | 2.8 | 11.04 | 13.35 | 7    | 30.9 | 0.02 | 105.3 | - | 391 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 4  | 3.7 | 8.49  | 10.92 | 6.78 | 30.5 | 0.02 | 76.6  | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 3  | 3.7 | 8.43  | 10.9  | 6.78 | 30.3 | 0.02 | 76    | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 2  | 4.7 | 5.4   | 9.84  | 6.77 | 33.7 | 0.02 | 47.5  | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 1  | 4.7 | 5.43  | 9.84  | 6.78 | 33.7 | 0.02 | 47.7  | - | 398 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 6  | 0.3 | 10.59 | 21.76 | 8.31 | 35.9 | 0.02 | ***** | - | 370 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 5  | 0.8 | 10.62 | 21.56 | 7.85 | 36   | 0.02 | ***** | - | 377 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 4  | 1.8 | 11.49 | 17.39 | 7.29 | 35.3 | 0.02 | ***** | - | 398 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 3  | 2.7 | 11.51 | 14.86 | 7.09 | 33.8 | 0.02 | ***** | - | 408 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 2  | 3.6 | 9.61  | 13.13 | 6.92 | 33.1 | 0.02 | 90.9  | - | 417 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 1  | 4.6 | 5.87  | 11.66 | 6.9  | 36   | 0.02 | 53.6  | - | 421 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 10 | 0.4 | 10.59 | 23.51 | 9.06 | 37.8 | 0.02 | ***** | - | 356 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 9  | 0   | 10.66 | 23.29 | 9.02 | 37.8 | 0.02 | ***** | - | 358 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 8  | 1   | 10.58 | 23.37 | 9.03 | 37.8 | 0.02 | ***** | - | 358 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 7  | 1.9 | 10.62 | 23.2  | 8.9  | 37.5 | 0.02 | ***** | - | 361 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 6  | 2   | 10.67 | 23.01 | 8.72 | 37.6 | 0.02 | ***** | - | 366 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 5  | 3   | 11.46 | 18.22 | 7.41 | 34   | 0.02 | ***** | - | 407 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 4  | 3   | 11.38 | 18.08 | 7.31 | 33.6 | 0.02 | ***** | - | 409 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 3  | 4   | 10.54 | 15.01 | 7.05 | 33   | 0.02 | ***** | - | 421 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 2  | 4   | 9.11  | 14.54 | 6.99 | 34.9 | 0.02 | 88.9  | - | 426 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 1  | 4.8 | 5.06  | 13.37 | 6.97 | 40.5 | 0.03 | 48.2  | - | 437 |

|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 5  | 0.3 | 11.3  | 26.45 | 9.38 | 42.4 | 0.03 | 138.4 | - | 259 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 4  | 1.6 | 11.69 | 24.11 | 9.41 | 42.5 | 0.03 | 137.3 | - | 256 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 3  | 2.8 | 10.7  | 22.54 | 8.41 | 39.8 | 0.03 | 122   | - | 276 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 2  | 3.6 | 10.2  | 20.44 | 7.28 | 37.3 | 0.02 | 111.7 | - | 309 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 1  | 4.7 | 2.78  | 15.59 | 7.06 | 47.9 | 0.03 | 27.2  | - | 337 |
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 9  | 0.4 | 8.91  | 23.87 | 8.9  | 42   | 0.03 | 104.8 | - | 260 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 8  | 0.9 | 8.91  | 23.86 | 8.9  | 42   | 0.03 | 104.8 | - | 256 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 7  | 1.4 | 8.89  | 23.86 | 8.91 | 42   | 0.03 | 104.6 | - | 251 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 6  | 1.9 | 8.89  | 23.76 | 8.83 | 42.1 | 0.03 | 104.4 | - | 250 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 5  | 2.3 | 8.88  | 23.75 | 8.74 | 42   | 0.03 | 104.3 | - | 248 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 4  | 2.9 | 7.94  | 22.77 | 8.28 | 41.6 | 0.03 | 91.6  | - | 254 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 3  | 3.4 | 8.65  | 20.97 | 7.3  | 41.2 | 0.03 | 96.3  | - | 277 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 2  | 3.9 | 8.53  | 20.2  | 7.12 | 41.6 | 0.03 | 93.4  | - | 277 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 1  | 4.4 | 2.15  | 18.22 | 6.93 | 49.6 | 0.03 | 22.7  | - | 293 |
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 10 | 0.3 | 9.39  | 22.33 | 8.27 | 46.4 | 0.03 | 107.6 | - | 271 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 9  | 0.8 | 9.41  | 22.18 | 8.36 | 46.5 | 0.03 | 107.5 | - | 264 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 8  | 1.3 | 9.42  | 22.08 | 8.35 | 46.6 | 0.03 | 107.3 | - | 263 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 7  | 1.8 | 9.46  | 21.85 | 8.37 | 46.6 | 0.03 | 107.3 | - | 260 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 6  | 2.3 | 9.64  | 21.58 | 8.49 | 46.6 | 0.03 | 108.8 | - | 253 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 5  | 2.8 | 9.6   | 21.49 | 8.45 | 46.6 | 0.03 | 108.1 | - | 250 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 4  | 3.3 | 9.59  | 21.44 | 8.39 | 46.7 | 0.03 | 108   | - | 248 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 3  | 3.7 | 9.2   | 21.32 | 8.03 | 46.6 | 0.03 | 103.3 | - | 248 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 2  | 4.3 | 3.62  | 19.73 | 6.74 | 54.2 | 0.03 | 39.4  | - | 292 |
| Benewah Lake | 3497 | DBSS | 8/27/97 | - | 3.8 | 1  | 4.8 | 1.44  | 18.97 | 6.69 | 57.1 | 0.04 | 15.5  | - | 280 |
|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3797 | DBAS | 9/16/97 | - | 1.5 | 10 | 0.2 | 8.58  | 17.07 | 7.22 | 46.1 | 0.03 | 88.5  | - | 308 |
| Benewah Lake | 3797 | DBAS | 9/16/97 | - | 1.5 | 9  | 0.5 | 8.57  | 17.07 | 7.22 | 46   | 0.03 | 88.6  | - | 308 |
| Benewah Lake | 3797 | DBAS | 9/16/97 | - | 1.5 | 8  | 1   | 8.56  | 17.05 | 7.22 | 46.1 | 0.03 | 88.4  | - | 307 |
| Benewah Lake | 3797 | DBAS | 9/16/97 | - | 1.5 | 7  | 1.5 | 8.53  | 17.07 | 7.23 | 46   | 0.03 | 88.3  | - | 306 |



|        |      |       |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------|------|-------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 12 | 0.2  | 11.4  | 8.54  | 6.99 | 25.9 | 0.02 | 96.7  | - | 356 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 11 | 1    | 11.41 | 8.46  | 6.96 | 25.9 | 0.02 | 96.6  | - | 364 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 10 | 1.9  | 11.44 | 8.36  | 6.97 | 25.9 | 0.02 | 96.6  | - | 365 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 9  | 3    | 11.44 | 8.36  | 6.96 | 25.9 | 0.02 | 96.4  | - | 365 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 8  | 4.1  | 11.44 | 8.36  | 6.96 | 25.9 | 0.02 | 96.5  | - | 366 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 7  | 5    | 11.44 | 8.35  | 6.95 | 25.9 | 0.02 | 96.6  | - | 367 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 6  | 6    | 11.46 | 8.35  | 6.96 | 26   | 0.02 | 96.7  | - | 367 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 5  | 6.9  | 11.43 | 8.38  | 6.96 | 25.9 | 0.02 | 96.5  | - | 368 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 4  | 8    | 11.43 | 8.38  | 6.96 | 26   | 0.02 | 96.6  | - | 369 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 3  | 9.2  | 11.39 | 8.4   | 6.97 | 26   | 0.02 | 96.3  | - | 369 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 2  | 10   | 11.44 | 8.4   | 6.99 | 25.9 | 0.02 | 96.7  | - | 371 |
| ST.JOE | 1997 | DB SS | 5/16/97 | - | 0.4 | 1  | 11.2 | 11.45 | 8.4   | 6.99 | 25.9 | 0.02 | 96.8  | - | 373 |
|        |      |       |         |   |     |    |      |       |       |      |      |      |       |   |     |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 10 | 0.3  | 11.53 | 8.97  | 7.08 | 26.3 | 0.02 | 99.1  | - | 363 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 9  | 0.3  | 11.53 | 8.99  | 7.09 | 26.4 | 0.02 | 99.1  | - | 362 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 8  | 0.3  | 11.51 | 8.99  | 7.08 | 26.6 | 0.02 | 98.9  | - | 363 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 7  | 0.3  | 11.53 | 8.97  | 7.09 | 26.1 | 0.02 | 99.1  | - | 363 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 6  | 0.3  | 11.53 | 8.99  | 7.09 | 26.4 | 0.02 | 99.1  | - | 363 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 5  | 9.1  | 11.53 | 8.97  | 7.06 | 26.5 | 0.02 | 98.7  | - | 366 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 4  | 10   | 11.51 | 8.97  | 7.06 | 26.5 | 0.02 | 98.9  | - | 366 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 3  | 10.1 | 11.51 | 8.99  | 7.07 | 26.7 | 0.02 | 98.8  | - | 366 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 2  | 10.5 | 11.5  | 8.97  | 7.06 | 26.4 | 0.02 | 98.9  | - | 366 |
| ST.JOE | 2197 | DBSS  | 5/29/97 | - | 1.1 | 1  | 10.6 | 11.49 | 8.97  | 7.06 | 26.6 | 0.02 | 99.1  | - | 366 |
|        |      |       |         |   |     |    |      |       |       |      |      |      |       |   |     |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 10 | 0.8  | 11.35 | 10.28 | 7.13 | 28.1 | 0.02 | 100.9 | - | 394 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 9  | 0.8  | 11.33 | 10.3  | 7.13 | 28.1 | 0.02 | 100.9 | - | 394 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 8  | 0.9  | 11.33 | 10.3  | 7.13 | 28.1 | 0.02 | 100.9 | - | 394 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 7  | 0.9  | 11.33 | 10.3  | 7.13 | 28.1 | 0.02 | 100.9 | - | 394 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 6  | 0.8  | 11.35 | 10.3  | 7.12 | 28.1 | 0.02 | 101   | - | 395 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 5  | 0.9  | 11.35 | 10.3  | 7.13 | 28.1 | 0.02 | 101   | - | 395 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 4  | 0.9  | 11.36 | 10.3  | 7.13 | 28.1 | 0.02 | 101.1 | - | 395 |
| ST.JOE | 2397 | DBAS  | 6/11/97 | - | 1.7 | 3  | 0.9  | 11.37 | 10.28 | 7.13 | 28.1 | 0.02 | 101.1 | - | 395 |

|        |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| ST.JOE | 2397 | DBAS | 6/11/97 | - | 1.7 | 2  | 1.1  | 11.37 | 10.28 | 7.14 | 28.1 | 0.02 | 101.2 | - | 395 |
| ST.JOE | 2397 | DBAS | 6/11/97 | - | 1.7 | 1  | 0.9  | 11.38 | 10.3  | 7.11 | 28   | 0.02 | 101.3 | - | 398 |
|        |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 11 | 0.2  | 11.21 | 10.68 | 7.18 | 32.1 | 0.02 | 100.6 | - | 399 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 10 | 1    | 11.17 | 10.66 | 7.17 | 32.1 | 0.02 | 100.1 | - | 400 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 9  | 1.9  | 11.2  | 10.64 | 7.17 | 32.1 | 0.02 | 100.4 | - | 402 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 8  | 2.9  | 11.17 | 10.64 | 7.17 | 32.1 | 0.02 | 100.2 | - | 402 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 7  | 4    | 11.19 | 10.64 | 7.17 | 32   | 0.02 | 100.3 | - | 402 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 6  | 4.9  | 11.17 | 10.64 | 7.17 | 32.1 | 0.02 | 100.2 | - | 402 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 5  | 6.1  | 11.19 | 10.64 | 7.17 | 32   | 0.02 | 100.3 | - | 402 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 4  | 7.1  | 11.18 | 10.61 | 7.18 | 32   | 0.02 | 100.1 | - | 401 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 3  | 8.1  | 11.18 | 10.62 | 7.17 | 32.1 | 0.02 | 100   | - | 401 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 2  | 9    | 11.18 | 10.62 | 7.18 | 32   | 0.02 | 100.2 | - | 400 |
| ST.JOE | 2597 | DBSS | 6/26/97 | - | 3.1 | 1  | 11.2 | 11.2  | 10.67 | 7.19 | 32   | 0.02 | 100.4 | - | 400 |
|        |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 13 | 0.4  | 9.9   | 15.06 | 7.14 | 34.6 | 0.02 | 97.6  | - | 414 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 12 | 1.3  | 9.89  | 15.04 | 7.16 | 34.6 | 0.02 | 97.5  | - | 414 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 11 | 2.4  | 9.9   | 15.07 | 7.15 | 34.6 | 0.02 | 97.6  | - | 415 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 10 | 3.5  | 9.92  | 15.02 | 7.15 | 34.6 | 0.02 | 97.7  | - | 415 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 9  | 4.4  | 9.9   | 15.09 | 7.15 | 34.6 | 0.02 | 97.8  | - | 415 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 8  | 5.4  | 9.9   | 15.06 | 7.15 | 34.6 | 0.02 | 97.7  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 7  | 6.4  | 9.89  | 15.02 | 7.15 | 34.6 | 0.02 | 97.5  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 6  | 7.4  | 9.88  | 15.02 | 7.16 | 34.6 | 0.02 | 97.4  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 5  | 8.3  | 9.9   | 15.02 | 7.16 | 34.6 | 0.02 | 97.6  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 4  | 9.4  | 9.89  | 15.06 | 7.16 | 34.6 | 0.02 | 97.5  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 3  | 10.4 | 9.91  | 15.06 | 7.18 | 34.6 | 0.02 | 97.7  | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 2  | 11.4 | 9.93  | 15.06 | 7.19 | 34.6 | 0.02 | 98    | - | 416 |
| ST.JOE | 2797 | DBSS | 7/8/97  | - | 2.5 | 1  | 12.3 | 10.07 | 15.07 | 7.22 | 34.6 | 0.02 | 99.3  | - | 416 |
|        |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3   | 11 | 0.3  | 9.37  | 18.94 | 7.2  | 38.3 | 0.02 | ***** | - | 430 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3   | 10 | 1.3  | 9.42  | 18.47 | 7.2  | 38.2 | 0.02 | 99.9  | - | 431 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3   | 9  | 2.3  | 9.4   | 18.22 | 7.21 | 38.1 | 0.02 | 99.2  | - | 432 |



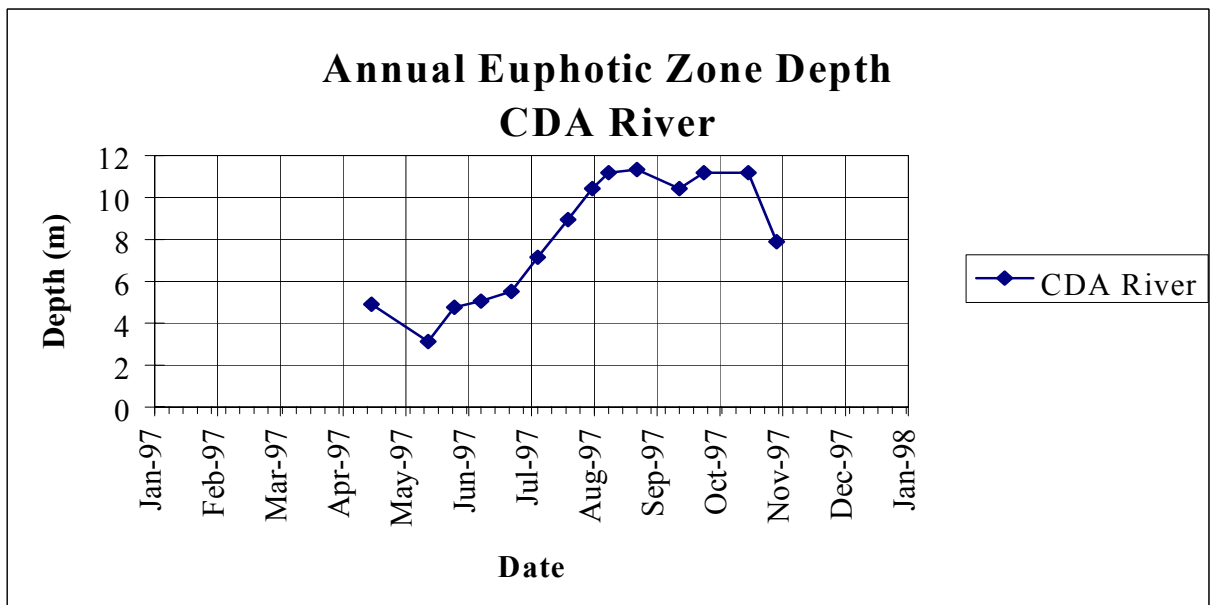
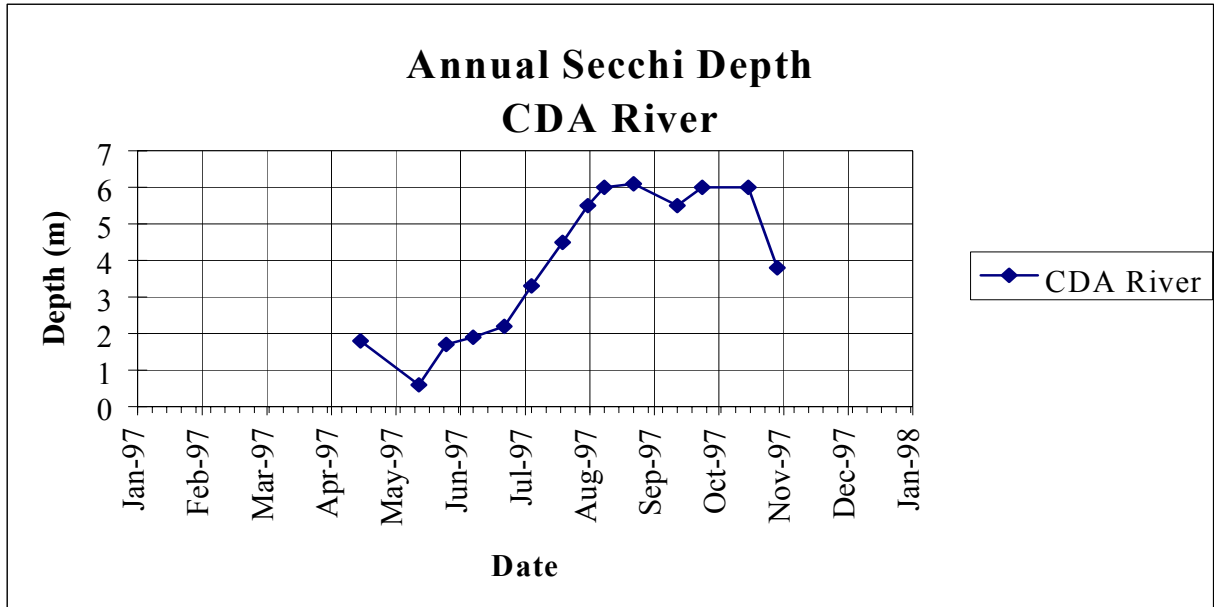
|        |      |      |         |   |   |    |      |      |       |      |      |      |       |   |     |
|--------|------|------|---------|---|---|----|------|------|-------|------|------|------|-------|---|-----|
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 8  | 3.3  | 9.42 | 18.15 | 7.21 | 38.2 | 0.02 | 99.2  | - | 432 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 7  | 4.3  | 9.43 | 18.22 | 7.21 | 38.2 | 0.02 | 99.5  | - | 433 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 6  | 5.3  | 9.41 | 18.22 | 7.2  | 38.1 | 0.02 | 99.3  | - | 434 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 5  | 6.3  | 9.4  | 18.1  | 7.21 | 38.1 | 0.02 | 99    | - | 433 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 4  | 7.3  | 9.41 | 18.22 | 7.21 | 38.1 | 0.02 | 99.3  | - | 434 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 3  | 8.3  | 9.4  | 18.18 | 7.22 | 38.1 | 0.02 | 99.1  | - | 434 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 2  | 9.3  | 9.43 | 18.1  | 7.25 | 38.1 | 0.02 | 99.3  | - | 434 |
| ST.JOE | 2997 | SSJD | 7/24/97 | - | 3 | 1  | 10.3 | 9.47 | 18.08 | 7.29 | 38.1 | 0.02 | 99.7  | - | 434 |
|        |      |      |         |   |   |    |      |      |       |      |      |      |       |   |     |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 13 | 0.3  | 9.07 | 23.01 | 7.3  | 44.8 | 0.03 | 104.4 | - | 351 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 12 | 1.5  | 9.11 | 21.37 | 7.26 | 44.5 | 0.03 | 101.5 | - | 354 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 11 | 2.4  | 9    | 21.28 | 7.25 | 44.7 | 0.03 | 100.2 | - | 355 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 10 | 3.4  | 8.91 | 20.55 | 7.22 | 44.3 | 0.03 | 97.8  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 9  | 4.5  | 8.64 | 19.94 | 7.19 | 44   | 0.03 | 93.7  | - | 357 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 8  | 5.5  | 8.62 | 19.71 | 7.19 | 44.1 | 0.03 | 93.1  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 7  | 6.5  | 8.54 | 19.57 | 7.19 | 43.7 | 0.03 | 92    | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 6  | 7.5  | 8.51 | 19.45 | 7.19 | 43.7 | 0.03 | 91.3  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 5  | 8.5  | 8.46 | 19.33 | 7.19 | 43.8 | 0.03 | 90.6  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 4  | 9.5  | 8.44 | 19.26 | 7.19 | 43.6 | 0.03 | 90.3  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 3  | 10.5 | 8.43 | 19.25 | 7.19 | 43.7 | 0.03 | 90.2  | - | 356 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 2  | 11.5 | 8.4  | 19.25 | 7.19 | 43.7 | 0.03 | 89.8  | - | 355 |
| ST.JOE | 3197 | DBSS | 8/5/97  | - | 3 | 1  | 12.1 | 8.4  | 19.3  | 7.2  | 43.6 | 0.03 | 89.9  | - | 354 |
|        |      |      |         |   |   |    |      |      |       |      |      |      |       |   |     |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 14 | 0.2  | 8.47 | 20.77 | 7.23 | 55.6 | 0.04 | 94.3  | - | 317 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 13 | 0.8  | 8.49 | 20.27 | 7.25 | 55.8 | 0.04 | 93.4  | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 12 | 1.7  | 8.47 | 20.2  | 7.24 | 55.7 | 0.04 | 93.1  | - | 317 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 11 | 2.7  | 8.49 | 20.2  | 7.25 | 55.6 | 0.04 | 93.2  | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 10 | 3.7  | 8.49 | 20.2  | 7.24 | 55.7 | 0.04 | 93.3  | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 9  | 4.6  | 8.48 | 20.15 | 7.24 | 55.6 | 0.04 | 93    | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 8  | 5.6  | 8.46 | 20.09 | 7.24 | 55.6 | 0.04 | 92.9  | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 7  | 6.4  | 8.41 | 20.08 | 7.23 | 55.5 | 0.04 | 92.3  | - | 316 |
| ST.JOE | 3497 | DBSS | 8/27/97 | - | 4 | 6  | 7.5  | 8.4  | 20.02 | 7.23 | 55.6 | 0.04 | 92    | - | 316 |



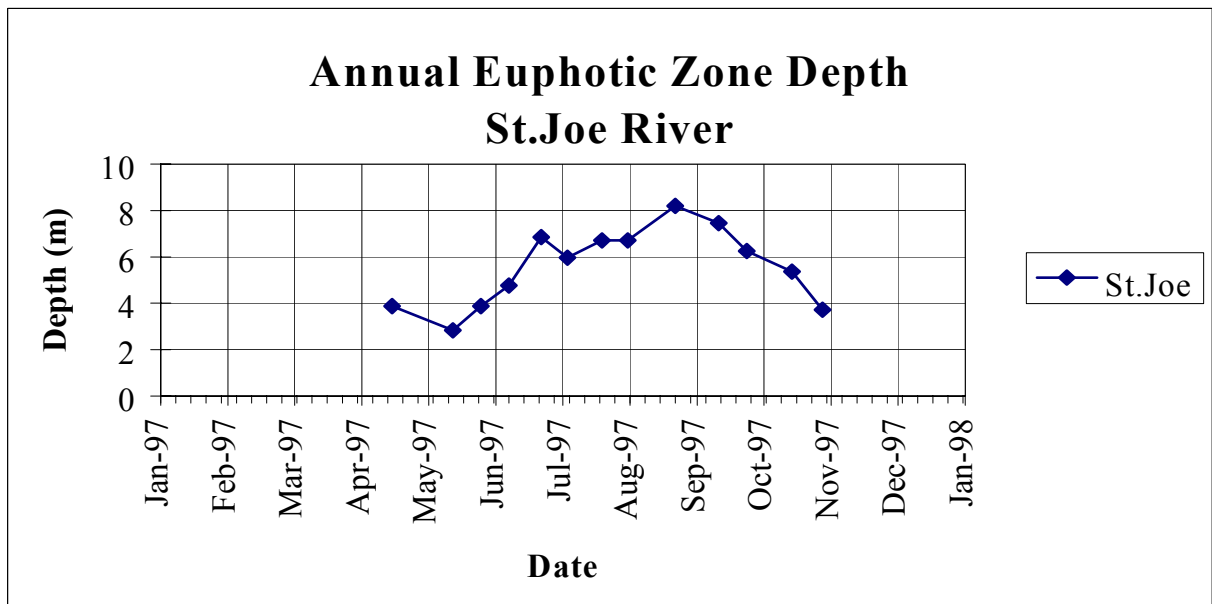
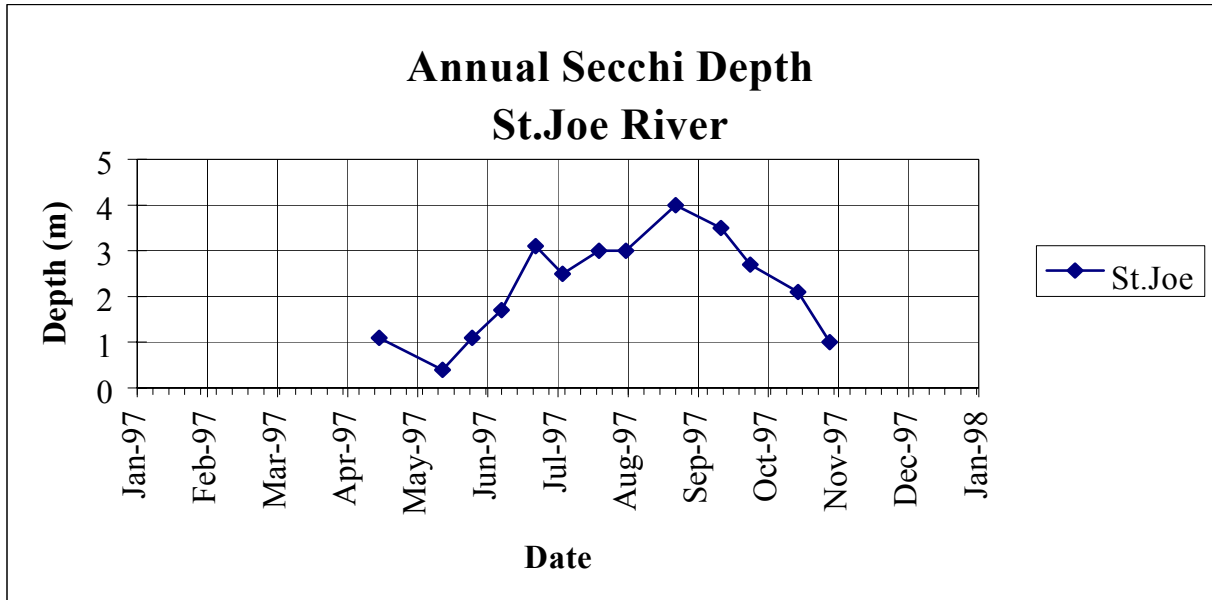
|        |      |       |          |   |     |    |      |       |      |      |      |      |      |       |     |
|--------|------|-------|----------|---|-----|----|------|-------|------|------|------|------|------|-------|-----|
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 11 | 0.3  | 10.77 | 9.06 | 6.97 | 56.7 | 0.04 | 92.3 | -     | 399 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 10 | 1.3  | 10.75 | 9.07 | 6.95 | 56.7 | 0.04 | 92.1 | -     | 402 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 9  | 2.3  | 10.76 | 9.06 | 6.93 | 56.7 | 0.04 | 92.1 | -     | 403 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 8  | 3.3  | 10.76 | 9.06 | 6.91 | 56.7 | 0.04 | 92.1 | -     | 404 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 7  | 4.3  | 10.74 | 9.06 | 6.91 | 56.6 | 0.04 | 92   | -     | 405 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 6  | 5.3  | 10.76 | 9.06 | 6.88 | 56.6 | 0.04 | 92.1 | -     | 406 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 5  | 6.3  | 10.76 | 9.04 | 6.87 | 56.7 | 0.04 | 92.2 | -     | 406 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 4  | 7.3  | 10.78 | 9.02 | 6.85 | 56.6 | 0.04 | 92.3 | -     | 407 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 3  | 8.3  | 10.78 | 9.04 | 6.83 | 56.6 | 0.04 | 92.3 | -     | 408 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 2  | 9.3  | 10.8  | 9.02 | 6.81 | 56.5 | 0.04 | 92.4 | -     | 408 |
| ST.JOE | 4297 | DBAS  | 10/20/97 | - | 2.1 | 1  | 10.3 | 10.83 | 9.04 | 6.76 | 56.7 | 0.04 | 92.8 | -     | 411 |
|        |      |       |          |   |     |    |      |       |      |      |      |      |      |       |     |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 11 | 0.4  | 11.9  | 7.03 | 6.63 | 39.7 | 0.03 | 96.8 | ***** | 436 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 10 | 1.8  | 11.93 | 7.02 | 6.61 | 39.9 | 0.03 | 96.9 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 9  | 2.8  | 11.9  | 7.03 | 6.63 | 39.8 | 0.03 | 96.8 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 8  | 3.8  | 11.92 | 7.03 | 6.61 | 39.8 | 0.03 | 96.9 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 7  | 4.8  | 11.89 | 7.04 | 6.61 | 39.9 | 0.03 | 96.6 | ***** | 437 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 6  | 5.8  | 11.93 | 7.06 | 6.61 | 39.8 | 0.03 | 96.9 | ***** | 436 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 5  | 6.8  | 11.93 | 7.06 | 6.61 | 39.9 | 0.03 | 96.9 | ***** | 436 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 4  | 7.8  | 11.95 | 7.07 | 6.61 | 40.2 | 0.03 | 97.1 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 3  | 8.8  | 11.96 | 7.08 | 6.61 | 39.9 | 0.03 | 97.3 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 2  | 9.8  | 12    | 7.08 | 6.61 | 39.9 | 0.03 | 97.6 | ***** | 438 |
| ST.JOE | 4497 | DB AS | 11/3/97  | - | 1   | 1  | 10.7 | 11.99 | 7.09 | 6.63 | 39.9 | 0.03 | 97.5 | ***** | 441 |

## **Appendix B**

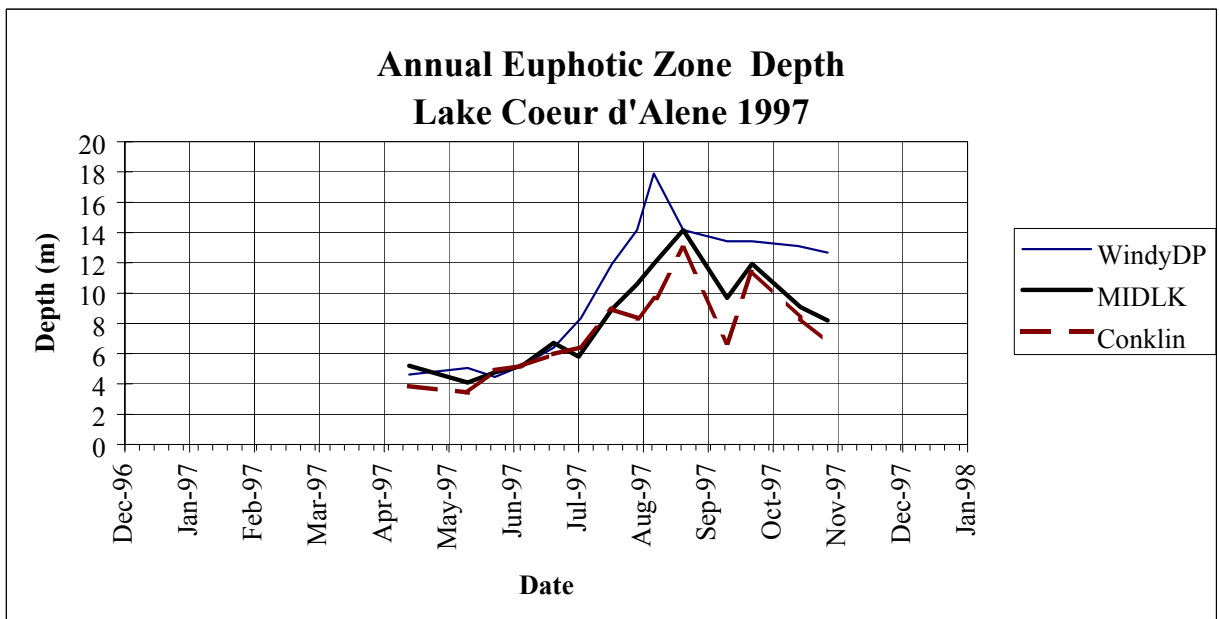
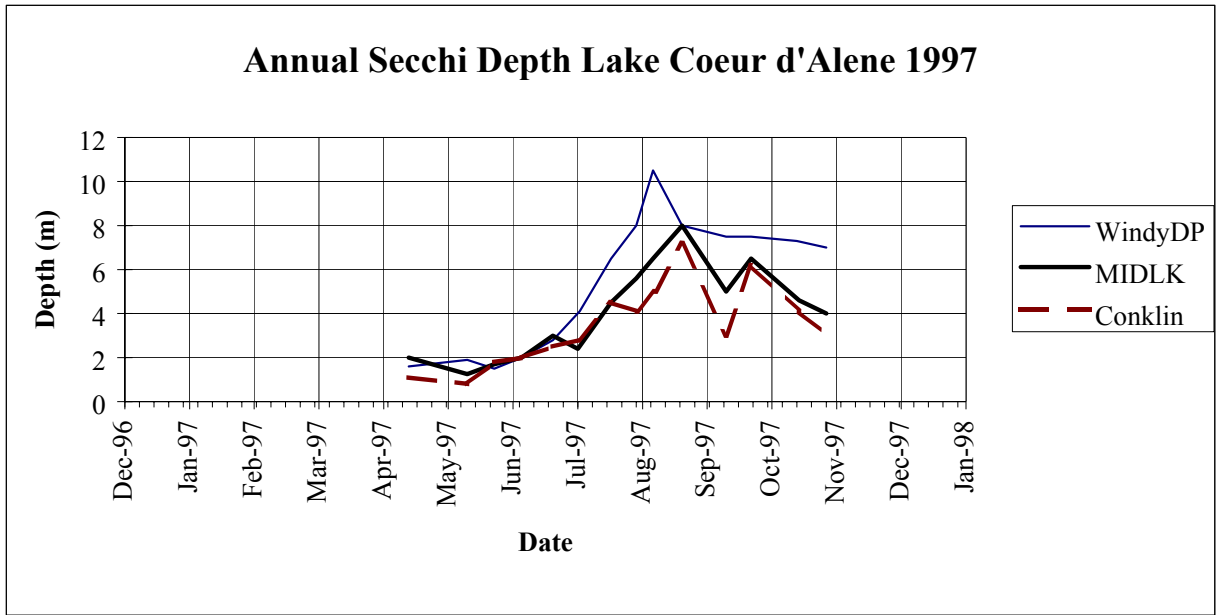
Secchi readings and empirically derived estimates of euphotic zone depth for thirteen stations in  
Coeur d'Alene Lake, 1997.



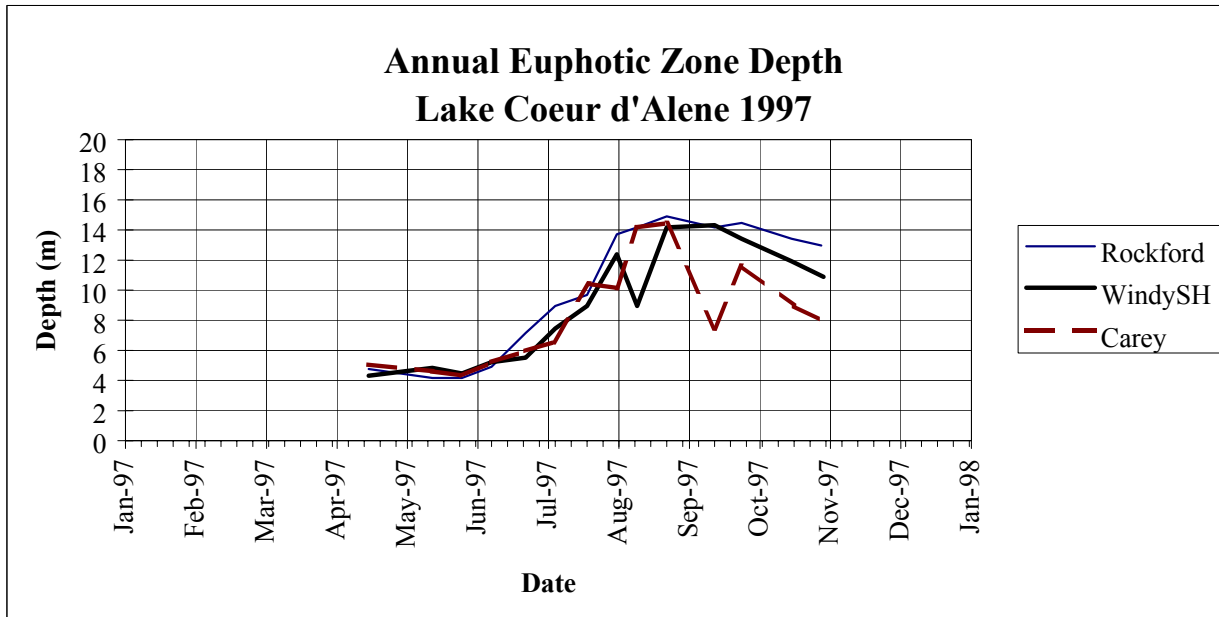
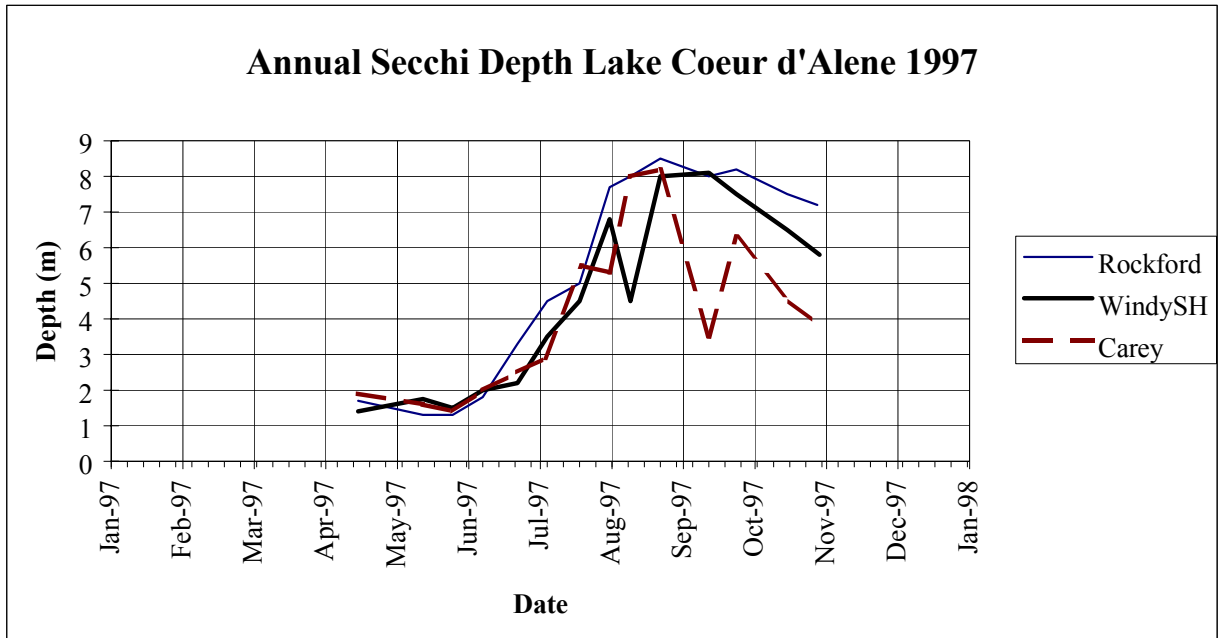
**(a)** Annual secchi readings (SD) for the mouth of the Coeur d'Alene River, 1997. **(b)** Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD=2.2303 + 1.4914SD$  ( $r^2 = .78$ )



**(a)** Annual secchi readings (SD) for the mouth of the St. Joe River, 1997. **(b)** Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ )

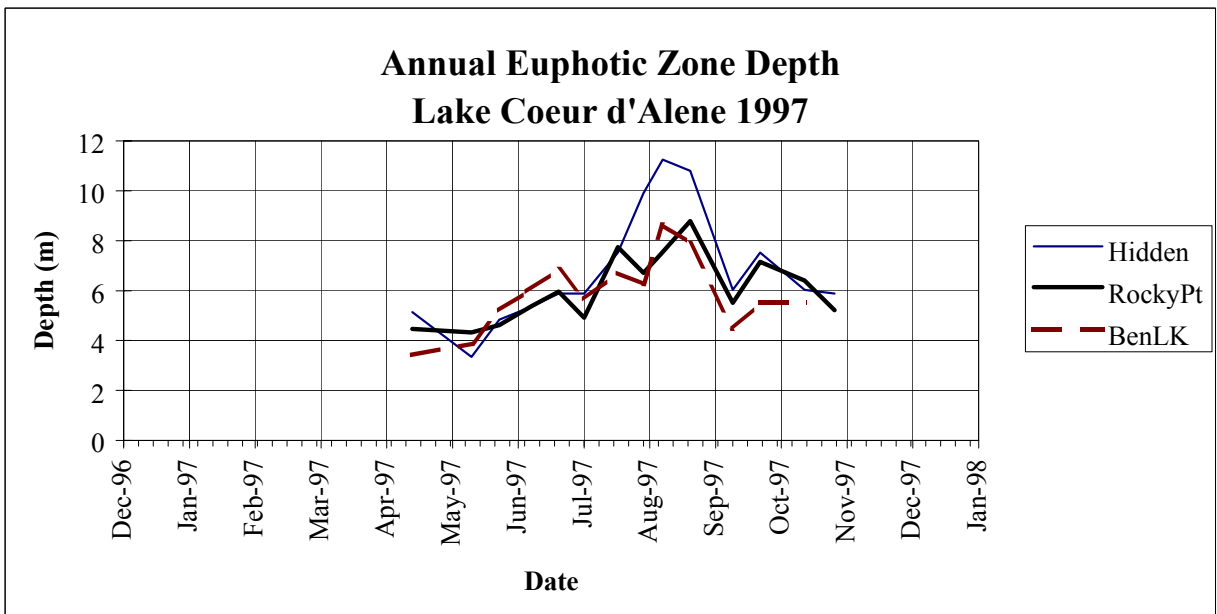
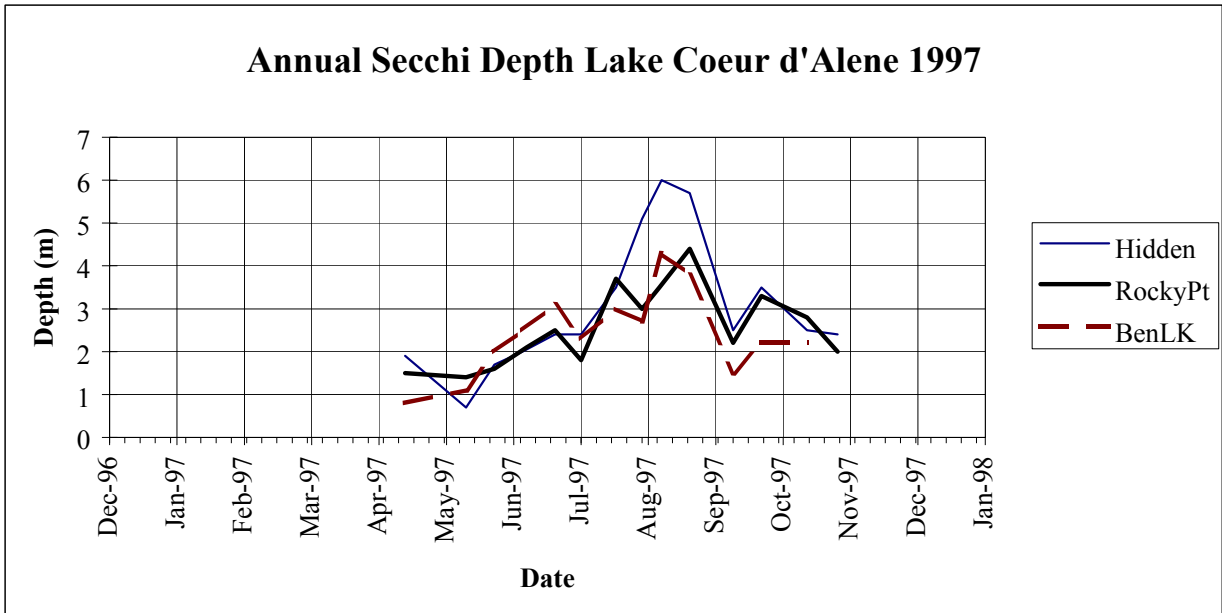


**(a)** Annual secchi readings (SD) for three geomorphically similar sampling locations on Coeur d'Alene Lake, 1997. **(b)** Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ )

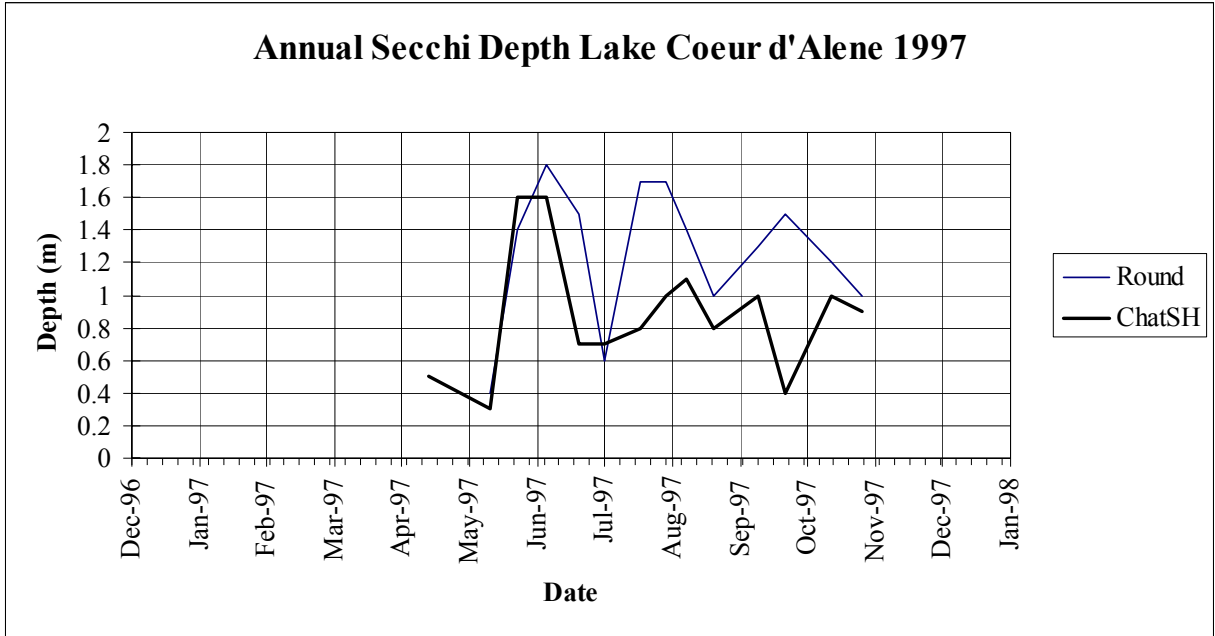


**(a)** Annual secchi readings (SD) for three geomorphically similar sampling locations on Coeur d'Alene Lake, 1997. **(b)** Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ )





**(a)** Annual secchi readings (SD) for three geomorphically similar sampling locations on Coeur d'Alene Lake, 1997. **(b)** Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ )



**(a)** Annual secchi readings (SD) for two geomorphically similar sampling locations on Coeur d'Alene Lake, 1997. Due to the shallowness of these lakes, the euphotic zone depth encompasses the entire water column.

## **Appendix C**

Cohort analysis of growth for cutthroat trout and brook trout, 1996-1997.

Mean back-calculated lengths at annulus formation by age class and cohort - 1996

| SPECIES         |        | Age Class |      |      |      |      |      |      |      |      |  |
|-----------------|--------|-----------|------|------|------|------|------|------|------|------|--|
| Cutthroat Trout |        |           |      |      |      |      |      |      |      |      |  |
| LOCATION        | Cohort | Number    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER           | 1995   | 1         | 81   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 31        | 84   | 120  | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 20        | 81   | 116  | 147  | #### | #### | #### | #### | #### |  |
|                 | 1992   | 2         | 90   | 119  | 160  | 192  | #### | #### | #### | #### |  |
| ALDER Total     |        | 54        | 83   | 118  | 148  | 192  | #### | #### | #### | #### |  |
| BENEWAH         | 1996   | 6         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 28        | 85   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 153       | 83   | 120  | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 38        | 83   | 123  | 155  | #### | #### | #### | #### | #### |  |
|                 | 1992   | 7         | 88   | 131  | 167  | 206  | #### | #### | #### | #### |  |
|                 | 1991   | 1         | 78   | 115  | 152  | 199  | 236  | #### | #### | #### |  |
|                 | 1990   | 2         | 91   | 136  | 181  | 222  | 269  | 316  | #### | #### |  |
|                 | 1989   | 1         | 84   | 117  | 151  | 191  | 224  | 265  | 318  | #### |  |
| BENEWAH Total   |        | 236       | 83   | 121  | 158  | 207  | 250  | 299  | 318  | #### |  |
| CHERRY          | 1994   | 1         | 83   | 133  | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 1         | 89   | 137  | 168  | #### | #### | #### | #### | #### |  |
|                 | 1990   | 6         | 87   | 135  | 177  | 214  | 256  | 298  | #### | #### |  |
| CHERRY Total    |        | 8         | 86   | 135  | 176  | 214  | 256  | 298  | #### | #### |  |
| EVANS           | 1996   | 7         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 35        | 84   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 54        | 84   | 121  | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 31        | 84   | 129  | 162  | #### | #### | #### | #### | #### |  |
|                 | 1992   | 6         | 86   | 131  | 169  | 207  | #### | #### | #### | #### |  |
| EVANS Total     |        | 133       | 84   | 124  | 163  | 207  | #### | #### | #### | #### |  |
| LAKE            | 1996   | 59        | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 54        | 79   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 72        | 86   | 127  | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 30        | 85   | 131  | 161  | #### | #### | #### | #### | #### |  |
|                 | 1992   | 6         | 88   | 134  | 173  | 211  | #### | #### | #### | #### |  |
|                 | 1990   | 21        | 94   | 142  | 184  | 228  | 269  | 314  | #### | #### |  |
|                 | 1989   | 6         | 91   | 132  | 172  | 216  | 251  | 291  | 336  | #### |  |
|                 | 1988   | 1         | 88   | 133  | 170  | 201  | 246  | 283  | 328  | 366  |  |
| LAKE Total      |        | 249       | 85   | 131  | 171  | 222  | 264  | 308  | 335  | 366  |  |
| Grand Total     |        | 680       | 84   | 124  | 163  | 216  | 261  | 306  | 333  | 366  |  |

Standard deviation of mean back-calculated lengths at annulus formation - 1996

| SPECIES         |        | Age Class |      |      |      |      |      |      |      |      |  |
|-----------------|--------|-----------|------|------|------|------|------|------|------|------|--|
| Cutthroat Trout |        |           |      |      |      |      |      |      |      |      |  |
| LOCATION        | Cohort | Number    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER           | 1995   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 31        | 3    | 6    | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 20        | 4    | 7    | 9    | #### | #### | #### | #### | #### |  |
|                 | 1992   | 2         | 1    | 2    | 11   | 9    | #### | #### | #### | #### |  |
| ALDER Total     |        | 54        | 4    | 7    | 10   | 9    | #### | #### | #### | #### |  |
| BENEWAH         | 1996   | 6         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 28        | 6    | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 153       | 6    | 10   | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 38        | 5    | 8    | 19   | #### | #### | #### | #### | #### |  |
|                 | 1992   | 7         | 4    | 6    | 6    | 7    | #### | #### | #### | #### |  |
|                 | 1991   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1990   | 2         | 8    | 10   | 25   | 32   | 18   | 4    | #### | #### |  |
|                 | 1989   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
| BENEWAH Total   |        | 236       | 6    | 10   | 19   | 15   | 25   | 30   | #### | #### |  |
| CHERRY          | 1994   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1990   | 6         | 4    | 7    | 12   | 15   | 21   | 22   | #### | #### |  |
| CHERRY Total    |        | 8         | 4    | 6    | 12   | 15   | 21   | 22   | #### | #### |  |
| EVANS           | 1996   | 7         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 35        | 6    | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 54        | 8    | 20   | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 31        | 6    | 8    | 23   | #### | #### | #### | #### | #### |  |
|                 | 1992   | 6         | 5    | 5    | 5    | 4    | #### | #### | #### | #### |  |
| EVANS Total     |        | 133       | 7    | 16   | 21   | 4    | #### | #### | #### | #### |  |
| LAKE            | 1996   | 59        | #### | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1995   | 54        | 7    | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1994   | 72        | 5    | 9    | #### | #### | #### | #### | #### | #### |  |
|                 | 1993   | 30        | 5    | 9    | 23   | #### | #### | #### | #### | #### |  |
|                 | 1992   | 6         | 8    | 10   | 13   | 17   | #### | #### | #### | #### |  |
|                 | 1990   | 21        | 3    | 7    | 8    | 9    | 11   | 11   | #### | #### |  |
|                 | 1989   | 6         | 4    | 5    | 11   | 10   | 13   | 13   | 17   | #### |  |
|                 | 1988   | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
| LAKE Total      |        | 249       | 7    | 10   | 20   | 13   | 14   | 15   | 16   | #### |  |
| Grand Total     |        | 680       | 6    | 12   | 20   | 15   | 17   | 18   | 16   | #### |  |

**Mean back-calculated lengths at annulus formation by age class and cohort - 1996**

| SPECIES       |        | Age Class |      |      |      |      |      |      |      |      |  |
|---------------|--------|-----------|------|------|------|------|------|------|------|------|--|
| Brook Trout   |        |           |      |      |      |      |      |      |      |      |  |
| LOCATION      | Cohort | Number    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER         | 1996   | 19        | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1995   | 3         | 84   | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1994   | 28        | 83   | 116  | #### | #### | #### | #### | #### | #### |  |
|               | 1993   | 34        | 85   | 121  | 152  | #### | #### | #### | #### | #### |  |
|               | 1992   | 21        | 85   | 126  | 150  | 200  | #### | #### | #### | #### |  |
|               | 1991   | 8         | 81   | 118  | 150  | 192  | 227  | #### | #### | #### |  |
| ALDER Total   |        | 113       | 84   | 120  | 151  | 198  | 227  | #### | #### | #### |  |
| BENEWAH       | 1996   | 4         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1995   | 4         | 76   | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1993   | 4         | 112  | 119  | 161  | #### | #### | #### | #### | #### |  |
|               | 1992   | 8         | 83   | 126  | 162  | 211  | #### | #### | #### | #### |  |
|               | 1991   | 6         | 87   | 134  | 176  | 220  | 268  | #### | #### | #### |  |
| BENEWAH Total |        | 26        | 88   | 127  | 166  | 215  | 268  | #### | #### | #### |  |
| Grand Total   |        | 139       | 85   | 121  | 155  | 203  | 244  | #### | #### | #### |  |

**Standard deviation of mean back-calculated lengths at annulus formation - 1996**

| SPECIES       |        | Age Class |      |      |      |      |      |      |      |      |  |
|---------------|--------|-----------|------|------|------|------|------|------|------|------|--|
| Brook Trout   |        |           |      |      |      |      |      |      |      |      |  |
| LOCATION      | Cohort | Number    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER         | 1996   | 19        | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1995   | 3         | 5    | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1994   | 28        | 5    | 9    | #### | #### | #### | #### | #### | #### |  |
|               | 1993   | 34        | 6    | 7    | 12   | #### | #### | #### | #### | #### |  |
|               | 1992   | 21        | 5    | 6    | 31   | 10   | #### | #### | #### | #### |  |
|               | 1991   | 8         | 4    | 8    | 9    | 16   | 21   | #### | #### | #### |  |
| ALDER Total   |        | 113       | 5    | 8    | 20   | 12   | 21   | #### | #### | #### |  |
| BENEWAH       | 1996   | 4         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1995   | 4         | 1    | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1993   | 4         | 54   | 18   | 18   | #### | #### | #### | #### | #### |  |
|               | 1992   | 8         | 4    | 5    | 12   | 15   | #### | #### | #### | #### |  |
|               | 1991   | 6         | 8    | 13   | 20   | 16   | 21   | #### | #### | #### |  |
| BENEWAH Total |        | 26        | 24   | 12   | 17   | 16   | 21   | #### | #### | #### |  |
| Grand Total   |        | 139       | 12   | 9    | 20   | 16   | 29   | #### | #### | #### |  |

Mean back-calculated lengths at annulus formation by age class and cohort - 1997

| SPECIES         |               | Age Class |      |      |      |      |      |      |      |      |  |
|-----------------|---------------|-----------|------|------|------|------|------|------|------|------|--|
| Cutthroat Trout |               |           |      |      |      |      |      |      |      |      |  |
| LOCATION        | Cohort Number | 0         | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER           | 1997          | 6         | 55   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1996          | 7         | #### | 67   | #### | #### | #### | #### | #### | #### |  |
|                 | 1995          | 6         | #### | 65   | 110  | #### | #### | #### | #### | #### |  |
|                 | 1994          | 6         | #### | 71   | 118  | 159  | #### | #### | #### | #### |  |
| ALDER Total     | 25            | 55        | 68   | 114  | 159  | #### | #### | #### | #### | #### |  |
| BENEWAH         | 1997          | 5         | 60   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1996          | 7         | #### | 69   | #### | #### | #### | #### | #### | #### |  |
|                 | 1995          | 20        | #### | 67   | 113  | #### | #### | #### | #### | #### |  |
|                 | 1994          | 9         | #### | 80   | 128  | 183  | #### | #### | #### | #### |  |
|                 | 1993          | 6         | #### | 74   | 124  | 177  | 220  | #### | #### | #### |  |
|                 | 1992          | 1         | #### | 72   | 135  | 189  | 242  | 295  | #### | #### |  |
|                 | 1990          | 1         | #### | 73   | 117  | 179  | 214  | 258  | 302  | 337  |  |
| 1989            | 1             | ####      | 68   | 107  | 154  | 193  | 224  | 271  | 302  | 349  |  |
| BENEWAH Total   | 50            | 60        | 71   | 119  | 180  | 219  | 259  | 286  | 320  | 349  |  |
| EVANS           | 1997          | 3         | 60   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1996          | 9         | #### | 70   | #### | #### | #### | #### | #### | #### |  |
|                 | 1995          | 7         | #### | 68   | 121  | #### | #### | #### | #### | #### |  |
|                 | 1994          | 8         | #### | 69   | 115  | 160  | #### | #### | #### | #### |  |
|                 | 1993          | 7         | #### | 66   | 113  | 164  | 202  | #### | #### | #### |  |
|                 | 1992          | 1         | #### | 64   | 106  | 148  | 190  | 232  | #### | #### |  |
| EVANS Total     | 35            | 60        | 68   | 116  | 161  | 200  | 232  | #### | #### | #### |  |
| LAKE            | 1997          | 5         | 57   | #### | #### | #### | #### | #### | #### | #### |  |
|                 | 1996          | 5         | #### | 68   | #### | #### | #### | #### | #### | #### |  |
|                 | 1995          | 22        | #### | 65   | 114  | #### | #### | #### | #### | #### |  |
|                 | 1994          | 14        | #### | 62   | 110  | 151  | #### | #### | #### | #### |  |
|                 | 1993          | 3         | #### | 67   | 105  | 155  | 205  | #### | #### | #### |  |
|                 | 1991          | 1         | #### | 63   | 125  | 176  | 227  | 268  | 320  | #### |  |
|                 | 1990          | 2         | #### | 67   | 122  | 167  | 207  | 248  | 297  | 338  |  |
| LAKE Total      | 52            | 57        | 65   | 112  | 154  | 210  | 255  | 305  | 338  | #### |  |
| Grand Total     | 162           | 58        | 68   | 115  | 164  | 210  | 253  | 297  | 329  | 349  |  |

Standard deviation of mean back-calculated lengths at annulus formation - 1997

| SPECIES       |               | Age Class |      |      |      |      |      |      |      |      |  |
|---------------|---------------|-----------|------|------|------|------|------|------|------|------|--|
|               |               |           |      |      |      |      |      |      |      |      |  |
| Location      | Cohort Number | 0         | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER         | 1997          | 6         | 7    | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1996          | 7         | #### | 17   | #### | #### | #### | #### | #### | #### |  |
|               | 1995          | 6         | #### | 6    | 6    | #### | #### | #### | #### | #### |  |
|               | 1994          | 6         | #### | 10   | 17   | 26   | #### | #### | #### | #### |  |
| ALDER Total   | 25            | 7         | 12   | 13   | 26   | #### | #### | #### | #### | #### |  |
| BENEWAH       | 1997          | 5         | 2    | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1996          | 7         | #### | 7    | #### | #### | #### | #### | #### | #### |  |
|               | 1995          | 20        | #### | 11   | 14   | #### | #### | #### | #### | #### |  |
|               | 1994          | 9         | #### | 8    | 15   | 27   | #### | #### | #### | #### |  |
|               | 1993          | 6         | #### | 10   | 12   | 23   | 27   | #### | #### | #### |  |
|               | 1992          | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1990          | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
| 1989          | 1             | ####      | #### | #### | #### | #### | #### | #### | #### |      |  |
| BENEWAH Total | 50            | 2         | 11   | 15   | 24   | 25   | 35   | 22   | 25   | #### |  |
| EVANS         | 1997          | 3         | 14   | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1996          | 9         | #### | 11   | #### | #### | #### | #### | #### | #### |  |
|               | 1995          | 7         | #### | 8    | 9    | #### | #### | #### | #### | #### |  |
|               | 1994          | 8         | #### | 6    | 7    | 11   | #### | #### | #### | #### |  |
|               | 1993          | 7         | #### | 4    | 4    | 6    | 6    | #### | #### | #### |  |
|               | 1992          | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
| EVANS Total   | 35            | 14        | 7    | 8    | 9    | 7    | #### | #### | #### | #### |  |
| LAKE          | 1997          | 5         | 3    | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1996          | 5         | #### | 5    | #### | #### | #### | #### | #### | #### |  |
|               | 1995          | 22        | #### | 6    | 10   | #### | #### | #### | #### | #### |  |
|               | 1994          | 14        | #### | 5    | 9    | 10   | #### | #### | #### | #### |  |
|               | 1993          | 3         | #### | 3    | 8    | 8    | 12   | #### | #### | #### |  |
|               | 1991          | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | 1990          | 2         | #### | 0    | 1    | 1    | 8    | 2    | 9    | 3    |  |
| LAKE Total    | 52            | 3         | 6    | 10   | 11   | 12   | 12   | 14   | 3    | #### |  |
| Grand Total   | 162           | 6         | 9    | 12   | 20   | 18   | 24   | 18   | 18   | #### |  |

Mean back-calculated lengths at annulus formation by age class and cohort - 1997

SPECIES



Brook Trout

| Location      | Cohort | Number | 0    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |
|---------------|--------|--------|------|------|------|------|------|------|------|------|------|
| ALDER         | 1997   | 3      | 58   | #### | #### | #### | #### | #### | #### | #### | #### |
|               | 1996   | 5      | #### | 69   | #### | #### | #### | #### | #### | #### | #### |
|               | 1995   | 16     | #### | 58   | 104  | #### | #### | #### | #### | #### | #### |
|               | 1994   | 10     | #### | 65   | 107  | 143  | #### | #### | #### | #### | #### |
|               | 1993   | 4      | #### | 63   | 102  | 133  | 167  | #### | #### | #### | #### |
|               | 1992   | 3      | #### | 73   | 128  | 183  | 244  | 299  | #### | #### | #### |
| ALDER Total   |        | 41     | 58   | 63   | 107  | 148  | 200  | 299  | #### | #### | #### |
| BENEWAH       | 1996   | 1      | #### | 60   | #### | #### | #### | #### | #### | #### | #### |
|               | 1995   | 4      | #### | 55   | 90   | #### | #### | #### | #### | #### | #### |
|               | 1994   | 2      | #### | 80   | 117  | 148  | #### | #### | #### | #### | #### |
|               | 1993   | 2      | #### | 94   | 142  | 190  | 224  | #### | #### | #### | #### |
|               | 1992   | 1      | #### | 96   | 163  | 230  | 313  | 347  | #### | #### | #### |
| BENEWAH Total |        | 10     | #### | 72   | 115  | 181  | 254  | 347  | #### | #### | #### |
| Grand Total   |        | 51     | 58   | 65   | 109  | 155  | 216  | 311  | #### | #### | #### |

Standard deviation of mean back-calculated lengths at annulus formation - 1997

SPECIES



Brook Trout

| Location      | Cohort | Number | 0    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |
|---------------|--------|--------|------|------|------|------|------|------|------|------|------|
| ALDER         | 1997   | 3      | 7    | #### | #### | #### | #### | #### | #### | #### | #### |
|               | 1996   | 5      | #### | 3    | #### | #### | #### | #### | #### | #### | #### |
|               | 1995   | 16     | #### | 17   | 22   | #### | #### | #### | #### | #### | #### |
|               | 1994   | 10     | #### | 10   | 20   | 30   | #### | #### | #### | #### | #### |
|               | 1993   | 4      | #### | 6    | 19   | 28   | 48   | #### | #### | #### | #### |
|               | 1992   | 3      | #### | 7    | 14   | 13   | 18   | 22   | #### | #### | #### |
| ALDER Total   |        | 41     | 7    | 13   | 21   | 31   | 55   | 22   | #### | #### | #### |
| BENEWAH       | 1996   | 1      | #### | #### | #### | #### | #### | #### | #### | #### | #### |
|               | 1995   | 4      | #### | 4    | 17   | #### | #### | #### | #### | #### | #### |
|               | 1994   | 2      | #### | 5    | 10   | 2    | #### | #### | #### | #### | #### |
|               | 1993   | 2      | #### | 1    | 8    | 15   | 3    | #### | #### | #### | #### |
|               | 1992   | 1      | #### | #### | #### | #### | #### | #### | #### | #### | #### |
| BENEWAH Total |        | 10     | #### | 19   | 30   | 35   | 51   | #### | #### | #### | #### |
| Grand Total   |        | 51     | 7    | 15   | 23   | 34   | 57   | 30   | #### | #### | #### |

# **Coeur d'Alene Tribe Fish, Water, and Wildlife Program**

## **BPA Annual Report, 1998**

---

Coeur d'Alene Tribe Department of Natural Resources  
Fish, Water, and Wildlife Program  
850 A Street, P.O. Box 408  
Plummer, ID 83851-0408





# **Coeur d'Alene Tribe Fish, Water, and Wildlife Program**

## **BPA Annual Report, 1998**

**Prepared By**

**Angelo J. Vitale  
Dee Ann Bailey  
Ron Peters  
And  
Kelly Lillengreen  
Program Manager**

**May 1999**

**Coeur d'Alene Tribe Department of Natural Resources  
Fish, Water, and Wildlife Program  
850 A Street, P.O. Box 408  
Plummer, ID 83851-0408**

**PHONE: (208) 686-5302  
FAX: (208) 686-3021**

## Abstract

As part of an ongoing project to restore fisheries resources in tributaries located on the Coeur d'Alene Indian Reservation, this report details the activities of the Coeur d'Alene Tribe's Fisheries Program for FY 1997 and 1998. This report 1) analyses the effect introduced species and water quality have on the abundance of native trout in Coeur d'Alene Lake and selected target tributaries; 2) details results from an ongoing mark-recapture study on predatory game fish; 3) characterizes spawning habitats in target tributaries and evaluates the effects of fine sediment on substrate composition and estimated emergence success; and 4) provides population estimates for westslope cutthroat trout in target tributaries.

Low dissolved oxygen values in the hypolimnion of Coeur d'Alene Lake continue to be a cause for concern with regard to available fisheries habitat. Four sample sites in 1997 and eight sample sites in 1998 had measured levels of dissolved oxygen below what is considered optimum (6.0 mg/L) for cutthroat trout. As well, two sample points located north of the Coeur d'Alene River showed hypolimnetic dissolved oxygen deficits. This could lead to a more serious problem associated with the high concentration of heavy metals bound up in the sediment north of the Coeur d'Alene River. Most likely these oxygen deficits are a result of allochthonous input of organic matter and subsequent decomposition.

Sediment loading from tributaries continues to be a problem in the lake. The build up of sediments at the mouths of all incoming tributaries results in the modification of existing wetlands and provides ideal habitat for predators of cutthroat trout, such as northern pike and largemouth bass. Furthermore, increased sediment deposition provides additional substrate for colonization by aquatic macrophytes, which serve as forage and habitat for other non-native species.

There was no significant difference in the relative abundance of fishes in Coeur d'Alene Lake from 1997 to 1998. Four out of the six most commonly sampled species are non-native. Northern pikeminnow and largescale suckers are the only native species among the six most commonly sampled. Northern pikeminnow comprise 8-9% of the electroshocking catch and 18-20% of the gillnet catch. Largescale suckers comprise 24-28% of the electroshocking catch and 9-21% of the gillnet catch. Cutthroat trout and mountain whitefish, on the other hand, comprise less than 1% of the catch when using electroshocking methods and about 1.4% of the gillnet catch.

Since 1994, the Coeur d'Alene Tribe Fish, Water and Wildlife Program has conducted an extensive mark-recapture study (Peters et al. 1999). To date, 636 fish have been tagged and 23 fish have been recaptured. We are finding that northern pike have a tendency to migrate from the original sampling site, while largemouth bass appear very territorial, rarely moving from the site where they were tagged. Both species are most commonly associated with shallow, near-shore habitats, where the potential for encountering seasonal migrations of cutthroat trout is maximized.

Low-order tributaries provide the most important spawning habitat for cutthroat trout on the Reservation. The mapped distribution of potentially suitable spawning gravel was patchy and did not vary considerably within reaches or between watersheds. Furthermore, the quantity of spawning gravel was low, averaging just 4.1% of measured stream area. The lack of a strong association between spawning gravel abundance and several reach characteristics (gradient, proportion of gravel and pea gravel) corroborates the findings of other authors who suggest that local hydrologic features influence spawning gravel availability. Although the distribution of spawning substrate was patchy within the target watersheds, there is probably adequate habitat to support resident and adfluvial spawners because of currently depressed numbers.

Spawning gravels in target tributaries of the Reservation contained proportions of fine sediments comparable to those in egg pockets of salmonid redds in the Rocky Mountain region. At 23 of 29 sample sites, low levels of fine sediment led to high predictions of overall embryo survival (mean = 28.4%). The estimates of fry production potential at sample sites ranged widely (0.0 to 31.2 fry/100 square meters) due, primarily, to the quantity of suitable gravels present. Only in the mainstem of Lake Creek were the proportions of both small and coarse fines considered above the levels for these particle sizes (10% and 30%, respectively) shown to adversely affect salmonid emergence success. Of the 6 sites where high levels of small or coarse fines were recorded, only the sites located in the mainstem of Lake Creek showed supporting evidence for low recruitment. The other sites had juvenile and adults densities (range = 12.4 – 33.8/100 square meters) that were notably higher than the average density (9.2/100 square meters) reported for seven other westslope cutthroat trout populations in Idaho and Montana. The habitat areas that supported higher than average trout densities, however, comprise relatively small fractions of the available habitat in the respective watersheds.

Due to the persistence of adverse water quality and habitat conditions in Reservation streams, cutthroat trout populations are thought to be at least moderately damaged. The average spawning escapements fall between the minimum viable population and the number of adults needed to produce 50% of the carrying capacity of the stream environment. None of the populations on the Coeur d'Alene Reservation are thought to be greater than or equal to 50% of the historic potential. These weakened populations may be particularly susceptible to normal environmental variability (such as temperature and stream discharge patterns) and to the frequency of extreme events such as wildfires, floods, or debris torrents. On the other hand, very little hybridization has occurred, suggesting that preservation of the existing genetic diversity may continue to allow for genetic combinations that permit survival in highly variable environments.

## **Acknowledgements**

Many individuals contributed to the completion of this report. A special thanks to:

George Aripa

Abel C. Sanchez

Richard Allen

John LaSarte

Sam Sanchez

Jeff Jordon

Kevin Brown

Darryl Trevino

Jim Whistocken

Charles Peone

Brian Harper

Carla Moore

And the rest of the Coeur d'Alene Tribe Fish, Water and Wildlife Program Staff

And the Coeur d'Alene Tribal Council, Ernie Stensgar Tribal Chairman

Funding was provided by the U.S. Department of Energy, Bonneville Power Administration.

# Table of Contents

|   |           |
|---|-----------|
| <b>1.0 Introduction .....</b>   | <b>1</b>  |
| Fisheries Management History of the Coeur d'Alene System .....  | 1         |
| Current Study Objectives .....  | 3         |
| Study Area.....   | 3         |
| <br>  |           |
| <b>2.0 Methods and Materials .....</b>  | <b>4</b>  |
| 2.1 Lake Studies .....  | 4         |
| 2.1.1 Water Quality .....   | 4         |
| 2.1.2 Fisheries .....   | 6         |
| 2.2 Stream Studies .....  | 8         |
| 2.2.1 Water Quality .....   | 8         |
| 2.2.2 Spawning Gravel Survey and Analysis .....   | 9         |
| 2.2.3 Population Surveys .....  | 10        |
| <br>  |           |
| <b>3.0 Results .....</b>  | <b>16</b> |
| 3.1 Lake Studies .....  | 16        |
| 3.1.1 Water Quality .....   | 16        |
| 3.1.2 Fisheries .....   | 33        |
| 3.2 Stream Studies .....  | 38        |
| 3.2.1 Water Quality .....   | 38        |
| 3.2.2 Spawning Gravel Survey and Analysis .....   | 41        |
| 3.2.3 Population Surveys.....   | 41        |
| <br>  |           |
| <b>4.0 Discussion .....</b>   | <b>48</b> |
| 4.1 Lake Studies .....  | 48        |
| 4.2 Stream Studies.....   | 51        |
| <br>  |           |
| <b>5.0 Bibliography .....</b>   | <b>55</b> |
| <br>  |           |
| <b>Appendix A. Vertical hydrolab profiles of thirteen stations on Coeur d'Alene Lake, 1997 and 1998.</b>  |           |
| <b>Appendix B. Water quality results from Coeur d'Alene Lake during 1997 and 1998.</b>  |           |
| <b>Appendix C. Secchi readings and empirically derived estimates of euphotic zone depth for thirteen stations on Coeur d'Alene Lake, 1997 and 1998.</b>                   |           |
| <b>Appendix D. Age distribution, length and weight frequencies and length and weight regression graphs for aged species sampled in Coeur d'Alene Lake, 1997 and 1998.</b> |           |
| <b>Appendix E. Statistics of mean length and weight, standard deviation and range for all aged species sampled in Coeur d'Alene Lake, 1997 and 1998.</b>                  |           |

**Appendix F. Analysis of Spawning gravel beds in select tributaries of the Coeur d'Alene Reservation, 1998.**

**Appendix G. Cohort analysis of growth for cutthroat trout and brook trout in select tributaries of the Coeur d'Alene Reservation, 1998.**

## List of Tables

|  |    |
|--|----|
| Table 2.1 Water quality sample sites on Coeur d'Alene Lake grouped by habitat area .....   | 6  |
| Table 2.2 Stream water quality sites and monitoring parameters.....  | 9  |
| Table 2.3 Juvenile and adult fish traps were operated at eleven different locations in the four target drainages .....   | 16 |
| Table 3.1 Water quality stations on Coeur d'Alene Lake that had dissolved oxygen readings less than 6.0 mg/L .....   | 17 |
| Table 3.2 Average annual and seasonal secchi measurements taken at thirteen stations on Coeur d'Alene Lake, 1997.....  | 31 |
| Table 3.3 Average annual and seasonal secchi measurements taken at thirteen stations on Coeur d'Alene Lake, 1998.....  | 32 |
| Table 3.4 Electroshocking relative abundance results from 1997 and 1998 .....  | 33 |
| Table 3.5 Electroshocking catch per unit effort during 1997 .....  | 35 |
| Table 3.6 Gillnetting relative abundance results from 1997 and 1998.....   | 35 |
| Table 3.7 Gillnetting catch per unit effort during 1997.....   | 36 |
| Table 3.8 Electroshocking catch per unit effort during 1998 .....  | 36 |
| Table 3.9 Gillnetting catch per unit effort during 1998.....   | 37 |
| Table 3.10 Habitat suitability index for lucustrine cutthroat trout 1997 .....   | 39 |
| Table 3.11 Habitat suitability index for lucustrine cutthroat trout 1998 .....   | 40 |
| Table 3.12 Means of habitat features in spawning reaches of the four target watersheds .....   | 41 |
| Table 3.13 Number of cores, substrate composition of cores, predicted mean emergence success and mean estimated production potential for reaches of the four target watersheds ..... | 42 |
| Table 3.14 Cutthroat trout abundance and distribution 1998 .....   | 44 |
| Table 3.15 Brook trout abundance and distribution 1998 .....   | 45 |
| Table 3.16 Age frequency distribution of cutthroat trout 1998.....   | 45 |
| Table 4.1 Mean portion of fines and mean fredle index of egg pockets in salmonid redds.....  | 52 |
| Table 4.2 Estimated mean annual variance in the infinitesimal rate of growth and probabilities of persistence over 100 years .....   | 53 |

## List of Figures

|  |    |
|--|----|
| Figure 2.1 Coeur d'Alene Lake water quality sample sites .....   | 5  |
| Figure 2.2 Coeur d'Alene Lake fisheries sample sites.....  | 7  |
| Figure 2.3 Lake Creek shock sites.....   | 12 |
| Figure 2.4 Benewah Creek shock sites.....  | 13 |
| Figure 2.5 Evans Creek shock sites.....  | 14 |
| Figure 2.6 Alder Creek shock sites.....  | 15 |
| Figure 3.1 Temperature profiles for Chatcolet Lake Shallow and Round Lake, 1997 and 1998 .....                     | 19 |
| Figure 3.2 Temperature profiles for Benewah Lake, Chatcolet Lake Deep and Hidden Lake, 1997 and 1998 .....         | 21 |
| Figure 3.3 Temperature profiles for Rockford Bay, Carey Bay, Windy Bay Shallow, 1997 and 1998 ....                 | 22 |
| Figure 3.4 Temperature profiles for Conkling Point, Mid Lake Coeur d'Alene and Windy Bay Deep, 1997 and 1998 ..... | 23 |
| Figure 3.5 Temperature profiles for Coeur d'Alene River, 1997 and 1998.....  | 24 |
| Figure 3.6 Temperature profiles for St. Joe River, 1997 and 1998 .....   | 25 |
| Figure 3.7 pH profiles for Benewah Lake, Chatcolet Lake Deep and Hidden Lake, 1997 and 1998 .....                  | 27 |
| Figure 3.8 pH profile for Chatcolet Lake Shallow, 1998 .....   | 28 |
| Figure 3.9 Conductivity profile Coeur d'Alene River, 1997 .....  | 29 |
| Figure 3.10 Regression equations of body length and age for cutthroat and brook trout, 1998.....                   | 46 |
| Figure 3.11 Cutthroat trout migration timing and age frequency in Benewah Creek, 1998.....                         | 47 |
| Figure 3.12 Cutthroat trout migration timing and age frequency in Lake Creek, 1998 .....                           | 47 |



## 1.0 Introduction

The recent decline in native salmonid stocks, in particular cutthroat trout (*Oncorhynchus clarki lewisi*, Richardson) and bull trout (*Salvelinus confluentus*, Walbaum) throughout the Coeur d'Alene basin has eliminated the subsistence fisheries for the Coeur d'Alene Tribe. Historically, the Coeur d'Alene Tribe depended on runs of anadromous salmon and steelhead along the Spokane River and Hangman Creek, as well as resident and adfluvial forms of trout and char in Coeur d'Alene Lake. Dams constructed on the Spokane River at Monroe Street in the City of Spokane and at Little Falls further downstream, were the first dams that initially cut-off the anadromous fish runs from the Coeur d'Alene Tribe. These fisheries were further removed by the construction of Chief Joseph and Grand Coulee Dams on the Columbia River. Together, these actions forced the Tribe to rely solely on the resident fish resources of Coeur d'Alene Lake (Staff Communication).

Many factors contributed to the decline of native salmonid fish stocks in the Coeur d'Alene Basin. These factors include: construction of dams along the Spokane and Columbia Rivers, pollution caused by mining and municipal waste, increases in logging and agricultural activities and species interactions between introduced and native fish species in Coeur d'Alene Lake. Post Falls Dam, built on an existing natural migration barrier in 1906 to supply hydroelectric power to the growing community, raised the water level of Coeur d'Alene Lake by approximately 3.6 meters. This higher lake level inundated traditional Coeur d'Alene Tribal fish traps like the one operated on the St. Joe River at Mission Point for over 50 years (Scholz et al. 1985) and forever changed the physical and hydrologic characteristics of Coeur d'Alene Lake. The higher lake level created conditions that favored introduced species, as well as, native species other than salmonids.

### *Fisheries Management History of the Coeur d'Alene System*

Scholz et al. (1985) estimated that historically the Coeur d'Alene Indian Tribe harvested around 42,000 cutthroat per year. In 1967, Mallet (1968, 1969) reported that 3,329 cutthroat were harvested from the St. Joe River, and a catch of 887 was reported from Coeur d'Alene Lake. This catch is far less than the 42,000 fish per year the tribe harvested historically. Today, only limited opportunities exist to harvest cutthroat trout. Within the Reservation boundaries Lake Creek and its tributaries and Benewah Creek and its tributaries are closed to all fishing until further notice. Both of the above closures extend 100 yards into Coeur d'Alene Lake from the mouth of the respective streams. Bull trout were listed as threatened under the Endangered Species Act in 1998, eliminating harvest opportunities for this species. In other portions of the Spokane River basin, including Coeur d'Alene Lake, only one cutthroat trout over 14" may be harvested per day. The Coeur d'Alene Lake system has three very distinct stocks of cutthroat trout. They are adfluvial-lacustrine, fluvial and resident. Peters et al. (1999) and Graves et al. (1990) have a more detailed description of the different stock types.

In an attempt to diversify the fishery in Coeur d'Alene Lake and offset losses to native trout fisheries, kokanee salmon (*Oncorhynchus nerka*, Walbaum) were introduced into the Coeur d'Alene Lake system in 1937. As kokanee salmon numbers began to increase and stocking was no longer necessary they soon became the most abundant and sought after fish in the lake. By 1967, total harvest of kokanee salmon from Coeur d'Alene Lake was second only to Pend Oreille Lake within the State of Idaho. However, kokanee soon reached their carrying capacity within the lake and in 1982, the Idaho Department of Fish and Game introduced chinook salmon (*Oncorhynchus tshawytscha*, Walbaum) in an attempt to control the burgeoning yet stunted kokanee population. A hatchery run of chinook salmon was established in Wolf Lodge Creek. Attempts were made to control the establishment of naturally reproducing populations of chinook salmon in the lakes' tributaries, however, chinook salmon became established in both the Coeur d'Alene and St. Joe River systems. It is unclear what impacts spawning and rearing chinook salmon have or will have on native trout species.

In 1987, the Northwest Power Planning Council (NPPC) amended the Columbia River Basin Fish and Wildlife Program to include baseline stream survey of tributaries located on the Coeur d'Alene Indian Reservation [section 903 (g)(1)(B)]. Initial studies conducted in 1990 by the Coeur d'Alene Tribe identified twenty-one Reservation streams flowing into Coeur d'Alene Lake, the St. Joe River and the St. Maries River as having potentially suitable habitat for trout species. Data obtained from a subsequent aerial survey prioritized ten of the original twenty-one survey streams for further study, based geographic location, accessibility by trout, and acceptable gradient and habitat (Graves et al. 1990). After compiling data on trout abundance and distribution, growth rates, and benthic macroinvertebrate densities, four tributary streams including Lake, Benewah, Evans and Alder Creeks were selected as having the highest potential for improvement and enhancement. See Graves et al. (1990) and Lillengreen et al. (1993) for further discussions of the past history of the study area.

In 1994, the NPPC adopted the recommendations set forth by the Coeur d'Alene Tribe to improve the Reservation fisheries. These actions included: 1.) Implement habitat restoration and enhancement measures in Lake, Benewah, Evans, and Alder Creeks; 2.) Purchase critical watershed areas for protection of fisheries habitat; 3.) Conduct an educational/outreach program for the general public within the Coeur d'Alene Indian Reservation to facilitate a "holistic" watershed protection process; 4.) Develop an interim fishery for Tribal and non-Tribal members on the Reservation through construction, operation and maintenance of five trout ponds; 5.) Design, construct, operate and maintain a trout production facility; and 6.) Implement a five-year monitoring program to evaluate the effectiveness of the hatchery and habitat improvement projects. These principles, priorities, and objectives were also adopted into the 1995 Fish and Wildlife Program (Section 10 Resident Fish, see paragraph 10.8B.20).

Subsequent studies by the Coeur d'Alene Tribe have determined that the following factors effectively limit the population of cutthroat trout in natal streams and in Coeur d'Alene Lake:

- Stochastic events that result in increased mortality of embryo, fry, and juvenile lifestages (e.g. peak flow events) have been exacerbated by land use practices during the last 60 years;
- Competition for limited space and food during base flow conditions cause displacement of juveniles into water quality limited stream reaches;
- Competitive interactions with introduced salmonids may result in replacement of native trout in Alder Creek and Benewah Creek;
- Water temperatures in the upper ten meters of the water column in Coeur d'Alene Lake exceed the optimum as described in the Habitat Suitability Index for cutthroat trout;
- Sediment loading from tributaries in combination with large quantities of aquatic macrophyte growth and low dissolved oxygen concentrations in the hypolimnion promote conditions more favorable for introduced fish species in the lake; and
- Competitive interactions with introduced species for food, living space, and through predation limit cutthroat trout in both the littoral and limnetic zones of Coeur d'Alene Lake.

Spikes in peak flows exceeded the sheer stress for spawning gravels (5-cm geometric mean particle diameter) during the cutthroat trout egg incubation period in some tributaries during the last few years (Peters et al. 1999). It is conceivable that flow events of this magnitude could scour trout redds and result in complete year class failures. Water quality has been sampled on the four target tributaries since 1990 and studies show that conditions have remained fairly consistent for dissolved oxygen, temperature, pH and conductivity (Graves et al. 1990; Lillengreen et al. 1993 and 1996; Peters et al. 1999). Temperature appears to be the factor that is most influencing cutthroat trout production in the streams. All streams

exceeded the optimum limit of 15.5° C for cutthroat trout during low flow periods in each of the last nine years. Today, there are 21 identified fish species in Coeur d'Alene Lake. Of these 21, only seven are native to Coeur d'Alene Lake. They include cutthroat trout, bull trout, mountain whitefish (Prosopium williamsoni, Girard), sculpin (Cottus sp.), (there may be more than one species of sculpin in the lake), northern pikeminnow (Ptychocheilus oregonensis, Richardson), largescale sucker (Catostomus macrocheilus, Girard), and longnose sucker (Catostomus catostomus, Forster). For a list of the 14 non-native species and their scientific names see Peters et al. (1999).

### Current Study Objectives

- Objective 1:** Estimate fisheries relative abundance in Coeur d'Alene Lake using electroshocking and gillnetting.
- Tasks 1.1* Measure all fish species for length and weight, and take scale samples from game fish.
- Objective 2:** Examine the effects of introduced fishes on native species of cutthroat and bull trout in Coeur d'Alene Lake.
- Tasks 2.1* Correlate population statistics with data on habitat types.
- Objective 3:** Conduct population estimates in Coeur d'Alene Lake using mark-recapture techniques.
- Tasks 3.1* Game fish that are greater than 300g and/or 300mm mark with a floy tag to obtain density estimates and monitor movement.
- Objective 4:** Examine the age structure of sampled fish from scales taken during sampling periods.
- Tasks 4.1* Scale data will be used to monitor the status of successive cohorts.
- Objective 5:** Continue collecting baseline water quality data on Coeur d'Alene Lake.
- Tasks 5.1* Nutrients, total kjeldahl nitrogen, total phosphorus, turbidity, total suspended solids, metals and chlorophyll<sub>a</sub> will be sampled.
- Tasks 5.2* Dissolved oxygen, temperature, pH, conductivity, depth and secchi readings will be taken.
- Objective 6:** Examine the effects of water quality parameters on the abundance and distribution of westslope cutthroat trout in target tributaries.
- Tasks 6.1* Continue monitoring stream discharge, dissolved oxygen, temperature, pH and conductivity at selected sites in Alder, Benewah, Evans and Lake creeks.
- Objective 7:** Determine if fine sediment is limiting recruitment of cutthroat trout populations in target tributaries.
- Tasks 7.1* Characterize spawning habitats of cutthroat trout in areas annually sampled for population estimation.
- Tasks 7.2* Evaluate effects of sedimentation on substrate composition and estimated emergence success.
- Objective 8:** Characterize the abundance and distribution of westslope cutthroat trout and brook trout in target tributaries and compare results with previous sample efforts.
- Tasks 8.1* Develop population estimates for selected sites using electroshocking techniques.
- Tasks 8.2* Examine age class structure of resident trout.
- Tasks 8.3* Monitor immigration and emigration of cutthroat trout in Lake and Benewah Creeks.

### Study Area

The Coeur d'Alene Indian Reservation encompasses 139,005 hectares (343,478 acres) in the panhandle section of north Idaho. Many lakes, streams, and rivers are located on the Reservation. The principle waterbody on the Reservation is Coeur d'Alene Lake. Coeur d'Alene Lake is the second largest lake in Idaho and is located within the 1,730,023-hectare (4,274,888 acres) Spokane River drainage basin. The lake lies in a naturally dammed river valley with the outflow currently controlled by Post Falls Dam. Post Falls Dam also controls the level of the St. Joe River upstream to the town of St. Maries. At full pool

(lake elevation 648.7 meters) the lake covers 129 square kilometers and at minimum pool level (lake elevation of 646.2 meters) the lake covers 122 square kilometers. The lake is 43 kilometers long and anywhere from 2 to 10 kilometers wide. Mean depth is 22 meters with a maximum depth of 63.7 meters (Woods and Berenbrock 1994). The lake is located in two Idaho counties: Kootenai and Benewah. Population centers around the lake are located on the northern most shoreline (Coeur d'Alene) and at the mouth of the Coeur d'Alene River (Harrison). The city of Coeur d'Alene is the largest in Kootenai County and Harrison is the second largest in Benewah County. The largest town in Benewah County (St. Maries) lies about 12 miles upstream of Coeur d'Alene Lake on the St. Joe River.

The two main tributaries of Coeur d'Alene Lake, the Coeur d'Alene and St. Joe Rivers, drain the Coeur d'Alene and St. Joe Mountains. These mountains are composed of primarily metasedimentary rocks of the belt group with local intrusions of granitics. Lower elevations are composed primarily of glaciofluvial deposits. This area receives some of the highest amounts of precipitation in Idaho. The lake receives about 25.4 inches of precipitation annually with more in the higher elevations around the lake (38.3 inches Wallace, ID). The southern end of Coeur d'Alene Lake is made up of four shallow lakes (Hidden, Round, Chatcolet, and Benewah) flooded as a result of construction of Post Falls Dam. Geological data was taken from U.S. Department of Agriculture (1984).

The four tributaries selected for restoration are located almost exclusively on the Reservation and have a combined basin area of 34,853 hectares (86,123 acres) and include 529 kilometers (328 miles) of intermittent and perennial stream channels. The climate and hydrology of the target watersheds are similar in that they are influenced by the maritime air masses from the pacific coast, which are modified by continental air masses from Canada. Summers are mild and relatively dry, while fall, winter, and spring brings abundant moisture in the form of both rain and snow. A seasonal snowpack generally covers the landscape at elevations above 4,500 feet from late November to May. Snowpack between elevations of 3,000 and 4,500 feet falls within the "rain-on-snow zone" and may accumulate and deplete several times during a given winter due to mild storms (US Forest Service 1998). The precipitation that often accompanies these mild storms is added directly to the runoff, since the soils are either saturated or frozen, causing significant flooding.

Anthropogenic disturbances are common in the target watersheds. Recent clearing of land for homesteads, logging, pasture, and agriculture has enhanced the rain-on-snow phenomenon and accelerated the rate of snow pack depletion. In Lake Creek for example, where nearly 40 percent of the basin area has been cleared for agriculture, peak discharges have increased by an estimated 55% for 100-year events when compared with the pre-settlement period (Peters et al. 1999). Lesser amounts of forest clearing have occurred in the other target watersheds, suggesting measurable increases in peak discharges for these areas as well. One of the more profound disturbances that the watersheds have been subjected to is from road construction. The road network includes five state highways, numerous county and municipal roads, and an extensive network of unimproved roads. Those areas with the highest density of roads occur on lands managed primarily for timber production. This road system has been constructed in many of the most sensitive locations (floodplains, and unstable land types) within the watersheds. The density of unimproved roads exceeds 2.5-miles/mile<sup>2</sup> in each of the affected watersheds.

## **2.0 Materials and Methods**

### **2.1 Lake Studies**

#### **2.1.1 Water Quality**

There are five distinct habitat areas in the southern third of Coeur d'Alene Lake that can be distinguished based on geomorphologic condition. The first of these areas is comprised of two shallow water stations

## Coeur d'Alene Lake Water Quality Sites

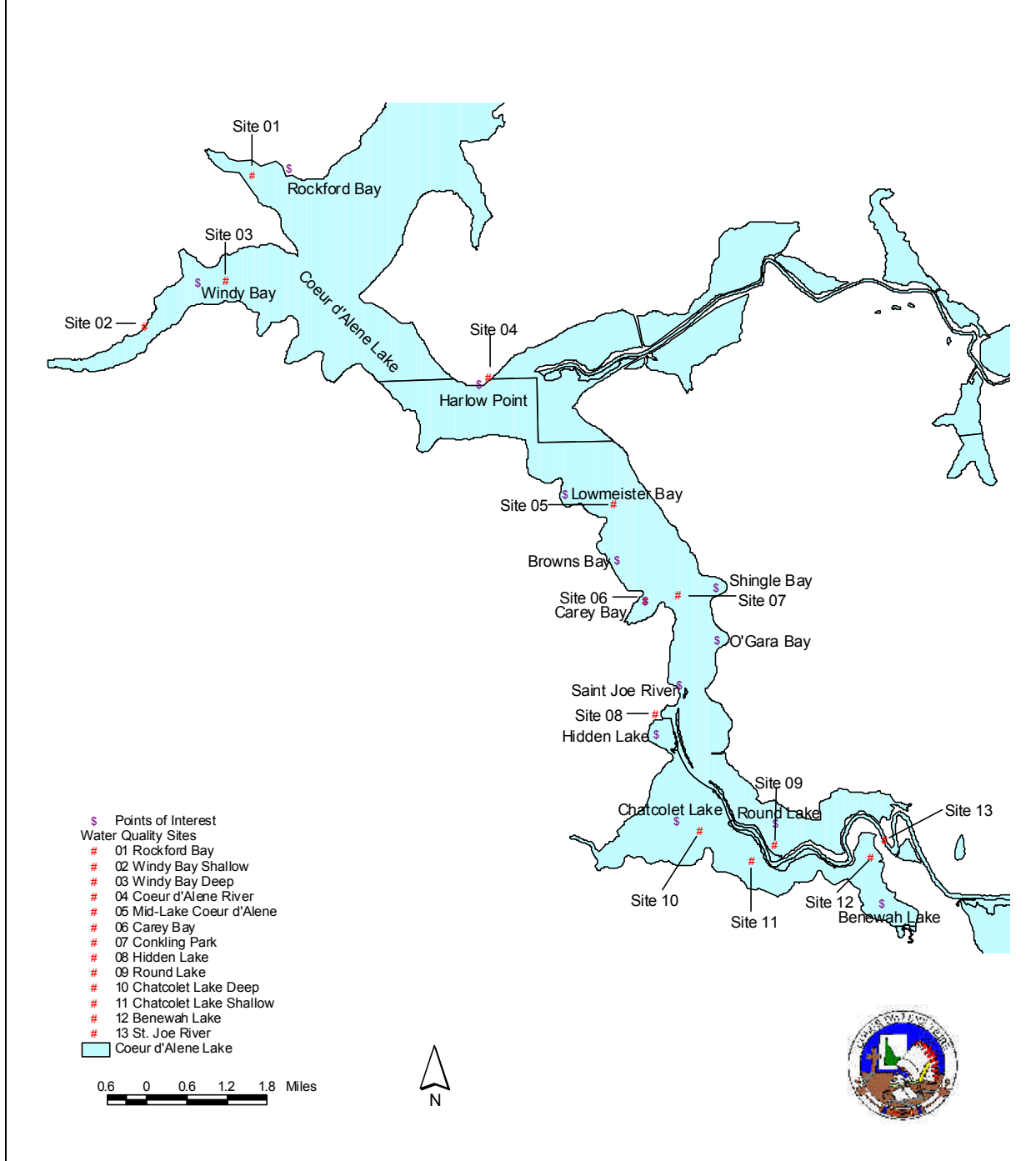


Figure 2.1 Water quality sites on Coeur d'Alene Lake, 1997 and 1998.

created entirely by inundation from Post Falls Dam. This area is dry during the drawdown period and wetted at full pool. The second habitat area is comprised of the three shallow, southern chain lakes of the St. Joe River. These lakes were separated from the Coeur d'Alene Lake system until the completion of Post Falls Dam. The third area consists of three deep, open water sections within the main body of Coeur d'Alene Lake. These areas are considered pelagic in nature. The fourth habitat area consists of three semi-isolated shallow bay areas located in the main Coeur d'Alene Lake. The fifth area is riverine habitat inundated by waters from Post Falls Dam.

A total of thirteen monitoring stations have been established to cover each distinct habitat area (Figure 2.1 and Table 2.1). Combinations of field and laboratory analytical methods were used to sample Coeur d'Alene Lake. See Peters et al. (1999) for a further description on the physical and chemical parameters being sampled as well as, methods used for sampling. A Hydrolab (trade name for a multi-parameter testing probe) was used to get vertical profiles of dissolved oxygen, temperature, pH and conductivity. A kemmerer bottle was used for sampling all other parameters except for water clarity, where a secchi disk was used.

### 2.1.2 Fisheries

The southern section of Coeur d'Alene Lake from Windy Bay south has twelve transect sampling areas that can be grouped based on similar geomorphology (Figure 2.2). Each transect is broken into sample reaches that encompass all habitat types within the transect. Each transect has from 1 to 4 sample reaches. The reach locations were determined by visual habitat characteristics and transect size. These reaches were chosen in order to best represent the shoreline habitat within the transect area. The collective submerged habitat within all reaches spans the range of conditions within each transect area (Peters et al. 1999).

Table 2.1 Water quality sample stations grouped by habitat area.

| Habitat Area         | Stations                                       |
|----------------------|--|
| Shallow Water        | Round Lake<br>Chatcolet Lake Shallow           |
| Southern Chain Lakes | Benewah Lake<br>Chatcolet Lake<br>Hidden Lake  |
| Interior Bays        | Carey Bay<br>Windy Bay Shallow<br>Rockford Bay |
| Open Water           | Conkling Point<br>Mid Lake<br>Windy Bay Deep   |
| Inundated Rivers     | Coeur d'Alene River<br>St. Joe River           |

### Shoreline Electrofishing

The sampling schedule for electroshocking and gillnetting is designed to capture data related to significant changes in relative abundance of various fish species throughout the year. A custom-built aluminum boat equipped with a Smith Root 7.5 GPP electroshock unit was used to conduct shoreline electrofishing. This technique is most effective in water less than 8 feet deep. Each reach was sampled for 5-10 minute, depending on fish density. After completion of the shocking effort, all captured fishes were measured for total length and weight, with scale samples taken only from game species.

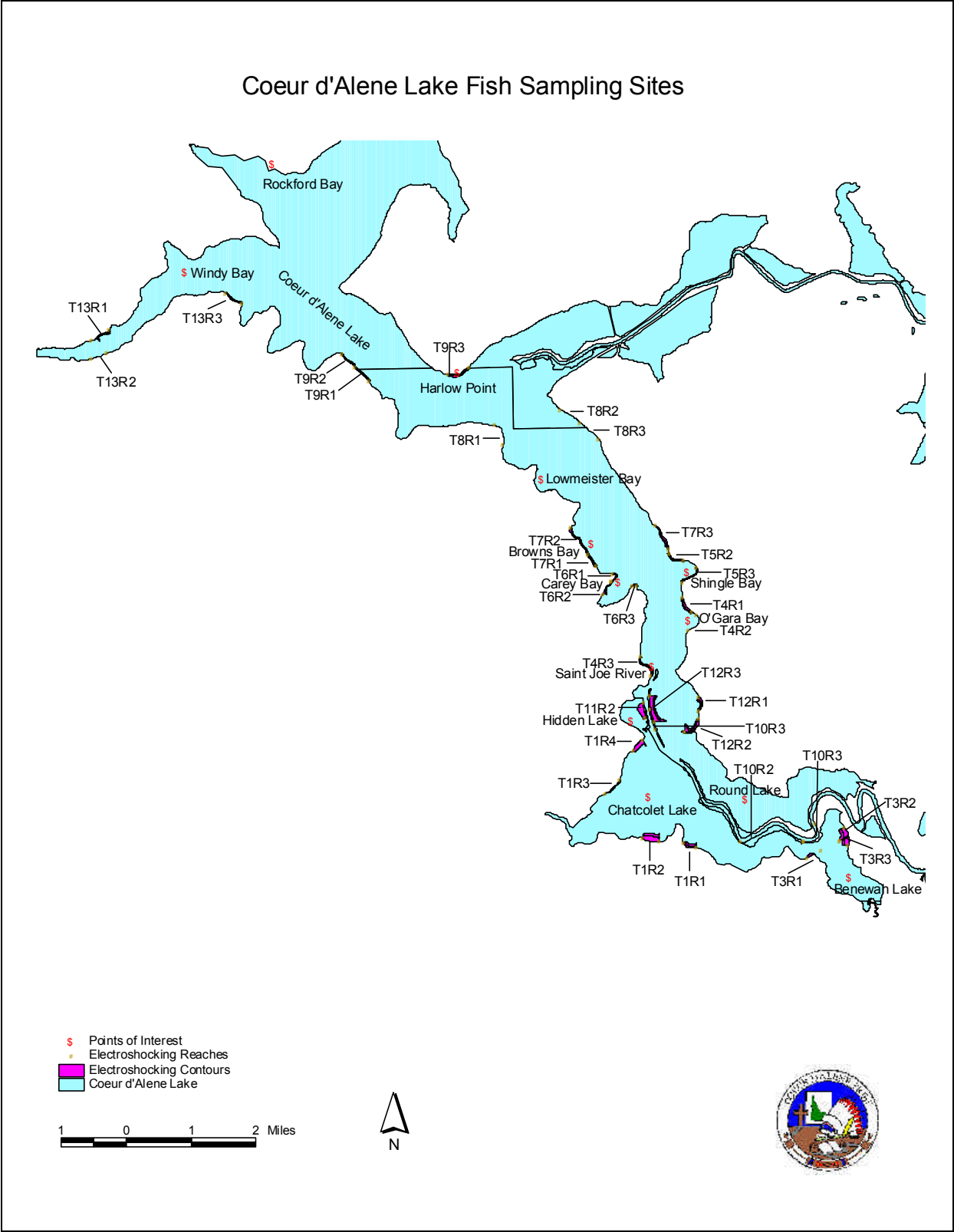


Figure 2.2 Electroshocking and gillnet sites on Coeur d'Alene Lake, 1997 and 1998.

### Gillnet Sampling

Nets were used to sample areas that could not be efficiently sampled using the electrofishing technique, and helped examine species composition as well as seasonal usage in the littoral and limnetic zones. Two different types of nets were used for sampling fish in different habitat areas. Vertical gillnets were used in areas with water depths greater than 30 feet (pelagic zones), while horizontal gillnets were used in shallow areas (8-30 feet deep). The vertical gillnets were 6 feet wide and 120 feet in length with mesh sizes of 2.5", 3", 4", and 5 inches. Horizontal gillnets were made of graded mesh monofilament 8'X200', 10'X200' and 12'X150'. Data recorded following the sample effort included depth of capture, total length, weight and scale samples.

### Age Analysis

Age was determined for all game species captured using methods described by Jearld (1983). Age frequency distributions were plotted for each species to describe the status of successive cohorts. By examining age frequency over time, instances of mortality as well as overall population trends will be identifiable.

### Habitat Suitability Index

The lacustrine habitat suitability index model developed by Hickman and Raleigh (1982) was used to predict suitable habitat for cutthroat trout in Coeur d'Alene Lake. The model consists of a water quality and a reproduction component. The lacustrine water quality component uses three variables: maximum temperature ( $V_1$ ), average minimum dissolved oxygen ( $V_3$ ) and pH ( $V_{13}$ ). Individual suitability index (SI) values were calculated for each variable using curves published in Hickman and Raleigh (1982). The following equation was used to ascertain the final SI values:

$$C_{wq} = (V_1 \times V_3 \times V_{13})^{1/3}$$

Where  $C_{wq}$  = HSI for the water quality component, and  
 $V_n$  = suitability index for water quality variables.

## **2.2 Stream Studies**

### **2.2.1 Water Quality**

Water quality monitoring was conducted on 13 streams in 1997 (Table 2.2). Each stream was sampled for the same parameters as described above for lake studies, except for chlorophyll<sub>a</sub>. Metals analyses were only completed at the two Fighting Creek sample sites. Additional monitoring parameters are described below.

A stage/discharge relationship was developed for each stream based on a linear regression of staff gauge height vs. stream discharge. The rating curve was used to determine the annual water budget for each stream sampled. Staff gauge heights were recorded to the nearest 0.002 of a foot. Discharge measurements were taken at low, medium, and high flows in order to complete the rating curve. Discharge measurements were taken in accordance with standard IFIM methodologies (Bovee 1982). The wetted stream channel was divided into 20 equal cells and water velocity was measured in each cell using a Price model 622 digital flow meter. Discharge for each cell was calculated by multiplying the cell width by depth and velocity. All individual cell discharges were summed to determine total discharge in cubic feet per second. Channel profiles were measured to evaluate changing flow dynamics over time.

### **2.2.2 Spawning Gravel Survey and Analysis**

Potential spawning tributaries were identified over the past six years based on trapping results and population surveys. Active migration into tributaries by adult fish and/or presence of young-of-the-year trout was used as an indication of spawning activity for the purposes of this survey. Habitat features and the area of potential spawning gravel were measured at established shock sites in each tributary.



Table 2.2 Stream water quality sites and monitoring parameters, 1998.

| Location                   | Discharge | Temperature | DO | pH | Conductivity | Turbidity | TSS | Metals | Nutrients |
|----------------------------|-----------|-------------|----|----|--------------|-----------|-----|--------|-----------|
| Lower Plummer Creek        | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Little Plummer Creek       | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Lower Lake Creek           | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Upper Lake Creek           | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Lower Fighting Creek       |           | X           | X  | X  | X            | X         | X   | X      | X         |
| Upper Fighting Creek       |           | X           | X  | X  | X            | X         | X   | X      | X         |
| Evans Creek <sup>a</sup>   | X         | X           | X  | X  | X            | X         | X   |        | X         |
| East Fork Evans            | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Cherry Creek               | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Benewah Creek <sup>b</sup> | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Windfall Creek             | X         | X           | X  | X  | X            | X         | X   |        | X         |
| School House Creek         | X         | X           | X  | X  | X            | X         | X   |        | X         |
| West Fork Benewah Creek    | X         | X           | X  | X  | X            |           |     |        |           |
| North Fork Alder           | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Alder Creek                | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Hangman Creek              | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Little Hangman Creek       | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Indian Creek               | X         | X           | X  | X  | X            | X         | X   |        | X         |
| Mocileme Creek             | X         | X           | X  | X  | X            | X         | X   |        | X         |
| North Fork Rock Creek      | X         | X           | X  | X  | X            | X         | X   |        | X         |

<sup>a</sup> Evans Creek had two sampling stations, Evans and Upper Evans.

<sup>b</sup> Benewah Creek had two sampling stations, Three Mile and Upper Benewah.

Habitat features were measured according to the procedures described by Hankin and Reeves (1988). Proceeding upstream, we described each channel unit (pool, riffle, etc.) and measured its length, mean wetted width, and mean and maximum depths. The area of potential spawning gravel was measured with a meter stick and expressed as a percent of the total surface area at bank full flow. Potential spawning gravel was defined according to Magee et al. (1996) as patches of substrate at least 0.25 m<sup>2</sup> in area with particles 2-35 mm in diameter. In pool tailouts and riffles, we measured substrate composition with the pebble count technique (Wolman 1954), using 100 data points at each site. Substrate was classified as sand-silt (<2 mm in diameter), pea gravel (2-16 mm), gravel (17-64 mm), rubble (65-128 mm), cobble (129-256 mm), or boulder (>256 mm).

To quantify substrate conditions near the time of emergence, we collected core samples using a McNeil hollow-core sampler (Platts et al. 1983) from pockets of potentially suitable spawning gravel. Cores were taken to a depth of 10-20 cm to mimic egg pocket depths observed in cutthroat trout redds (Magee et al. 1996). Oven-dried samples were weighed after being mechanically shaken through sieves of 63.5, 31.5, 16, 8, 6.3, 4.75, 2.0, 1.0, 0.85, 0.5, 0.125, and 0.25 mm. As recommended by Chapman (1988) and Young et al. (1991), we expressed substrate composition by two methods: the percentage of fine substrate smaller than a given size (6.3 and 0.85 mm), and by a measure of central tendency, the fredle index ( $F_i = D_g/S_o$ ; where  $D_g$  is the geometric mean particle diameter and  $S_o$  is a sorting coefficient).

We predicted fry emergence success for each core sample using the equation developed by Weaver and Fraley (1993) for westslope cutthroat trout:

$$\% \text{ emergence} = -0.7512 (\text{arcsine } \% \text{ SP}_{6.3}) + 39.67$$

where,  $\text{SP}_{6.3}$  = percentage of substrate particles smaller than 6.3 mm.

The maximum fry production potential for each shock reach was estimated by combining data on potential redd number, estimated egg deposition, and average substrate composition of core samples. Egg deposition ( $E$ ) was calculated using a length-fecundity relation for westslope cutthroat trout:  $E = 82.63e^{0.007958L}$ , where  $L$  is fork length in millimeters (C. Downs and B. Shepard, Montana State University, personal communication). The average fork length for adfluvial females (310 mm), as measured at

migration traps over the last four years, was used in the equation. Fry production was estimated by multiplying total estimated egg deposition by estimated emergence success.

### 2.2.3 Population Survey

The channel types delineated during previous surveys (Lillengreen, 1993) served as the basic geomorphic units for selecting sample sites for conducting fish population surveys. In these early channel type surveys, stream reaches were classified into relatively homogeneous types according to broad geomorphologic characteristics of stream morphology as defined by Rosgen (1994). Sample sites within each reach were selected to include habitat types representative of the reach as a whole (Figures 2.3 – 2.6). The length of each sample unit was defined as twenty times the average stream width, with a minimum sample distance of 100 meters. Longer stream reaches were sampled more intensively than shorter reaches. Sample sites were also selected in each of the tributaries known to have spawning activity, regardless of whether channel type surveys had been completed. In these cases, sample sites were distributed evenly at approximately 1,500 meter intervals.

Sites were sampled in the summer to quantify the abundance and distribution of fishes during base flow conditions (June 26 – August 12). Trout populations were estimated using the removal-depletion method (Seber and LeCren 1967, Zippin 1958). Blocknets were placed at the upstream and downstream boundaries to prevent immigration and emigration. Each sample site was electrofished using the standard guidelines and procedures described by Reynolds (1983). Fish were collected by spot shocking using a Smith-Root Type VII pulsed-DC backpack electrofisher. Two electrofishing passes were made for each sample site. Salmonid species, including cutthroat trout, brook trout, and bull trout, were the target species for this study. Captured fish were identified, enumerated, measured (TL to nearest mm), and weighed. Cutthroat trout greater than 200 mm in length were tagged with a Floy FD-6B numbered anchor tag. Other species such as longnose dace, redbreast shiner, longnose sucker, and sculpin (sp.) were considered incidental catch and were only counted.

Population estimates were calculated using the following equation (Armour et al. 1983):

$$N = \frac{U_1}{1 - (U_2 / U_1)}$$

where:

N = estimated population size;

U<sub>1</sub> = number of fish collected in the first pass; and

U<sub>2</sub> = number of fish collected in the second pass.

The standard error of the estimate was calculated as:

$$se(N) = \sqrt{\frac{M(1 - M / N)}{A - [(2p)^2 (U_2 / U_1)]}}$$

where:

se(N) = standard error of the population estimate;

M = U<sub>1</sub> + U<sub>2</sub>;

A = (M/N)<sup>2</sup>; and

p =  $1 - \frac{U_2}{U_1}$ .

The population estimates were converted into density values (# fish/100 square meters) for each sample site then extrapolated to the reach in which the samples were collected. The confidence intervals were converted in the same manner. Total reach lengths were obtained from the digital data layer maintained by the Tribal GIS Program.

### Age Analysis

Raw scales were used for age determination and calculating growth rates. Salmonid scales were taken from the side of the body just behind the dorsal fin and above the lateral line (Jearld 1983). Scale samples were sorted by watershed to allow for independent determination of age and growth rate. In the laboratory, several dried scales were mounted between two glass microscope slides and viewed using a Realist, Inc., Vantage 5 microfiche reader. Age was determined by counting the number of annuli (Lux 1971, Jearld 1983). Simultaneous to age determination, a measurement was made from the center of the focus to the furthest edge of the scale. Along this line, measurements were made to each annulus under a constant magnification. Annual growth was then back calculated using the Lee method as described by Carlander (1981). The formula used:

$$L_i = a + \left( \frac{L_c - a}{S_c} \right) S_i$$

where:

- Li = Length of fish (in mm) at each annulus;
- a = intercept of the body scale regression line;
- Lc = length of fish (in mm) at time of capture;
- Sc = distance (in mm) from the focus to the edge of the scale; and
- Si = scale measurement to each annulus.

The intercept (a) was obtained from the linear regression of body length versus scale length at time of capture. The proportional method of back-calculation was used for species with small sample sizes due to poor regression results. The following equation was used:

$$L_i = \left( \frac{S_i}{S_c} \right) L_c$$

This formula does not take into account the size of fish at scale formation as does the Lee method. A linear regression of body length versus age was calculated independently for fish from each subject watershed and the resulting equation was used to determine the age of fish for which scale samples were not taken.

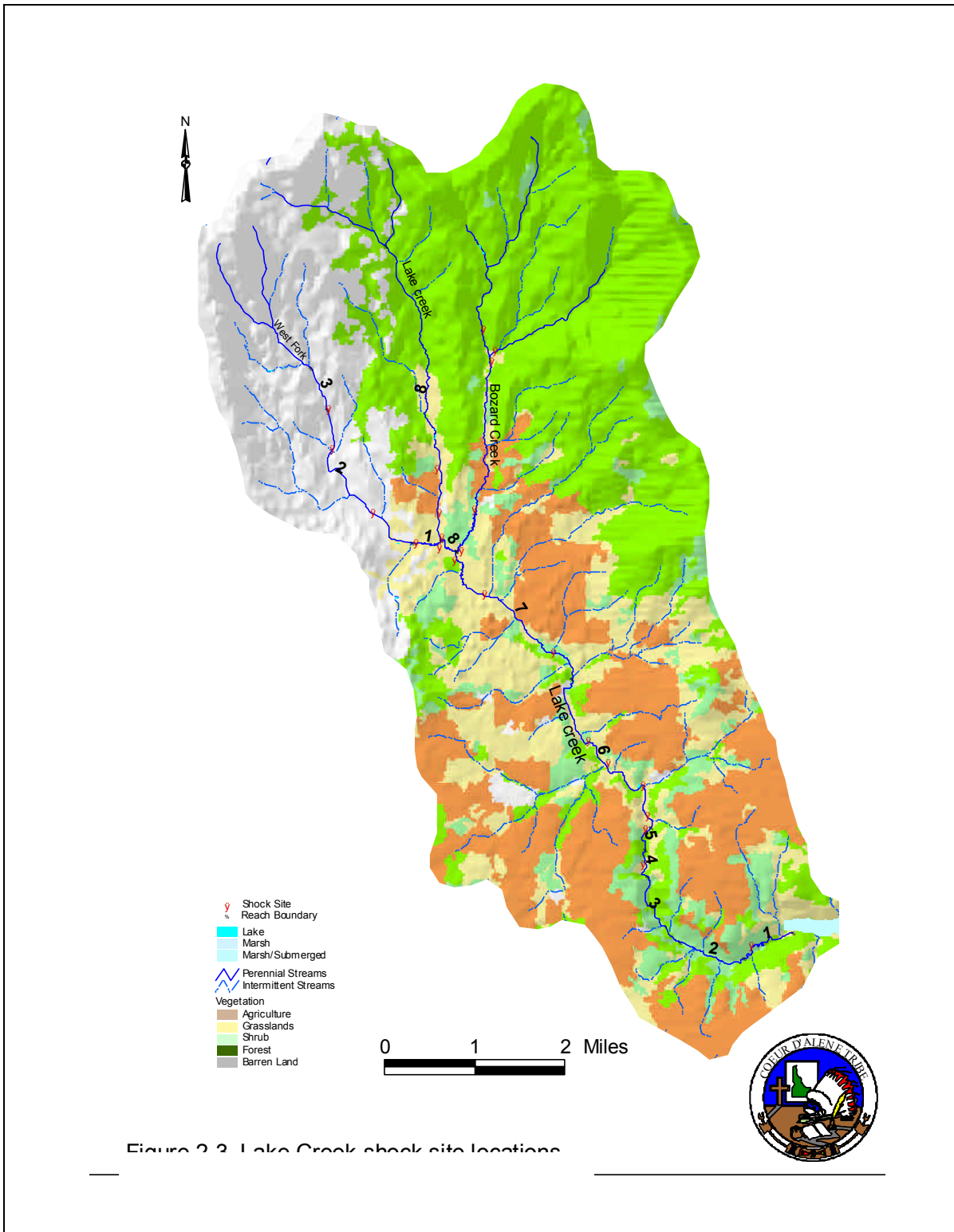


Figure 2.3 Lake Creek shock site locations.

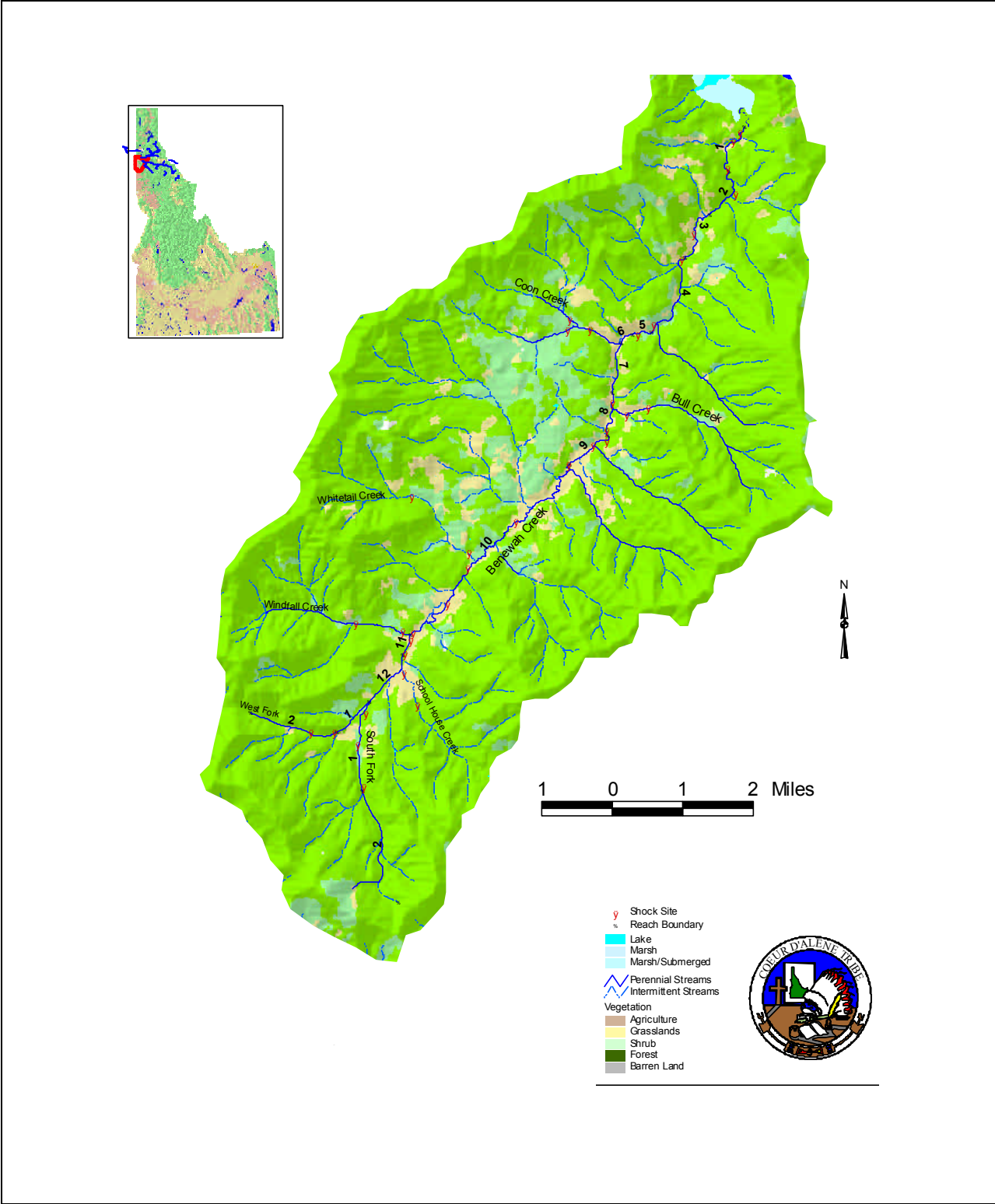


Figure 2.4 Benewah Creek shock site locations.

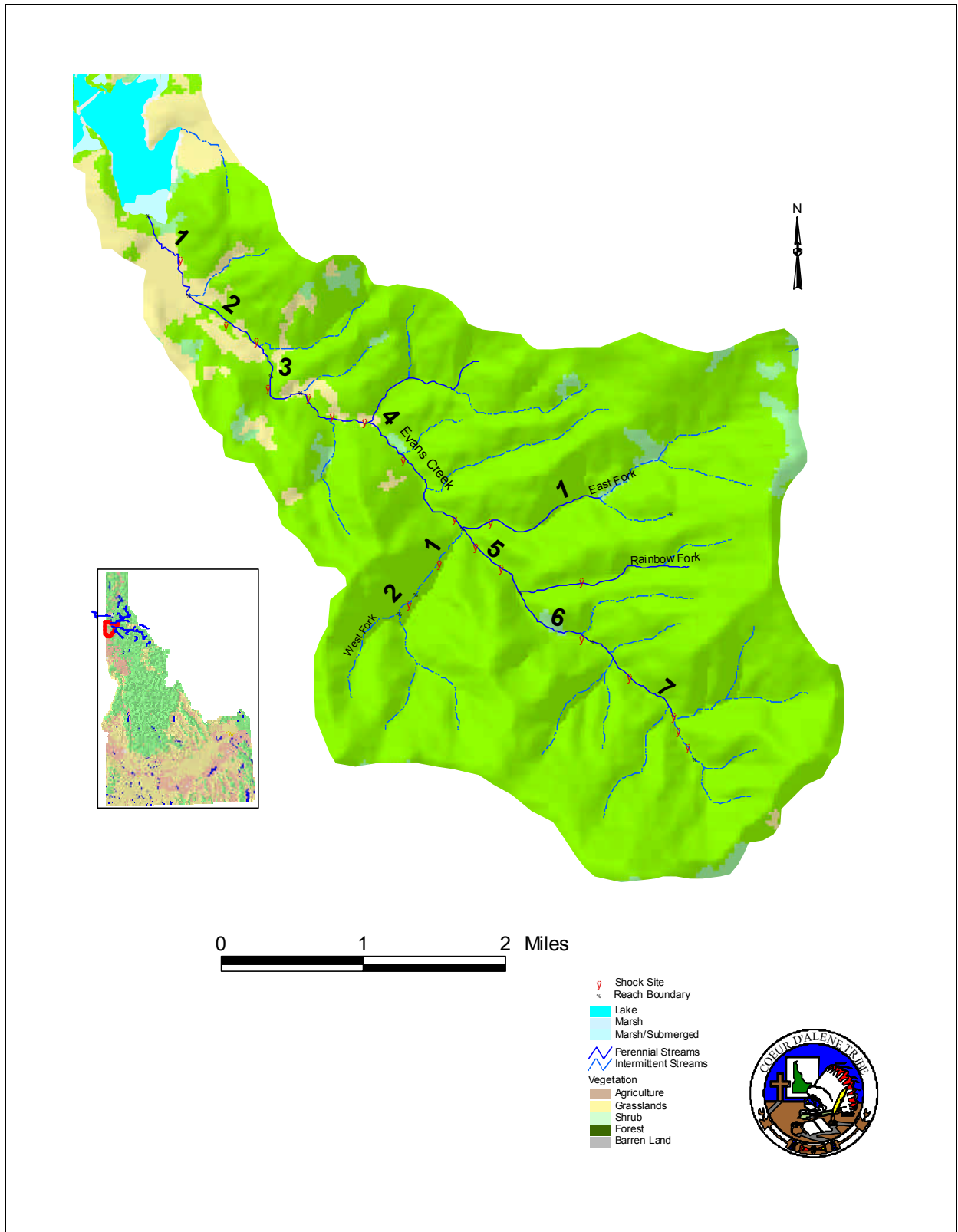


Figure 2.5 Evans Creek shock site locations.

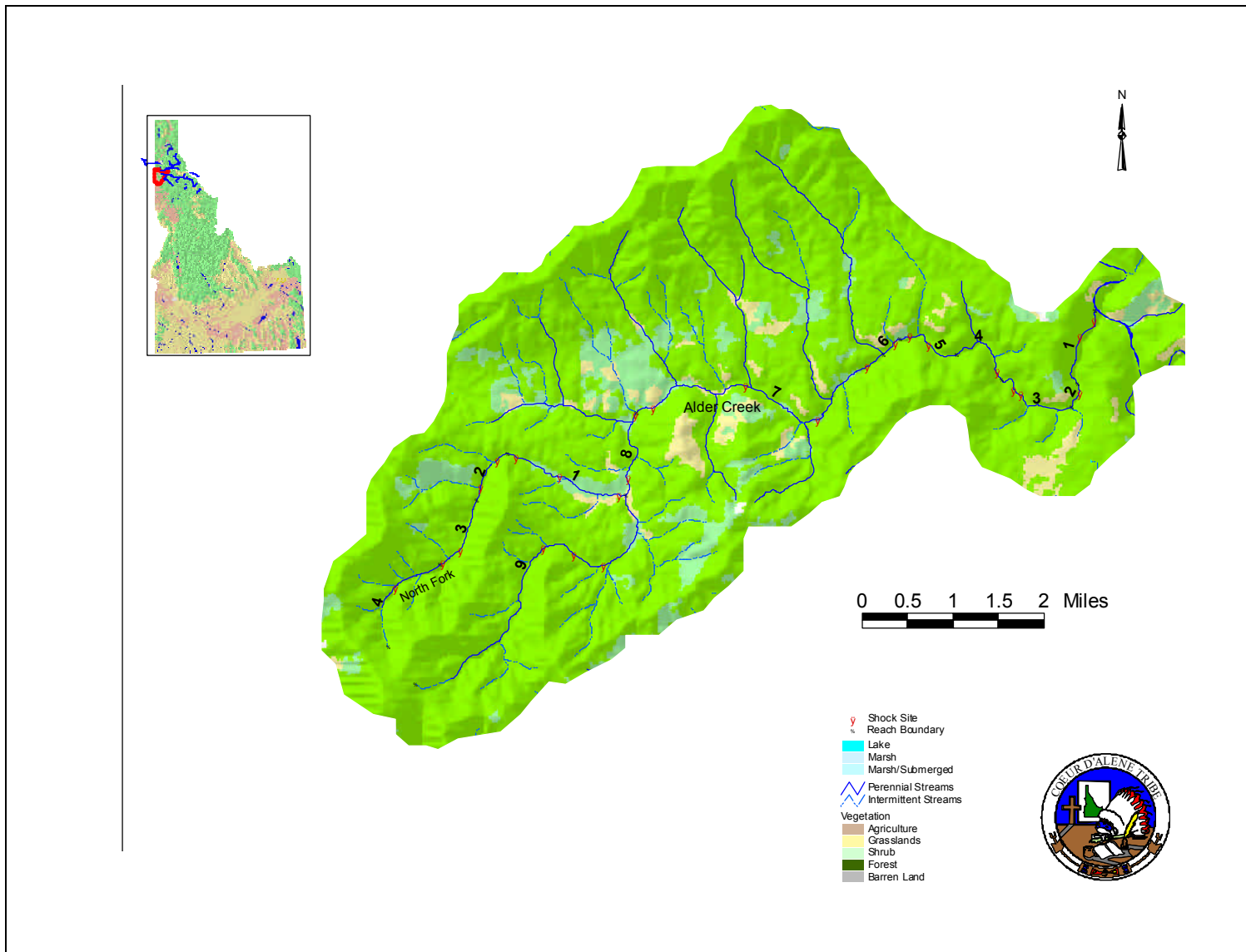


Figure 2.6 Alder Creek shock site locations.

### Trout Migration

Migration traps were installed in Lake Creek, Benewah Creek and Cherry Creek to assess migratory patterns, reproductive cycles, distribution, and relative abundance. Traps were functional from March 24<sup>th</sup> – June 8<sup>th</sup>, 1996 and March 6<sup>th</sup> – June 5<sup>th</sup>, 1997, except during periods of high stream flow. The traps consisted of a weir, runway and a holding box. The design was a modification of the juvenile downstream trap found in Conlin and Tuty (1979). Traps were installed in pairs at some locations to allow monitoring of both upstream and downstream movements of fish (Table 2.3). Paired traps were placed approximately 20 meters apart.

Table 2.3 Fish trap locations in target watersheds, 1998.

| Lake Creek             | Benewah Creek         | Evans Creek       |
|------------------------|-----------------------|-------------------|
| At Hyw 95 crossing (p) | Mouth Benewah (p)     | At mouth of creek |
|                        | Mouth W.F. Benewah    |                   |
|                        | Mouth S.F. Benewah    |                   |
|                        | Mouth Coon Creek      |                   |
|                        | Mouth Whitetail Creek |                   |
|                        | Mouth Windfall Creek  |                   |

(p) Indicates a paired trap location.

Traps were checked at least once daily during peak spawning periods from mid-March through the middle of May and once daily afterwards until late June, when traps were removed. Fish captured in the traps were identified, counted, measured, and weighed. A scale sample was taken to assess the age, growth, and condition of the fish. Trap efficiency was calculated to allow for comparisons among years and was determined as catch per unit effort (CPUE), where one unit effort was defined as one 24-hour period.

## **3.0 Results**

### **3.1 Lake Studies**

#### **3.1.1 Water Quality**

Seven physical/chemical properties of water, two essential nutrients, and 11 dissolved metals were tested for on Coeur d'Alene Lake in 1997 and 1998 on a monthly basis. The testing program was directed primarily at the properties of water that most affected fish production and distribution within the lake. Vertical profiles were also taken for dissolved oxygen, temperature, pH and conductivity, using a multi-parameter water quality testing probe. Samples were taken on a bi-weekly basis during 1997 and 1998.

#### Dissolved Oxygen

Vertical profiles of dissolved oxygen were taken at each of the thirteen stations (Appendix A) during 1997 and 1998. Dissolved oxygen, when at low concentrations is an indication of high levels of organic decomposition. The surface water of every sampling station remained at acceptable levels (>6.0 mg/L) over the course of both years.

In 1997, four of the thirteen sampling stations showed dissolved oxygen concentrations of less than 6.0 mg/L in the hypolimnion, and in 1998, eight out of the thirteen water quality sites had dissolved oxygen



concentrations below 6.0 mg/L in the hypolimnion (Table 3.1). Sites that were similar in geomorphology showed similar results in dissolved oxygen concentrations.

Table 3.1 Water quality stations on Coeur d'Alene Lake that had dissolved oxygen readings less than 6.0 mg/L in 1997 and 1998.

| Location            | Depth (m) | 1997                    | 1998                    |
|---------------------|-----------|-------------------------|-------------------------|
|                     |           | Dissolved Oxygen (mg/L) | Dissolved Oxygen (mg/L) |
| Hidden Lake         | 9         | 0.25                    | 0.12                    |
| Chatcolet Lake Deep | 11        | 0.50                    | 0.10                    |
| Benewah Lake        | 5         | 1.25                    | 0.14                    |
| Mid Lake            | 13        | 4.5                     | 3.75                    |
| Conkling Point      | 16        | -                       | 2.78                    |
| Windy Bay Deep      | 32        | -                       | 5.68                    |
| Carey Bay           | 14        | -                       | 4.6                     |
| St. Joe River       | 13        | -                       | 5.75                    |

During 1997 and 1998, the lowest dissolved oxygen concentrations were recorded in the shallow southern lakes stations. In 1997, the lowest dissolved oxygen concentrations were recorded in Hidden Lake, where a reading of 0.25 mg/L was recorded on ( ) in the lowest one meter of the lake. Dissolved oxygen concentrations in Hidden Lake were in violation of the 6.0 mg/L standard from 6.0 to 8.5 meters or the bottom from ( ). Chatcolet Lake Deep had the second lowest reading 0.50 mg/L on ( ) and was in violation of the 6.0 mg/L standard from 7 to 13 meters or the bottom from ( ). Benewah Lake was also in violation of the 6.0 mg/L standard in the bottom 1 meter with a reading of 1.25 mg/L on ( ). In 1998, Chatcolet Lake Deep had the lowest dissolved oxygen level at 0.10 mg/L on ( ). Dissolved oxygen values were below the 6.0 mg/L standard in the lower 2-4 meters from ( ). Hidden Lake had a reading of .12 mg/L on ( ) and the lowest 1-2 meters were in violation of the 6.0 mg/L standard from ( ). Dissolved oxygen in Benewah Lake was in violation of the 6.0 mg/L standard in the bottom 1 meter with a reading of .14 mg/L on ( ).

During 1997, the only station found in violation of the 6.0 mg/L standard within the main body of Coeur d'Alene Lake was the Mid-Lake Coeur d'Alene station where the 6.0 mg/L standard was violated in the bottom 1 meter with a reading of 4.5 mg/L on ( ). This drop in the dissolved oxygen level is a general indicator of increasing trophic status or eutrophication. The general trend was an increase in trophic status on a north to south axis. However, the Conkling Point sample station, which is between Mid-Lake Coeur d'Alene and Hidden Lake, did not violate the 6.0 mg/L dissolved oxygen standard.

In contrast to 1997, all three sites in the open water zone during 1998 had dissolved oxygen concentrations below 6.0 mg/L. Conkling Point had a dissolved oxygen reading of 2.78 mg/L on ( ) and the lower 2 to 3 meters had dissolved oxygen levels below the 6.0 mg/L standard from ( ). Mid-Lake Coeur d'Alene was in violation of the 6.0 mg/L standard with a reading of 3.75 mg/L in the lower 1 to 4 meters on ( ). Windy Bay Deep had only one day ( ) where the dissolved oxygen dropped below 6.0 mg/L and it was in the lower 4 meters.

During 1998, Carey Bay was the only interior bay site in which dissolved oxygen values dropped below 6.0 mg/L. A value of 4.6 mg/L was recorded in the lower 1 meter on ( ).

In 1998, the St. Joe River had a dissolved oxygen reading of 5.75 mg/L on ( ).

By October of 1998 all but two sites were isothermal with dissolved oxygen values greater than 8.0 mg/L. Windy Bay Deep and Mid- Lake Coeur d'Alene were the only two sites that had not turned over at this time. Turn over is when the thermocline begins to weaken and the epilimnion increases in depth as it decreases in temperature. The stratified structure is lost and the lake becomes homothermous; that is, it has a uniform temperature from surface to bottom.

### Temperature

Optimum temperature requirements for cutthroat and bull trout range between 10° C to 15° C. Vertical profiles of temperature were taken at all thirteen stations during both years (Appendix A).

Geomorphologically similar stations showed similarities in the temperature profiles in both timing of stratification and magnitude of the warming. Shallow stations heated up sooner than deeper water stations. The shallow southern lakes had more variability in the timing and magnitude of the change when compared to the other sample stations (Peters et al. 1999).

In 1997, the shallowest stations, (Round Lake and Chatcolet Lake Shallow @1.5m deep), the epilimnion extended to the bottom before May 16<sup>th</sup> (Figure 3.1). Temperatures greater than the 15.0° C temperature standard existed from July 8<sup>th</sup> through September 29<sup>th</sup> from surface to bottom. The temperature peaked at 26.5° C on August 8<sup>th</sup>.

In 1998, Round Lake and Chatcolet Lake Shallow (Figure 3.1) had temperatures above 15° C from June 22<sup>nd</sup> through September 17<sup>th</sup>, and the epilimnion extended to the bottom. Round Lake had its warmest temperature of 23.86° C on August 11<sup>th</sup> and Chatcolet Lake Shallow reached a high of 26.62° C on July 21<sup>st</sup>.

In 1997, the shallow southern lake stations (Chatcolet Lake Deep, Benewah Lake, and Hidden Lake) the thermocline was present by May 16<sup>th</sup> (Figure 3.2) and reached its deepest point on September 17<sup>th</sup>. The temperature peaked at 26.5° C on August 8<sup>th</sup>.

In 1998, the thermocline varied between the three shallow southern lakes (Figure 3.2). By June 22<sup>nd</sup> the thermocline was in place at Hidden Lake and the upper 3.5 meters had temperatures above 15° C. The temperature peaked on August 11<sup>th</sup> with a surface reading of 25.06° C. Chatcolet Lake Deep had temperatures above 15° C on June 8<sup>th</sup> with a well-established thermocline. Peak temperature was reached on July 21<sup>st</sup> with a reading of 25.45° C at the surface. Benewah Lake had exceeded the optimum temperature requirements by May 21<sup>st</sup> in the upper 2 meters. A thermocline was established at this time. Peak temperature was reached on July 21<sup>st</sup> with a value of 25.92° C at the surface.

Vertical profiles for the interior bay stations (Windy Bay Shallow, Rockford Bay, and Carry Bay) during 1997 showed that the thermocline had started to build in by May 16<sup>th</sup> and the water nearest to the bottom had warmed from 5.5° C to 7.0° C (Figure 3.3). By June 26<sup>th</sup> the thermocline had reached the bottom. The surface temperature peaked on August 5<sup>th</sup> at 23.5° C.

During 1998, all three interior bays sample sites had temperatures above 15° C (Figure 3.3) and a well-established thermocline by June 8<sup>th</sup>. Temperature in the interior bays peaked at 24.79° C on July 7<sup>th</sup>.

Temperature profiles for the 1997 deep-water stations (Conkling Point, Windy Bay Deep, Mid-Lake Coeur d'Alene) showed a definite thermocline by May 16<sup>th</sup> (Figure 3.4) with complete thermal stratification by August 5<sup>th</sup>. The thermocline was still present in the deepest sampling station on October 22<sup>nd</sup> however; it was nearly broken down in the other two deep-water stations. The temperature criterion was never exceeded from surface to bottom however, the temperature standard was exceeded and habitat was limited.

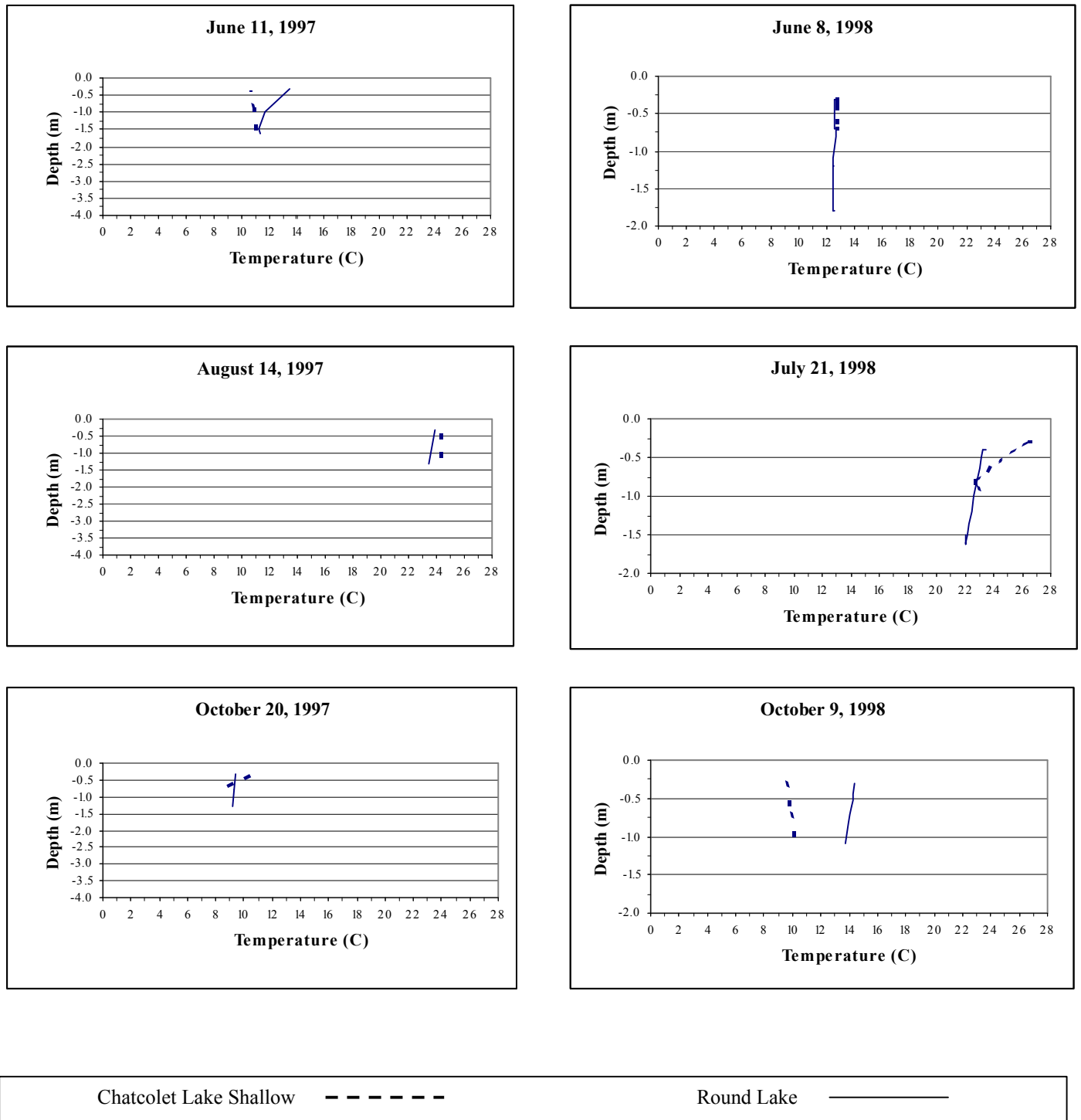


Figure 3.1 Peak spring, summer and fall temperature profiles vs. depth for two geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.

By June 8<sup>th</sup> 1998, the deep water zones had temperatures that violated the standards and had well-established thermoclines (Figure 3.4). The depth of the thermoclines varied by site, Conkling Point's extended down 1.5 meters while Mid-Lake Coeur d'Alene's consisted of the upper 3 meters and Windy Bay Deep had the most pronounced thermocline, which encompassed the upper 8.5 meters. Peak temperature was reached on July 7<sup>th</sup> with a reading of 24.51° C.

In 1997, Coeur d'Alene River's water temperature had exceeded 15° C by June 11<sup>th</sup> and a thermocline was forming (Figure 3.5). Water temperature peaked on August 5<sup>th</sup> (25.18° C) and the surface to bottom temperature was greater than 15° C. In 1998, water temperature in the Coeur d'Alene River had exceeded 15° C by June 8<sup>th</sup> (Figure 3.5). Maximum water temperature was reached on August 10<sup>th</sup> (24.59° C).

By July 8<sup>th</sup> 1997 the water temperature in the St. Joe River had exceeded 15° C (Figure 3.6). Water temperature peaked in the St. Joe River on August 5<sup>th</sup> with a reading of 23.01° C. In 1998, the St. Joe River exceeded 15° C by June 22<sup>nd</sup> in the upper 2 meters (Figure 3.6). Peak temperature was reached on July 21<sup>st</sup> (22.77° C).

In 1997, by the first of October all water quality sites were isothermal at 15° C. By mid October in 1998 water temperature at all thirteen stations was between the optimum temperature requirement of 10° C to 15° C.

### pH

Cutthroat and bull trout have optimum pH limits of 6-8. The pH of Coeur d'Alene Lake does not change very much from season to season or on an annual basis. Observed values ranged from 6.8 to 8.0 (Appendix A), all within the optimal tolerance limits for cutthroat production.

In 1997, the only variation occurred at three stations in the southern lakes section (Chatcolet Lake Deep, Benewah Lake, and Hidden Lake) where the pH rose to 9.4 (Figure 3.7) in the upper 2 meters of the water column. The three stations that showed similar changes in pH were in geomorphologically similar areas.

During 1998, eleven out of the thirteen sites had pH values that exceeded the upper limit. pH values increased on a north to south axis as did the duration. The interior bays and the deep-water zones had pH values that exceeded eight from June 24<sup>th</sup> through July 8<sup>th</sup>. The shallow southern lakes had pH values above eight from July 7<sup>th</sup> through September 3<sup>rd</sup>. The peak pH during this time was 9.17 on August 11<sup>th</sup> (Figure 3.7). On October 9<sup>th</sup> and 19<sup>th</sup> Round Lake exceed the upper pH limit with a reading of 8.63. Chatcolet Lake Shallow exceeded the upper pH limit on and off from July 7<sup>th</sup> through October 19<sup>th</sup>, with the maximum pH being 9.36 on October 19<sup>th</sup> (Figure 3.8).

### Conductivity

Vertical profiles of the specific conductance were taken concurrently with the other physical/chemical parameters during 1997 and 1998 (Appendix A). In 1997, all conductivities were within the tolerance limits for cutthroat trout production and ranged from 15.3 µg/L to 81.0 µg/L. Any variability to the conductivity was due to natural environmental conditions except at the mouth of the Coeur d'Alene River (Figure 3.9) where higher than normal levels of dissolved metals are flowing into the lake. Values remained very stable during the fall, winter, and spring months. Only during the peak summer conditions did variation occur.

In 1998, all thirteen water quality sites fell within the conductivity limits. The shallow southern lakes and Coeur d'Alene River had the highest values. Hidden Lake (129.1µs/cm), Chatcolet Lake Deep (92.6µs/cm), Benewah Lake (75.7µs/cm) and the Coeur d'Alene River (85µs/cm). The rest of the sites had values ranging from 30-65 µs/cm.

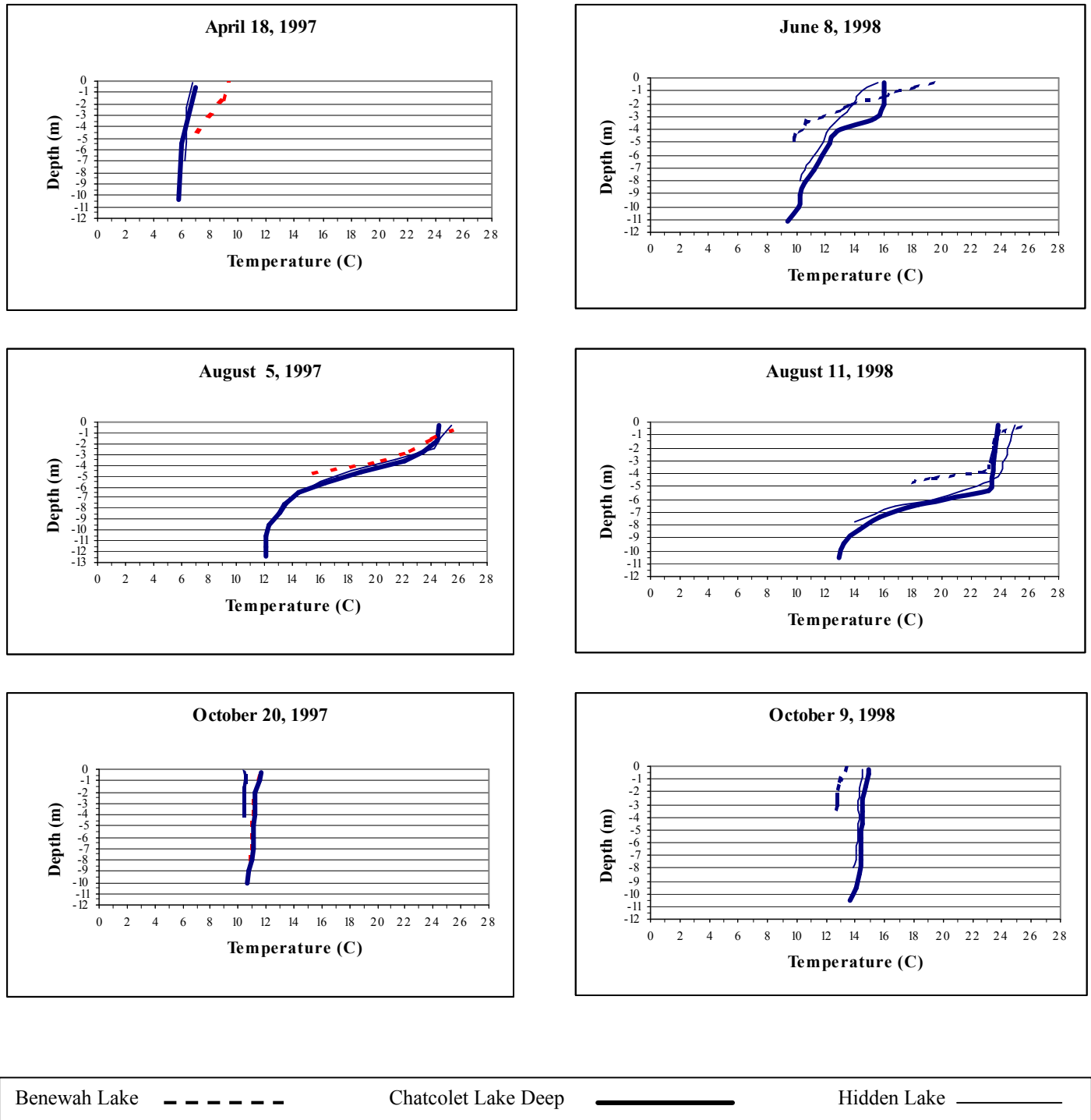


Figure 3.2 Peak spring, summer and fall temperature profiles vs depth for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.

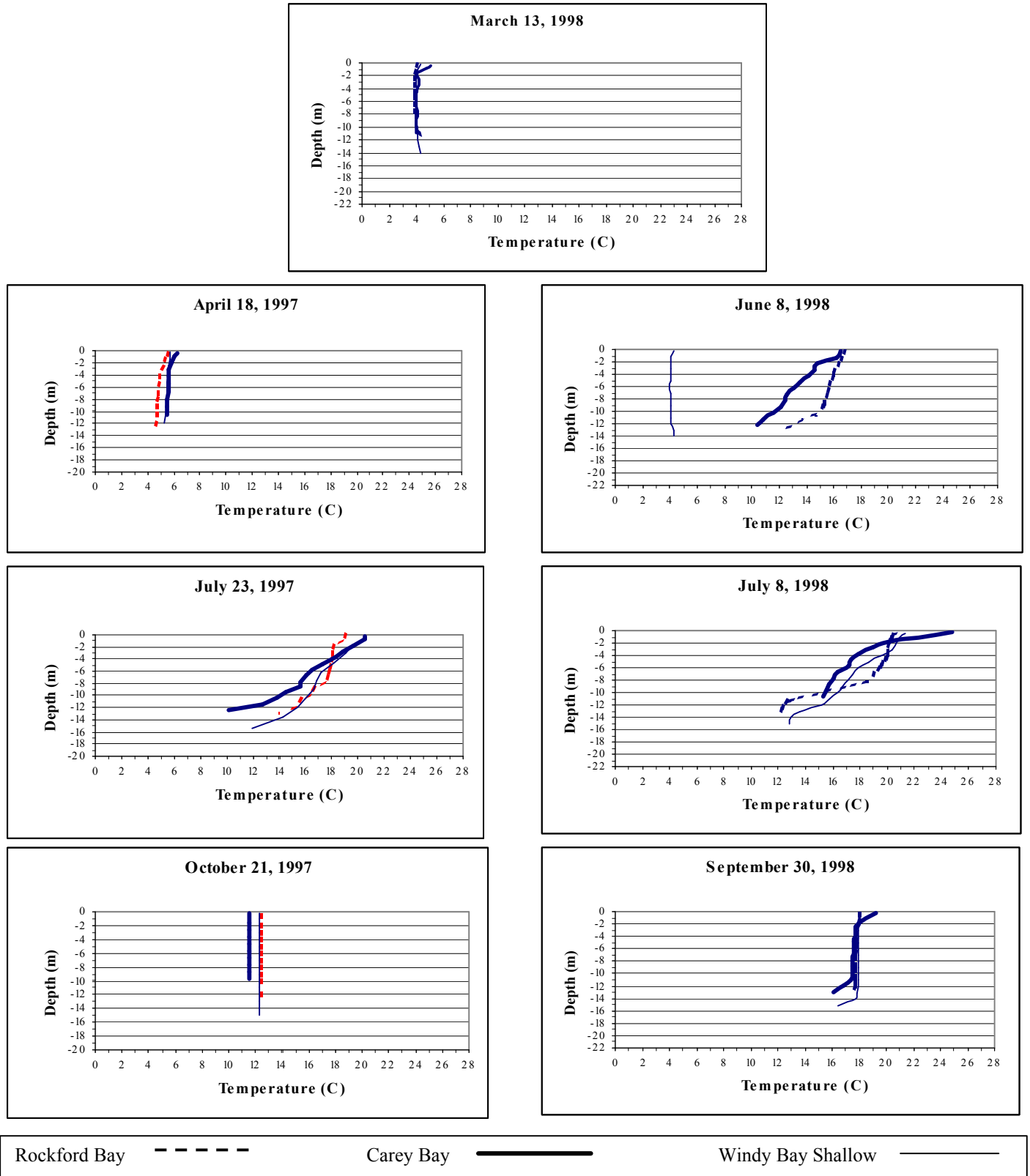


Figure 3.3 Peak winter, spring, summer and fall temperature profiles vs depth for three geomorphology similar stations on Coeur d'Alene Lake during 1997 and 1998. A winter sample was not taken in 1997.

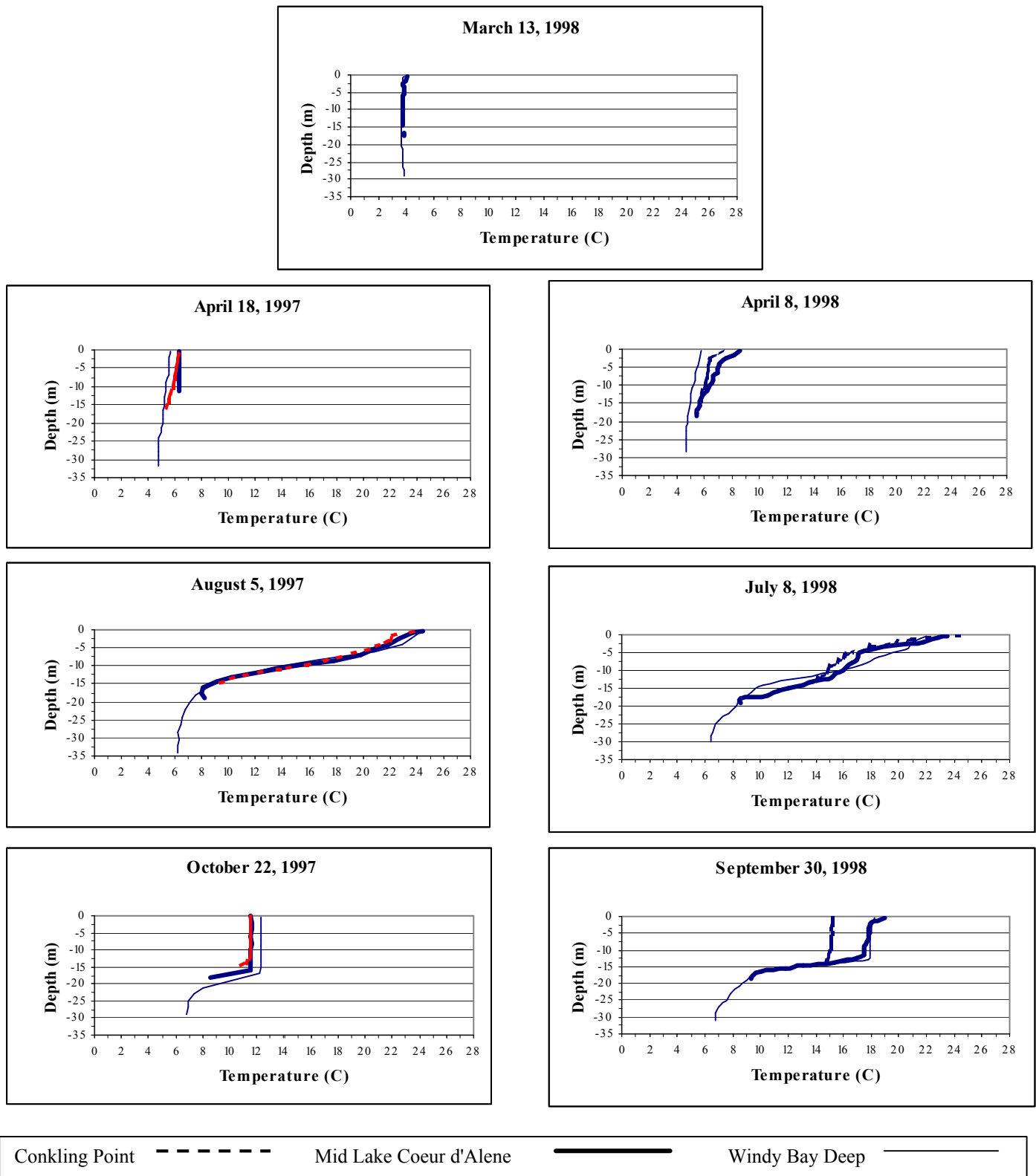


Figure 3.4 Peak winter, spring, summer and fall temperature profiles vs depth for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998. A winter sample was not taken in 1997 and Conkling Point had no winter sample taken in 1998.

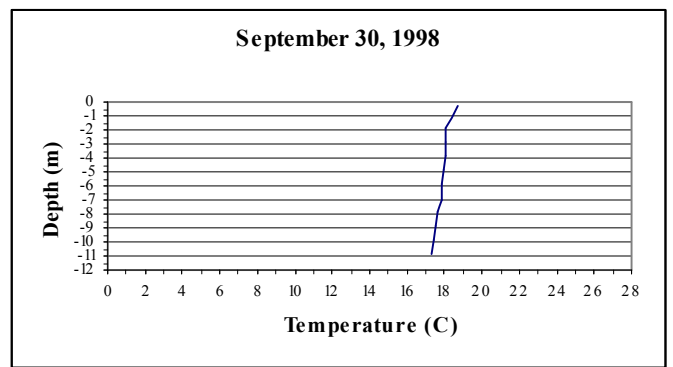
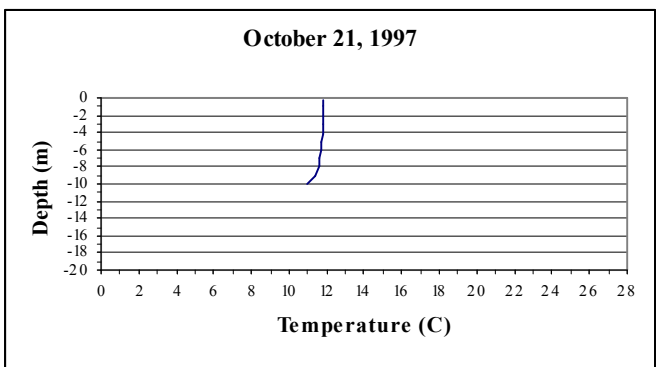
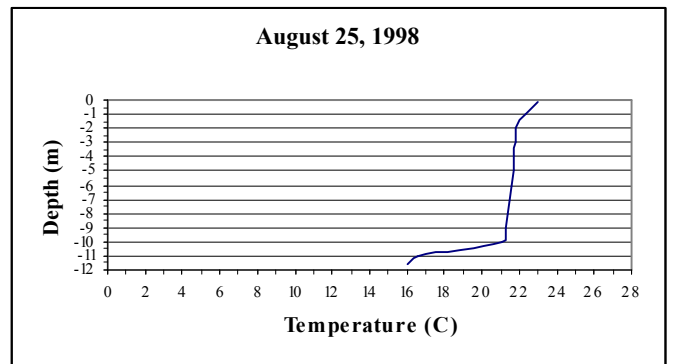
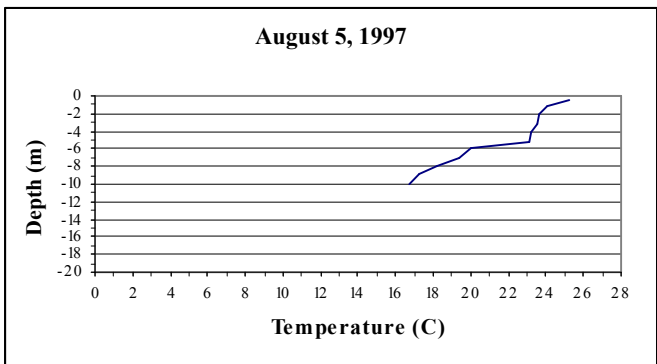
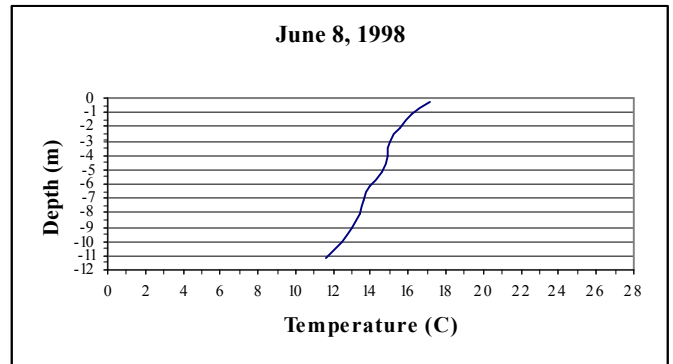
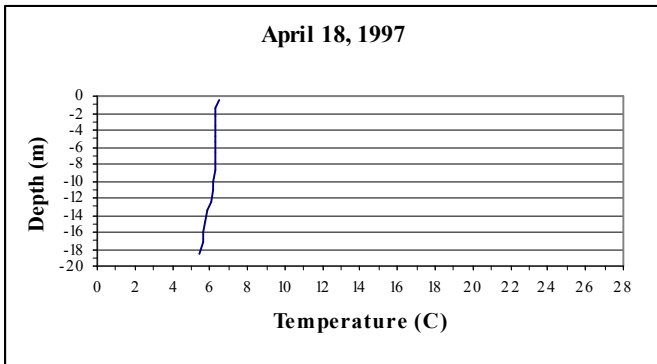
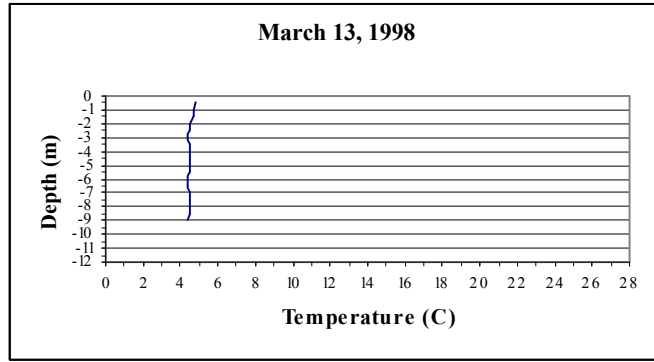


Figure 3.5 Peak winter, spring, summer and fall temperature profiles vs depth for the Coeur d'Alene River sampling location on Coeur d'Alene Lake during 1997 and 1998. There was no winter sample taken in 1997.



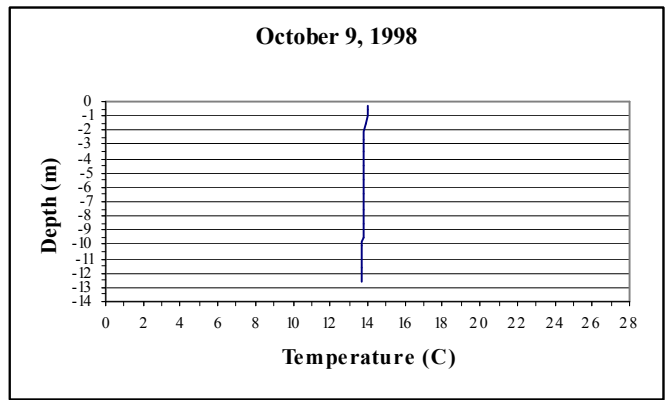
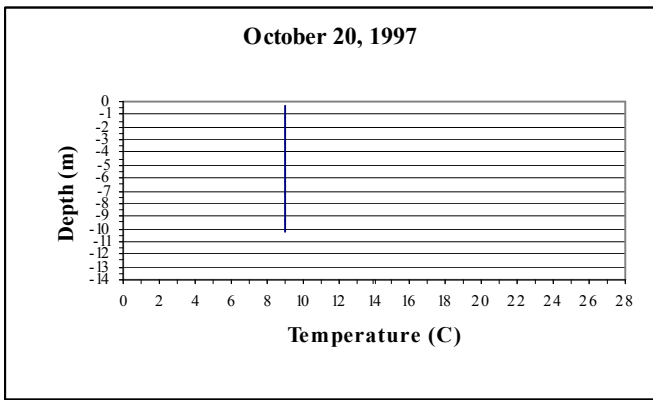
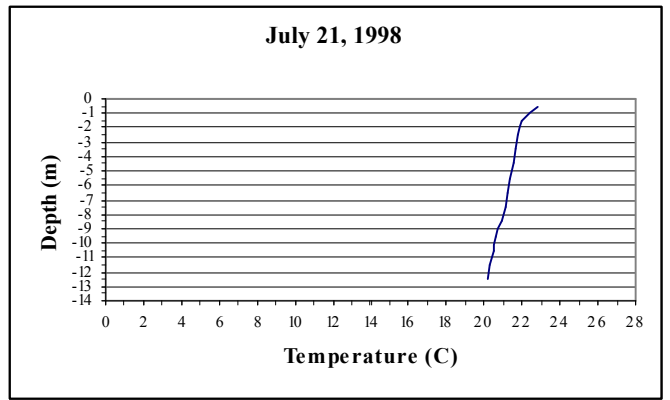
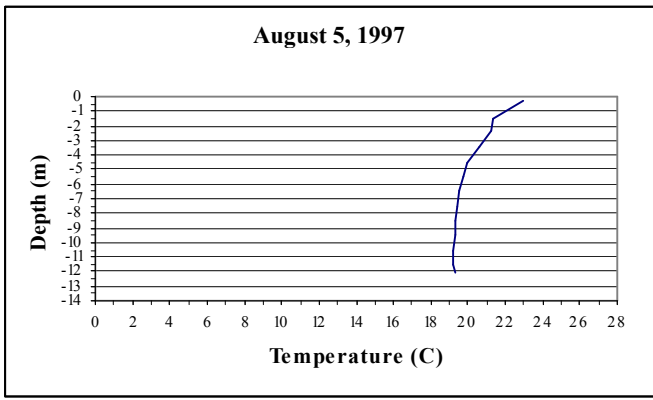
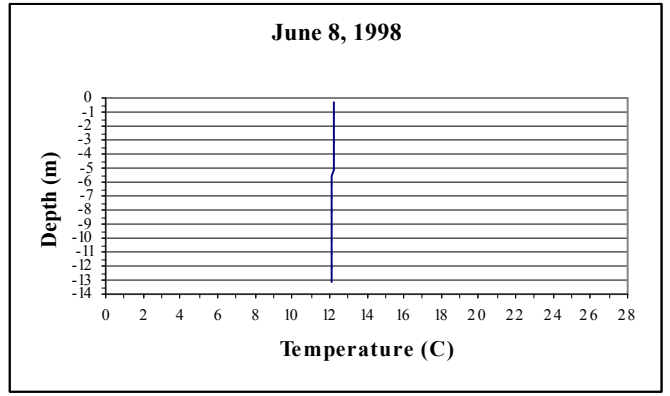
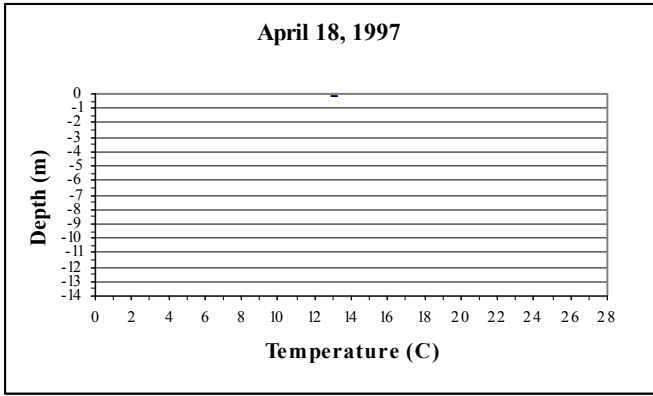


Figure 3.6 Peak spring, summer and fall temperature profiles vs depth for the St. Joe River sampling during 1997 and 1998.

### Nitrogen

Nitrogen compounds follow regular seasonal patterns. For the most part nitrate concentrations sampled at the 13 stations (Appendix B) in 1997 and 1998, were below the detection limit (0.01 µg/L) thus, all data below detection limits were reported as half the detection limit. In only a few instances during 1997 (both stations in Windy Bay, Rockford Bay, and the two shallow stations) did nitrate concentrations exceed the detection limit in the epilimnion, reaching a high of 0.113 µg/L at the Rockford Bay sample station in the October sample period. However, in the hypolimnion concentrations of nitrate generally increased in the November sample over the October sample period reaching a high of 0.131 µg/L in the Windy Bay Deep sample station.

In the epilimnion during the months of April and November in 1998 half of the sites were above the detection limit. From June through October all sites were below the detection limit. Hidden Lake (0.210 µg/L) had the highest reading. In the hypolimnion, four out of the seven sampling months had values above the detection limit. Mid-Lake Coeur d'Alene (0.078 µg/L) had the highest reading in the hypolimnion.

Nitrite, the form of nitrogen found in the smallest quantities were all below the detection limit of 0.01 µg/L in the epilimnion and hypolimnion for both samples (Appendix B) at all thirteen stations in 1997. All data measured below the detection limit were recorded as half the detection limit.

During the sample months of April and June in 1998, nitrite values were <0.029 µg/L in both the epilimnion and hypolimnion (Appendix B). The rest of the time the values were below the detection limit. All data measured below the detection limit were recorded as half the detection limit.

Organic nitrogen is defined functionally as organically bound nitrogen in the tri-negative oxidation state. Analytically, organic nitrogen and ammonia can be determined together and have been referred to as "kjeldahl nitrogen", a term that reflects the technique used in their determination. Total kjeldahl nitrogen was added as a new water quality sampling method during the 1998 field season. The values in both the epilimnion and hypolimnion were fairly uniform throughout the thirteen stations (Appendix B).

### Phosphate

During 1997 and 1998, ortho-phosphate concentrations were below the detection limit (0.026 µg/L) at all sites in both the epilimnion and hypolimnion (Appendix B). The data were reported as half of the detection limit so data point loss would not occur.

Total phosphorus was a new parameter that was added in 1998. Total phosphorus values increased on a north to south axis (Appendix B). All sites north of Hidden Lake in both the epilimnion and hypolimnion were below the detection limit of 0.005 µg/L. All data measured below the detection limit were recorded as half the detection limit. A few sites south of Hidden Lake had values slightly higher than the detection limit. Values in the epilimnion ranged from <0.005 µg/L in the northern sites to 0.065 µg/L at Chatcolet Lake Shallow. In the hypolimnion the values ranged from <0.005 µg/L in the northern sites to 0.026 µg/L at Benewah Lake.

### Primary Productivity

Chlorophyll<sub>a</sub> is the pigment aquatic plants use for photosynthesis. In 1997 Chlorophyll<sub>a</sub> values ranged from 0.005 µg/L to 25.790 µg/L in the epilimnion while values ranged from 0.005 µg/L to 34.14 µg/L in the hypolimnion (Appendix B). The general trend is increasing values on a north to south axis. Higher levels of chlorophyll<sub>a</sub> is an indicator of increasing trophic status.



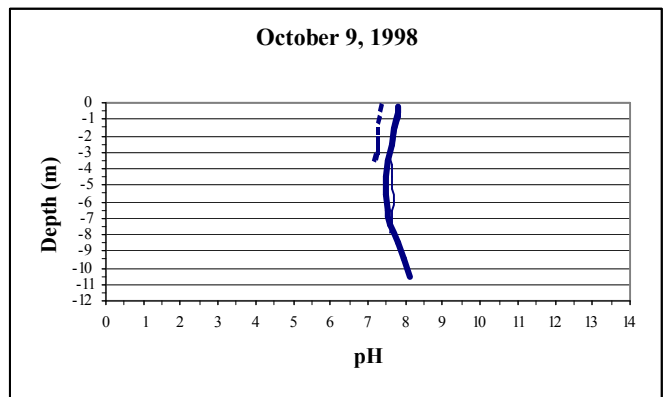
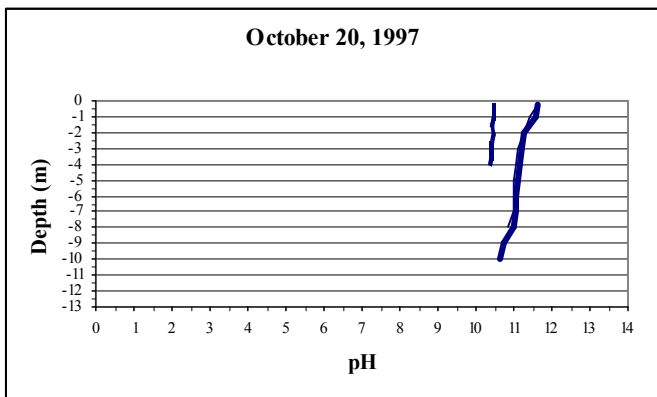
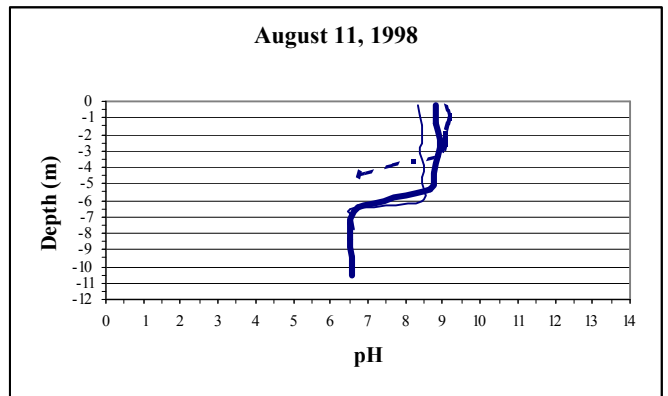
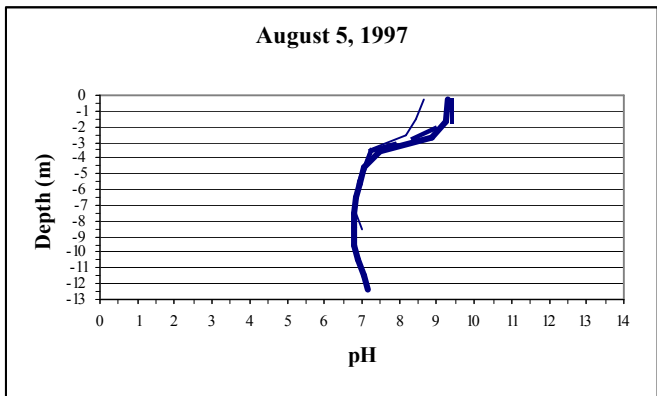
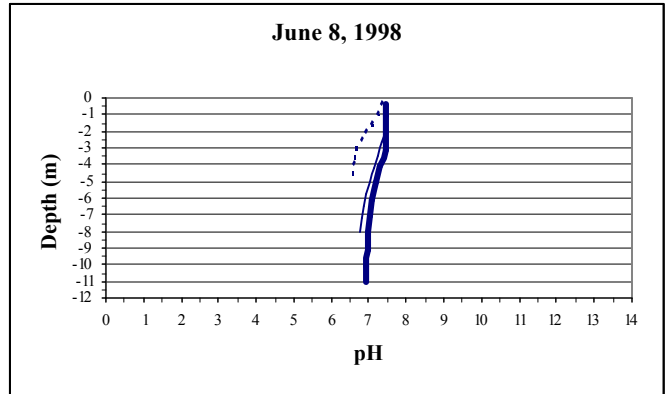
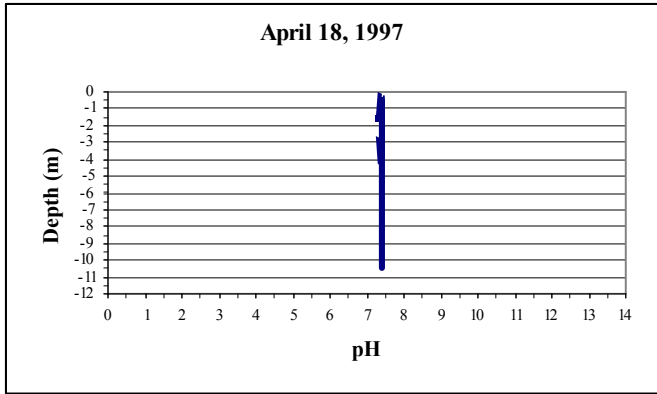


Figure 3.7 Peak spring, summer and fall pH profiles vs depth for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.

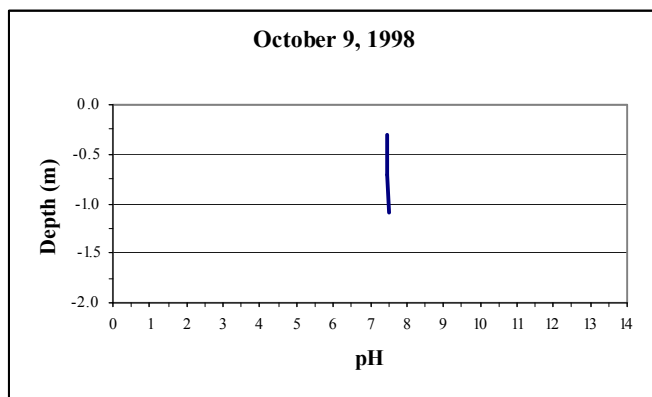
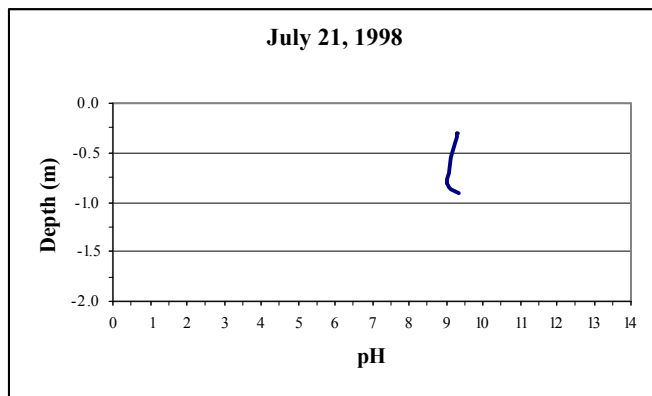
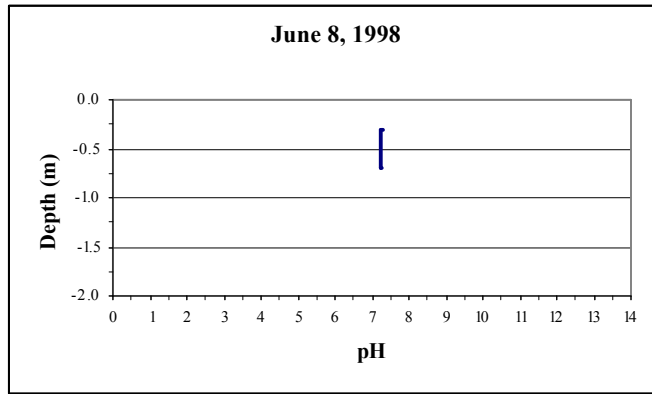


Figure 3.8 Peak spring, summer and fall pH profiles vs depth for the Chatcolet Lake Shallow sampling location on Coeur d'Alene Lake during 1998.

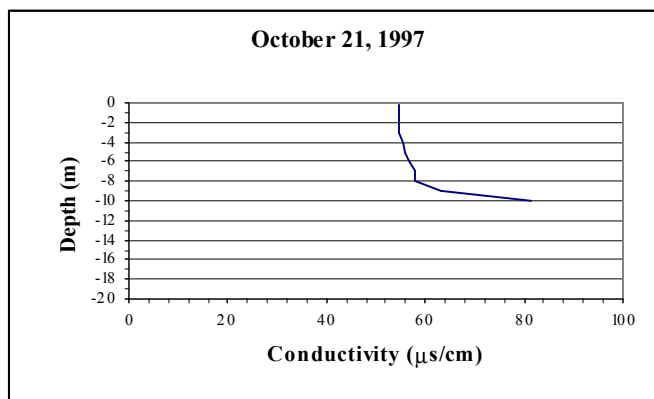
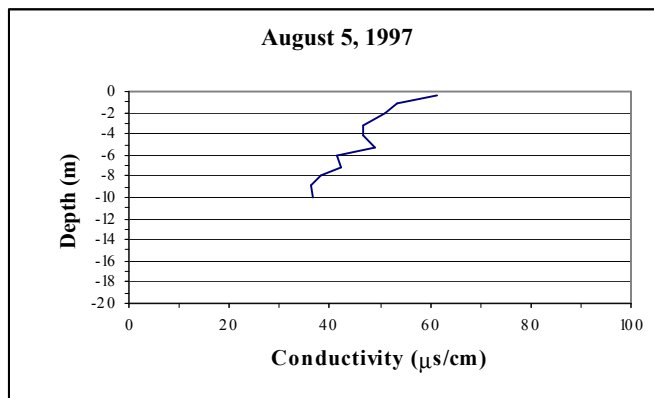
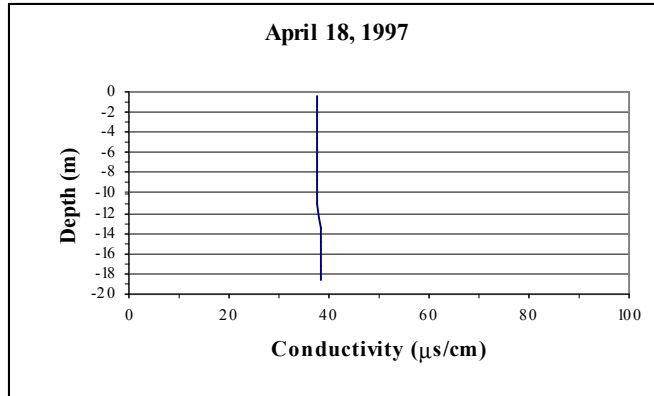


Figure 3.9 Peak spring, summer and fall conductivity profiles vs depth for the Coeur d'Alene River sampling location on Coeur d'Alene Lake during 1997.

In 1998, the same trend occurred as in 1997, Chlorophyll<sub>a</sub> values increased on a north to south axis. The values in the epilimnion ranged from <0.001 µg/L to 61.410 µg/L and the values in the hypolimnion ranged from <0.010 µg/L to 80.920 µg/L (Appendix B).

### Physical

Total Suspended Solids (TSS) is defined as the residue left on a filter paper of 2 µm or smaller after a portion of the sample has been filtered through (APHA, 1992). The detection limit for TSS is 4.0 mg/L. In 1997, total suspended solids were fairly uniform throughout the epilimnion of the lake with only a few differences found (Appendix B). The shallow stations ran slightly higher than the other stations with Round Lake reaching a high of 16.0 mg/L. The southern lakes also ran slightly higher with Hidden Lake and Benewah Lake reaching a high of 10 mg/L. The hypolimnion was quite variable throughout the lake. Windy Bay Deep reached a high of 27.0 mg/L. Drinking water standards are set at 500 mg/L however; levels much lower can impart a foul taste to the water. No analysis of the composition of the suspended solids was completed. However, it was noted that high levels of suspended sediment was present in the southern lakes sample stations and that most likely the high suspended solids present in the Windy Bay Deep station was related to decomposing algae not sediment.

During 1998, none of the water quality sites exceeded the drinking water standard but some sites did exceed the detection limit (Appendix B). Chatcolet Lake Shallow had a high reading of 10.500 mg/L in the epilimnion. TSS increased in the hypolimnion, with more sites having values over 4.0 mg/L. The southern lakes had some of the highest values. Chatcolet Lake Deep had a reading of 141.000 mg/L in June and Hidden Lake had a reading of 47.000 mg/L in April.

Turbidity is the amount of suspended inorganic and organic matter throughout the water column. In 1997, turbidity in the epilimnion and hypolimnion showed the same trend as suspended solids (Appendix B) with a general increase on a north to south axis. The highest turbidity reading was recorded at the Chatcolet Lake Shallow sample station (11.2 NTU). The next highest was Benewah Lake (5.810 NTU) followed by Hidden Lake (4.760 NTU). The lowest value was recorded at the Windy Bay Deep sample station (0.230 NTU). The high turbidity readings were due to suspended sediments flowing in from Plummer Creek a tributary to Chatcolet Lake.

In 1998, all of the turbidity readings in the epilimnion were above the detection limit (Appendix B). Benewah Lake (6.840 NTU) had some of the highest readings throughout the season. Windy Bay Deep (0.403 NTU) had the lowest turbidity reading. Turbidity values were higher in the hypolimnion than in the epilimnion. In June Chatcolet Lake Deep had a reading of 43.100 NTU. Windy Bay Deep (0.335 NTU) had the lowest turbidity value. As the field season continued there was a gradual decrease in turbidity values in the hypolimnion.

### Dissolved Metals

Little variation in the concentrations of the eleven metals (Appendix B) sampled was observed during 1997 and 1998. Potassium (K) values in the hypolimnion during October of 1998 were slightly higher than October of 1997. The 1998 values ranged from 1.02 mg/L to 1.4 mg/L compared to 1997, which had values that were <0.850 mg/L.

Secchi disk readings were taken at each of the thirteen stations throughout 1997 and 1998 they were used to determine the euphotic zone depth. Graphs showing the measured secchi disk readings and the empirically derived euphotic zone depth are located in (Appendix C). In 1997 and 1998 the empirically derived euphotic zone depth was variable throughout the lake (Table 3.2 & 3.3) with each station having different depths based on variable site-specific conditions. Similarities existed between the five distinct

habitat areas however; the general trend was decreasing secchi and euphotic zone depths on a north to south axis within the lake.



Table 3.2 Average annual and seasonal secchi measurements taken at thirteen stations on Coeur d'Alene Lake in 1997. Euphotic zone depths were empirically derived using the regression equation  $EZD=2.2302+1.4914(SD)R^2=.78$  published by Alaska Fish and Game 1987. All measurements are in meters.

| Location            | Station Depth | Average                    |                           | Spring                       |                        | Summer                       |                        | Fall                         |                        |
|---------------------|---------------|----------------------------|---------------------------|------------------------------|------------------------|------------------------------|------------------------|------------------------------|------------------------|
|                     |               | Annual Secchi <sup>a</sup> | Annual Euphotic Zone      | Seasonal Secchi <sup>b</sup> | Seasonal Euphotic Zone | Seasonal Secchi <sup>c</sup> | Seasonal Euphotic Zone | Seasonal Secchi <sup>d</sup> | Seasonal Euphotic Zone |
| Round Lake          | 1.50          | 1.27                       | Bottom (1.5) <sup>e</sup> | 1.20                         | Bottom (1.5)           | 1.31                         | Bottom (1.5)           | 1.23                         | Bottom (1.5)           |
| Chatcolet Shallow   | 1.50          | .89                        | Bottom (1.5)              | 1.00                         | Bottom (1.5)           | 0.87                         | Bottom (1.5)           | 0.77                         | Bottom (1.5)           |
| Rockford Bay        | 13.00         | 5.29                       | 10.11                     | 153                          | 4.50                   | 6.43                         | 11.81                  | 7.63                         | 13.61                  |
| Windy Bay Shallow   | 14.00         | 4.58                       | 9.05                      | 1.66                         | 4.71                   | 5.37                         | 10.24                  | 6.60                         | 12.07                  |
| Carey Bay           | 12.00         | 4.10                       | 8.34                      | 1.73                         | 4.80                   | 5.13                         | 9.87                   | 4.87                         | 9.48                   |
| Windy Bay Deep      | 33.00         | 5.44                       | 10.35                     | 1.75                         | 4.84                   | 6.77                         | 12.33                  | 7.27                         | 13.07                  |
| Mid Lake            | 18.00         | 4.08                       | 8.31                      | 1.74                         | 4.82                   | 5.00                         | 9.68                   | 5.03                         | 9.73                   |
| Conkling Point      | 16.00         | 3.44                       | 7.36                      | 1.43                         | 4.36                   | 4.16                         | 8.43                   | 4.47                         | 8.89                   |
| Hidden Lake         | 8.00          | 3.02                       | 6.81                      | 1.58                         | 4.65                   | 3.94                         | 8.18                   | 2.80                         | 6.48                   |
| Chatcolet Lake      | 11.50         | 2.56                       | 6.04                      | 1.65                         | 4.69                   | 3.01                         | 6.73                   | 2.70                         | 6.26                   |
| Benewah Lake        | 4.50          | 2.42                       | 5.84                      | 1.60                         | 4.62                   | 2.96                         | 6.64                   | 2.20                         | 5.51                   |
| Coeur d'Alene River | 10.00         | 3.92                       | 8.08                      | 1.50                         | 4.47                   | 4.73                         | 9.28                   | 5.27                         | 10.08                  |
| St. Joe River       | 12.50         | 2.25                       | 5.58                      | 1.08                         | 3.83                   | 3.18                         | 6.98                   | 1.93                         | 5.11                   |

<sup>a</sup> Average annual secchi was calculated from April 18, 1997 to November 4, 1997.

<sup>b</sup> Annual Spring secchi is from April 18, 1997 to June 11, 1997.

<sup>c</sup> Annual Summer secchi is from June 26, 1997 to September 17, 1997.

<sup>d</sup> Annual Fall secchi is from September 29, 1997 to November 4, 1997.

<sup>e</sup> Numbers in parenthesis represent the bottom in meters.

Table 3.3 Average annual and seasonal secchi measurements taken at thirteen stations on Coeur d'Alene Lake in 1998. Euphotic zone depths were empirically derived using the regression equation  $EZD=2.2302+1.4914(SD)R^2=.78$  published by Alaska Fish and Game 1987. All measurements are in meters.

| Location            | Station Depth | Average                    |                            | Spring                       |                            | Summer                       |                            | Fall                         |                            | Winter                       |                        |
|---------------------|---------------|----------------------------|----------------------------|------------------------------|----------------------------|------------------------------|----------------------------|------------------------------|----------------------------|------------------------------|------------------------|
|                     |               | Annual Secchi <sup>a</sup> | Annual Euphotic Zone       | Seasonal Secchi <sup>b</sup> | Seasonal Euphotic Zone     | Seasonal Secchi <sup>c</sup> | Seasonal Euphotic Zone     | Seasonal Secchi <sup>d</sup> | Seasonal Euphotic Zone     | Seasonal Secchi <sup>e</sup> | Seasonal Euphotic Zone |
| Round Lake          | 2.00          | 1.27                       | Bottom (2.00) <sup>f</sup> | 1.36                         | Bottom (2.00) <sup>f</sup> | 1.29                         | Bottom (2.00) <sup>f</sup> | 0.7                          | Bottom (2.00) <sup>f</sup> | -                            | -                      |
| Chatcolet Shallow   | 1.10          | 0.85                       | Bottom (1.10)              | 0.7                          | Bottom (1.10)              | 0.86                         | Bottom (1.10)              | 1                            | Bottom (1.10)              | -                            | -                      |
| Rockford Bay        | 13.00         | 6.06                       | 11.27                      | 3.88                         | 8.01                       | 6.87                         | 12.47                      | 8.40                         | 14.76                      | 3.00                         | 6.70                   |
| Windy Bay Shallow   | 15.00         | 5.47                       | 10.39                      | 3.43                         | 7.34                       | 6.22                         | 11.50                      | 7.87                         | 13.96                      | 2.00                         | 5.21                   |
| Carey Bay           | 12.00         | 5.07                       | 9.80                       | 3.08                         | 6.82                       | 6.43                         | 11.82                      | 5.67                         | 10.68                      | 1.80                         | 4.91                   |
| Windy Bay Deep      | 30.00         | 6.25                       | 11.55                      | 3.43                         | 7.34                       | 7.28                         | 13.09                      | 9.20                         | 15.95                      | 2.50                         | 5.96                   |
| Mid Lake            | 18.00         | 5.27                       | 10.09                      | 2.90                         | 6.56                       | 6.93                         | 12.57                      | 6.23                         | 11.53                      | 1.90                         | 5.06                   |
| Conkling Point      | 15.00         | 4.42                       | 8.82                       | 2.92                         | 6.59                       | 5.47                         | 10.39                      | 4.47                         | 8.89                       | -                            | -                      |
| Chatcolet Lake      | 10.00         | 3.37                       | 7.26                       | 2.60                         | 6.11                       | 4.50                         | 8.94                       | 2.57                         | 6.06                       | 1.00                         | 3.72                   |
| Hidden Lake         | 7.00          | 3.52                       | 7.48                       | 2.14                         | 5.42                       | 4.77                         | 9.35                       | 2.90                         | 6.56                       | -                            | -                      |
| Benewah Lake        | 4.50          | 2.51                       | 5.98                       | 1.53                         | 4.51                       | 3.29                         | 7.13                       | 2.00                         | 5.21                       | -                            | -                      |
| Coeur d'Alene River | 10.00         | 5.05                       | 9.76                       | 3.15                         | 6.93                       | 6.02                         | 11.20                      | 6.67                         | 12.17                      | 2.00                         | 5.21                   |
| St. Joe River       | 11.00         | 3.75                       | 7.83                       | 3.88                         | 8.02                       | 3.84                         | 7.95                       | 3.10                         | 6.85                       | -                            | -                      |

<sup>a</sup>Average annual secchi was calculated from March 13, 1998 through November 17, 1998.

<sup>b</sup>Annual Spring secchi is from March 22, 1998 to June 21, 1998.

<sup>c</sup>Annual Summer secchi is from June 22, 1998 to September 21, 1998.

<sup>d</sup>Annual Fall secchi is from September 22, 1998 to December 21, 1998.

<sup>e</sup>Annual Winter secchi is from December 22, 1998 to March 21, 1998.

<sup>f</sup>Numbers in parenthesis represent the bottom in meters.

### 3.1.2 Fisheries

Coeur d'Alene Lake has been sampled on a regular basis since 1994. In 1994, 1,418 fish were sampled by electrofishing and 211 fish were sampled using horizontal gillnets. During 1995, 1,727 fish were sampled by electrofishing and 78 fish were sampled using horizontal gillnets. In 1996, 536 fish were sampled by electroshocking and 286 fish were sampled using horizontal gillnets. Largescale suckers and yellow perch were the two most frequently sampled fish for all three years by electroshocking. Yellow perch, northern pikeminnow and largescale suckers were the most frequently sampled fish in the horizontal gillnets during all three years. Refer to Peters et al. (1999) for a detailed analysis of relative abundance data for all three years.

In 1994, 20 hours and 56 minutes were spent electroshocking and gillnets were set for 69 hours. During 1995, 24 hours and 53 minutes were spent electroshocking and gillnets were set for 79 hours. In 1996, 12 hours were spent electroshocking and gillnets were set for 170 hours and 50 minutes. Refer to Peters et al. (1999) for a detailed analysis of the CPUE data from 1994-1996.

During 1997 and 1998, Coeur d'Alene Lake was sampled a total of eleven times. 11,806 fish were sampled by electroshocking and 1,372 fish were sampled using horizontal and vertical gillnets. During both years largescale suckers were the most abundant fish sampled by electroshocking and yellow perch were the most abundant fish sampled in gillnets.

Table 3.4 Electroshocking relative abundance results from 1997 and 1998 for Coeur d'Alene Lake.

| 1997    |               | 1998    |              |
|---------|---------------|---------|--------------|
| Species | n=4599        | Species | n=7207       |
| LSS     | 28.55% (1313) | LSS     | 24.6% (1771) |
| YP      | 22.94% (1055) | LMB     | 16.4% (1179) |
| PSS     | 10.55% (485)  | YP      | 16% (1155)   |
| LMB     | 09.92% (456)  | BC      | 14.9% (1073) |
| SQW     | 09.35% (430)  | PSS     | 09.4% (676)  |
| BC      | 08.52% (392)  | SQW     | 08.1% (585)  |
| BBH     | 04.02% (185)  | BBH     | 04.6% (333)  |
| TCH     | 03.50% (161)  | TCH     | 04.1% (297)  |
| CTT     | 0.78% (36)    | CTT     | 0.6% (40)    |
| PIK     | 0.50% (23)    | PIK     | 0.5% (36)    |
| SCP     | 0.43% (20)    | CHN     | 0.3% (21)    |
| KOK     | 0.33% (15)    | SMB     | 0.2% (17)    |
| CHN     | 0.30% (14)    | MWF     | 0.1% (10)    |
| MWF     | 0.15% (7)     | SCP     | 0.1% (9)     |
| CCF     | 0.09% (4)     | CCF     | 0.1% (5)     |
| SMB     | 0.04% (2)     |         |              |
| RBT     | 0.02% (1)     |         |              |
| BLT     | 0             |         |              |

In 1997, 4,599 fish were sampled over a four-month period (July – October). It's known that there are twenty-one different species of fish in the lake. Seventeen of the twenty-one fish species were sampled by electroshocking (Table 3.4). Largescale suckers were the most frequently sampled fish. They made up more than ¼ of the catch at 28.55% (1313). Yellow perch were second at 22.94% (1055), followed by pumpkinseed at 10.55% (485), largemouth bass 9.92% (456), northern pikeminnow 9.35% (430), black crappie 8.52% (392), brown bullhead 4.02% (185), tench 3.50% (161), cutthroat trout 0.78% (36), northern pike 0.50% (23), sculpin 0.43% (20), kokanee trout 0.33% (15), chinook salmon 0.30% (14), mountain whitefish 0.15% (7), channel catfish 0.09% (4), smallmouth bass 0.04% (2), and rainbow trout 0.02% (1).

In 1997, exotic species made up 60.73% of the electrofishing catch while native species made up the rest. Non-game native species made up 38.33% of the catch while native game fish made up only 0.93% of the catch.

During 1997, 42 hours and 27 minutes were spent electroshocking (Table 3.5). Largescale suckers had the highest CPUE (fish/hr) at 30.93. Yellow perch were next at 24.85, pumpkinseed 11.43, largemouth bass 10.74, northern pikeminnow 10.13, black crappie 9.23, brown bullhead 4.36, tench 3.79, cutthroat trout 0.85, northern pike 0.54, sculpin 0.47, kokanee salmon 0.35, chinook salmon 0.33, mountain whitefish 0.16, channel catfish 0.09, smallmouth bass 0.05, and rainbow trout 0.02.

In 1997, horizontal gillnets were used at all of the sites and a total of 636 fish were sampled (Table 3.6). Thirteen out of the twenty-one species were sampled. The species most frequently sampled was yellow perch at 34.28% (218). Largescale suckers were next at 21.23% (135), followed by northern pikeminnow 18.55% (118), kokanee salmon 10.06% (64), brown bullhead 7.39% (47), black crappie 2.52% (16), cutthroat trout 1.42% (9), tench 1.26% (8), pumpkinseed 0.94% (6), chinook salmon 0.79% (5), channel catfish 0.63% (4), and mountain whitefish 0.16% (1).

In 1997, exotic species made up 58.66% of the gillnet catch while native species made up the rest. Non-game native species made up 39.78% of the catch while native game fish made up 1.58% of the catch.

Horizontal gillnets were set for 677 hours in 1997 (Table 3.7). Yellow perch had the highest CPUE at 0.32 followed by largescale suckers 0.20. Next were northern pikeminnow 0.17, kokanee salmon 0.09, brown bullhead 0.07, black crappie 0.02, cutthroat trout 0.01, tench 0.01, pumpkinseed 0.01, chinook salmon 0.01, northern pike 0.01, channel catfish 0.01 and mountain whitefish 0.001.

Coeur d'Alene Lake was sampled seven times (April-October) during 1998. 7,207 fish were sampled during 1998 (Table 3.4) and fifteen out of the twenty-one species were sampled. The most frequently sampled fish was the largescale sucker 24.6% (1771). Largemouth bass typically not as abundant was the second most frequently taken fish 16.4% (1179). Next were yellow perch 16.0% (1155), black crappie 14.9% (1073), pumpkinseed 9.4% (676), northern pikeminnow 8.1% (585), brown bullhead 4.6% (333), tench 4.1% (297), cutthroat trout 0.6% (40), northern pike 0.5% (36), chinook salmon 0.3% (21), smallmouth bass 0.2% (17), mountain whitefish 0.1% (10), sculpin 0.1% (9), and channel catfish 0.1% (5).

In 1998, exotic species made up 66.5% of the electrofishing catch while native species made up the rest. Non-game native species made up 32.80% of the catch while native game fish made up 0.7% of the catch.

Table 35 Electroshocking catch per unit effort during 1997.

| 1997    |        |                          |
|---------|--------|--------------------------|
| Species | n=4599 | Hr./min.<br><b>42.27</b> |
| LSS     | 1313   | 30.93                    |
| YP      | 1055   | 24.85                    |
| PSS     | 485    | 11.43                    |
| LMB     | 456    | 10.74                    |
| SQW     | 430    | 10.13                    |
| BC      | 392    | 9.23                     |
| BBH     | 185    | 4.36                     |
| TCH     | 161    | 3.79                     |
| CTT     | 36     | 0.85                     |
| PIK     | 23     | 0.54                     |
| SCP     | 20     | 0.47                     |
| KOK     | 15     | 0.35                     |
| CHN     | 14     | 0.33                     |
| MWF     | 7      | 0.16                     |
| CCF     | 4      | 0.09                     |
| SMB     | 2      | 0.05                     |
| RBT     | 1      | 0.02                     |
| BLT     | 0      | 0.00                     |

Table 3.6 Gillnetting relative abundance results from 1997 and 1998 for Coeur d'Alene Lake.

| 1997    |              | 1998    |             |
|---------|--------------|---------|-------------|
| Species | n=636        | Species | n=636       |
| YP      | 34.28% (218) | YP      | 52.6% (335) |
| LSS     | 21.23% (135) | SQW     | 20.4% (130) |
| SQW     | 18.55% (118) | LSS     | 09.3% (59)  |
| KOK     | 10.06% (64)  | BBH     | 05.7% (36)  |
| BBH     | 07.39% (47)  | KOK     | 03.8% (24)  |
| BC      | 02.52% (16)  | BC      | 02.4% (15)  |
| CTT     | 01.42% (9)   | PIK     | 02.4% (15)  |
| TCH     | 01.26% (8)   | CTT     | 01.4% (9)   |
| PSS     | 0.94% (6)    | CHN     | 01.3% (8)   |
| CHN     | 0.79% (5)    | TCH     | 0.5% (3)    |
| PIK     | 0.79% (5)    | RBT     | 0.2% (1)    |
| CCF     | 0.63% (4)    | LMB     | 0.2% (1)    |
| MWF     | 0.16% (1)    |         |             |

In 1998, 64 hours and 36 minutes were spent electrofishing (Table 3.8). Largescale suckers had the highest CPUE at 27.52 fish/hour. Largemouth bass were second at 18.32 followed by yellow perch 17.95, black crappie 16.67, pumpkinseed 10.50, northern pikeminnow 9.09, brown bullhead 5.17, tench 4.61, cutthroat trout 0.62, northern pike 0.56, chinook salmon 0.33, smallmouth bass 0.26, mountain whitefish 0.16, sculpin 0.14, and channel catfish 0.08.

During 1998, vertical gillnets were used to sample deep open water and horizontal nets were used in shallow lakes and bays. In 1998, 636 fish were sampled using the vertical and horizontal gillnets (Table 3.5). Yellow perch were the most abundant fish species making up more than half of the catch 52.6% (335). Northern pikeminnow were the second most frequently sampled fish at 20.4% (130), followed by largescale suckers 9.3% (59), brown bullhead 5.7% (36), kokanee salmon 3.8% (24), black crappie 2.4% (15), northern pike 2.4% (15), cutthroat trout 1.4% (9), chinook salmon 1.39% (8), tench 0.5% (3), rainbow trout 0.2% (1), and largemouth bass 0.2% (1).

Table 3.7 Gillnetting catch per unit effort during 1997.

| 1997    |         |                     |
|---------|---------|---------------------|
| Species | (n=636) | Hours<br><b>677</b> |
| YP      | 218     | 0.32                |
| LSS     | 135     | 0.20                |
| SQW     | 118     | 0.17                |
| KOK     | 64      | 0.09                |
| BBH     | 47      | 0.07                |
| BC      | 16      | 0.02                |
| CTT     | 9       | 0.01                |
| TCH     | 8       | 0.01                |
| PSS     | 6       | 0.01                |
| CHN     | 5       | 0.01                |
| PIK     | 5       | 0.01                |
| CCF     | 4       | 0.01                |
| MWF     | 1       | 0.00                |
| BLT     | 0       | 0.00                |
| LMB     | 0       | 0.00                |
| RBT     | 0       | 0.00                |
| SCP     | 0       | 0.00                |
| SMB     | 0       | 0.00                |

Table 3.8 Electroshocking catch per unit effort during 1998.

| 1998    |        |                          |
|---------|--------|--------------------------|
| Species | N=7202 | Hr./min.<br><b>64.36</b> |
| LSS     | 1771   | 27.52                    |
| LMB     | 1179   | 18.32                    |
| YP      | 1155   | 17.95                    |
| BC      | 1073   | 16.67                    |
| PSS     | 676    | 10.50                    |
| SQW     | 585    | 9.09                     |
| BBH     | 333    | 5.17                     |
| TCH     | 297    | 4.61                     |
| CTT     | 40     | 0.62                     |
| PIK     | 36     | 0.56                     |
| CHN     | 21     | 0.33                     |
| SMB     | 17     | 0.26                     |
| MWF     | 10     | 0.16                     |
| SCP     | 9      | 0.14                     |
| CCF     | 5      | 0.08                     |

In 1998, exotic species made up 69.1% of the gillnet catch while native species made up the rest. Non-game native species made up 29.7% of the catch while native game fish made up 1.4% of the catch.

In 1998, 968 hours were spent having vertical and horizontal gillnets in the water (Table 3.9). Yellow perch had the highest CPUE at 0.346. Northern pikeminnow was second with 0.134, followed by largescale suckers 0.061, brown bullhead 0.037, kokanee salmon 0.025, black crappie 0.015, northern pike 0.015, cutthroat trout 0.009, chinook salmon 0.008, tench 0.003, rainbow trout 0.001, and largemouth bass 0.001.

Age Analysis

Total number of fish scaled in was 1,499 in 1997; and 2,582 in 1998. Graphs showing the age distribution, length and weight frequency distribution and length and weight regression analysis are in Appendix D. In order to perform a regression analysis on aged fish we needed at least four different age classes. Chinook salmon, mountain whitefish, and smallmouth bass were lacking the four-age class requirement, so the data was combined from 1997 and 1998 for these three species. There was no regression analysis done on rainbow trout because only two fish have been sampled. Channel catfish were sampled in 1997 but they were not aged. In 1998, Dr. Scarnecchia, professor at University of Idaho aged the five channel catfish spines.

When comparing the mean length and weight, standard deviation and range for aged fish during the two sample years, the results vary slightly (Appendix E). Roughly the same number of yellow perch were sampled in 1997 and 1998. The statistical results were almost identical for each age group. On the other hand the statistical analysis for kokanee salmon showed an increase in mean length and weight from 1997 to 1998.

Habitat Suitability Index

When comparing the 1997 and 1998 (Table 3.10 & 3.11) habitat suitability indexes, Carey Bay and Conkling Point provided the same cutthroat habitat both years and other sites saw a dramatic decrease in habitat suitability. Rockford Bay and Windy Bay Shallow saw a decrease in the depth range of 7-11 meters. Rockford Bay went from 0.845 SI to 0.0 SI and Windy Bay Shallow from 0.94 SI to 0.0 SI. Hidden Lake and Chatcolet Lake Deep both saw an improvement in their bottom depth zone. Hidden Lake went from 0.0 SI to 0.67 SI and Chatcolet Lake Deep from 0.0 SI to 0.88 SI.

Table 3.9 Gillnetting catch per unit effort during 1998.

| Species | 1998  |                     |
|---------|-------|---------------------|
|         | n=636 | Hours<br><b>968</b> |
| YP      | 335   | 0.346               |
| SQW     | 130   | 0.134               |
| LSS     | 59    | 0.061               |
| BBH     | 36    | 0.037               |
| KOK     | 24    | 0.025               |
| BC      | 15    | 0.015               |
| PIK     | 15    | 0.015               |
| CTT     | 9     | 0.009               |
| CHN     | 8     | 0.008               |
| TCH     | 3     | 0.003               |
| RBT     | 1     | 0.001               |
| LMB     | 1     | 0.001               |

Except for only a few areas it appears that the upper ten meters of Coeur d'Alene Lake is unsuitable for cutthroat trout (Table 3.10 & 3.11). This is based primarily on summertime daily maximum temperatures exceeding what is considered optimum for cutthroat trout survival. HSI's calculated in 1995 and 1996 for the same sample areas showed similar results (Peters et. al. 1999). In areas where temperatures are within

the tolerance limits for cutthroat trout (primarily the hypolimnion) dissolve oxygen content becomes a factor. When temperatures are reaching their maximum dissolved oxygen in the hypolimnion drops below what is considered optimum for cutthroat trout. Thus, cutthroat trout are unable to seek refuge from the high water temperatures in the lower depths of the lake due to suboptimal dissolved oxygen concentrations.

## **3.2 Stream Studies**

### **3.2.1 Water Quality**

#### **Lake Creek**

Water temperature measured at the lower Lake Creek station indicated an instantaneous maximum of 25.5° C on July 25<sup>th</sup>, 1998. The 7-day moving average at the same station exceeded 16° C from July 4<sup>th</sup> through September 10<sup>th</sup>. Daily temperature fluctuations ranged from 6.2° C to 8.6° C during this period, indicating that substantial over-night cooling did take place. Temperatures recorded at the upper Lake Creek station were considerably lower. The 7-day moving average exceeded 16° C from July 23<sup>rd</sup> to August 6<sup>th</sup>, with an instantaneous maximum temperature of 19° C. Gaps in data prior to July 23<sup>rd</sup>, however, may have missed several days of high temperatures. Measured levels of dissolved oxygen generally did not drop below 9 mg/L at the lower station, where the highest temperatures were recorded.

#### **Benewah Creek**

A maximum temperature of 26° C was recorded at the mouth of Benewah Creek on July 25<sup>th</sup> and 26<sup>th</sup>, 1998. The 7-day moving average exceeded 16° C from June 29<sup>th</sup> through September 6<sup>th</sup>, and minimum temperature did not fall below 16° C between July 8<sup>th</sup> and September 2<sup>nd</sup>. Dissolved oxygen did not fall below 9 mg/l during the period of highest water temperatures at this same site. Discharge at the 3 mile station ranged from 5.2 to 1.1 cfs from July through September.

Tributary habitats generally provided more suitable water temperatures than mainstem reaches however diminished water quality was frequently noted. In S.F. Benewah Creek, maximum instantaneous water temperature was 21° C and 7-day moving average exceeded 16° C for just 12 days. Measured levels of dissolved oxygen did not drop below 9 mg/l during the period of highest temperatures. Discharge, however, was less than 1.0 cfs from mid July through early November. School House Creek was also a source of cold water and the 7-day moving average exceeded 16° C for just 13 days. There were periods, however, where habitat consisted of stagnant pools from July through August and dissolved oxygen decreased from 5.8 to 2.3 mg/l during this time. The average daily temperature in Windfall Creek exceeded 16° C from July through mid August. A maximum temperature of 27.5° C was recorded on July 27. There was little flow in Windfall Creek from late July through August and dissolved oxygen dropped below 9 mg/l during this period. In Whitetail Creek, the average daily water temperature exceeded 16° C from mid-July through mid-August and stream flow decreased to zero.

#### **Evans Creek**

A maximum instantaneous water temperature of 18.5° C was recorded several times between July 26<sup>th</sup> and August 5<sup>th</sup>, 1998. The 7-day moving average exceeded 16° C for 12 days in late July and early August. Average daily water temperature may have exceeded 16° C in mid July as well, but temperature monitors were not deployed until July 23<sup>rd</sup>. Measured levels of dissolved oxygen did not drop below 9 mg/l during the period of highest water temperatures. Stream flow ranged from 4.5 to 2.7 cfs from mid-August through early October.



**Alder Creek**

The maximum instantaneous water temperature was 24° C on July 27<sup>th</sup> and 28<sup>th</sup>, 1998. The 7-day moving average exceeded 16° C from July through mid August. Measured levels of dissolved oxygen did not drop below 9 mg/l during the period of highest water temperatures and base flows ranged from 2.3 to 0.4 cfs during the same period.

Table 3.10 Habitat suitability index for lacustrine Cutthroat Trout based on water quality for 1997.

| Location               | Depth                       | HSI <sup>a</sup>                 | Suitability Index |
|------------------------|-----------------------------|----------------------------------|-------------------|
| Rockford Bay           | 0-7 Meters                  | $(0.25 \times 1 \times 1)^{1/3}$ | = 0.25 SI         |
|                        | 7-11 Meters                 | $(0.60 \times 1 \times 1)^{1/3}$ | = 0.845 SI        |
|                        | 11-Bottom (14) <sup>b</sup> | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Windy Bay Shallow      | 0-7 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 7-10 Meters                 | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 10-Bottom (15)              | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Windy Bay Deep         | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 10-15 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 15-Bottom (33)              | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Coeur d'Alene River    | 0-Bottom                    | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
| Mid-Lake Coeur d'Alene | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 10-13 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 13-Bottom (17)              | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Carey Bay              | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 10-12 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 12-Bottom (13)              | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Conkling Point         | 0-10 Meters                 | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 10-13 Meters                | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 13-Bottom (16)              | $(1 \times 1 \times 1)^{1/3}$    | = 1.0 SI          |
| Hidden Lake            | 0-5 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 5-7 Meters                  | $(0.8 \times 1 \times 1)^{1/3}$  | = 0.92 SI         |
|                        | 7-Bottom (10)               | $(1 \times 0.0 \times 1)^{1/3}$  | = 0.0 SI          |
| Round Lake             | 0-Bottom (1.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
| Chatcolet Lake Deep    | 0-6 Meters                  | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
|                        | 6-9 Meters                  | $(0.85 \times 1 \times 1)^{1/3}$ | = 0.94 SI         |
|                        | 9-Bottom (11)               | $(1 \times 0.0 \times 1)^{1/3}$  | = 0.0 SI          |
| Chatcolet Lake Shallow | 0-Bottom (1.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
| Benewah Lake           | 0-Bottom (4.5)              | $(0.0 \times 1 \times 1)^{1/3}$  | = 0.0 SI          |
| St. Joe River          | 0-Bottom (12.5)             | $(0.4 \times 1 \times 1)^{1/3}$  | = 0.4 SI          |

<sup>a</sup> Habitat Suitability Index (HSI).

<sup>b</sup> Numbers in parenthesis represent the bottom in meters.

Table 3.11 Habitat suitability index for lacustrine cutthroat trout based on water quality for 1998.

| Location               | Depth                       | HSI <sup>a</sup>                  | Suitability Index |
|------------------------|-----------------------------|-----------------------------------|-------------------|
| Rockford Bay           | 0-7 meters                  | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 7-11 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 11-bottom (14) <sup>b</sup> | $(.7 \times 1 \times 1)^{1/3}$    | = 0.89 SI         |
| Windy Bay Shallow      | 0-7 meters                  | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 7-10 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 10-bottom (16)              | $(.45 \times 1 \times 1)^{1/3}$   | = 0.77 SI         |
| Windy Bay Deep         | 0-10 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 10-15 meters                | $(.46 \times 1 \times 1)^{1/3}$   | = 0.77 SI         |
|                        | 15-bottom (32)              | $(1 \times 1 \times 1)^{1/3}$     | = 1.0 SI          |
| Coeur d'Alene River    | 0-bottom (12)               | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
| Mid-Lake Coeur d'Alene | 0-10 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 10-13 meters                | $(.95 \times 1 \times 1)^{1/3}$   | = 0.98 SI         |
|                        | 13-bottom (19)              | $(1 \times .85 \times 1)^{1/3}$   | = 0.95 SI         |
| Carey Bay              | 0-10 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 10-12 meters                | $(.84 \times 1 \times 1)^{1/3}$   | = 0.94 SI         |
|                        | 12-bottom (14)              | $(1 \times 1 \times .98)^{1/3}$   | = 0.99 SI         |
| Conkling Point         | 0-10 meters                 | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 10-13 meters                | $(.75 \times 1 \times 1)^{1/3}$   | = 0.91 SI         |
|                        | 13-bottom (16)              | $(1 \times .86 \times 1)^{1/3}$   | = 0.95 SI         |
| Hidden Lake            | 0-5 meters                  | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 5-7 meters                  | $(.24 \times .8 \times 1)^{1/3}$  | = 0.58 SI         |
|                        | 7-bottom (9)                | $(1 \times .30 \times 1)^{1/3}$   | = 0.67 SI         |
| Round Lake             | 0-bottom (2)                | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
| Chatcolet Lake Deep    | 0-6 meters                  | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
|                        | 6-9 meters                  | $(.65 \times .79 \times 1)^{1/3}$ | = 0.80 SI         |
|                        | 9-bottom (11)               | $(1 \times .68 \times 1)^{1/3}$   | = 0.88 SI         |
| Chatcolet Lake Shallow | 0-bottom (1.1)              | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |
| Benewah Lake           | 0-bottom (5)                | $(0 \times 0 \times 1)^{1/3}$     | = 0.0 SI          |
| St. Joe River          | 0-bottom (13)               | $(0 \times 1 \times 1)^{1/3}$     | = 0.0 SI          |

<sup>a</sup> Habitat Suitability Index (HSI).

<sup>b</sup> Numbers in parenthesis represent the bottom in meters.

### 3.2.2. Spawning Gravel Survey and Analysis

Cutthroat trout primarily reside and spawn in reaches of small (1-4 meters wide) tributaries with moderate gradients (1.0-4.4%) and suitable spawning gravels (Table 3.12). Substrate embeddedness was high at the sampled sites, averaging about 50 percent. Proportion of potential spawning gravel was low and did not vary much among sites (mean = 4.1±2.1). We found no association between abundance of suitable spawning gravels and reach gradient, proportion of riffle habitat, proportion of pea gravel, or proportion of gravel substrate.

Table 3.12 Means of habitat features in spawning reaches of Alder (A), Benewah (B), Evans (E), and Lake (L) creeks and their major tributaries.

| Reach          | Gradient (%) | Width (ft.) | Depth (ft.) | Embeddedness <sup>a</sup> | Percent by Area |        |      |            |        |        |
|----------------|--------------|-------------|-------------|---------------------------|-----------------|--------|------|------------|--------|--------|
|                |              |             |             |                           | Spawning gravel | Riffle | Silt | Pea gravel | Gravel | Rubble |
| <b>Alder</b>   |              |             |             |                           |                 |        |      |            |        |        |
| A9             | 2.3          | 6.3         | 1.1         | 2.4                       | 5.5             | 30     | 10   | 65         | 25     | 0      |
| North Fork     | 2.3          | 11.6        | 1.5         | 2.6                       | 2.8             | 9      | 10   | 51         | 33     | 6      |
| <b>Benewah</b> |              |             |             |                           |                 |        |      |            |        |        |
| B12            | 1.0          | 16.0        | 2.4         | 3.4                       | 1.5             | 14     | 8    | 45         | 40     | 7      |
| Bull           | 3.0          | 6.3         | 1.4         | 3.0                       | 1.1             | 42     | 4    | 58         | 38     | 0      |
| School House   | 1.6          | 5.9         | 1.5         | 4.0                       | 5.7             | 14     | 7    | 64         | 28     | 1      |
| South Fork     | 3.9          | 7.3         | 1.0         | 3.2                       | 3.3             | 27     | 5    | 34         | 31     | 13     |
| West Fork      | 4.4          | 5.0         | 0.9         | 2.6                       | 6.9             | 20     | 9    | 61         | 27     | 2      |
| Whitetail      | 4.2          | 6.3         | 1.0         | 2.2                       | 8.9             | 31     | 7    | 34         | 49     | 10     |
| Windfall       | 1.6          | 6.6         | 1.1         | 2.4                       | 5.1             | 10     | 5    | 41         | 38     | 16     |
| <b>Evans</b>   |              |             |             |                           |                 |        |      |            |        |        |
| R5             | 6.0          | 10.0        | 1.5         | 1.0                       | 1.3             | 0      | 7    | 41         | 27     | 15     |
| R6             | 7.5          | 3.5         | 0.9         | 1.5                       | 2.4             | 0      | 10   | 46         | 33     | 8      |
| R7             | 7.5          | 4.0         | 0.9         | 3.0                       | 3.5             | 0      | 7    | 47         | 35     | 7      |
| East Fork      | 4.0          | 4.5         | 1.5         | 1.0                       | 3.9             | 0      | 6    | 46         | 48     | 0      |
| South Fork     | 9.5          | 3.0         | 0.9         | 3.0                       | 4.1             | 0      | 3    | 26         | 53     | 15     |
| <b>Lake</b>    |              |             |             |                           |                 |        |      |            |        |        |
| L8             | 1.2          | 5.5         | 1.6         | 3.8                       | 3.5             | 32     | 6    | 66         | 28     | 0      |
| Bozard         | 2.4          | 5.7         | 1.4         | 3.3                       | 4.9             | 40     | 5    | 36         | 48     | 11     |
| West Fork      | 3.9          | 5.1         | 0.8         | 2.9                       | 5.0             | 44     | 7    | 43         | 46     | 4      |

<sup>a</sup> Rated on a scale from 1 to 4 for percent of substrate embedded: 1 =>75%; 2 = 50 – 75%; 3 = 25 – 50%; 4 = 0 – 25% (Platts et al. 1983).

Predicted emergence success was generally high, averaging 28.4 percent for all sampled sites (Table 3.13). Emergence success was positively correlated to the  $F_i$  at each site ( $r = 0.79$ ). The lowest values were observed in upper Lake Creek, where silt and sand sized particles comprised 63 and 89 percent of the core samples, respectively at the two sampled sites. Fry production was positively correlated to the availability of spawning gravel in the sampled reaches ( $r = 0.78$ ). The production potentials for sampled sites ranged from 0 to 31.2 fry/100 square meters (mean=13.7±8.2). The highest calculated values occurred in several tributaries to Benewah Creek.

### 3.2.3 Population Surveys

General patterns of cutthroat trout abundance and distribution vary among the target watersheds but are consistent from year to year and seem to be highly correlated to seasonal changes in water quality and quantity. As reported in past years (Peters et al. 1999; Lillengreen et al. 1996), cutthroat trout were sporadically distributed in the Lake, Benewah, and Alder Creek watersheds during base flow condition in the summer (Tables 3.14). Abundance in the second order tributaries was consistently much higher than in adjacent mainstem reaches for both Lake Creek (mean = 19.4 fish/100 sq. m. versus 1.2 fish/100 sq. m.) and Benewah Creek (mean = 14.5 fish/100 sq. m. versus 0.5 fish/100 sq. m.), despite the effects of low flow conditions. During base flow conditions, for example, cutthroat trout were observed crowding (>15 fish/sq. m.) into small, isolated pools located in cool tributaries, rather than face conditions of high

water temperatures in mainstem reaches. In contrast, favorable temperature and flow conditions in the Evans Creek watershed resulted in a more even distribution of cutthroat trout (mean = 2.7 fish/100 sq. m., range = 1.0-4.8) in the lower mainstem reaches. The highest recorded abundance still occurred in the upper mainstem and primary tributaries where the majority of spawning activity is thought to take place. Of all the target watersheds, abundance has been consistently lowest in Alder Creek, with much of the upper watershed and the North Fork devoid of cutthroat.

Brook trout have been found only in the Alder Creek and Benewah Creek watersheds--their respective dates of introduction are unknown. In Alder Creek, brook trout are found in greater numbers than cutthroat trout in all but the lowermost stream reaches (Table 3.15). Fish were distributed fairly evenly (mean = 3.6 fish/100 sq. m.) in 11 of 13 sampled stream reaches. The remaining two reaches are influenced by vast beaver dam complexes and supported significantly higher populations (17.0 and 18.8 fish/100 sq. m., respectively). Distribution in the Benewah Creek watershed was limited to the upper mainstem and Windfall Creek. Abundance of brook trout ranged from 0.4-17.7 fish/100 square meters, and was much lower than for cutthroat trout in all sampled reaches.

Analysis of the age frequency of cutthroat trout caught during population surveys indicated that 88% of the catch consisted of juveniles (age 0-3) when averaged across all watersheds (Table 3.16). Young of the year cutthroat comprised an unusually high percentage of the catch (50.2%) in the Lake Creek watershed.

Table 3.13 Number of cores, mean fredle index ( $F_i$ ) of substrate composition of cores (range in parentheses), predicted mean emergence success, and mean estimated production potential for reaches in Alder (A), Benewah (B), Evans (E), and Lake (L) creek subbasins and their primary tributaries.

| Reach or subbasin | Number of cores | $F_i$                 | Emergence success | Production potential (# of fry/100 square meters) |
|-------------------|-----------------|-----------------------|-------------------|---|
| <b>Alder</b>      | <b>5</b>        | <b>7.8 (6.0-11.2)</b> | <b>31.9</b>       | <b>11.8</b>                                       |
| A9                | 1               | 6.0                   | 30.6              | 19.8  |
| North Fork        | 4               | 8.2 (7.0-11.2)        | 32.3              | 10.6  |
| <b>Benewah</b>    | <b>9</b>        | <b>7.1 (2.4-16.6)</b> | <b>29.6</b>       | <b>15.3</b>                                       |
| B12               | 1               | 3.7                   | 31.2              | 5.5   |
| Bull              | 1               | 16.6                  | 38.7              | 4.9   |
| School House      | 1               | 5.3                   | 28.4              | 19.0  |
| South Fork        | 2               | 2.8 (2.4-3.1)         | 24.0              | 9.1   |
| West Fork         | 2               | 5.9 (4.7-7.1)         | 26.9              | 24.0  |
| Whitetail         | 1               | 7.6                   | 29.6              | 31.2  |
| Windfall          | 1               | 13.8                  | 36.9              | 22.2  |
| <b>Evans</b>      | <b>8</b>        | <b>8.8 (2.7-13.7)</b> | <b>31.5</b>       | <b>11.1</b>                                       |
| R5                | 1               | 7.1                   | 30.6              | 4.9   |
| R6                | 2               | 11.1 (8.5-13.7)       | 34.7              | 9.8   |
| R7                | 2               | 10.4 (10.0-10.7)      | 34.7              | 14.5  |
| East Fork         | 1               | 2.7                   | 14.7              | 6.8   |
| South Fork        | 2               | 8.8 (6.6-11.0)        | 33.8              | 16.0  |
| <b>Lake</b>       | <b>7</b>        | <b>4.5 (0.8-7.1)</b>  | <b>20.9</b>       | <b>13.6</b>                                       |
| L8                | 2               | 0.9 (0.8-0.9)         | 0.0               | 0.0   |
| Bozard            | 2               | 5.4 (4.1-6.7)         | 28.2              | 16.6  |
| West Fork         | 3               | 6.4 (5.9-7.1)         | 29.9              | 17.7  |
| <b>Totals</b>     | <b>29</b>       | <b>7.1</b>            | <b>28.4</b>       | <b>13.0</b>                                       |

### Age and Growth

A total of 614 cutthroat and brook trout scales were examined for age and growth determination in 1998. Growth and potential maximum size of cutthroat trout varied from stock to stock (Table 3.17). As in past years, the adfluvial stocks found in Benewah Creek and Lake Creek exhibit the maximum growth potential for this species. The largest cutthroat recorded in Lake Creek for 1998 measured 16.3 inches (415 mm TL) and 1.5 pounds (667 grams), while the largest fish in Benewah Creek measured 15.2 inches (385 mm TL) and 1.3 pounds (570 grams). Maximum growth for resident fish stocks ranged from 222 to 280mm total length and 102 to 180 grams. A complete tabular analysis of growth for cutthroat trout is provided in (Appendix F).

Table 3.14 Cutthroat trout abundance and distribution by watershed, 1998.

| Lake Creek      |       | Summer 98    |          |             |         |  |
|-----------------|-------|--------------|----------|-------------|---------|--|
| Cutthroat Trout |       |              |          |             |         |  |
| Tributary       | Reach | Area (sq. m) | N±(SE)   | #/100 sq. m | Total # |  |
| Mainstem        | 1     | 5396         | 0.0      | 0.0         | 0       |  |
|                 | 4     | 2696         | 0.0      | 0.0         | 0       |  |
|                 | 5     | 2555         | 0.0      | 0.0         | 0       |  |
|                 | 6     | 11668        | 28(2.6)  | 4.9         | 572     |  |
|                 | 7     | 13284        | 7(0.0)   | 1.1         | 146     |  |
| 8               | 9715  | 26(0.5)      | 6.1      | 593         |         |  |
| West Fork       | 9,10  | 6270         | 125(1.7) | 42.0        | 2633    |  |
| Bozard          | 1,2,3 | 11085        | 28(2.5)  | 10.0        | 1109    |  |
| Total           |       | 62669        | 214(4.0) |             | 5053    |  |

| Evans Creek     |       | Summer 98    |          |             |         |  |
|-----------------|-------|--------------|----------|-------------|---------|--|
| Cutthroat Trout |       |              |          |             |         |  |
| Tributary       | Reach | Area (sq. m) | N±(SE)   | #/100 sq. m | Total # |  |
| Mainstem        | 1     | 4977         | 4(0.0)   | 1.0         | 50      |  |
|                 | 2     | 7227         | 21(2.2)  | 2.2         | 159     |  |
|                 | 3     | 1970         | 8(0.0)   | 2.7         | 53      |  |
|                 | 4     | 10127        | 20(1.6)  | 2.8         | 284     |  |
|                 | 5     | 2692         | 16(2.2)  | 4.8         | 129     |  |
|                 | 6     | 1178         | 12(0.8)  | 9.2         | 108     |  |
|                 | 7     | 2231         | 12(0.6)  | 8.1         | 181     |  |
| E. F. Evans     | 1     | 3990         | 23(3.6)  | 12.4        | 495     |  |
| R. F. Evans     | 1     | 2099         | 6(0.8)   | 8.1         | 170     |  |
| S. F. Evans     | 1,2   | 1126         | 14(3.0)  | 12.6        | 142     |  |
| Total           |       | 37616        | 136(3.0) |             | 1771    |  |

| Benewah Creek   |                 | Summer 98    |          |             |         |      |  |
|-----------------|-----------------|--------------|----------|-------------|---------|------|--|
| Cutthroat Trout |                 |              |          |             |         |      |  |
| Tributary       | Reach           | Area (sq. m) | N±(SE)   | #/100 sq. m | Total # |      |  |
| Mainstem        | 1               | 7422         | 0        | 0.0         | 0       |      |  |
|                 | 2               | 9419         | 1(0.0)   | 0.1         | 9       |      |  |
|                 | 3               | 5588         | 10(0.0)  | 1.3         | 73      |      |  |
|                 | 4               | 16104        | 0        | 0.0         | 0       |      |  |
|                 | 5               | 2318         | 0        | 0.0         | 0       |      |  |
|                 | 8               | 5656         | 10(2.9)  | 1.1         | 62      |      |  |
|                 | 9               | 5648         | 6(0.0)   | 1.1         | 62      |      |  |
|                 | 10              | 25981        | 3(0.0)   | 0.3         | 78      |      |  |
|                 | 11              | 1399         | 1(0.0)   | 0.3         | 4       |      |  |
|                 | South East Fork | 13           | 6915     | 69(3.2)     | 17.7    | 1224 |  |
|                 | West Fork       | 14           | 3205     | 44(2.7)     | 33.8    | 1083 |  |
| Bull            | 1               | 3685         | 14(2.2)  | 18.8        | 693     |      |  |
| Coon            | 1,2             | 2149         | 0        | 0.0         | 0       |      |  |
| School House    | 1               | 2741         | 2(0.0)   | 0.9         | 25      |      |  |
| Whitetail       | 1,2             | 5204         | 10(0.8)  | 6.7         | 349     |      |  |
| Windfall        | 2               | 5531         | 35(4.5)  | 23.5        | 1300    |      |  |
| Total           |                 | 108966       | 205(7.2) |             | 4962    |      |  |

| Alder Creek     |       | Summer 98    |         |             |         |  |
|-----------------|-------|--------------|---------|-------------|---------|--|
| Cutthroat Trout |       |              |         |             |         |  |
| Tributary       | Reach | Area (sq. m) | N±(SE)  | #/100 sq. m | Total # |  |
| Mainstem        | 1     | 7052         | 0       | 0.0         | 0       |  |
|                 | 2     | 1825         | 0       | 0.0         | 0       |  |
|                 | 3     | 9446         | 2(0.0)  | 3.6         | 340     |  |
|                 | 4     | 4158         | 6(0.0)  | 2.2         | 91      |  |
|                 | 5     | 5064         | 2(0.0)  | 0.4         | 20      |  |
|                 | 6     | 1823         | 2(0.0)  | 0.7         | 13      |  |
|                 | 7     | 16860        | 13(1.0) | 1.4         | 236     |  |
|                 | 8     | 4916         | 5(0.0)  | 1.4         | 69      |  |
|                 | 9     | 12635        | 0       | 0.0         | 0       |  |
| N. Fork Alder   | 1     | 4475         | 0       | 0.0         | 0       |  |
|                 | 2     | 1403         | 0       | 0.0         | 0       |  |
|                 | 3     | 2058         | 0       | 0.0         | 0       |  |
|                 | 4     | 2503         | 0       | 0.0         | 0       |  |
| Total           |       | 74216        | 30(1.0) |             | 769     |  |

Table 3.15 Brook trout abundance and distribution by watershed, 1998.

| <b>Benewah Creek<br/>Brook Trout</b> |                 | <b>Summer 98</b>    |                |                    |                |     |
|--------------------------------------|-----------------|---------------------|----------------|--------------------|----------------|-----|
| <b>Tributary</b>                     | <b>Reach</b>    | <b>Area (sq. m)</b> | <b>N±(SE)</b>  | <b>#/100 sq. m</b> | <b>Total #</b> |     |
| Mainstem                             | 1               | 7422                | 0              | 0.0                | 0              |     |
|                                      | 2               | 9419                | 0              | 0.0                | 0              |     |
|                                      | 3               | 5588                | 0              | 0.0                | 0              |     |
|                                      | 4               | 16104               | 0              | 0.0                | 0              |     |
|                                      | 5               | 2318                | 0              | 0.0                | 0              |     |
|                                      | 8               | 5656                | 0              | 0.0                | 0              |     |
|                                      | 9               | 5648                | 0              | 0.0                | 0              |     |
|                                      | 10              | 25981               | 4(3.5)         | 0.4                | 104            |     |
|                                      | 11              | 1399                | 4(0.0)         | 1.2                | 17             |     |
|                                      | South East Fork | 13                  | 6915           | 11(1.4)            | 2.8            | 194 |
|                                      | West Fork       | 14                  | 3205           | 23(2.2)            | 17.7           | 567 |
| Bull                                 | 1               | 3685                | 0              | 0.0                | 0              |     |
| Coon                                 | 1,2             | 2149                | 0              | 0.0                | 0              |     |
| School House                         | 1               | 2741                | 0              | 0.0                | 0              |     |
| Whitetail                            | 1,2             | 5204                | 0              | 0.0                | 0              |     |
| Windfall                             | 2               | 5531                | 2(0.0)         | 1.3                | 72             |     |
| <b>Total</b>                         |                 | <b>108966</b>       | <b>44(4.3)</b> |                    | <b>954</b>     |     |

| <b>Alder Creek<br/>Brook Trout</b> |              | <b>Summer 98</b>    |                  |                    |                |  |
|------------------------------------|--------------|---------------------|------------------|--------------------|----------------|--|
| <b>Tributary</b>                   | <b>Reach</b> | <b>Area (sq. m)</b> | <b>N±(SE)</b>    | <b>#/100 sq. m</b> | <b>Total #</b> |  |
| Mainstem                           | 1            | 7052                | 0                | 0.0                | 0              |  |
|                                    | 2            | 1825                | 0                | 0.0                | 0              |  |
|                                    | 3            | 9446                | 2(0.0)           | 3.6                | 340            |  |
|                                    | 4            | 4158                | 4(3.5)           | 2.2                | 91             |  |
|                                    | 5            | 5064                | 16(0.6)          | 3.2                | 162            |  |
|                                    | 6            | 1823                | 9(13.4)          | 3.2                | 58             |  |
|                                    | 7            | 16860               | 64(12.8)         | 7.0                | 1180           |  |
|                                    | 8            | 4916                | 60(5.3)          | 17.0               | 836            |  |
|                                    | 9            | 12635               | 26(3.0)          | 7.4                | 935            |  |
| N.F.Alder                          | 1            | 4475                | 30(2.9)          | 8.5                | 380            |  |
|                                    | 2            | 1403                | 16(3.0)          | 1.9                | 27             |  |
|                                    | 3            | 2058                | 14(0.6)          | 3.1                | 64             |  |
|                                    | 4            | 2503                | 14(2.2)          | 18.8               | 471            |  |
| <b>Total</b>                       |              | <b>74216</b>        | <b>255(20.4)</b> |                    | <b>4544</b>    |  |

Table 3.16 Age frequency distribution of cutthroat trout in electrofishing samples by watershed, 1998.

| <b>Count by Age Class</b> | <b>Age Class</b> |          |          |          |          |          |          |
|---------------------------|------------------|----------|----------|----------|----------|----------|----------|
|                           | <b>0</b>         | <b>1</b> | <b>2</b> | <b>3</b> | <b>4</b> | <b>5</b> | <b>6</b> |
| <b>Watershed</b>          |                  |          |          |          |          |          |          |
| Alder                     | 2                | 1        | 17       | 7        | 5        |          |          |
| Benewah                   | 8                | 35       | 91       | 16       | 5        | 3        |          |
| Evans                     | 5                | 28       | 45       | 31       | 16       | 3        |          |
| Lake                      | 88               | 18       | 22       | 22       | 15       | 7        | 3        |



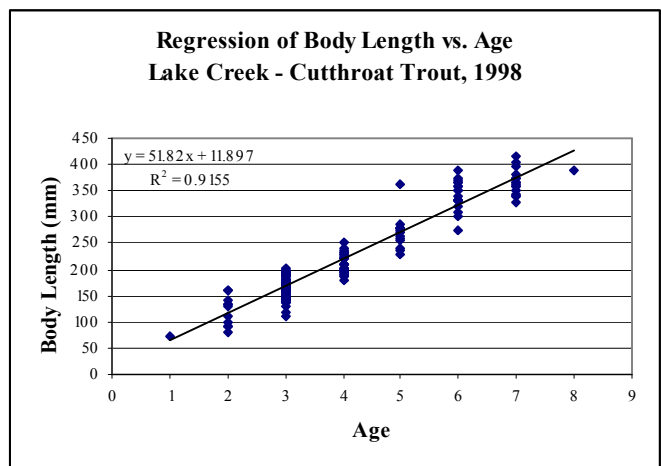
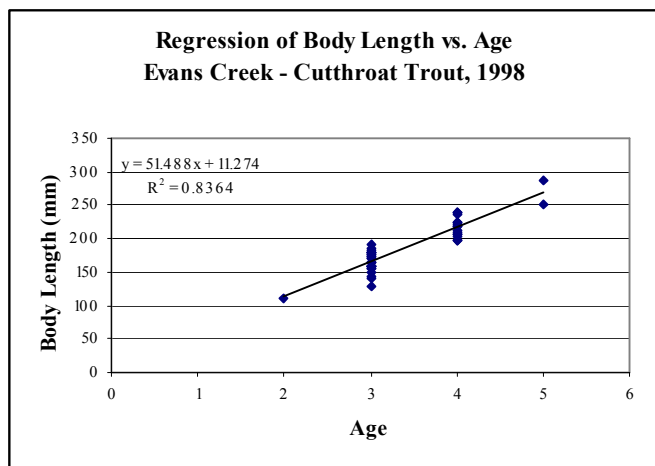
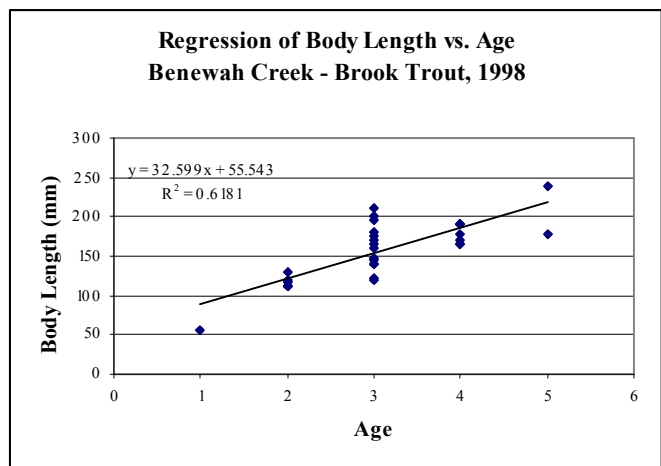
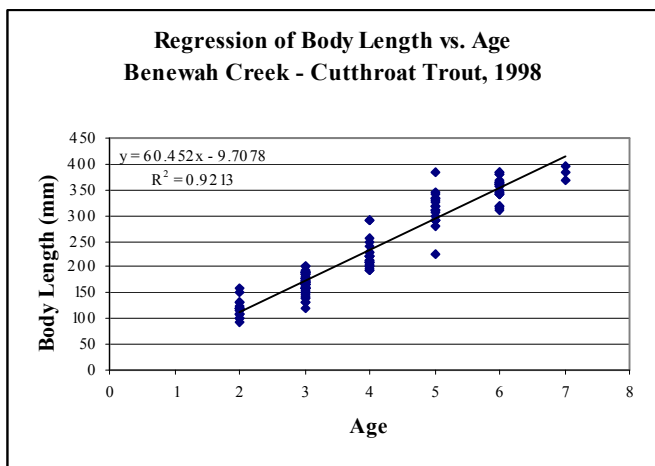
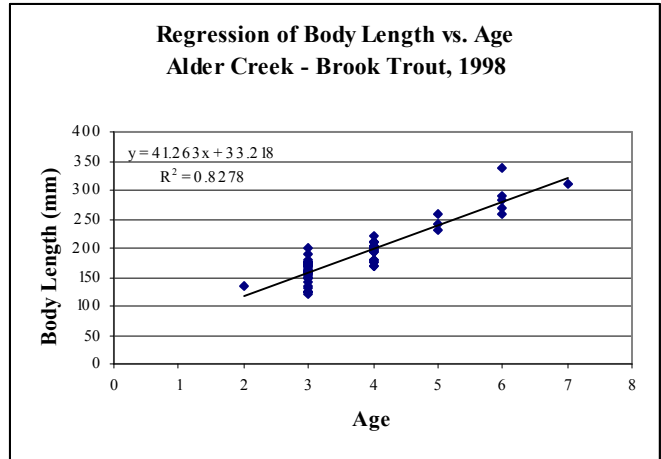
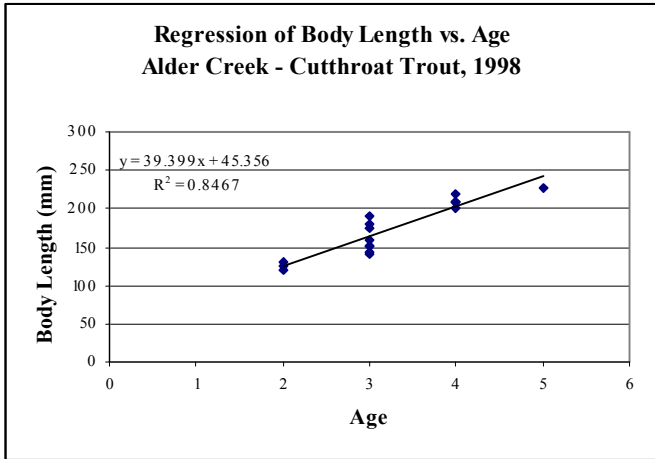


Figure 3.10 Regression equations of body length versus age for cutthroat and brook trout in four target watersheds, 1998.

Trout Migration

A total of 1277 cutthroat trout were caught in the lower Lake Creek trap in 1998 at a rate of 13.6 fish/day (Figure 3.10). Adult fish (age IV or older) accounted for 11.5% of the catch. Trapping success in Benawah Creek was considerably lower than in Lake Creek. A total of 535 cutthroat trout were caught in 1998 at a rate of 5.9 fish/day. Adult fish (age IV or older) accounted for 24.1% of the catch. Trapping

efficiency for smolts was estimated to be high (>75%) at both trap locations, with the exception of a five day period at the end of May when the traps were covered with water and fish would have been able to swim around the weirs.

Upstream migration of adult fish into Benewah Creek and Lake Creek was documented for the period March 11 through April 24, 1998. Average daily water temperature during the documented period of upstream migration ranged from 3.1°C to 11.4°C (mean = 5.5°C) and 3.1°C to 9.9°C (mean = 5.1°C) in Benewah and Lake creeks, respectively. Residence time for adults in Benewah Creek ranged from 22 to 34 days (mean = 27 days), while residence time in Lake Creek averaged 44 days.

Outmigrant behavior, was observed predominantly in age II and III fish (Figures 3.11 and 3.12). These fish ranged in size from 9 to 54 grams and 120 to 162 mm long. More than 70 percent of these fish were greater than 18 grams and 125mm long. Migration by smaller fish was observed, however, mortality both from predation and physical damage likely increased with decreasing size and weight. Migration by smolts to Coeur d'Alene Lake began in mid-April, as water temperatures stabilized around 9°C, and was completed by early June. Peak migrations coincided with average daily water temperatures ranging from 9°C to 11°C and were often correlated with small spikes in the hydrograph.

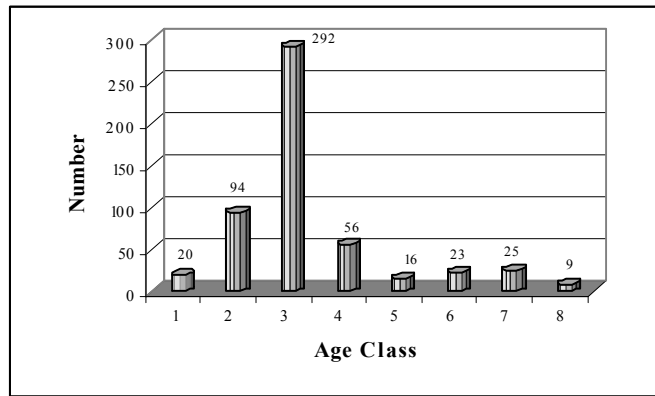
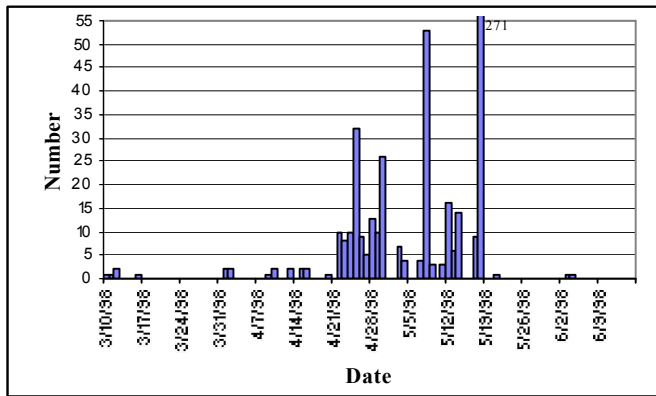


Figure 3.11 Analysis of cutthroat trout migration showing run timing and age frequency of migrants in Benewah Creek, 1998.

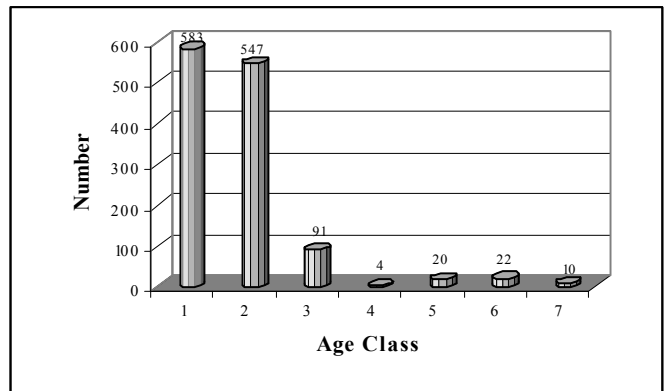
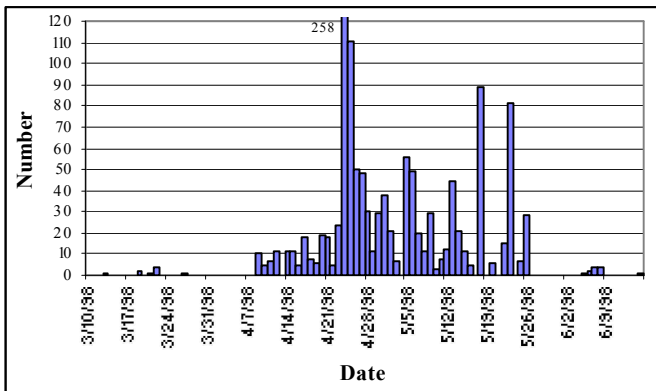


Figure 3.12 Analysis of cutthroat trout migration showing run timing and age frequency of migrants in Lake Creek, 1998.

## 4.0 Discussion

### 4.1 Lake Studies

#### Water Quality

Results from water quality sampling in 1998 did not differ much from results obtained in 1997. The main difference was in the number of sites with dissolved oxygen values below 6.0 mg/L. There were eight sites with values below 6.0 mg/L in 1998 and only four in 1997. A reason for the increase in the number of sites with dissolved oxygen readings below the optimum required for cutthroat trout survival was indirectly related to the weather. In 1998, from July to September there were 32 days with temperatures above 32.2° C (90° F), compared to only 10 days in 1997 (Personal communication with the National Weather Service). The more days with temperatures above 32.2° C (90° F) the warmer the water temperature will be. Upon comparing the water temperatures recorded in 1998 with the results from 1997 there was a noticeable difference. The thermoclines in both years seem to start around the end of May but it's the duration and depth at which the thermoclines occurred that was different. The thermoclines were much stronger and had a longer duration in 1998 compared to 1997. The surface temperatures also reached their peak at the end of July in 1998 compared to mid August in 1997. Due to the warmer air and water temperatures, and increased light penetration aquatic macrophyte growth most likely increased. Even though no direct measurements were taken. This in turn, led to more dead and decomposing material that sank and accumulated in the hypolimnion. Respiration and decomposition can easily deplete dissolved oxygen in the hypolimnion. Woods and Beckwith (1996) reported seeing low dissolved oxygen values in Coeur d'Alene Lake during the summer months of 1991 and 1992.

The low dissolved oxygen values that we are seeing in the hypolimnion is cause for concern when related to available fisheries habitat. The low dissolved oxygen values recorded in 1998 are not considered completely limiting for cutthroat trout suitability. The low dissolved oxygen values recorded, however, are thought to have indirect affects on cutthroat trout suitability in the southern lakes area (Peters et al. 1999). Peters et al. (1999) examined the habitat suitability and determined the lower 10 m in most areas is unsuitable habitat for cutthroat trout due to low dissolved oxygen values. This is not to say that cutthroat trout will not be found in these areas, but they most likely will have a hard time sustaining themselves for long periods of time.

Another more serious problem associated with decreases in dissolved oxygen values is the high concentration of metals that are bound up in the sediment north of the Coeur d'Alene River. Sediments to the south of the Coeur d'Alene River are relatively unaffected by heavy metal deposition from the Coeur d'Alene River. The deposition of trace elements in the sediments of Coeur d'Alene Lake is well documented by (Funk 1973; Rieman 1980; Woods 1989; Woods and Beckwith 1996). Lakebed geochemistry analyses revealed that most of the trace elements in surficial and subsurface sediments are associated with a ferric oxide phase and, thus, under reducing (anoxic) conditions, the trace elements would be readily solubilized and available for release to the overlying water column (Woods and Beckwith, 1996). The fact that trace metals are found in the sediments at the mouth of the river and north causes us to be concerned when sites like Windy Bay Deep and Mid-Lake Coeur d'Alene have dissolved oxygen values below 6.0 mg/L.

One hypothesis for the low dissolved oxygen values recorded is the presence of zinc in the water, which strongly inhibits phytoplankton growth. As phytoplankton drifts northward towards the mouth of the Coeur d'Alene River, it is thought to produce a sestonic "rain" of dead and dying phytoplankton that settles into the hypolimnion of the lake's northern basin and produces the hypolimnetic dissolved-oxygen deficit when the lake is thermally stratified (Woods and Beckwith, 1996). Monitoring dissolved oxygen levels in the hypolimnion is and will continue to be a high priority so that heavy metal contamination of the over lying water column will not occur.

Sediment transport to the lake from the tributaries continues to be a problem. The build up of these sediments at the mouths of all incoming tributaries to Coeur d'Alene Lake results in the loss of existing wetlands and provides prime habitat for predators of cutthroat trout. The increase in sediment also gives aquatic macrophytes more room to establish themselves along with providing more suitable forage habitat for non-native species. Sediment accumulation near the mouth of Plummer Creek, in Chatcolet Lake, has occurred at an average rate of 2.4 cm per year since the Mt. St. Helen's eruption of May 18, 1980 (Breithaupt, 1990). To examine the amount of sediment that has entered the Coeur d'Alene system aerial photos from 1933 and 1995 were compared. It was noted that the size of the deltas on four main tributaries that feed into Coeur d'Alene Lake are increasing in size. Another interesting point to be made is the change in vegetation over the years. In 1933, the tributaries at their mouths had well defined channels and were lined with cottonwoods. Now the mouths of the tributaries and bays are overrun with submerged and emerged aquatic macrophytes.

### **Fisheries**

In 1998, we hypothesized that there could be a significant difference in relative abundance results if we increased the sampling effort. Twenty-two more hours were spent electroshocking and two hundred and ninety-one more hours were spent gillnetting in 1998 than in 1997. We sampled approximately 7,800 fish in 1998 and 5,200 in 1997. The 1998 data reiterated the findings from 1994-1996 (Peters et. al. 1999) and 1997. Thus, we saw no significant difference in relative abundance by increasing our sampling effort. Species composition remained the same even though more fish were sampled.

Four out of the top six sampled species are non-native species. Northern pikeminnow and largescale suckers are the only native fish sampled in the top six. Northern pikeminnow make up 8-9% of the electroshocking catch and 18-20% of the gillnet catch. Largescale suckers make up 24-28% of the electroshocking catch and 9-21% of the gillnet catch. Cutthroat trout and mountain whitefish make up less than 1% of the catch when electroshocking and about 1.4% of the gillnet catch. When Post Falls Dam went on line in 1903 it inundated over 10 square kilometers of shoreline. This created prime habitat for exotic and non-salmonid native fish species.

Since 1994 the Coeur d'Alene Tribe Fish, Water and Wildlife program has conducted an extensive mark-recapture study (Peters et al. 1999). To date 636 fish have been tagged, there have been 23 recaptures between the Fish, Water and Wildlife program and anglers and anglers have harvested 15 fish. Eleven out of the 21 species have been tagged. The mark-recapture study relies heavily on the effectiveness of our sample gear and on angler's participation. We realize these results are biased because not all anglers report their catch, tags can be removed by anglers and they can fall out. During the 1997 and 1998 field season all game fish that were 300 mm and 300 g received a floy tag. Through the mark-recapture study we are finding that northern pike have a tendency to migrate from the original sampling site and largemouth bass are very territorial rarely moving from the site where they were tagged. It appears that both northern pike and largemouth tend to occupy shoreline habitat areas.

In general, northern pike are fairly sedentary, establishing a vague territory where cover and food are adequate (Scott and Crossman, 1973). Normally, this would be true because northern pike found on the northern end of Coeur d'Alene Lake tend not to migrate (Personal communication with Ned Horner IDF&G). However, northern pike in the southern 1/3 of Coeur d'Alene Lake appear to migrate over larger distances than would be normally expected. For example, a northern pike was sampled in September of 1997 in Chatcolet Lake and was harvested in March of 1998 in waters off of Harlow Point. This fish traveled approximately eight miles up the lake. Another northern pike was sampled in October of 1998 in O'Gara Bay and was harvested in March of 1999 in Chatcolet Lake; this fish migrated approximately 2-3 miles in a southerly direction. The reason we might be seeing this movement is due to the size of the littoral zone in the southern 1/3 of Coeur d'Alene Lake relative to the rest of the lake. When Post Falls Dam was built it raised Coeur d'Alene Lake by 3.6 m. This affected the southern 1/3 of

Coeur d'Alene Lake by turning small isolated lakes into a continuous body of water with Coeur d'Alene Lake. Because the inundated area has an average depth of 3.6 m or less, this led to the growth of aquatic macrophytes and suitable habitat for introduced non-native species. Due to the shallowness and the warm water temperatures in the southern 1/3 and near by bays of Coeur d'Alene Lake, northern pike tend not to restrict themselves to a specific forage area.

Scott and Crossman (1973) state that largemouth bass movement is not extensive usually less than five miles and summer territories are small. The majority of the largemouth bass in the southern 1/3 of Coeur d'Alene Lake appear to migrate less than one mile. A largemouth bass was originally sampled using electroshocking equipment in October of 1994 in Shingle Bay. The Fish, Water and Wildlife program sampled the same largemouth bass again in August of 1996 and June of 1998 in Shingle Bay. An example where anglers have played an important role in our mark-recapture study is with a largemouth that was sampled by the Fish, Water and Wildlife program in August of 1996 in O'Gara Bay. The same bass was caught and released in September of 1996 by an angler and was recently harvested in June of 1999 in O'Gara Bay by another angler.

Out of the 21 species inhabiting Coeur d'Alene Lake seven species have been shown to actively feed on cutthroat trout. They are chinook salmon, northern pike, northern pikeminnow, largemouth and smallmouth bass, bull trout and channel catfish. Of the non-native predators smallmouth bass are the newest addition. Smallmouth were illegally introduced into the northern part of Coeur d'Alene Lake. Peters et al. (1999) showed that during 1994- 1996 smallmouth bass were not sampled by electroshocking or gillnetting methods. In 1997 two smallmouth bass were sampled in waters off of Harlow Point. Current data is showing a gradual increase in smallmouth bass numbers in the littoral zone from Windy Bay south to O'Gara Bay where seventeen smallmouth were sampled during the 1998 field season. Most likely these fish will be found throughout the Coeur d'Alene system in a few years.

Smallmouth bass actively feed on crayfish and small fish from the age of two and older (Simpson and Wallace 1982). This could prove detrimental to cutthroat trout since they will occupy some of the same habitat as the smallmouth bass upon entry into the lake. The introduction of smallmouth bass may also have an effect on the existing largemouth bass population to what extent we do not know. This will be monitored in the future. Scott and Crossman (1973) state that the habitats of the smallmouth and largemouth bass seldom overlap even though the two species often occur in the same lake. Smallmouth prefer colder water than largemouth and they prefer rocky and sandy areas to the largemouth's vegetation habitat. Prior to the introduction of smallmouth into Coeur d'Alene Lake largemouth bass occupied some of the habitat currently occupied by smallmouth bass. This expansion of the smallmouth bass population could reduce the abundance of largemouth bass in certain parts of its former range within Coeur d'Alene Lake. The program will continue to monitor this in the future.

Kokanee salmon are not considered a predator on cutthroat trout but they may compete with the cutthroat for habitat and, on occasion, food. Kokanee in Coeur d'Alene Lake inhabit the middle to bottom layers of the limnetic zone, with the majority being sampled in the middle of the water column. Kokanee usually inhabit the upper and middle layers of the open lake (Scott and Crossman, 1973). According to our data cutthroat trout inhabit the bottom and middle layers of Coeur d'Alene Lake, thus it appears that these two species habitats overlap. Kokanee are mainly a pelagic, plankton feeder but it may derive a significant portion of its food from bottom organisms (Scott and Crossman, 1973). Cutthroat trout tend to be opportunistic in their food habits and are not highly piscivorous; instead they tend to specialize as invertebrate feeders (Roscoe 1974; Behnke 1979) thus, it is unlikely that they compete for food. However, Marnell (1988) determined that declines in westslope cutthroat trout populations in lakes in Glacier National Park where kokanee were introduced were caused by competition for planktivorous food. Unlike Marnell who showed that cutthroat trout numbers declined in the presence of kokanee salmon, in Coeur d'Alene Lake cutthroat trout and kokanee salmon are co-existing. The productivity of

the low elevation lakes compared to the non-productive high mountain lakes is a possible reason for co-existence. The low elevation lakes provide greater abundance and more diverse food than the high mountain lakes.

The majority of cutthroat trout are sampled at night as they come inshore to feed. A possible reason for sampling more cutthroat trout at night is that the water temperature cools enough to allow them to come to the surface to forage for food. The age structure of the cutthroat trout we sample in the littoral zone range from 2-6 and the age structure of the cutthroat trout in the limnetic zone are 3-5. Our age data is slightly biased towards older fish. Age 0 & 1 fish from all species is lacking representation in our sampling efforts. Juvenile cutthroat trout rear in local tributaries for two years so it is not surprising that they are, absent from our lucustrine sampling efforts. In order to examine this hypothesis we are initiating a beach seine sampling project around the lake in order to try and capture fish not susceptible to our electrofishing or gillnet capture methods.

## **4.2 Stream Studies**

### **Spawning Habitat**

Other studies have documented low-order tributaries as important spawning habitat for cutthroat trout (Rieman and Apperson 1989, Magee et al. 1996). Previous work on the Coeur d'Alene Reservation suggests that this holds true for our target tributaries as well (Peters et al 1999). In a 1997 study, adfluvial cutthroat trout were fitted with radio tags and tracked to spawning areas located in 2<sup>nd</sup> order tributaries within the Benewah Creek watershed. Spawning activity has been observed in the West Fork, a 2<sup>nd</sup> order tributary of Benewah Creek, for the past 4 years. Additionally, young of the year cutthroat are most often found in association with 2<sup>nd</sup> order tributaries in all target watersheds. The close association between redd distribution and density of juvenile and adults has been reported by other investigators (Beard and Carline 1991). Therefore, we feel that the effort to map the abundance and distribution of gravels within 2<sup>nd</sup> order tributaries should provide a good index of spawning potential for the respective watersheds.

The mapped distribution of potentially suitable spawning substrate within the spawning reaches of four target tributaries on the Coeur d'Alene Reservation was patchy and did not vary considerably within reaches or between watersheds (Figure 3.12). Furthermore, the quantity of spawning gravel was low, averaging just 4.1% of measured stream area. These results lie in contrast with those of Magee (1996), who reported a wide variance in proportion of spawning gravel for a Montana stream basin, even among nearby reaches, and documented much higher proportions of suitable spawning substrate (up to 25%). The lack of a strong association between spawning gravel abundance and several reach characteristics (gradient, proportion of gravel and pea gravel) in this study, however, corroborates the findings of Magee (1996), who suggested that local hydrologic features influenced spawning gravel availability. Although the distribution of spawning substrate was patchy within the target watersheds, there is probably adequate habitat to support resident and adfluvial spawners because of currently depressed numbers.

Abundance of potential spawning gravels has been reported as a key factor influencing redd density for westslope cutthroat trout at the reach scale (Magee et al. 1996). Similar relations between redd density and availability of spawning gravel were noted by Cope (1957) for Yellowstone cutthroat trout and by Beard and Carline (1991) for brown trout. Other important factors in determining redd density probably include temperature and fish density. Current population estimates for the target tributaries indicate wide variation in the abundance of juvenile fish (Figure 3.14), even though water temperatures and availability of spawning gravels are similar. Although redd densities were not estimated as part of this study, it is thought that fish density overrides other variables in determining redd density across the entire basin. Given the observed abundance and distribution of suitable spawning gravels in the target watersheds, redd superimposition could occur at relatively low population densities. This could have profound implications on projections for juvenile recruitment if the number of spawning adults increases over time,

either as a result of improved juvenile to adult survival or initiation of a supplementation program. In any case, it would be prudent to monitor spawning activity to look for evidence of superimposition.

### Is fine sediment limiting recruitment?

Spawning gravels in target tributaries of the Reservation contained proportions of fine sediments comparable to those in egg pockets of salmonid redds in the Rocky Mountain region (Table 4.1). Sites in Alder, Benewah and Evans Creeks contained low proportions of fines that were reflective of stable habitats and low impact landuses. The erosive geology of the Lake Creek watershed, on the other hand, results in substrate conditions similar to those reported by Magee et al. (1996) in Montana. The proportion of small fines (<0.85 mm) averaged two to 20 times higher than in spawning gravels elsewhere on the Reservation.

Table 4.1 Mean proportion of fines smaller than 6.35 mm and 0.85 mm and the mean freddie index ( $F_i$ ) of egg pockets in salmonid redds from the Rocky Mountain region, USA.

| Target species | Location      | Time of sampling | Percent fines<br>Smaller than: |         | $F_i$ | N  | Reference                              |
|----------------|---------------|------------------|--------------------------------|---------|-------|----|--|
|                |               |                  | 6.35 mm                        | 0.85 mm |       |    |  |
| Cutthroat      | Alder Creek   | Post emergence   | 17.8                           | 1.0     | 7.8   | 5  | This study*                            |
| Cutthroat      | Benewah Creek | Post emergence   | 22.8                           | 3.7     | 7.1   | 9  | This study*                            |
| Cutthroat      | Evans Creek   | Post emergence   | 18.6                           | 0.5     | 8.8   | 8  | This study*                            |
| Cutthroat      | Lake Creek    | Post emergence   | 44.0                           | 10.1    | 4.5   | 7  | This study*                            |
| Cutthroat      | Idaho         | Near emergence   | 24.4                           | 6.5     | 7.6   | 13 | R. Thurow and J. King unpublished data |
| Cutthroat      | Montana       | Near emergence   | 41.6                           | 17.9    | 2.0   | 11 | Magee et al. (1996)                    |
| Cutthroat      | Montana       | Near emergence   | 27.4                           | 5.9     |       | 13 | Weaver and Fraley (1993)               |
| Steelhead      | Idaho         | Near emergence   | 15.4                           | 3.7     | 10.5  | 9  | R. Thurow and J. King unpublished data |
| Brook          | Wyoming       | Near spawning    | 12.1                           | 6.4     |       | 31 | Young et al. (1989)                    |
| Brook          | Wyoming       | Near spawning    | 15.0                           | 3.0     |       | 69 | Grost et al. (1991)                    |

\*Means calculated from suitable spawning gravels rather than egg pockets for this study.

Several factors may account for reported differences in gravel characteristics, including sampling method, geology, and sediment transport. Other investigators have sampled egg pockets, noting significantly lower concentrations of fines in the upper strata when compared with nearby gravels (Chapman 1988, Young et al. 1989). Because we were unable to identify actual redds to sample, our results are most likely representative of unaltered spawning substrate or unstratified redd samples, as described by Young et al. (1989). The timing of sampling effort also may have resulted in some particle size differences. We sampled spawning gravels as late as four weeks after emergence, whereas some authors sampled egg pockets near spawning or at the time of emergence. Streams on the Reservation, however, carry little sediment during the incubation period for embryos because flows (and hence sediment transport) rapidly decline during spring and early summer.

At 23 of 29 sample sites, low levels of fine sediment led to high predictions of overall embryo survival (mean = 28.4%). The estimates of fry production potential for all sample sites ranged widely (0.0 to 31.2 fry/100 square meters) due, primarily, to the quantity of suitable gravels present (Table 3.13). Our estimates of emergence success were made by using proportions of fine sediment less than 6.35 mm in diameter, based on the experiments of Weaver and Fraley (1993) with artificial egg pockets planted in stream substrate. Actual emergence success may be lower in tributaries of Lake Creek and portions of South East Benewah Creek, where high proportions of small fines (<0.85 mm) were measured. Small fines are more detrimental to survival of incubating eggs and intrude deeper into spawning gravels than coarser fines (Reiser and White 1988, Beschta and Jackson 1979). Only in the mainstem of Lake Creek were the proportions of both small and coarse fines considered above the levels for these particle sizes

(10% and 30%, respectively) shown to adversely affect salmonid emergence success (McNeil and Ahnell 1964, Shepard et al. 1984a, Cederholm and Reid 1987, Reiser and White 1988).

Previous studies have demonstrated that reduced emergence success from high sedimentation can result in low juvenile densities and low adult recruitment (Cederholm and Reid 1987, Scrivener and Brownlee 1989). Of the 6 sites where high levels of small or coarse fines were recorded, only the sites located in the mainstem of Lake Creek showed supporting evidence for low recruitment. The other sites had juvenile and adults densities (range = 12.4 – 33.8/100 square meters) that were notably higher than the average density (9.2/100 square meters) reported for seven other westslope cutthroat trout populations in Idaho and Montana (Shepard et al. 1984b, Ireland 1993). Several factors may buffer the effects of high sedimentation of redds. High fry mortality is common in resident salmonids, but low fry densities may result in decreased competition, increased growth, and compensatory survival (McFadden 1969). Three additional spawning tributaries (School House Creek, Whitetail Creek, and Rainbow Fork Evans Creek) showed evidence of low recruitment; however this is probably attributable to low numbers of spawning adults, low quantity of suitable gravels, or poor water quality conditions, rather than sedimentation of redds.

### **Population Trends**

Due to the persistence of adverse water quality and habitat conditions in Reservation streams, cutthroat trout populations are thought to be at least moderately damaged (i.e. average spawning escapements fall between the minimum viable population and the number of adults needed to produce 50% of the carrying capacity of the stream environment). Reiman and Apperson (1989) estimated that populations considered as “strong” (greater than or equal to 50% of historic potential) by Idaho Department of Fish and Game (IDFG) remained in only 11% of the historic range within the State of Idaho. In contrast, none of the populations on the Coeur d’Alene Indian Reservation are considered “strong”.

Mean annual population estimates from 1996-1998 rank Benewah Creek, Lake Creek, Evans Creek, and then Alder Creek in order of decreasing population strength (Table 4.2). In Benewah Creek, four tributaries, comprising 32% of the usable stream area in the watershed, have cutthroat trout densities that are notably higher than the average density (9.2/100 square meters) reported for seven other westslope cutthroat trout populations in Idaho and Montana (Shepard et al. 1984b, Ireland 1993). In Lake Creek, 41% of the perennial stream area supports higher than average densities, while 26% of the stream area in Evans Creek supports higher than average densities. All reaches in the Alder Creek drainage have significantly lower than average densities.

The probability of persistence, as shown in table 4.2, was calculated based on methods described by Reiman and McIntyre (1993), but should be used only as an index of population resiliency. When used in this manner, probability of persistence improves as population size increases or when the inter-annual variance in population size decreases. Populations in Lake Creek and Benewah Creek demonstrate the most resiliency of our target stocks, because of both higher numbers of individuals and relatively low variances.

Despite the apparent instability of cutthroat trout populations, preliminary genetic analyses show that relatively pure stocks exist in Reservation waters. A study by Spruell et al. (1999) differentiated westslope cutthroat trout, rainbow trout and their hybrids at 16 sample sites. Six sites contained samples of westslope cutthroat trout with no evidence of hybridization. The remaining ten sites included at least one hybrid individual. When present, hybridization occurs at a low level and most likely represents episodic events of migration into these systems by rainbow trout or hybrid individuals. Thus, it appears that even though the populations are not “strong”, they are not threatened to a large extent with hybridization. These weakened populations may be particularly susceptible to normal environmental variability (such as temperature and stream discharge patterns) and to the frequency of extreme events



such as wildfires, floods, or debris torrents. However, the existing genetic diversity may continue to allow for genetic combinations that permit survival in highly variable environments. Implications are that if the effect of limiting factors can be reduced, then genetically pure populations would have a chance to recover.

Table 4.2. Mean annual population estimates, the estimated mean annual variance in the infinitesimal rate of population growth, and probabilities of persistence over 100 years for westslope cutthroat trout populations monitored on the Coeur d'Alene Reservation. The 95% confidence interval is shown in parentheses.

| Stream        | Years | Mean Annual Population Estimate | Variance         | Probability Of Persistence |
|---------------|-------|---------------------------------|------------------|----------------------------|
| Alder Creek   | 3     | 808                             | 0.03 (0.02-0.04) | 0.58                       |
| Benewah Creek | 3     | 5,553                           | 0.16 (0.04-0.36) | 0.67                       |
| Evans Creek   | 3     | 2,675                           | 0.33 (0.05-0.71) | 0.45                       |
| Lake Creek    | 3     | 4,946                           | 0.14 (0.02-0.26) | 0.70                       |

## 5.0 Bibliography

- Alaska Department of Fish and Game. 1987. Limnology field and laboratory manual: Methods for assessing aquatic production. Juneau, Alaska. p. 212.
- American Public Health Association. 1992. Standard methods for the examination of water and wastewater.
- Armour, C.L., K.P. Burnham, and W.S. Platts. 1983. Field methods and statistical analyses for monitoring small salmonid streams. USDI, Fish and Wildlife Service. FWS/OBS-83/33.
- Beard, T.D., and R.F. Carline. 1991. Influence of spawning and other stream habitat features on spatial variability of wild brown trout. Transactions of the American Fisheries Society 120:711-722.
- Behnke, R.J. 1979. The native trouts of the genus *Salmo* of western North America. Report to U.S. Fish and Wildlife Service, Denver, Colorado.
- Beschta, R.L. and W.L. Hackson. 1979. The intrusion of fine sediments into a stable gravel bed. Journal of the Fisheries Research Board of Canada 36:204-210.
- Bovee, K.D. 1982. A guide to stream habitat analysis using the Instream Flow Incremental Methodology. Instream Flow Information Paper No. 12. U.S. Fish Wildl. Serv. FWS/OBS-82/26. 248 pp.
- Breithaupt, S.A. 1990. Plummer Creek and Chatcolet Lake Benewah and Kootenai Counties. Water Quality Status Report No. 94. Idaho Dept. of Health and welfare Division of Environmental Quality Water Quality Bureau.
- Carlander, K.D. 1981. Caution on the use of the regression method of back-calculating lengths from scale measurements. Fisheries 6:2-4.
- Cederholm, C.J. and L.M. Reid. 1987. Impact of forest management on coho salmon (*Oncorhynchus kisutch*) populations of the Clearwater River, Washington: a project summary. Pages 373-398 in E. Salo and T. Cundy, editors. Streamside management: forestry and fishery interactions. University of Washington, College of Forest Resources, Contribution 57, Seattle.
- Chambers, M. 1999. Personal communication. National Weather Service.
- Chapman, D.W. 1988. Critical review of variables used to define effects of fines in redds of large salmonids. Transactions of the American Fisheries Society 117:1-21.
- Conlin, K. and B.D. Tuty. 1979. Juvenile salmon field trapping manual. Dept. of Fisheries and Oceans. Fisheries and Marine Service Resource Services Branch, Habitat Protection Division, Vancouver, B.C. 136p.
- Cope, O.B. 1957. The choice of spawning sites by cutthroat trout. Proceedings of the Utah Academy of Sciences, Arts, and Letters 34:73-79.
- Downs, C. and B. Shepard. 1999. Personal communication. Montana State University, Bozeman, Montana.
- Funk, W.H., F.W. Rabe, and F. Royston. 1973. The biological impact of combined metallic and organic pollution in the Coeur d'Alene-Spokane River drainage system. Water and Energy Resources Research Institute. Moscow, Idaho. 202p.
- Graves, S., K.L. Lillengreen, D.C. Johnson, and A.T. Scholz. 1990. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation. Bonneville Power Administration. Division of Fish and Wildlife. Portland Oregon. Project #90-44. 68p.

- Grost, R.T., W.A. Hubert, and T.A. Wesche. 1991. Description of brown trout redds in a mountain stream. *Transactions of the American Fisheries Society* 120:582-588.
- Hankin, D.G. and G.H. Reeves. 1988. Estimating total fish abundance and total habitat area in small streams based on visual estimation methods. *Canadian Journal of Fisheries and Aquatic Sciences* 45:834-844.
- Hickman, T., and R.F. Raleigh. 1982. Habitat suitability index models: Cutthroat trout. U.S.D.I. Fish and Wildlife Service. FWS/OBS-82/10.5. 38p.
- Horner, N. 1998. Personal communication. Idaho Department of Fish and Game, Coeur d'Alene, Idaho.
- Ireland, S.C. 1993. Seasonal distribution and habitat use of westslope cutthroat trout in a sediment-rich basin in Montana. Master's thesis. Montana State University, Bozeman.
- Jearld, T. 1983. Age determination. *In*: Nielsen, L.A. and D.L. Johnson (eds.), *Fisheries Techniques*. American Fisheries Society, Bethesda, MD. 468p.
- Lillengreen, K.L., A.J. Vitale, and R. Peters. 1996. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation, 1993-1994 annual report. USDE, Bonneville Power Administration, Portland, OR. 260p.
- Lillengreen, K.L., T. Skillingstad, and A.T. Scholz. 1993. Fisheries habitat evaluation on tributaries of the Coeur d'Alene Indian Reservation. Bonneville Power Administration. Division of Fish and Wildlife. Portland Or. Project # 90-44. 218p.
- Lux, F.E. 1971. Age determination of fishes (revised). NMFS, Fishery leaflet 637. 7p.
- Magee, J.P., T.E. McMahon, and R.F. Thurow. 1996. Spatial variation in spawning habitat of cutthroat trout in a sediment-rich stream basin. *Transactions of the American Fisheries Society* 126:768-779.
- Mallet, J. 1968. Coeur d'Alene Lake fisheries investigations. Idaho Department of Fish and Game. Job Performance Report.
- Mallet, J. 1969. The Coeur d'Alene Lake fishery. *Idaho Wildlife Review*. May-June, 1969. 3-6 p.
- Marnell, L.F. 1998. Status of the westslope cutthroat trout in glacier national park, Montana. *American Fisheries Society Symposium* 4:61-70.
- McFadden, J.T. 1969. Dynamics and regulation of salmonid populations in streams. Pages 313-329 in T.G. Northcote, editor. *Symposium on salmon and trout in streams*. University of British Columbia, Vancouver.
- McNeil, W.J. and W.H. Ahnell. 1964. Success of pink salmon spawning relative to size of spawning bed materials. U.S. Fish and Wildlife Service Special Scientific Report Fisheries 469.
- Peters, R., A.J. Vitale, and K.L. Lillengreen. 1999. Supplementation feasibility report Coeur d'Alene Tribe Fish, Water and Wildlife Program. USDE, Bonneville Power Administration, Portland, OR.
- Platts, W.S., W.F. Megahan, and G.W. Minshall. 1983. Methods for evaluating stream, riparian, and biotic conditions. USDA Forest Service General Technical Report INT-138.
- Rabe, W., and D.C. Flaherty. 1974. The river of green and gold: A pristine wilderness dramatically affected by man's discovery of gold. Idaho Research Foundation, Inc. Natural Resource Series Number 4. 93 pp.
- Reiser, D.W. and R.G. White. 1988. Effects of two sediment size-classes on survival of steelhead and chinook salmon eggs. *North American Journal of Fisheries Management* 8:432-437.

- Reynolds, J.B. 1983. Electrofishing. *In*: Nielsen, L.A. and D.L. Johnson (eds.), Fisheries Techniques. American Fisheries Society, Bethesda, MD. 468p.
- Rieman, B.E. 1980. Coeur d'Alene Lake limnology: Idaho Department of Fish and Game, Lake and Reservoir Investigations, Job Performance Report F-72-R-2. 27-68p.
- Rieman, B.E., and K.A. Apperson. 1989. Status and analysis of salmonid fisheries: westslope cutthroat trout synopsis and analysis of fishery information. Idaho Department of Fish and Game, Federal Aid in Sport Fish Restoration, Project F-73-R-11, Boise, Idaho.
- Rieman, B.E. and J.D. McIntyre. 1993. Demographic and habitat requirements for conservation of bull trout. USDA Forest Service, Intermountain Research Station, General Technical Report INT-302.
- Roscoe, J.W. 1974. Systematics of the westslope cutthroat trout. M.S. Thesis, Colorado State University, Ft. Collins, CO. 74p.
- Rosgen, D.L. 1994. A classification of natural rivers. *Catena* 22:169-199.
- Scott, W.B., and E.J. Crossman. 1973. Freshwater fishes of Canada. Fisheries Research Board of Canada, Ottawa. Bulletin 184. 966p.
- Scholz, A.T., D.R. Geist, and J.K. Uehara. 1985. Feasibility report on restoration of Coeur d'Alene Tribal Fisheries. Upper Columbia United Tribes Fisheries Center. Cheney, WA. 85 pp.
- Scrivener, J.C. and M.J. Brownlee. 1989. Effects of forest harvesting on spawning gravel and incubation survival of chum (*Oncorhynchus keta*) and coho salmon (*O. kisutch*) in Carnation Creek, British Columbia. *Canadian Journal of Fisheries and Aquatic Sciences* 46:681-696.
- Seber, G.A.F., and E.D. LeCren. 1967. Estimating population parameters from catches large relative to the population. *Journal Animal Ecology* 36:631-643.
- Shepard, B.B., S.A. Leathe, T.M. Weaver, and M.D. Enk. 1984a. Monitoring levels of fine sediment within tributaries to Flathead Lake, and impacts of fine sediment on bull trout recruitment. Pages 146-156 *in* F. Richardson and F.H. Hamre, editors. Wild trout III. Trout Unlimited, Vienna, Virginia.
- Shepard, B.B., K.L. Pratt, and P.J. Graham. 1984b. Life histories of westslope cutthroat trout and bull trout in the upper Flathead River basin, Montana. Montana Department of Fish, Wildlife and Parks, Kalispell.
- Simpson, J.C. and R.L. Wallace. 1982. Fishes of Idaho. University of Idaho Press, Moscow, Idaho.
- Spruell, P., K.L. Knudsen, J. Miller, and F.W. Allendorf. 1999. Genetic analysis of westslope cutthroat trout in tributaries of Coeur d'Alene Lake. Progress report WTSFL99-101 to the Coeur d'Alene Tribe.
- Thurow, R.F., and J.G. King. 1994. Attributes of Yellowstone cutthroat trout redds in a tributary of the Snake River, Idaho. *Transactions of the American Fisheries Society* 123:37-50.
- U.S. Department of Agriculture. 1984. General soil map and landform provinces of Idaho: Boise, Idaho. Soil Conservation Service. 1p.
- U.S. Department of Agriculture. 1998. Biological Assessment St. Joe and North Fork Clearwater Rivers. US Forest Service.
- Weaver, T.M. and J.J. Fraley. 1993. A method to measure emergence success of westslope cutthroat trout fry from varying substrate compositions in a natural stream channel. *North American Journal of Fisheries Management* 13:817-822.

- Wolman, G.M. 1954. A method of sampling coarse river-bed material. Transactions, American Geophysical Union 35:951-956.
- Woods, P.F. 1989. Hypolimnetic concentrations of dissolved oxygen, nutrients, and trace elements. USGS- Water Resources Investigations Report 89-4032. 56p.
- Woods, P. F. and C. Barenbrock. 1994. Bathymetric map of Coeur d'Alene Lake, Idaho. U.S. Geological Survey Water Resources Investigations Report 94-xxx. 1p.
- Woods, P.F., and M.A. Beckwith. 1996. Nutrient and trace-element enrichment of Coeur d'Alene Lake, Idaho. Open file report 95-740.
- Young, M.K., W.A. Hubert, and T.A. Wesche. 1989. Substrate alteration by spawning brook trout in a southeastern Wyoming stream. Transactions of the American Fisheries Society 118:379-385.
- Young, M.K., W.A. Hubert, and T.A. Wesche. 1991. Selection of measures of substrate composition to estimate survival to emergence of salmonids and to detect changes in stream substrates. North American Journal of Fisheries Management 11:339-346.
- Zippen, C. 1958. The removal method of population estimation. Journal of Wildlife Management 22:82-90.

## **Appendix A**

Vertical hydrograph profiles of thirteen stations on Coeur d'Alene Lake, 1997 and 1998.

Appendix A.1 Vertical hydrograph profiles for thirteen sampling station on Coeur d'Alene Lake, 1997.

| Location     | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 11       | 0.4       | 12.58                   | 5.64            | 7.34 | 42.4                 | 0.03 | *****                      | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 10       | 1.3       | 12.51                   | 5.42            | 7.33 | 42.2                 | 0.03 | 99                         | -                  | 373   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 9        | 2.6       | 12.54                   | 5.24            | 7.34 | 41.9                 | 0.03 | 98.8                       | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 8        | 4         | 12.48                   | 4.98            | 7.32 | 41.6                 | 0.03 | 97.6                       | -                  | 373   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 7        | 5.4       | 12.5                    | 4.94            | 7.31 | 41.8                 | 0.03 | 97.7                       | -                  | 373   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 6        | 6.7       | 12.47                   | 4.87            | 7.31 | 42.5                 | 0.03 | 97.3                       | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 5        | 7.8       | 12.47                   | 4.81            | 7.31 | 42.7                 | 0.03 | 97.2                       | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 4        | 9         | 12.48                   | 4.74            | 7.3  | 43.2                 | 0.03 | 97                         | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 3        | 10.3      | 12.47                   | 4.69            | 7.29 | 43.2                 | 0.03 | 96.9                       | -                  | 372   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 2        | 11.8      | 12.79                   | 4.65            | 7.29 | 43.4                 | 0.03 | 99.2                       | -                  | 370   |
| Rockford Bay | 1597  | ASSS     | 4/18/97 | -    | 1.7        | 1        | 13        | 12.19                   | 4.61            | 7.28 | 43.2                 | 0.03 | 94.4                       | -                  | 369   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 11       | 0.2       | 11.36                   | 13.57           | 7.09 | 36.1                 | 0.02 | 108.4                      | -                  | 409   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 10       | 2         | 11.31                   | 10.51           | 7.09 | 34.2                 | 0.02 | 100.6                      | -                  | 412   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 9        | 3.4       | 11.39                   | 10.46           | 7.08 | 34.7                 | 0.02 | 101.3                      | -                  | 412   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 8        | 5         | 11.54                   | 10.41           | 7.06 | 37.4                 | 0.02 | 102.4                      | -                  | 413   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 7        | 6.5       | 11.53                   | 10.22           | 7.05 | 37.5                 | 0.02 | 101.9                      | -                  | 413   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 6        | 8         | 11.53                   | 10              | 7.05 | 38                   | 0.02 | 101.3                      | -                  | 414   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 5        | 9.5       | 11.49                   | 9.74            | 7.05 | 37.6                 | 0.02 | 100.3                      | -                  | 413   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 4        | 11        | 11.43                   | 9.55            | 7.04 | 37.6                 | 0.02 | 99.4                       | -                  | 413   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 3        | 12.5      | 11.37                   | 9.32            | 7.02 | 38.2                 | 0.02 | 98.3                       | -                  | 414   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 2        | 14        | 11.27                   | 8.83            | 7.01 | 38.7                 | 0.02 | 96.3                       | -                  | 414   |
| Rockford Bay | 1997  | DBSS     | 5/16/97 | -    | 1.3        | 1        | 15.4      | 11.21                   | 7.66            | 7.01 | 39.9                 | 0.03 | 93.1                       | -                  | 414   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 15       | 0.6       | 10.57                   | 16.8            | 6.91 | 32.1                 | 0.02 | 108.4                      | -                  | 387   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 14       | 1.6       | 10.6                    | 16.43           | 6.88 | 32                   | 0.02 | 107.9                      | -                  | 388   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 13       | 2.5       | 10.42                   | 12.02           | 6.82 | 32.5                 | 0.02 | 96.2                       | -                  | 393   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 12       | 3.5       | 10.75                   | 11.31           | 6.76 | 31                   | 0.02 | 97.7                       | -                  | 395   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 11       | 4.6       | 10.54                   | 10.92           | 6.73 | 30.8                 | 0.02 | 94.8                       | -                  | 396   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 10       | 5.5       | 10.49                   | 10.36           | 6.72 | 30.4                 | 0.02 | 93.2                       | -                  | 396   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 9        | 6.6       | 10.49                   | 10.18           | 6.7  | 30.4                 | 0.02 | 92.8                       | -                  | 397   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 8        | 7.6       | 10.62                   | 9.96            | 6.69 | 29                   | 0.02 | 93.4                       | -                  | 397   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 7        | 8.7       | 10.61                   | 9.82            | 6.66 | 30.2                 | 0.02 | 93.1                       | -                  | 397   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 6        | 9.7       | 10.63                   | 9.71            | 6.64 | 30.2                 | 0.02 | 93                         | -                  | 398   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 5        | 10.6      | 10.62                   | 9.69            | 6.62 | 30.4                 | 0.02 | 92.8                       | -                  | 398   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 4        | 11.6      | 10.51                   | 9.64            | 6.6  | 31.3                 | 0.02 | 91.9                       | -                  | 400   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 3        | 12.6      | 10.51                   | 9.36            | 6.57 | 32                   | 0.02 | 91.1                       | -                  | 401   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 2        | 13.6      | 10.5                    | 9.12            | 6.53 | 32.8                 | 0.02 | 90.5                       | -                  | 401   |
| Rockford Bay | 2197  | DBSS     | 5/29/97 | -    | 1.3        | 1        | 14.8      | 10.53                   | 8.63            | 6.48 | 34                   | 0.02 | 89.7                       | -                  | 404   |

|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 14 | 0.6  | 10.87 | 14.41 | 7.14 | 34.4 | 0.02 | 106.3 | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 13 | 1.5  | 11.11 | 12.51 | 7.13 | 33.1 | 0.02 | 104.1 | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 12 | 2.5  | 11.13 | 12.08 | 7.13 | 32.4 | 0.02 | 103.1 | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 11 | 3.4  | 11.09 | 11.62 | 7.13 | 32   | 0.02 | 101.8 | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 10 | 4.6  | 10.9  | 10.8  | 7.13 | 30.9 | 0.02 | 98.1  | - | 416 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 9  | 5.5  | 10.81 | 10.41 | 7.14 | 30   | 0.02 | 96.5  | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 8  | 6.4  | 10.72 | 10.33 | 7.14 | 29.6 | 0.02 | 95.5  | - | 415 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 7  | 7.6  | 10.65 | 10.13 | 7.14 | 29.9 | 0.02 | 94.4  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 6  | 8.5  | 10.69 | 10.08 | 7.15 | 30.1 | 0.02 | 94.6  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 5  | 9.7  | 10.68 | 9.97  | 7.15 | 30.1 | 0.02 | 94.3  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 4  | 10.6 | 10.66 | 9.92  | 7.16 | 30   | 0.02 | 94.1  | - | 414 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 3  | 11.7 | 10.64 | 9.92  | 7.18 | 30   | 0.02 | 93.8  | - | 413 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 2  | 12.6 | 10.63 | 9.76  | 7.19 | 30.5 | 0.02 | 93.4  | - | 412 |
| Rockford Bay | 2397 | DBAS | 6/11/97 | - | 1.8 | 1  | 13.6 | 10.2  | 10.12 | 7.24 | 30.5 | 0.02 | 90.3  | - | 411 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 14 | 0.2  | 10.64 | 16.29 | 7.11 | 33.8 | 0.02 | 108.2 | - | 442 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 13 | 0.9  | 10.66 | 16.23 | 7.1  | 33.9 | 0.02 | 108.3 | - | 443 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 12 | 1.9  | 10.66 | 16.04 | 7.08 | 33.8 | 0.02 | 107.8 | - | 444 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 11 | 2.9  | 10.63 | 15.71 | 7.07 | 33.6 | 0.02 | 106.7 | - | 444 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 10 | 3.9  | 10.7  | 15.04 | 7.05 | 33.5 | 0.02 | 105.9 | - | 446 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 9  | 4.9  | 10.78 | 14.13 | 7.02 | 33.1 | 0.02 | 104.6 | - | 447 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 8  | 6    | 10.78 | 13.55 | 7    | 33.1 | 0.02 | 103.3 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 7  | 7    | 10.77 | 13.29 | 7    | 32.9 | 0.02 | 102.6 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 6  | 8    | 10.73 | 13.09 | 6.98 | 32.8 | 0.02 | 101.7 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 5  | 8.9  | 10.78 | 12.96 | 6.96 | 32.8 | 0.02 | 101.9 | - | 448 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 4  | 10   | 10.76 | 12.86 | 6.94 | 32.8 | 0.02 | 101.5 | - | 449 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 3  | 11   | 10.69 | 12.77 | 6.91 | 32.8 | 0.02 | 100.7 | - | 450 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 2  | 12   | 10.55 | 12.56 | 6.87 | 32.9 | 0.02 | 98.9  | - | 451 |
| Rockford Bay | 2597 | DBSS | 6/26/97 | - | 3.3 | 1  | 12.9 | 9.92  | 11.23 | 6.81 | 34   | 0.02 | 90.1  | - | 454 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 14 | 0.4  | 10.09 | 17.81 | 7.28 | 36.2 | 0.02 | ****  | - | 381 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 13 | 1.2  | 10.26 | 16.91 | 7.27 | 36.1 | 0.02 | ****  | - | 381 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 12 | 2.2  | 10.33 | 16.58 | 7.21 | 36.2 | 0.02 | ****  | - | 382 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 11 | 3.2  | 10.35 | 16.38 | 7.17 | 36.1 | 0.02 | ****  | - | 383 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 10 | 4.2  | 10.37 | 15.79 | 7.15 | 36.1 | 0.02 | ****  | - | 383 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 9  | 5.2  | 10.36 | 15.4  | 7.09 | 35.8 | 0.02 | ****  | - | 384 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 8  | 6.2  | 10.3  | 14.94 | 7.05 | 35.7 | 0.02 | ****  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 7  | 7.2  | 10.01 | 13.61 | 7.02 | 34.8 | 0.02 | 96.9  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 6  | 8.2  | 9.88  | 13.17 | 7    | 34.9 | 0.02 | 94.7  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 5  | 9.2  | 9.74  | 12.26 | 6.99 | 34.6 | 0.02 | 91.5  | - | 385 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 4  | 10.2 | 9.68  | 11.16 | 7    | 35.2 | 0.02 | 88.6  | - | 384 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 3  | 11.1 | 9.79  | 10.66 | 7.03 | 35.5 | 0.02 | 88.5  | - | 382 |
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 2  | 12.3 | 9.81  | 10.31 | 7.04 | 35.9 | 0.02 | 88    | - | 380 |



|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 2797 | DBSS | 7/9/97  | - | 4.5 | 1  | 13.2 | 10    | 10.02 | 7.09 | 36.4 | 0.02 | 89.1  | - | 377 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 14 | 0.2  | 10.78 | 19.13 | 7.56 | 37.2 | 0.02 | 109.3 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 13 | 0.9  | 10.82 | 18.87 | 7.56 | 37.1 | 0.02 | 109.2 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 12 | 1.9  | 10.96 | 18.2  | 7.56 | 36.8 | 0.02 | 109.1 | - | 398 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 11 | 2.8  | 11.01 | 18.05 | 7.57 | 36.8 | 0.02 | 109.3 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 10 | 3.9  | 11.02 | 18.03 | 7.53 | 36.7 | 0.02 | 109.3 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 9  | 5    | 11    | 18.01 | 7.48 | 36.7 | 0.02 | 109.1 | - | 395 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 8  | 5.9  | 11.01 | 17.9  | 7.44 | 36.8 | 0.02 | 109   | - | 395 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 7  | 6.9  | 10.99 | 17.74 | 7.38 | 36.8 | 0.02 | 108.5 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 6  | 7.9  | 10.87 | 17.49 | 7.32 | 36.8 | 0.02 | 106.7 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 5  | 8.9  | 11    | 16.68 | 7.24 | 36.6 | 0.02 | 106.1 | - | 395 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 4  | 9.9  | 10.77 | 16.38 | 7.16 | 36.5 | 0.02 | 103.2 | - | 396 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 3  | 10.9 | 10.68 | 15.74 | 7.08 | 36.3 | 0.02 | 101   | - | 394 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 2  | 11.9 | 10.59 | 15.24 | 7    | 36.2 | 0.02 | 99.1  | - | 389 |
| Rockford Bay | 2997 | SSAS | 7/23/97 | - | 5   | 1  | 12.9 | 10.21 | 14.03 | 6.91 | 35.8 | 0.02 | 93    | - | 385 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 14 | 0.5  | 9.03  | 23.66 | 7.42 | 41.2 | 0.03 | 105.1 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 13 | 1.1  | 9.11  | 23.37 | 7.42 | 41   | 0.03 | 105.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 12 | 2.1  | 9.24  | 23.08 | 7.39 | 40.8 | 0.03 | 106.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 11 | 3.1  | 9.29  | 22.99 | 7.36 | 40.6 | 0.03 | 106.5 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 10 | 4.1  | 9.3   | 22.86 | 7.34 | 40.4 | 0.03 | 106.7 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 9  | 5.1  | 9.34  | 22.66 | 7.29 | 40.3 | 0.03 | 106.8 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 8  | 6.1  | 9.53  | 21.58 | 7.27 | 39.4 | 0.03 | 106.7 | - | 365 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 7  | 7.1  | 9.52  | 21.44 | 7.14 | 39.5 | 0.03 | 106.3 | - | 369 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 6  | 8.1  | 9.85  | 19.73 | 7.05 | 37.9 | 0.02 | 106.4 | - | 372 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 5  | 9.1  | 9.99  | 15.56 | 6.85 | 34.6 | 0.02 | 99    | - | 379 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 4  | 10.1 | 9.29  | 12.25 | 6.69 | 34   | 0.02 | 85.5  | - | 382 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 3  | 11.1 | 9.15  | 10.74 | 6.63 | 34.4 | 0.02 | 81.3  | - | 383 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 2  | 12.1 | 9.18  | 10.3  | 6.61 | 34.8 | 0.02 | 80.8  | - | 381 |
| Rockford Bay | 3197 | DBSS | 8/5/97  | - | 7.7 | 1  | 13.1 | 9.08  | 10.02 | 6.57 | 35.4 | 0.02 | 79.4  | - | 385 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 14 | 0.3  | 8.54  | 22.49 | 7.43 | 42.7 | 0.03 | 97.9  | - | 378 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 13 | 1.7  | 8.55  | 22.45 | 7.37 | 42.7 | 0.03 | 97.9  | - | 379 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 12 | 2.7  | 8.65  | 22.25 | 7.35 | 41.9 | 0.03 | 98.8  | - | 380 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 11 | 3.7  | 8.68  | 22.06 | 7.33 | 41.1 | 0.03 | 98.7  | - | 379 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 10 | 4.7  | 8.67  | 21.88 | 7.28 | 41.2 | 0.03 | 98.2  | - | 381 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 9  | 5.7  | 8.62  | 21.69 | 7.19 | 41.3 | 0.03 | 97.3  | - | 384 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 8  | 6.7  | 8.64  | 21.35 | 7.13 | 41.3 | 0.03 | 97    | - | 385 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 7  | 7.7  | 8.72  | 20.58 | 7.04 | 39.5 | 0.03 | 96.3  | - | 388 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 6  | 8.7  | 9.23  | 17.2  | 6.92 | 36.6 | 0.02 | 95.2  | - | 395 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 5  | 9.7  | 9.25  | 16.24 | 6.87 | 35.9 | 0.02 | 93.6  | - | 397 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 4  | 10.7 | 9.12  | 14.81 | 6.84 | 35.3 | 0.02 | 89.4  | - | 398 |
| Rockford Bay | 3297 | DBSS | 8/14/97 | - | 8   | 3  | 11.7 | 9.02  | 13.12 | 6.82 | 34.9 | 0.02 | 85.2  | - | 399 |

|              |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Rockford Bay | 3297 | DBSS | 8/14/97  | - | 8   | 2  | 12.8 | 8.61  | 12.44 | 6.79 | 35.3 | 0.02 | 80.2  | - | 401 |
| Rockford Bay | 3297 | DBSS | 8/14/97  | - | 8   | 1  | 13.7 | 8.22  | 10.67 | 6.81 | 36.1 | 0.02 | 73.5  | - | 401 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 14 | 0.3  | 9.16  | 21.23 | 7.43 | 47.7 | 0.03 | 102.6 | - | 372 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 13 | 1.2  | 9.17  | 21.23 | 7.39 | 47.6 | 0.03 | 102.7 | - | 373 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 12 | 2.2  | 9.19  | 21.16 | 7.37 | 47.8 | 0.03 | 102.9 | - | 374 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 11 | 3.1  | 9.23  | 21.11 | 7.35 | 47.8 | 0.03 | 103.1 | - | 374 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 10 | 4.1  | 9.24  | 21.06 | 7.31 | 47.9 | 0.03 | 103.4 | - | 375 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 9  | 5.2  | 9.21  | 21    | 7.26 | 47.7 | 0.03 | 102.8 | - | 377 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 8  | 6.1  | 9.19  | 20.91 | 7.18 | 47.8 | 0.03 | 102.3 | - | 380 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 7  | 7.2  | 9.17  | 20.76 | 7.09 | 47.7 | 0.03 | 101.8 | - | 383 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 6  | 8.2  | 9.26  | 18.17 | 6.96 | 42.3 | 0.03 | 97.6  | - | 389 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 5  | 9.2  | 9.23  | 16.83 | 6.9  | 40.5 | 0.03 | 94.8  | - | 392 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 4  | 10.2 | 9.09  | 15.66 | 6.82 | 39.7 | 0.03 | 90.9  | - | 396 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 3  | 11.3 | 8.94  | 14.06 | 6.79 | 39.1 | 0.03 | 86.3  | - | 398 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 2  | 12.3 | 8.74  | 11.82 | 6.76 | 39   | 0.03 | 80.4  | - | 400 |
| Rockford Bay | 3497 | DBSS | 8/27/97  | - | 8.5 | 1  | 13.2 | 8.8   | 11.23 | 6.78 | 39.2 | 0.03 | 79.8  | - | 400 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 13 | 0.2  | 9.33  | 16.91 | 7.31 | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 12 | 1.8  | 9.34  | 16.9  | 7.31 | 48.5 | 0.03 | 97.5  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 11 | 2.8  | 9.33  | 16.89 | 7.31 | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 10 | 3.8  | 9.33  | 16.86 | 7.3  | 48.4 | 0.03 | 97.3  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 9  | 4.8  | 9.35  | 16.85 | 7.31 | 48.4 | 0.03 | 97.5  | - | 391 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 8  | 5.8  | 9.33  | 16.83 | 7.29 | 48.5 | 0.03 | 97.2  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 7  | 6.8  | 9.34  | 16.82 | 7.29 | 48.3 | 0.03 | 97.4  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 6  | 7.8  | 9.35  | 16.8  | 7.28 | 48.5 | 0.03 | 97.6  | - | 392 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 5  | 8.8  | 9.34  | 16.63 | 7.27 | 47.7 | 0.03 | 96.9  | - | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 4  | 9.8  | 9.31  | 16.6  | 7.25 | 47.6 | 0.03 | 96.5  | - | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 3  | 10.8 | 9.37  | 16.55 | 7.25 | 47.4 | 0.03 | 97.1  | - | 394 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 2  | 11.8 | 9.34  | 16.36 | 7.26 | 46.8 | 0.03 | 96.5  | - | 393 |
| Rockford Bay | 3797 | ASRP | 9/17/97  | - | 8   | 1  | 12.8 | 9.38  | 16.33 | 7.25 | 46.6 | 0.03 | 96.8  | - | 393 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 10 | 0.2  | 9.72  | 15.81 | 7.19 | 43.5 | 0.03 | 99.1  | - | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 9  | 1    | 9.68  | 15.72 | 7.18 | 43.6 | 0.03 | 98.5  | - | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 8  | 2.5  | 9.66  | 15.67 | 7.15 | 43.6 | 0.03 | 98.2  | - | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 7  | 4    | 9.65  | 15.61 | 7.12 | 43.6 | 0.03 | 98    | - | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 6  | 5.5  | 9.58  | 15.54 | 7.1  | 43.5 | 0.03 | 97.1  | - | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 5  | 7    | 9.52  | 15.36 | 7.06 | 43.6 | 0.03 | 96    | - | 448 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 4  | 8.5  | 9.48  | 15.19 | 7.04 | 43.6 | 0.03 | 95.4  | - | 447 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 3  | 10   | 9.44  | 15.13 | 6.98 | 43.6 | 0.03 | 94.9  | - | 446 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 2  | 11.3 | 9.51  | 14.94 | 6.95 | 43.6 | 0.03 | 95.1  | - | 444 |
| Rockford Bay | 3997 | ASRA | 9/29/97  | - | 8.2 | 1  | 12.8 | 9.42  | 14.71 | 6.89 | 43.5 | 0.03 | 93.8  | - | 441 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 12 | 0.4  | 10.03 | 12.41 | 7.16 | 48.2 | 0.03 | 92.9  | - | 396 |

|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|--------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 11 | 1.8  | 10.02 | 12.41 | 7.13 | 48.2 | 0.03 | 92.9 | -   | 398 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 10 | 2.8  | 10.02 | 12.41 | 7.13 | 48.2 | 0.03 | 92.9 | -   | 397 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 9  | 3.8  | 10.02 | 12.41 | 7.12 | 48.2 | 0.03 | 92.9 | -   | 398 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 8  | 4.8  | 10.02 | 12.41 | 7.09 | 48.3 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 7  | 5.9  | 10.02 | 12.41 | 7.09 | 48.2 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 6  | 6.8  | 10.02 | 12.41 | 7.07 | 48.3 | 0.03 | 92.9 | -   | 399 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 5  | 7.8  | 10.04 | 12.41 | 7.04 | 48.3 | 0.03 | 93   | -   | 400 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 4  | 8.8  | 10.04 | 12.41 | 7.02 | 48.3 | 0.03 | 92.9 | -   | 400 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 3  | 9.8  | 10.02 | 12.41 | 7    | 48.2 | 0.03 | 92.9 | -   | 401 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 2  | 11.8 | 10.03 | 12.4  | 6.93 | 48.2 | 0.03 | 92.9 | -   | 402 |
| Rockford Bay | 4297 | DBAS | 10/21/97 | - | 7.5 | 1  | 12.8 | 10.05 | 12.38 | 6.88 | 48.3 | 0.03 | 93   | -   | 404 |
|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 13 | 0.3  | 10.68 | 10.28 | 7.1  | 46.8 | 0.03 | 94.4 | 39  | 420 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 12 | 1.7  | 10.69 | 10.15 | 7.07 | 46.9 | 0.03 | 94.1 | 41  | 423 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 11 | 2.7  | 10.66 | 10.14 | 7.08 | 46.9 | 0.03 | 93.9 | 46  | 422 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 10 | 3.8  | 10.67 | 10.12 | 7.07 | 47   | 0.03 | 93.9 | 122 | 423 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 9  | 4.7  | 10.73 | 10.09 | 7.04 | 46.8 | 0.03 | 94.4 | 50  | 425 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 8  | 5.8  | 10.66 | 10.09 | 7.05 | 46.9 | 0.03 | 93.8 | 49  | 424 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 7  | 6.7  | 10.66 | 10.07 | 7    | 46.9 | 0.03 | 93.7 | 101 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 6  | 7.7  | 10.66 | 10.07 | 7    | 46.9 | 0.03 | 93.7 | 108 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 5  | 8.7  | 10.66 | 10.05 | 7.01 | 47   | 0.03 | 93.7 | 124 | 427 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 4  | 9.7  | 10.72 | 10    | 6.99 | 47   | 0.03 | 94   | 101 | 428 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 3  | 10.7 | 10.75 | 10    | 6.95 | 47.2 | 0.03 | 94.3 | 42  | 431 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 2  | 11.7 | 10.71 | 9.99  | 6.96 | 47.2 | 0.03 | 93.9 | 53  | 430 |
| Rockford Bay | 4497 | DBAS | 11/4/97  | - | 7.2 | 1  | 12.6 | 10.66 | 9.99  | 6.88 | 47.2 | 0.03 | 93.5 | 730 | 434 |

| Location          | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 10       | 0.2       | 12.97                   | 5.69            | 7.39 | 42.6                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 9        | 1.4       | 12.96                   | 5.67            | 7.39 | 42.7                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 8        | 2.5       | 12.99                   | 5.62            | 7.38 | 42.7                 | 0.03 | *****                      | -                  | 389   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 7        | 3.7       | 12.97                   | 5.6             | 7.37 | 42.9                 | 0.03 | *****                      | -                  | 389   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 6        | 5.2       | 12.93                   | 5.52            | 7.37 | 43.1                 | 0.03 | *****                      | -                  | 389   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 5        | 6.5       | 12.95                   | 5.52            | 7.36 | 43.2                 | 0.03 | *****                      | -                  | 390   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 4        | 7.8       | 12.95                   | 5.47            | 7.36 | 42.8                 | 0.03 | *****                      | -                  | 390   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 3        | 9.1       | 12.91                   | 5.46            | 7.36 | 43.1                 | 0.03 | *****                      | -                  | 390   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 2        | 10.5      | 12.89                   | 5.42            | 7.35 | 43.3                 | 0.03 | *****                      | -                  | 391   |
| Windy Bay Shallow | 1597  | ASSS     | 4/18/97 | -    | 1.4        | 1        | 11.9      | 12.85                   | 5.22            | 7.36 | 43.7                 | 0.03 | *****                      | -                  | 392   |
|                   |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Windy Bay Shallow | 1997  | DBSS     | 5/16/97 | -    | 1.75       | 13       | 0.2       | 11.48                   | 12.43           | 7.03 | 21                   | 0.01 | 106.8                      | -                  | 420   |
| Windy Bay Shallow | 1997  | DBSS     | 5/16/97 | -    | 1.75       | 12       | 1.1       | 11.6                    | 11.46           | 7    | 36.7                 | 0.02 | 105.4                      | -                  | 427   |
| Windy Bay Shallow | 1997  | DBSS     | 5/16/97 | -    | 1.75       | 11       | 2.6       | 11.6                    | 10.94           | 6.95 | 36                   | 0.02 | 104.2                      | -                  | 431   |
| Windy Bay Shallow | 1997  | DBSS     | 5/16/97 | -    | 1.75       | 10       | 4.1       | 11.39                   | 10.15           | 6.9  | 34.8                 | 0.02 | 100.4                      | -                  | 433   |

|                   |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 9  | 5.6  | 11.34 | 9.74  | 6.89 | 35.3 | 0.02 | 99    | - | 433 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 8  | 7.1  | 11.39 | 9.53  | 6.86 | 36.5 | 0.02 | 99    | - | 434 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 7  | 8.6  | 11.4  | 9.15  | 6.84 | 38.2 | 0.02 | 98.2  | - | 434 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 6  | 10.1 | 11.38 | 8.69  | 6.81 | 38.7 | 0.02 | 97.1  | - | 435 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 5  | 11.6 | 11.36 | 8.17  | 6.8  | 40.8 | 0.03 | 95.5  | - | 436 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 4  | 13.1 | 11.34 | 8.04  | 6.77 | 41.4 | 0.03 | 95    | - | 436 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 3  | 14.6 | 11.45 | 7.47  | 6.76 | 43.4 | 0.03 | 94.6  | - | 437 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 2  | 16.1 | 11.4  | 7.19  | 6.73 | 43.4 | 0.03 | 93.5  | - | 437 |
| Windy Bay Shallow | 1997 | DBSS | 5/16/97 | - | 1.75 | 1  | 17.6 | 11.5  | 7.02  | 6.69 | 43.6 | 0.03 | 94    | - | 436 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 16 | 0.5  | 10.7  | 16.3  | 7.03 | 31.5 | 0.02 | 108.5 | - | 382 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 15 | 1.5  | 10.69 | 16.23 | 7.01 | 31.5 | 0.02 | 108.3 | - | 382 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 14 | 2.4  | 10.74 | 15.76 | 6.99 | 31.5 | 0.02 | 107.7 | - | 383 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 13 | 3.5  | 10.76 | 14.11 | 6.93 | 30.9 | 0.02 | 104.1 | - | 386 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 12 | 4.5  | 9.97  | 12.69 | 6.86 | 31.2 | 0.02 | 93.5  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 11 | 5.6  | 10.42 | 11.36 | 6.88 | 29.9 | 0.02 | 94.7  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 10 | 6.5  | 10.43 | 10.68 | 6.87 | 29.5 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 9  | 7.4  | 10.48 | 10.5  | 6.87 | 29.4 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 8  | 8.5  | 10.52 | 10.33 | 6.87 | 29.6 | 0.02 | 93.3  | - | 389 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 7  | 9.5  | 10.5  | 9.97  | 6.85 | 29.4 | 0.02 | 92.4  | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 6  | 10.6 | 10.45 | 9.09  | 6.85 | 30.4 | 0.02 | 90    | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 5  | 11.5 | 10.37 | 8.97  | 6.84 | 30.6 | 0.02 | 89.1  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 4  | 12.6 | 10.26 | 8.83  | 6.84 | 31.1 | 0.02 | 87.8  | - | 390 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 3  | 13.6 | 10.34 | 8.61  | 6.84 | 31.7 | 0.02 | 88.1  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 2  | 14.6 | 10.31 | 8.42  | 6.84 | 32.7 | 0.02 | 87.4  | - | 391 |
| Windy Bay Shallow | 2197 | DBSS | 5/29/97 | - | 1.5  | 1  | 15.5 | 10.34 | 8.27  | 6.84 | 33.3 | 0.02 | 87.3  | - | 391 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 16 | 0.5  | 10.8  | 13.95 | 7.02 | 34.9 | 0.02 | 104.5 | - | 423 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 15 | 1.4  | 10.67 | 12.79 | 7.01 | 34.6 | 0.02 | 100.6 | - | 425 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 14 | 2.3  | 10.89 | 11.64 | 6.98 | 32.6 | 0.02 | 100   | - | 426 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 13 | 3.3  | 10.84 | 11.1  | 6.98 | 32   | 0.02 | 98.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 12 | 4.4  | 10.74 | 10.53 | 6.98 | 32.5 | 0.02 | 96.1  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 11 | 5.3  | 10.78 | 10.22 | 6.97 | 30.8 | 0.02 | 95.8  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 10 | 6.3  | 10.69 | 10.04 | 6.97 | 30.5 | 0.02 | 94.7  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 9  | 7.5  | 10.66 | 9.96  | 6.96 | 30.6 | 0.02 | 94.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 8  | 8.3  | 10.58 | 9.89  | 6.96 | 31   | 0.02 | 93.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 7  | 9.5  | 10.63 | 9.66  | 6.96 | 30.9 | 0.02 | 93.2  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 6  | 10.4 | 10.6  | 9.63  | 6.96 | 31.4 | 0.02 | 92.8  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 5  | 11.6 | 10.61 | 9.5   | 6.96 | 31.5 | 0.02 | 92.7  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 4  | 12.5 | 10.61 | 9.32  | 6.95 | 32   | 0.02 | 92.2  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 3  | 13.6 | 10.56 | 9.1   | 6.95 | 33   | 0.02 | 91.3  | - | 427 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 2  | 14.7 | 10.65 | 8.68  | 6.95 | 34.5 | 0.02 | 91.1  | - | 428 |
| Windy Bay Shallow | 2397 | DBAS | 6/11/97 | - | 2    | 1  | 15.7 | 10.64 | 8.66  | 6.96 | 34.7 | 0.02 | 91    | - | 427 |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 16 | 0.4  | 10.71 | 14.84 | 7.13 | 35.5 | 0.02 | 105.6 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 15 | 0.8  | 10.71 | 14.83 | 7.14 | 35.4 | 0.02 | 105.5 | - | 438 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 14 | 1.7  | 10.72 | 14.74 | 7.11 | 35.4 | 0.02 | 105.5 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 13 | 2.7  | 10.74 | 14.21 | 7.11 | 35.4 | 0.02 | 104.4 | - | 440 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 12 | 3.7  | 10.81 | 13.65 | 7.09 | 34.9 | 0.02 | 103.8 | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 11 | 4.7  | 10.83 | 13.22 | 7.08 | 34.3 | 0.02 | 103   | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 10 | 5.6  | 10.82 | 12.61 | 7.08 | 34   | 0.02 | 101.4 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 9  | 6.5  | 10.8  | 12.73 | 7.07 | 34   | 0.02 | 101.4 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 8  | 7.9  | 10.81 | 12.12 | 7.05 | 33.5 | 0.02 | 100.2 | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 7  | 8.5  | 10.77 | 12.05 | 7.05 | 33.6 | 0.02 | 99.8  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 6  | 9.7  | 10.81 | 11.61 | 7.04 | 33.5 | 0.02 | 99.1  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 5  | 10.6 | 10.82 | 11.56 | 7.04 | 33.4 | 0.02 | 99.1  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 4  | 11.5 | 10.79 | 11.41 | 7.03 | 33.7 | 0.02 | 98.4  | - | 442 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 3  | 12.5 | 10.78 | 11.2  | 7.03 | 33.7 | 0.02 | 97.9  | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 2  | 13.6 | 10.81 | 11.12 | 7.03 | 33.7 | 0.02 | 98    | - | 441 |
| Windy Bay Shallow | 2597 | DBSS | 6/26/97 | - | 2.2 | 1  | 14.8 | 10.78 | 11.02 | 7.05 | 33.5 | 0.02 | 97.5  | - | 440 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 16 | 0.4  | 10.28 | 16.6  | 7.25 | 37.3 | 0.02 | ****  | - | 411 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 15 | 1.1  | 10.28 | 16.59 | 7.24 | 37.3 | 0.02 | ****  | - | 412 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 14 | 2    | 10.29 | 16.55 | 7.2  | 37.3 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 13 | 2.9  | 10.28 | 16.53 | 7.18 | 37.3 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 12 | 3.9  | 10.28 | 16.32 | 7.15 | 37.4 | 0.02 | ****  | - | 414 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 11 | 5    | 10.39 | 15.56 | 7.12 | 36.7 | 0.02 | ****  | - | 416 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 10 | 6    | 10.4  | 15.31 | 7.1  | 36.5 | 0.02 | ****  | - | 416 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 9  | 7    | 10.35 | 15.03 | 7.06 | 36.8 | 0.02 | ****  | - | 417 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 8  | 8    | 10.35 | 14.51 | 7.06 | 36   | 0.02 | ****  | - | 418 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 7  | 9    | 10.3  | 14.21 | 7.03 | 36.3 | 0.02 | ****  | - | 418 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 6  | 10.1 | 10.19 | 13.65 | 6.99 | 35.2 | 0.02 | 98.7  | - | 419 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 5  | 11   | 10.04 | 13.32 | 6.94 | 35.3 | 0.02 | 96.5  | - | 420 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 4  | 12   | 9.95  | 12.79 | 6.93 | 34.7 | 0.02 | 94.5  | - | 421 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 3  | 13   | 9.66  | 11.92 | 6.9  | 34.9 | 0.02 | 90    | - | 423 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 2  | 14   | 9.5   | 10.12 | 6.9  | 36.4 | 0.02 | 84.8  | - | 424 |
| Windy Bay Shallow | 2797 | DBSS | 7/9/97  | - | 3.5 | 1  | 15   | 9.57  | 9.74  | 6.94 | 37.4 | 0.02 | 84.7  | - | 422 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 10 | 1.6  | 10.61 | 19.64 | 7.58 | 39.1 | 0.03 | 108.7 | - | 437 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 9  | 3    | 10.74 | 19.09 | 7.58 | 38.8 | 0.02 | 108.9 | - | 437 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 8  | 4.7  | 11.09 | 18.2  | 7.53 | 37.9 | 0.02 | 110.4 | - | 439 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 7  | 6.1  | 11.2  | 17.22 | 7.43 | 37.5 | 0.02 | 109.3 | - | 442 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 6  | 7.6  | 11.19 | 16.86 | 7.34 | 37.3 | 0.02 | 108.3 | - | 444 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 5  | 9    | 11.06 | 16.59 | 7.26 | 37.2 | 0.02 | 106.5 | - | 445 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 4  | 10.5 | 10.86 | 15.92 | 7.17 | 36.6 | 0.02 | 103.1 | - | 447 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 3  | 12   | 10.67 | 15.4  | 7.05 | 36.6 | 0.02 | 100   | - | 449 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 2  | 13.5 | 9.73  | 14.29 | 6.93 | 36.9 | 0.02 | 89.2  | - | 453 |
| Windy Bay Shallow | 2997 | SSJD | 7/23/97 | - | 4.5 | 1  | 15.5 | 9.38  | 11.93 | 6.87 | 36.7 | 0.02 | 81.7  | - | 455 |

|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 12 | 0.4  | 8.97 | 24.09 | 7.47 | 42.1 | 0.03 | 105.3 | - | 369 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 11 | 1    | 8.97 | 24.06 | 7.47 | 42   | 0.03 | 105.3 | - | 370 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 10 | 2.5  | 9.06 | 23.78 | 7.42 | 42.1 | 0.03 | 105.8 | - | 372 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 9  | 4    | 9.06 | 23.38 | 7.4  | 42   | 0.03 | 105   | - | 373 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 8  | 5.5  | 9.39 | 22.43 | 7.35 | 41.3 | 0.03 | 106.8 | - | 374 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 7  | 7    | 9.93 | 20.42 | 7.23 | 39.5 | 0.03 | 108.6 | - | 378 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 6  | 8.5  | 10   | 17.76 | 7.11 | 36.1 | 0.02 | 103.5 | - | 383 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 5  | 10   | 9.48 | 14.74 | 6.98 | 34.9 | 0.02 | 92.2  | - | 389 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 4  | 11.5 | 9.06 | 12.82 | 6.95 | 34.6 | 0.02 | 84.5  | - | 392 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 3  | 13   | 8.88 | 10.21 | 6.94 | 35.4 | 0.02 | 78    | - | 394 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 2  | 14.5 | 8.99 | 9.33  | 6.97 | 36.2 | 0.02 | 77.3  | - | 394 |
| Windy Bay Shallow | 3197 | DBSS | 8/5/97  | - | 6.8 | 1  | 16   | 8.8  | 8.9   | 7.01 | 37.6 | 0.02 | 75.1  | - | 394 |
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 14 | 0.2  | 8.65 | 22.04 | 7.52 | 41.5 | 0.03 | 98.3  | - | 378 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 13 | 1.6  | 8.65 | 22.01 | 7.41 | 41.5 | 0.03 | 98.3  | - | 381 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 12 | 2.6  | 8.64 | 22.02 | 7.4  | 41.6 | 0.03 | 98.2  | - | 381 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 11 | 3.6  | 8.64 | 21.99 | 7.32 | 41.6 | 0.03 | 98.2  | - | 384 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 10 | 4.6  | 8.64 | 21.92 | 7.26 | 41.6 | 0.03 | 98    | - | 386 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 9  | 5.5  | 8.64 | 21.78 | 7.19 | 41.5 | 0.03 | 97.7  | - | 388 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 8  | 6.6  | 8.6  | 21.32 | 7.05 | 41.6 | 0.03 | 96.4  | - | 394 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 7  | 7.6  | 8.99 | 16.09 | 7.02 | 35.7 | 0.02 | 90.6  | - | 398 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 6  | 8.5  | 8.96 | 15.22 | 7.01 | 35.6 | 0.02 | 88.8  | - | 399 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 5  | 9.6  | 8.77 | 14.04 | 7.01 | 35.3 | 0.02 | 84.7  | - | 400 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 4  | 10.5 | 8.59 | 12.81 | 7.03 | 35   | 0.02 | 80.6  | - | 400 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 3  | 11.5 | 8.43 | 12.08 | 7.07 | 35.2 | 0.02 | 77.9  | - | 399 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 2  | 12.6 | 8.19 | 11.67 | 7.14 | 35.7 | 0.02 | 74.9  | - | 396 |
| Windy Bay Shallow | 3297 | DBSS | 8/14/97 | - | 4.5 | 1  | 13.7 | 8.03 | 10.82 | 7.13 | 36.3 | 0.02 | 72    | - | 398 |
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 12 | 0.4  | 9.26 | 20.74 | 7.51 | 46.5 | 0.03 | 102.8 | - | 368 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 11 | 1    | 9.27 | 20.76 | 7.5  | 46.4 | 0.03 | 102.9 | - | 368 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 10 | 2.5  | 9.26 | 20.74 | 7.48 | 46.5 | 0.03 | 102.8 | - | 369 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 9  | 3.9  | 9.27 | 20.74 | 7.45 | 46.4 | 0.03 | 102.8 | - | 370 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 8  | 5.5  | 9.23 | 20.7  | 7.38 | 46.6 | 0.03 | 102.4 | - | 372 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 7  | 7.1  | 9.21 | 20.65 | 7.27 | 46.8 | 0.03 | 102   | - | 377 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 6  | 8.6  | 9.02 | 17.29 | 6.94 | 39.9 | 0.03 | 93.4  | - | 389 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 5  | 10   | 9.19 | 14.11 | 6.9  | 38.6 | 0.02 | 88.9  | - | 392 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 4  | 11.5 | 8.86 | 12.31 | 6.86 | 38.7 | 0.02 | 82.3  | - | 395 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 3  | 13.1 | 8.52 | 10.95 | 6.86 | 39.6 | 0.03 | 76.7  | - | 396 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 2  | 14.4 | 8.44 | 9.36  | 6.9  | 41.3 | 0.03 | 73.2  | - | 396 |
| Windy Bay Shallow | 3497 | DBSS | 8/27/97 | - | 8   | 1  | 16.1 | 8.59 | 8.64  | 7.06 | 42.5 | 0.03 | 73.2  | - | 392 |
|                   |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97 | - | 8.1 | 16 | 0.2  | 9.32 | 17.01 | 7.27 | 47.5 | 0.03 | 97.5  | - | 384 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97 | - | 8.1 | 15 | 1.7  | 9.32 | 17.01 | 7.26 | 47.6 | 0.03 | 97.5  | - | 385 |

|                   |      |      |          |   |     |    |      |       |       |      |      |      |      |   |     |
|-------------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|---|-----|
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 14 | 2.7  | 9.32  | 17    | 7.24 | 47.6 | 0.03 | 97.5 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 13 | 3.7  | 9.3   | 17.01 | 7.24 | 47.5 | 0.03 | 97.3 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 12 | 4.7  | 9.3   | 16.98 | 7.23 | 47.6 | 0.03 | 97.2 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 11 | 5.7  | 9.31  | 16.96 | 7.22 | 47.6 | 0.03 | 97.2 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 10 | 6.7  | 9.29  | 16.91 | 7.19 | 47.3 | 0.03 | 97   | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 9  | 7.7  | 9.28  | 16.81 | 7.18 | 47.2 | 0.03 | 96.8 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 8  | 8.7  | 9.29  | 16.75 | 7.15 | 47.1 | 0.03 | 96.8 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 7  | 9.7  | 9.31  | 16.58 | 7.13 | 46.5 | 0.03 | 96.5 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 6  | 10.7 | 9.33  | 16.52 | 7.1  | 46.3 | 0.03 | 96.6 | - | 385 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 5  | 11.7 | 9.31  | 16.48 | 7.07 | 46.1 | 0.03 | 96.3 | - | 386 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 4  | 12.7 | 9.3   | 16.38 | 7.05 | 45.9 | 0.03 | 96   | - | 386 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 3  | 13.7 | 9.24  | 16.04 | 7    | 45.6 | 0.03 | 94.7 | - | 387 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 2  | 14.7 | 9.23  | 15.25 | 6.96 | 43.7 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 3797 | ASRP | 9/17/97  | - | 8.1 | 1  | 15.7 | 9.13  | 15.04 | 6.94 | 43.9 | 0.03 | 91.8 | - | 388 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 11 | 0.3  | 9.65  | 14.71 | 7.17 | 46.5 | 0.03 | 96.1 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 10 | 2    | 9.62  | 14.67 | 7.18 | 46.6 | 0.03 | 95.7 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 9  | 3.5  | 9.61  | 14.64 | 7.17 | 46.6 | 0.03 | 95.5 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 8  | 5    | 9.59  | 14.61 | 7.17 | 46.6 | 0.03 | 95.3 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 7  | 6.5  | 9.55  | 14.54 | 7.16 | 46.9 | 0.03 | 94.8 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 6  | 8    | 9.54  | 14.52 | 7.15 | 46.9 | 0.03 | 94.7 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 5  | 9.5  | 9.5   | 14.44 | 7.15 | 46.5 | 0.03 | 94   | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 4  | 11   | 9.52  | 14.41 | 7.15 | 46.9 | 0.03 | 94.2 | - | 448 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 3  | 12.5 | 9.47  | 14.27 | 7.15 | 46.6 | 0.03 | 93.7 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 2  | 14   | 9.52  | 14.24 | 7.15 | 46.9 | 0.03 | 93.8 | - | 447 |
| Windy Bay Shallow | 3997 | ASRA | 9/29/97  | - | 7.5 | 1  | 15.2 | 9.48  | 14.16 | 7.15 | 46.8 | 0.03 | 93.2 | - | 447 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 16 | 0.3  | 10.06 | 12.32 | 7.24 | 49.9 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 15 | 1    | 10.07 | 12.33 | 7.25 | 49.8 | 0.03 | 93.1 | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 14 | 2    | 10.05 | 12.33 | 7.25 | 49.8 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 13 | 3    | 10.05 | 12.33 | 7.24 | 49.9 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 12 | 4    | 10.06 | 12.32 | 7.24 | 49.9 | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 11 | 5    | 10.05 | 12.33 | 7.23 | 50   | 0.03 | 93   | - | 387 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 10 | 6    | 10.05 | 12.33 | 7.21 | 49.9 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 9  | 7    | 10.05 | 12.33 | 7.2  | 49.9 | 0.03 | 93   | - | 388 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 8  | 8    | 10.04 | 12.33 | 7.19 | 49.8 | 0.03 | 93   | - | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 7  | 9    | 10.06 | 12.32 | 7.18 | 49.9 | 0.03 | 93   | - | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 6  | 10   | 10.07 | 12.32 | 7.15 | 49.9 | 0.03 | 93.1 | - | 390 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 5  | 11   | 10.07 | 12.33 | 7.16 | 49.9 | 0.03 | 93.1 | - | 389 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 4  | 12   | 10.08 | 12.32 | 7.13 | 49.9 | 0.03 | 93.2 | - | 390 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 3  | 13   | 10.09 | 12.33 | 7.12 | 49.9 | 0.03 | 93.3 | - | 391 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 2  | 14   | 10.1  | 12.32 | 7.1  | 49.9 | 0.03 | 93.3 | - | 391 |
| Windy Bay Shallow | 4297 | DBAS | 10/21/97 | - | 6.5 | 1  | 15   | 10.11 | 12.3  | 7.06 | 50   | 0.03 | 93.5 | - | 392 |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |      |     |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 11 | 0.4  | 10.55 | 10.17 | 7.05 | 47.6 | 0.03 | 93   | 57  | 409 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 10 | 1.7  | 10.53 | 9.89  | 7.14 | 47.7 | 0.03 | 92.2 | 105 | 406 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 9  | 3.2  | 10.53 | 9.78  | 7.1  | 48.2 | 0.03 | 91.9 | 144 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 8  | 4.7  | 10.53 | 9.74  | 7.12 | 48.6 | 0.03 | 91.8 | 117 | 407 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 7  | 6.2  | 10.51 | 9.66  | 7.14 | 49.1 | 0.03 | 91.5 | 141 | 406 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 6  | 7.7  | 10.52 | 9.55  | 7.1  | 49.7 | 0.03 | 91.4 | 134 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 5  | 9.2  | 10.55 | 9.51  | 7.12 | 49.8 | 0.03 | 91.5 | 100 | 407 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 4  | 10.7 | 10.56 | 9.46  | 7.09 | 50   | 0.03 | 91.5 | 138 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 3  | 12.2 | 10.49 | 9.42  | 7.09 | 50.2 | 0.03 | 90.8 | 152 | 408 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 2  | 13.7 | 10.53 | 9.33  | 7.08 | 50.2 | 0.03 | 90.9 | 115 | 409 |
| Windy Bay Shallow | 4497 | DBAS | 11/4/97 | - | 5.8 | 1  | 15.2 | 10.46 | 9.33  | 7.09 | 50.3 | 0.03 | 90.3 | 601 | 409 |

| Location       | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|----------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 27       | 0.4       | 12.89                   | 5.6             | 7.38 | 43.4                 | 0.03 | *****                      | -                  | 385   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 26       | 1         | 12.89                   | 5.6             | 7.36 | 43.5                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 25       | 2         | 12.95                   | 5.57            | 7.37 | 43.4                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 24       | 3.2       | 12.96                   | 5.54            | 7.37 | 43.5                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 23       | 4.2       | 12.77                   | 5.54            | 7.36 | 43.5                 | 0.03 | *****                      | -                  | 386   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 22       | 5.5       | 12.88                   | 5.51            | 7.35 | 43.6                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 21       | 6.7       | 12.83                   | 5.5             | 7.35 | 43.5                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 20       | 7.8       | 12.98                   | 5.39            | 7.34 | 43.6                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 19       | 9         | 12.7                    | 5.34            | 7.33 | 43.5                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 18       | 10.1      | 12.87                   | 5.32            | 7.33 | 43.5                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 17       | 11.4      | 12.92                   | 5.3             | 7.33 | 43.5                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 16       | 12.8      | 12.72                   | 5.24            | 7.33 | 43.5                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 15       | 14        | 12.75                   | 5.22            | 7.32 | 43.4                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 14       | 15.3      | 12.86                   | 5.17            | 7.32 | 43.3                 | 0.03 | *****                      | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 13       | 16.5      | 12.62                   | 5.12            | 7.32 | 43.2                 | 0.03 | 99                         | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 12       | 17.8      | 12.61                   | 5.09            | 7.31 | 43.2                 | 0.03 | 99                         | -                  | 389   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 11       | 19        | 12.59                   | 5.04            | 7.31 | 43.2                 | 0.03 | 98.7                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 10       | 20.2      | 12.63                   | 5.06            | 7.31 | 43.2                 | 0.03 | 99                         | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 9        | 21.5      | 12.77                   | 4.99            | 7.31 | 43.2                 | 0.03 | 99.9                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 8        | 22.7      | 12.77                   | 4.96            | 7.3  | 43.3                 | 0.03 | 99.9                       | -                  | 389   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 7        | 24        | 12.84                   | 4.74            | 7.31 | 43.9                 | 0.03 | 99.9                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 6        | 25.3      | 12.7                    | 4.73            | 7.3  | 43.9                 | 0.03 | 98.8                       | -                  | 389   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 5        | 26.5      | 12.79                   | 4.74            | 7.3  | 43.9                 | 0.03 | 99.5                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 4        | 27.9      | 12.85                   | 4.74            | 7.31 | 43.9                 | 0.03 | 99.7                       | -                  | 387   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 3        | 29.1      | 12.85                   | 4.71            | 7.3  | 43.9                 | 0.03 | 99.8                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 2        | 30.5      | 12.71                   | 4.71            | 7.3  | 44                   | 0.03 | 98.8                       | -                  | 388   |
| Windy Bay Deep | 1597  | ASSS     | 4/18/97 | -    | 1.6        | 1        | 31.8      | 12.93                   | 4.74            | 7.31 | 43.9                 | 0.03 | *****                      | -                  | 387   |
| Windy Bay Deep | 1997  | DBSS     | 5/16/97 | -    | 1.9        | 19       | 0.2       | 11.55                   | 12.53           | 7.06 | 36.7                 | 0.02 | 107.6                      | -                  | 413   |



|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 18 | 1.8  | 11.42 | 10.91 | 7.05 | 34.6 | 0.02 | 102.4 | - | 417 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 17 | 3.8  | 11.3  | 9.97  | 7.01 | 34.4 | 0.02 | 99.2  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 16 | 5.8  | 11.31 | 9.48  | 7    | 35   | 0.02 | 98.2  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 15 | 7.8  | 11.3  | 9.09  | 6.98 | 35.3 | 0.02 | 97.2  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 14 | 9.8  | 11.44 | 8.3   | 6.98 | 37.5 | 0.02 | 96.5  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 13 | 11.8 | 11.49 | 7.66  | 6.97 | 39.6 | 0.03 | 95.4  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 12 | 13.8 | 11.53 | 7.27  | 6.96 | 40.2 | 0.03 | 94.8  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 11 | 15.8 | 11.54 | 7.12  | 6.96 | 40.6 | 0.03 | 94.6  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 10 | 17.8 | 11.55 | 7.07  | 6.96 | 40.9 | 0.03 | 94.5  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 9  | 19.8 | 11.63 | 6.79  | 6.95 | 42.7 | 0.03 | 94.5  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 8  | 21.8 | 11.65 | 6.66  | 6.95 | 43.3 | 0.03 | 94.4  | - | 420 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 7  | 23.8 | 11.74 | 6.33  | 6.94 | 45.1 | 0.03 | 94.3  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 6  | 25.8 | 11.71 | 6.33  | 6.94 | 45.3 | 0.03 | 94.1  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 5  | 27.8 | 11.72 | 6.25  | 6.93 | 45.8 | 0.03 | 93.8  | - | 419 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 4  | 29.8 | 11.63 | 6.17  | 6.92 | 46   | 0.03 | 93    | - | 418 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 3  | 31.8 | 11.6  | 5.89  | 6.9  | 47   | 0.03 | 92.1  | - | 418 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 2  | 33.8 | 11.57 | 5.75  | 6.91 | 47.2 | 0.03 | 91.6  | - | 421 |
| Windy Bay Deep | 1997 | DBSS | 5/16/97 | - | 1.9 | 1  | 35.8 | 11.52 | 5.75  | 6.91 | 47.5 | 0.03 | 91.1  | - | 425 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 19 | 0.2  | 10.79 | 15.64 | 7.04 | 31.2 | 0.02 | 107.9 | - | 385 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 18 | 1    | 10.79 | 15.59 | 7.01 | 31.3 | 0.02 | 107.8 | - | 386 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 17 | 3    | 10.89 | 14.01 | 6.95 | 31.2 | 0.02 | 105.1 | - | 390 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 16 | 5    | 10.49 | 11.31 | 6.9  | 30.6 | 0.02 | 95.3  | - | 394 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 15 | 7    | 10.51 | 10.33 | 6.92 | 29.1 | 0.02 | 93.3  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 14 | 9    | 10.6  | 9.99  | 6.92 | 28.2 | 0.02 | 93.3  | - | 392 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 13 | 11.1 | 10.72 | 9.42  | 6.92 | 29.4 | 0.02 | 93.1  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 12 | 12.9 | 10.8  | 8.99  | 6.93 | 29.1 | 0.02 | 92.7  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 11 | 14.9 | 10.88 | 8.48  | 6.92 | 31.1 | 0.02 | 92.3  | - | 393 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 10 | 17   | 11.03 | 7.91  | 6.91 | 34.3 | 0.02 | 92.3  | - | 394 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 9  | 19   | 11.14 | 7.33  | 6.89 | 37.4 | 0.02 | 91.9  | - | 395 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 8  | 21   | 11.05 | 7.1   | 6.88 | 38.4 | 0.02 | 90.7  | - | 395 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 7  | 23   | 11.13 | 6.61  | 6.87 | 40.4 | 0.03 | 90.2  | - | 396 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 6  | 24.9 | 11.12 | 6.33  | 6.86 | 41.5 | 0.03 | 89.5  | - | 396 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 5  | 27.1 | 11.03 | 6.08  | 6.85 | 42.5 | 0.03 | 88.2  | - | 397 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 4  | 28.9 | 11.03 | 5.97  | 6.85 | 42.8 | 0.03 | 87.9  | - | 397 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 3  | 30.9 | 10.89 | 5.87  | 6.84 | 43.3 | 0.03 | 86.6  | - | 398 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 2  | 33.1 | 10.89 | 5.84  | 6.85 | 43.4 | 0.03 | 86.4  | - | 398 |
| Windy Bay Deep | 2197 | DBSS | 5/29/97 | - | 1.5 | 1  | 35   | 10.81 | 5.9   | 6.85 | 43.4 | 0.03 | 86.1  | - | 398 |
|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 17 | 0.8  | 10.58 | 14.84 | 6.97 | 37   | 0.02 | 104.4 | - | 422 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 16 | 2.8  | 10.68 | 13.55 | 6.94 | 35.7 | 0.02 | 102.6 | - | 424 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 15 | 4.8  | 10.67 | 10.86 | 6.95 | 31.8 | 0.02 | 96.2  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 14 | 6.9  | 10.7  | 10.12 | 6.95 | 30.1 | 0.02 | 94.8  | - | 424 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 13 | 8.8  | 10.61 | 9.91  | 6.94 | 30.1 | 0.02 | 93.5  | - | 425 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 10.8 | 10.67 | 9.2   | 6.94 | 31.8 | 0.02 | 92.5  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 12.8 | 10.66 | 9.02  | 6.93 | 33.2 | 0.02 | 92    | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 14.8 | 10.67 | 8.84  | 6.93 | 33.9 | 0.02 | 91.7  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 16.7 | 10.67 | 8.58  | 6.92 | 35.2 | 0.02 | 91.2  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 18.9 | 10.75 | 7.86  | 6.91 | 38.3 | 0.02 | 90.3  | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 20.7 | 10.81 | 6.91  | 6.89 | 42.8 | 0.03 | 88.6  | - | 427 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 22.6 | 10.81 | 6.76  | 6.89 | 43.7 | 0.03 | 88.2  | - | 427 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 24.8 | 10.82 | 6.61  | 6.89 | 44.3 | 0.03 | 88    | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 26.8 | 10.85 | 6.51  | 6.88 | 44.7 | 0.03 | 88    | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 28.7 | 10.92 | 6.33  | 6.88 | 45.4 | 0.03 | 88.2  | - | 426 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 30.5 | 10.76 | 6.18  | 6.88 | 46.2 | 0.03 | 86.4  | - | 425 |
| Windy Bay Deep | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 32.4 | 10.85 | 5.98  | 6.87 | 47.1 | 0.03 | 86.7  | - | 426 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 16 | 0.3  | 10.75 | 13.52 | 7.13 | 35   | 0.02 | 102.9 | - | 435 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 15 | 1.2  | 10.74 | 13.52 | 7.12 | 35   | 0.02 | 102.8 | - | 436 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 14 | 3.3  | 10.76 | 13.44 | 7.08 | 34.9 | 0.02 | 102.8 | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 13 | 5.3  | 10.79 | 12.84 | 7.05 | 34.5 | 0.02 | 101.6 | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 12 | 7.4  | 10.78 | 12.14 | 7.05 | 33.5 | 0.02 | 100   | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 11 | 9.4  | 10.7  | 11.99 | 7.02 | 33.2 | 0.02 | 99    | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 10 | 11.4 | 10.74 | 11.92 | 7    | 33.1 | 0.02 | 99.2  | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 9  | 13.4 | 10.77 | 11.27 | 6.98 | 33.1 | 0.02 | 98    | - | 437 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 8  | 15.3 | 10.75 | 10.79 | 6.96 | 33   | 0.02 | 96.7  | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 7  | 17.4 | 10.64 | 9.37  | 6.92 | 34.9 | 0.02 | 92.5  | - | 439 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 6  | 19.4 | 10.66 | 7.47  | 6.9  | 39.3 | 0.03 | 88.5  | - | 441 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 5  | 21.2 | 10.65 | 7.27  | 6.9  | 40.2 | 0.03 | 87.9  | - | 440 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 4  | 23.3 | 10.67 | 6.73  | 6.9  | 42.3 | 0.03 | 87    | - | 440 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 3  | 25.3 | 10.71 | 6.56  | 6.9  | 43   | 0.03 | 86.8  | - | 439 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 2  | 27.3 | 10.62 | 6.23  | 6.91 | 44.4 | 0.03 | 85.5  | - | 438 |
| Windy Bay Deep | 2597 | DBSS | 6/26/97 | - | 2.8 | 1  | 29.2 | 10.67 | 6.25  | 6.93 | 44.4 | 0.03 | 85.9  | - | 437 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 16 | 0.3  | 10.09 | 17.22 | 7.16 | 38.2 | 0.02 | ****  | - | 421 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 15 | 1    | 10.11 | 17.17 | 7.15 | 38.1 | 0.02 | ****  | - | 421 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 14 | 3    | 10.19 | 16.89 | 7.06 | 37.9 | 0.02 | ****  | - | 425 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 13 | 5    | 10.35 | 15.28 | 7.04 | 36.2 | 0.02 | ****  | - | 426 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 12 | 7    | 10.26 | 14.31 | 6.98 | 35.7 | 0.02 | ****  | - | 428 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 11 | 9    | 10.18 | 13.32 | 6.93 | 35.2 | 0.02 | 97.9  | - | 429 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 10 | 11   | 9.93  | 12.73 | 6.86 | 34.4 | 0.02 | 94.3  | - | 431 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 9  | 13   | 9.88  | 11.56 | 6.83 | 34.5 | 0.02 | 91.5  | - | 433 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 8  | 15   | 9.83  | 9.32  | 6.78 | 37.1 | 0.02 | 86.2  | - | 435 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 7  | 17   | 10.07 | 7.65  | 6.79 | 40.5 | 0.03 | 84.7  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 6  | 18.9 | 10.12 | 7.43  | 6.78 | 41   | 0.03 | 84.7  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 5  | 20.9 | 10.3  | 7.1   | 6.79 | 41.4 | 0.03 | 85.5  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 4  | 23   | 10.31 | 6.79  | 6.79 | 42.7 | 0.03 | 84.8  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1 | 3  | 25   | 10.32 | 6.69  | 6.79 | 43.1 | 0.03 | 84.8  | - | 436 |

|                |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1  | 2  | 27   | 10.34 | 6.4   | 6.78 | 44.4 | 0.03 | 84.3  | - | 436 |
| Windy Bay Deep | 2797 | DBSS | 7/9/97  | - | 4.1  | 1  | 29.1 | 10.13 | 6.13  | 6.78 | 45.6 | 0.03 | 82.1  | - | 436 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 15 | 0.4  | 9.68  | 20.43 | 7.4  | 38.9 | 0.02 | ***** | - | 417 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 14 | 1.5  | 9.88  | 19.74 | 7.38 | 38   | 0.02 | ***** | - | 418 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 13 | 5.6  | 10.3  | 17.73 | 7.24 | 35.3 | 0.02 | ***** | - | 423 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 12 | 7.6  | 10.39 | 16.87 | 7.14 | 35.4 | 0.02 | ***** | - | 426 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 11 | 9.6  | 10.36 | 16.08 | 6.99 | 35.3 | 0.02 | ***** | - | 431 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 10 | 11.6 | 9.99  | 14.64 | 6.85 | 34.2 | 0.02 | 97.7  | - | 437 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 9  | 13.6 | 9.49  | 13.24 | 6.75 | 33.4 | 0.02 | 90    | - | 441 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 8  | 15.6 | 8.99  | 10.54 | 6.69 | 34   | 0.02 | 80.2  | - | 444 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 7  | 17.6 | 9.71  | 8.29  | 6.67 | 37   | 0.02 | 82    | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 6  | 19.6 | 9.92  | 7.28  | 6.67 | 39.2 | 0.03 | 81.7  | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 5  | 21.6 | 9.99  | 6.87  | 6.66 | 40.4 | 0.03 | 81.4  | - | 447 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 4  | 23.6 | 10.1  | 6.66  | 6.66 | 41.1 | 0.03 | 81.8  | - | 447 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 3  | 25.5 | 10.09 | 6.51  | 6.66 | 41.5 | 0.03 | 81.5  | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 2  | 27.7 | 10.07 | 6.31  | 6.66 | 42.2 | 0.03 | 81    | - | 446 |
| Windy Bay Deep | 2997 | SSJD | 7/24/97 | - | 6.5  | 1  | 29.6 | 9.6   | 6.33  | 6.65 | 42.7 | 0.03 | 77.2  | - | 447 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 18 | 0.3  | 8.91  | 24.3  | 7.42 | 42.5 | 0.03 | 105   | - | 371 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 17 | 2.3  | 9.05  | 23.52 | 7.4  | 43.6 | 0.03 | 105.1 | - | 372 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 16 | 4.2  | 9.24  | 22.92 | 7.31 | 43   | 0.03 | 106.1 | - | 375 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 15 | 6.3  | 9.79  | 20.85 | 7.2  | 40   | 0.03 | 108.1 | - | 378 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 14 | 8.3  | 10.19 | 16.78 | 7.03 | 35.7 | 0.02 | 103.5 | - | 385 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 13 | 10.4 | 9.31  | 13.48 | 6.87 | 34.2 | 0.02 | 88    | - | 392 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 12 | 12.3 | 8.86  | 11.15 | 6.84 | 34.7 | 0.02 | 79.6  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 11 | 14.2 | 8.98  | 9.66  | 6.85 | 35.4 | 0.02 | 77.8  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 10 | 16.2 | 9.21  | 8.37  | 6.85 | 37.6 | 0.02 | 77.4  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 9  | 18.3 | 9.66  | 7.5   | 6.87 | 38.7 | 0.02 | 79.5  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 8  | 20.3 | 9.78  | 7.05  | 6.87 | 40.3 | 0.03 | 79.5  | - | 395 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 7  | 22.3 | 9.96  | 6.79  | 6.89 | 40.8 | 0.03 | 80.4  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 6  | 24.2 | 10.08 | 6.57  | 6.89 | 41.3 | 0.03 | 81    | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 5  | 26.3 | 10.1  | 6.38  | 6.91 | 42   | 0.03 | 80.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 4  | 28.3 | 10    | 6.23  | 6.92 | 42.7 | 0.03 | 79.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 3  | 30.3 | 9.74  | 6.26  | 6.93 | 43.2 | 0.03 | 77.7  | - | 394 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 2  | 32.3 | 9.57  | 6.21  | 6.94 | 43.2 | 0.03 | 76.2  | - | 393 |
| Windy Bay Deep | 3197 | DBSS | 8/5/97  | - | 8    | 1  | 34.3 | 9.56  | 6.18  | 6.98 | 44.8 | 0.03 | 76.1  | - | 392 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 16 | 0.3  | 8.72  | 22.94 | 7.3  | 41.9 | 0.03 | 100.8 | - | 382 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 15 | 1.3  | 8.77  | 22.33 | 7.2  | 42.6 | 0.03 | 100.2 | - | 388 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 14 | 3.3  | 8.8   | 22.03 | 7.13 | 43.1 | 0.03 | 100.1 | - | 391 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 13 | 5.3  | 8.9   | 21.82 | 7.43 | 42   | 0.03 | 100.7 | - | 384 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 12 | 7.3  | 8.87  | 21.61 | 7.31 | 41.5 | 0.03 | 100   | - | 388 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 11 | 9.3  | 8.96  | 20.73 | 7.4  | 41.3 | 0.03 | 99.4  | - | 383 |

|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|------|----|------|------|-------|------|------|------|-------|---|-----|
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 10 | 11.3 | 8.86 | 13.82 | 6.78 | 35.3 | 0.02 | 85.6  | - | 408 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 9  | 13.5 | 8.82 | 11.59 | 7.2  | 35.9 | 0.02 | 80.5  | - | 399 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 8  | 15.4 | 8.47 | 10.68 | 6.75 | 35.9 | 0.02 | 75.7  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 7  | 17.4 | 8.57 | 9.04  | 6.76 | 37.3 | 0.02 | 73.7  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 6  | 19.3 | 8.84 | 8.18  | 6.77 | 39   | 0.03 | 74.4  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 5  | 21.4 | 9.11 | 7.27  | 6.78 | 40.8 | 0.03 | 74.9  | - | 411 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 4  | 23.4 | 9.37 | 6.92  | 6.8  | 41.5 | 0.03 | 76.6  | - | 410 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 3  | 25.4 | 9.5  | 6.72  | 6.81 | 42.1 | 0.03 | 77.1  | - | 409 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 2  | 27.4 | 9.52 | 6.56  | 6.82 | 42.5 | 0.03 | 76.9  | - | 408 |
| Windy Bay Deep | 3297 | DBSS | 8/13/97 | - | 10.5 | 1  | 29.4 | 9.12 | 6.33  | 6.83 | 43.8 | 0.03 | 73.3  | - | 408 |
|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 16 | 0.3  | 9.22 | 20.87 | 7.41 | 47   | 0.03 | 102.6 | - | 372 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 15 | 1.5  | 9.23 | 20.83 | 7.39 | 47.1 | 0.03 | 102.7 | - | 372 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 14 | 3.5  | 9.23 | 20.78 | 7.33 | 46.9 | 0.03 | 102.5 | - | 374 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 13 | 5.4  | 9.24 | 20.67 | 7.26 | 46.8 | 0.03 | 102.4 | - | 375 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 12 | 7.5  | 9.15 | 20.69 | 7.08 | 46.7 | 0.03 | 101.5 | - | 382 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 11 | 9.5  | 9.28 | 15.93 | 6.87 | 38.8 | 0.02 | 93.5  | - | 394 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 10 | 11.5 | 8.91 | 12.38 | 6.82 | 38.3 | 0.02 | 82.9  | - | 397 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 9  | 13.6 | 8.48 | 9.58  | 6.8  | 40.6 | 0.03 | 73.9  | - | 399 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 8  | 15.5 | 9.07 | 7.68  | 6.8  | 43.2 | 0.03 | 75.5  | - | 401 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 7  | 17.2 | 9.43 | 7.07  | 6.83 | 44.3 | 0.03 | 77.4  | - | 400 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 6  | 21.6 | 9.72 | 6.64  | 6.88 | 45.3 | 0.03 | 78.8  | - | 399 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 5  | 23.5 | 9.72 | 6.5   | 6.89 | 45.6 | 0.03 | 78.6  | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 4  | 25.6 | 9.71 | 6.38  | 6.9  | 46.1 | 0.03 | 78.2  | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 3  | 27.5 | 9.58 | 6.3   | 6.92 | 46.4 | 0.03 | 77    | - | 398 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 2  | 29.5 | 9.47 | 6.28  | 6.96 | 46.5 | 0.03 | 76.2  | - | 397 |
| Windy Bay Deep | 3497 | DBSS | 8/27/97 | - | 8    | 1  | 31.7 | 9.48 | 6.28  | 7.05 | 46.7 | 0.03 | 76.2  | - | 395 |
|                |      |      |         |   |      |    |      |      |       |      |      |      |       |   |     |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 17 | 0.4  | 9.28 | 17.53 | 7.3  | 52   | 0.03 | 98.1  | - | 389 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 16 | 2.7  | 9.34 | 17.56 | 7.27 | 52   | 0.03 | 98.6  | - | 390 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 15 | 4.7  | 9.22 | 17.51 | 7.22 | 52   | 0.03 | 97.5  | - | 392 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 14 | 6.7  | 9.22 | 17.24 | 7.18 | 50.2 | 0.03 | 96.9  | - | 393 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 13 | 8.7  | 9.25 | 17.17 | 7.15 | 49.7 | 0.03 | 97.1  | - | 393 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 12 | 10.7 | 9.28 | 17.12 | 7.08 | 48.7 | 0.03 | 97.3  | - | 395 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 11 | 12.7 | 9.26 | 16.8  | 6.98 | 47.3 | 0.03 | 96.3  | - | 398 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 10 | 14.7 | 9.07 | 15.91 | 6.86 | 45   | 0.03 | 92.6  | - | 403 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 9  | 16.7 | 8.66 | 12.17 | 6.73 | 40.1 | 0.03 | 81.4  | - | 411 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 8  | 18.7 | 8.43 | 10.51 | 6.7  | 41.1 | 0.03 | 76.3  | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 7  | 20.7 | 8.47 | 9.46  | 6.71 | 41.9 | 0.03 | 74.8  | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 6  | 22.7 | 8.57 | 8.43  | 6.71 | 42.7 | 0.03 | 73.8  | - | 413 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 5  | 24.7 | 8.89 | 7.38  | 6.72 | 43.9 | 0.03 | 74.6  | - | 413 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 4  | 26.7 | 9.04 | 7.07  | 6.74 | 44.5 | 0.03 | 75.3  | - | 412 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 3  | 28.7 | 9.16 | 6.73  | 6.76 | 45.2 | 0.03 | 75.6  | - | 411 |
| Windy Bay Deep | 3797 | ASRP | 9/17/97 | - | 7.5  | 2  | 30.7 | 9.21 | 6.64  | 6.8  | 45.2 | 0.03 | 75.9  | - | 410 |

|                |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|----------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Windy Bay Deep | 3797 | ASRP | 9/17/97  | - | 7.5 | 1  | 32.7 | 9.23  | 6.64  | 6.84 | 45.2 | 0.03 | 76   | -   | 408 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 17 | 0.2  | 9.77  | 15.09 | 7.16 | 46.3 | 0.03 | 98.1 | -   | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 16 | 2    | 9.65  | 15.03 | 7.16 | 46.4 | 0.03 | 96.8 | -   | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 15 | 4    | 9.6   | 14.96 | 7.15 | 46.2 | 0.03 | 96.2 | -   | 444 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 14 | 6    | 9.59  | 14.78 | 7.11 | 46.6 | 0.03 | 95.7 | -   | 445 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 13 | 8    | 9.57  | 14.66 | 7.09 | 46.5 | 0.03 | 95.2 | -   | 446 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 12 | 10   | 9.57  | 14.66 | 7.07 | 46.6 | 0.03 | 95.2 | -   | 446 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 11 | 12   | 9.5   | 14.58 | 7.02 | 46.5 | 0.03 | 94.3 | -   | 447 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 10 | 14   | 9.39  | 14.35 | 6.97 | 45.9 | 0.03 | 92.8 | -   | 448 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 9  | 16   | 9.34  | 13.92 | 6.94 | 45.4 | 0.03 | 91.4 | -   | 449 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 8  | 18   | 9.28  | 13.6  | 6.91 | 45   | 0.03 | 90.2 | -   | 450 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 7  | 20   | 9.23  | 13.3  | 6.85 | 44.7 | 0.03 | 89.1 | -   | 451 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 6  | 22   | 8.88  | 11.2  | 6.78 | 42.2 | 0.03 | 81.7 | -   | 455 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 5  | 23.9 | 8.9   | 8.48  | 6.75 | 41.4 | 0.03 | 76.8 | -   | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 4  | 26   | 8.91  | 7.97  | 6.74 | 41.8 | 0.03 | 75.9 | -   | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 3  | 28   | 8.94  | 7.6   | 6.76 | 42   | 0.03 | 75.5 | -   | 457 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 2  | 30   | 8.93  | 7.28  | 6.8  | 42.2 | 0.03 | 74.8 | -   | 455 |
| Windy Bay Deep | 3997 | ASRA | 9/29/97  | - | 7.5 | 1  | 31.8 | 9.05  | 7     | 6.83 | 42.6 | 0.03 | 75.3 | -   | 454 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 16 | 0.3  | 9.92  | 12.28 | 7.14 | 49.7 | 0.03 | 91.7 | -   | 386 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 15 | 1    | 9.94  | 12.28 | 7.14 | 49.7 | 0.03 | 91.8 | -   | 387 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 14 | 3    | 9.92  | 12.3  | 7.14 | 49.7 | 0.03 | 91.6 | -   | 386 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 13 | 5    | 9.91  | 12.3  | 7.12 | 49.7 | 0.03 | 91.5 | -   | 387 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 12 | 7    | 9.91  | 12.28 | 7.1  | 49.9 | 0.03 | 91.6 | -   | 387 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 11 | 9    | 9.92  | 12.27 | 7.07 | 49.8 | 0.03 | 91.6 | -   | 388 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 10 | 11   | 9.91  | 12.27 | 7.03 | 49.7 | 0.03 | 91.5 | -   | 388 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 9  | 13   | 9.88  | 12.27 | 6.99 | 49.7 | 0.03 | 91.2 | -   | 389 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 8  | 15   | 9.82  | 12.27 | 6.91 | 49.8 | 0.03 | 90.7 | -   | 392 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 7  | 17   | 9.74  | 12.18 | 6.8  | 49.8 | 0.03 | 89.7 | -   | 395 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 6  | 19   | 8.21  | 10.1  | 6.64 | 48.8 | 0.03 | 72.1 | -   | 401 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 5  | 21   | 8.3   | 7.99  | 6.64 | 47.5 | 0.03 | 69.2 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 4  | 23   | 8.4   | 7.35  | 6.64 | 47.5 | 0.03 | 69   | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 3  | 25   | 8.27  | 6.96  | 6.65 | 47.6 | 0.03 | 67.3 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 2  | 27   | 8.28  | 6.91  | 6.65 | 47.9 | 0.03 | 67.2 | -   | 402 |
| Windy Bay Deep | 4297 | DBAS | 10/21/97 | - | 7.3 | 1  | 29.1 | 8.27  | 6.84  | 6.71 | 47.9 | 0.03 | 67   | -   | 401 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 16 | 0.6  | 10.6  | 10.33 | 7.11 | 46.8 | 0.03 | 93.7 | 57  | 411 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 15 | 2.8  | 10.56 | 10.23 | 7.1  | 46.9 | 0.03 | 93.1 | 142 | 413 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 14 | 4.8  | 10.5  | 10.22 | 7.11 | 46.9 | 0.03 | 92.6 | 207 | 413 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 13 | 6.8  | 10.43 | 10.07 | 7.09 | 47.1 | 0.03 | 91.6 | 56  | 414 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 12 | 8.8  | 10.46 | 10.07 | 7.08 | 47.1 | 0.03 | 91.9 | 46  | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 11 | 10.8 | 10.45 | 10.09 | 7.09 | 47.2 | 0.03 | 91.9 | 40  | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97  | - | 7   | 10 | 12.8 | 10.46 | 10.07 | 7.09 | 47.2 | 0.03 | 91.9 | 117 | 415 |

|                |      |      |         |   |   |   |      |       |       |      |      |      |      |     |     |
|----------------|------|------|---------|---|---|---|------|-------|-------|------|------|------|------|-----|-----|
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 9 | 14.8 | 10.47 | 10.05 | 7.09 | 47.2 | 0.03 | 92   | 159 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 8 | 16.8 | 10.42 | 10.04 | 7.08 | 47.4 | 0.03 | 91.5 | 132 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 7 | 18.8 | 10.44 | 10.04 | 7.08 | 47.3 | 0.03 | 91.7 | 102 | 415 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 6 | 20.8 | 10.45 | 10.02 | 7.07 | 47.3 | 0.03 | 91.7 | 112 | 416 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 5 | 22.8 | 10.43 | 9.97  | 7.07 | 47.5 | 0.03 | 91.5 | 52  | 416 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 4 | 24.8 | 10.39 | 9.91  | 7.04 | 47.7 | 0.03 | 91   | 55  | 417 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 3 | 26.8 | 10.34 | 9.71  | 7.03 | 48.3 | 0.03 | 90.1 | 119 | 418 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 2 | 28.8 | 10.24 | 9.33  | 7.02 | 49.5 | 0.03 | 88.5 | 111 | 419 |
| Windy Bay Deep | 4497 | DBAS | 11/4/97 | - | 7 | 1 | 30.8 | 10.05 | 9.02  | 7.01 | 50   | 0.03 | 86.1 | 631 | 421 |

| Location  | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-----------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 17       | 0.4       | 12.48                   | 6.49            | 7.48 | 37.4                 | 0.02 | ****                       | -                  | 372   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 16       | 1.4       | 12.52                   | 6.33            | 7.46 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 15       | 2.4       | 12.46                   | 6.3             | 7.46 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 14       | 3.5       | 12.47                   | 6.28            | 7.45 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 13       | 4.6       | 12.42                   | 6.28            | 7.44 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 12       | 5.5       | 12.4                    | 6.28            | 7.44 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 11       | 6.6       | 12.41                   | 6.26            | 7.43 | 37.5                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 10       | 7.8       | 12.39                   | 6.25            | 7.43 | 37.4                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 9        | 8.8       | 12.45                   | 6.23            | 7.42 | 37.5                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 8        | 10        | 12.42                   | 6.2             | 7.42 | 37.5                 | 0.02 | ****                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 7        | 11.1      | 12.38                   | 6.17            | 7.42 | 37.6                 | 0.02 | 99.9                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 6        | 12.4      | 12.34                   | 6.08            | 7.41 | 38                   | 0.02 | 99.3                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 5        | 13.5      | 12.32                   | 5.88            | 7.41 | 38.4                 | 0.02 | 98.6                       | -                  | 374   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 4        | 14.8      | 12.27                   | 5.77            | 7.4  | 38.5                 | 0.02 | 98                         | -                  | 375   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 3        | 16        | 12.23                   | 5.67            | 7.39 | 38.4                 | 0.02 | 97.4                       | -                  | 376   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 2        | 17.2      | 12.19                   | 5.59            | 7.39 | 38.3                 | 0.02 | 96.9                       | -                  | 380   |
| CDA River | 1597  | ASSS     | 4/18/97 | -    | 1.8        | 1        | 18.6      | 12.59                   | 5.44            | 7.38 | 38.4                 | 0.02 | 99.9                       | -                  | 382   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 15       | 0.2       | 10.61                   | 14.93           | 6.94 | 32                   | 0.02 | 104.3                      | -                  | 390   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 14       | 1.6       | 10.56                   | 12.51           | 6.86 | 32                   | 0.02 | 98.3                       | -                  | 399   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 13       | 2.7       | 10.43                   | 12.37           | 6.84 | 32.1                 | 0.02 | 96.8                       | -                  | 401   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 12       | 3.7       | 10.41                   | 12.27           | 6.83 | 32.3                 | 0.02 | 96.4                       | -                  | 402   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 11       | 4.7       | 10.33                   | 12.18           | 6.82 | 32.2                 | 0.02 | 95.5                       | -                  | 402   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 10       | 5.7       | 10.25                   | 11.91           | 6.82 | 32.4                 | 0.02 | 94.1                       | -                  | 404   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 9        | 6.7       | 10.25                   | 11.76           | 6.82 | 32.3                 | 0.02 | 93.8                       | -                  | 404   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 8        | 7.7       | 10.24                   | 11.64           | 6.85 | 32.3                 | 0.02 | 93.5                       | -                  | 403   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 7        | 8.7       | 10.33                   | 11.31           | 6.93 | 32                   | 0.02 | 93.7                       | -                  | 400   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 6        | 9.7       | 10.67                   | 10.4            | 6.99 | 31                   | 0.02 | 94.6                       | -                  | 396   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 5        | 10.7      | 11.02                   | 9.42            | 7.01 | 29.6                 | 0.02 | 95.5                       | -                  | 394   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 4        | 11.7      | 11                      | 9.06            | 7.01 | 31.9                 | 0.02 | 94.5                       | -                  | 394   |
| CDA River | 1997  | DBSS     | 5/16/97 | -    | 0.6        | 3        | 12.7      | 10.99                   | 8.88            | 7.01 | 32.5                 | 0.02 | 94                         | -                  | 393   |



|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 11 | 0.4  | 9.72  | 18.54 | 7.11 | 49.9 | 0.03 | ***** | - | 421 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 10 | 1.1  | 9.72  | 18.53 | 7.09 | 50   | 0.03 | ***** | - | 421 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 9  | 2.1  | 9.73  | 18.5  | 7.06 | 52.1 | 0.03 | ***** | - | 422 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 8  | 3    | 9.78  | 18.37 | 7    | 54.8 | 0.04 | ***** | - | 424 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 7  | 4.1  | 9.6   | 18.03 | 6.91 | 69.8 | 0.04 | ***** | - | 428 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 6  | 5    | 9.65  | 17.73 | 6.86 | 75.4 | 0.05 | ***** | - | 430 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 5  | 6    | 9.96  | 14.99 | -    | 43   | 0.03 | 99.4  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 4  | 7    | 10    | 13.1  | 6.94 | 34.1 | 0.02 | 95.7  | - | 425 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 3  | 9    | 9.76  | 11.94 | 6.91 | 33.4 | 0.02 | 90.9  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 2  | 10.1 | 9.75  | 11.82 | 6.92 | 33.4 | 0.02 | 90.6  | - | 426 |
| CDA River | 2797 | DBSS | 7/9/97  | - | 3.3 | 1  | 11   | 9.71  | 11.82 | 6.94 | 33.4 | 0.02 | 90.3  | - | 430 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 12 | 0.5  | 9.38  | 21.69 | 7.49 | 39.8 | 0.03 | ***** | - | 395 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 11 | 1.2  | 9.4   | 21.37 | 7.46 | 42.7 | 0.03 | ***** | - | 396 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 10 | 2.3  | 9.4   | 21.33 | 7.4  | 48.4 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 9  | 3.2  | 9.4   | 21.33 | 7.37 | 50.1 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 8  | 4.2  | 9.4   | 21.3  | 7.33 | 51.3 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 7  | 5.3  | 9.42  | 21.11 | 7.26 | 49.9 | 0.03 | ***** | - | 398 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 6  | 6.2  | 9.33  | 20.67 | 7.11 | 58.2 | 0.04 | ***** | - | 402 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 5  | 7.2  | 10.2  | 16.19 | 7.12 | 35.9 | 0.02 | ***** | - | 401 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 4  | 8.2  | 9.82  | 15.19 | 7.02 | 34.9 | 0.02 | 97.2  | - | 403 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 3  | 9.2  | 9.45  | 14.18 | 6.98 | 34   | 0.02 | 91.5  | - | 402 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 2  | 10.2 | 9.28  | 13.84 | 0    | 34   | 0.02 | 89.2  | - | 400 |
| CDA River | 2997 | SSJD | 7/24/97 | - | 4.5 | 1  | 11.2 | 8.84  | 13.68 | 7.01 | 34.2 | 0.02 | 84.7  | - | 397 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 11 | 0.4  | 9.38  | 25.18 | 7.65 | 61.4 | 0.04 | 112.4 | - | 359 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 10 | 1.1  | 9.29  | 24.06 | 7.58 | 53.3 | 0.03 | 109   | - | 361 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 9  | 2.1  | 9.23  | 23.67 | 7.54 | 50.8 | 0.03 | 107.6 | - | 362 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 8  | 3.2  | 9.1   | 23.53 | 7.5  | 46.6 | 0.03 | 105.8 | - | 363 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 7  | 4.2  | 9.14  | 23.26 | 7.46 | 46.6 | 0.03 | 105.8 | - | 365 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 6  | 5.2  | 9.22  | 23.15 | 7.39 | 49.1 | 0.03 | 106.3 | - | 366 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 5  | 6    | 10.06 | 19.97 | 7.38 | 41.6 | 0.03 | 109.2 | - | 367 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 4  | 7.1  | 9.71  | 19.4  | 7.31 | 42.1 | 0.03 | 0     | - | 369 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 3  | 8    | 9.81  | 18.17 | 7.33 | 38.3 | 0.02 | 102.7 | - | 368 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 2  | 8.9  | 10.49 | 17.23 | 7.35 | 36.1 | 0.02 | 107.7 | - | 366 |
| CDA River | 3197 | DBSS | 8/5/97  | - | 5.5 | 1  | 10   | 9.81  | 16.75 | 7.32 | 36.6 | 0.02 | 99.7  | - | 368 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 10 | 0.2  | 8.82  | 23.38 | 7.74 | 54.8 | 0.04 | 102.8 | - | 363 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 9  | 0.7  | 8.77  | 23.13 | 7.65 | 53.9 | 0.03 | 101.8 | - | 369 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 8  | 1.7  | 8.74  | 22.74 | 7.66 | 47.7 | 0.03 | 100.7 | - | 368 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 7  | 2.7  | 8.8   | 22.65 | 7.68 | 50.3 | 0.03 | 101.3 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 6  | 3.7  | 8.88  | 22.63 | 7.66 | 52.9 | 0.03 | 102.2 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 5  | 4.7  | 8.91  | 22.33 | 7.64 | 56   | 0.04 | 101.9 | - | 367 |
| CDA River | 3297 | DBSS | 8/13/97 | - | 6   | 4  | 5.6  | 8.95  | 22.24 | 7.58 | 61.9 | 0.04 | 102.2 | - | 368 |



|           |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| CDA River | 3297 | DBSS | 8/13/97  | - | 6   | 3  | 6.7  | 8.95  | 22.08 | 7.3  | 67.3 | 0.04 | 101.8 | - | 375 |
| CDA River | 3297 | DBSS | 8/13/97  | - | 6   | 2  | 7.7  | 8.67  | 19.88 | 7.15 | 46.5 | 0.03 | 94.5  | - | 381 |
| CDA River | 3297 | DBSS | 8/13/97  | - | 6   | 1  | 8.7  | 8.73  | 17.78 | 7.14 | 39.5 | 0.03 | 91.2  | - | 382 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 13 | 0.3  | 9.13  | 20.86 | 7.53 | 52.7 | 0.03 | 101.5 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 12 | 0.9  | 9.14  | 20.84 | 7.54 | 52.6 | 0.03 | 101.7 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 11 | 2    | 9.14  | 20.84 | 7.54 | 52.5 | 0.03 | 101.7 | - | 344 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 10 | 2.9  | 9.14  | 20.83 | 7.54 | 53.3 | 0.03 | 101.7 | - | 344 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 9  | 4    | 9.14  | 20.81 | 7.52 | 55.2 | 0.04 | 101.6 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 8  | 5.1  | 9.12  | 20.81 | 7.5  | 55.4 | 0.04 | 101.4 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 7  | 6.1  | 9.11  | 20.77 | 7.48 | 58   | 0.04 | 101.2 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 6  | 6.9  | 9.08  | 20.76 | 7.47 | 58.3 | 0.04 | 100.9 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 5  | 7.9  | 9.09  | 20.72 | 7.44 | 60   | 0.04 | 100.9 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 4  | 9    | 9.06  | 20.69 | 7.41 | 59.8 | 0.04 | 100.4 | - | 346 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 3  | 10   | 9.03  | 20.67 | 7.39 | 59.1 | 0.04 | 100.2 | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 2  | 10.9 | 8.92  | 20.56 | 7.37 | 59   | 0.04 | 98.7  | - | 345 |
| CDA River | 3497 | DBSS | 8/27/97  | - | 6.1 | 1  | 12   | 8.08  | 16.69 | 7.11 | 42.1 | 0.03 | 82.6  | - | 359 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 11 | 0.1  | 9.38  | 17.47 | 7.44 | 58.9 | 0.04 | 99.1  | - | 370 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 10 | 1    | 9.32  | 17.49 | 7.43 | 58.9 | 0.04 | 98.4  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 9  | 2    | 9.24  | 17.48 | 7.45 | 59.1 | 0.04 | 97.6  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 8  | 3    | 9.35  | 17.49 | 7.42 | 59   | 0.04 | 98.8  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 7  | 4    | 9.37  | 17.48 | 7.4  | 59.6 | 0.04 | 99    | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 6  | 5    | 9.35  | 17.49 | 7.39 | 59.3 | 0.04 | 98.8  | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 5  | 6    | 9.29  | 17.48 | 7.37 | 59.8 | 0.04 | 97.9  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 4  | 7    | 9.28  | 17.43 | 7.34 | 61   | 0.04 | 97.8  | - | 368 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 3  | 8    | 9.19  | 17.44 | 7.22 | 61   | 0.04 | 97    | - | 369 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 2  | 9    | 8.59  | 9.82  | 7.04 | 50.3 | 0.03 | 76.9  | - | 380 |
| CDA River | 3797 | ASRP | 9/17/97  | - | 5.5 | 1  | 10.1 | 8.38  | 9.79  | 6.96 | 52.8 | 0.03 | 74.6  | - | 385 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 11 | 0.2  | 9.72  | 16.01 | 7.37 | 52.6 | 0.03 | 99.4  | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 10 | 1.8  | 9.66  | 15.76 | 7.36 | 52.2 | 0.03 | 98.4  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 9  | 2.8  | 9.66  | 15.71 | 7.33 | 53.3 | 0.03 | 98.3  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 8  | 3.8  | 9.62  | 15.66 | 7.31 | 53.9 | 0.03 | 97.8  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 7  | 4.8  | 9.59  | 15.61 | 7.29 | 53.4 | 0.03 | 97.4  | - | 415 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 6  | 5.9  | 9.56  | 15.54 | 7.26 | 53.8 | 0.03 | 97    | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 5  | 6.8  | 9.62  | 15.52 | 7.23 | 56.7 | 0.04 | 97.5  | - | 414 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 4  | 7.8  | 9.74  | 15.12 | 7.15 | 64.5 | 0.04 | 97.9  | - | 416 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 3  | 8.8  | 9.62  | 14.99 | 7.03 | 68   | 0.04 | 96.5  | - | 419 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 2  | 9.8  | 7.74  | 12.13 | 6.9  | 48.6 | 0.03 | 72.8  | - | 425 |
| CDA River | 3997 | ASRA | 9/29/97  | - | 6   | 1  | 10.8 | 7.85  | 10.26 | 6.97 | 44.5 | 0.03 | 70.6  | - | 424 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |       |   |     |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 11 | 0.3  | 10.11 | 11.79 | 7.24 | 54.6 | 0.03 | 92.3  | - | 371 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 10 | 1    | 10.1  | 11.81 | 7.27 | 54.6 | 0.03 | 92.2  | - | 369 |

|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|-----------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 9  | 2    | 10.11 | 11.81 | 7.23 | 54.8 | 0.04 | 92.4 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 8  | 3    | 10.08 | 11.81 | 7.24 | 54.7 | 0.04 | 92.1 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 7  | 4.1  | 10.12 | 11.77 | 7.22 | 55.3 | 0.04 | 92.4 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 6  | 5    | 10.13 | 11.74 | 7.19 | 55.7 | 0.04 | 92.4 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 5  | 6    | 10.16 | 11.68 | 7.18 | 56.7 | 0.04 | 92.5 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 4  | 7    | 10.17 | 11.59 | 7.17 | 57.9 | 0.04 | 92.5 | -   | 369 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 3  | 8    | 10.17 | 11.59 | 7.13 | 57.9 | 0.04 | 92.5 | -   | 370 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 2  | 9    | 10.22 | 11.41 | 7.1  | 63.2 | 0.04 | 92.5 | -   | 371 |
| CDA River | 4297 | DBAS | 10/21/97 | - | 6   | 1  | 10   | 10.23 | 10.94 | 7.01 | 81.3 | 0.05 | 91.6 | -   | 379 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 11 | 0.4  | 10.87 | 10.02 | 7.35 | 53.9 | 0.03 | 95.5 | 54  | 392 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 10 | 1.5  | 10.94 | 9.79  | 7.34 | 53.7 | 0.03 | 95.5 | 50  | 393 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 9  | 2.5  | 10.87 | 9.79  | 7.37 | 53.7 | 0.03 | 94.9 | 146 | 392 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 8  | 3.5  | 10.92 | 9.74  | 7.33 | 53.6 | 0.03 | 95.3 | 51  | 393 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 7  | 4.5  | 10.88 | 9.68  | 7.28 | 53.7 | 0.03 | 94.8 | 102 | 395 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 6  | 5.5  | 10.9  | 9.4   | 7.21 | 54.5 | 0.03 | 94.3 | 59  | 398 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 5  | 6.5  | 10.91 | 9.38  | 7.26 | 54.6 | 0.03 | 94.3 | 121 | 395 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 4  | 7.5  | 11    | 9.25  | 7.19 | 54.9 | 0.04 | 94.8 | 49  | 398 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 3  | 8.5  | 11.23 | 8.55  | 7.14 | 57.3 | 0.04 | 95.1 | 42  | 401 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 2  | 9.5  | 11.29 | 8.25  | 7.12 | 58.2 | 0.04 | 95   | 59  | 402 |
| CDA River | 4497 | DBAS | 11/4/97  | - | 3.8 | 1  | 10.5 | 11.26 | 8.17  | 7.1  | 58.6 | 0.04 | 94.5 | 536 | 404 |

| Location     | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 10       | 0.4       | 12.33                   | 6.36            | 7.36 | 44.6                 | 0.03 | 99.9                       | -                  | 377   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 9        | 1.4       | 12.36                   | 6.3             | 7.35 | 45.2                 | 0.03 | ****                       | -                  | 378   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 8        | 2.6       | 12.27                   | 6.33            | 7.35 | 44.4                 | 0.03 | 99.4                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 7        | 3.8       | 12.31                   | 6.28            | 7.34 | 45.3                 | 0.03 | 99.6                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 6        | 5.1       | 12.32                   | 6.3             | 7.34 | 45.4                 | 0.03 | 99.7                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 5        | 6.4       | 12.35                   | 6.31            | 7.34 | 44.9                 | 0.03 | ****                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 4        | 7.5       | 12.43                   | 6.31            | 7.33 | 44.6                 | 0.03 | ****                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 3        | 8.7       | 12.4                    | 6.31            | 7.33 | 44.7                 | 0.03 | ****                       | -                  | 378   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 2        | 10.1      | 12.44                   | 6.26            | 7.32 | 44.8                 | 0.03 | ****                       | -                  | 379   |
| Mid Lake CDA | 1597  | ASSS     | 4/18/97 | -    | 2          | 1        | 11.2      | 12.57                   | 6.28            | 7.32 | 44.8                 | 0.03 | ****                       | -                  | 378   |
|              |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 15       | 0.2       | 12.04                   | 13.83           | 6.33 | 0.7                  | 0    | 115.5                      | -                  | 422   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 14       | 1.5       | 11.12                   | 10.32           | 7    | 28.2                 | 0.02 | 98.5                       | -                  | 392   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 13       | 3         | 11.1                    | 9.86            | 7    | 28.8                 | 0.02 | 97.3                       | -                  | 392   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 12       | 4.5       | 11.16                   | 9.79            | 6.99 | 28.3                 | 0.02 | 97.5                       | -                  | 392   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 11       | 6         | 11.1                    | 9.51            | 6.98 | 29                   | 0.02 | 96.4                       | -                  | 393   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 10       | 7.5       | 11.04                   | 9.25            | 6.98 | 29.4                 | 0.02 | 95.3                       | -                  | 392   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 9        | 9         | 11.04                   | 9.17            | 6.99 | 29.6                 | 0.02 | 95                         | -                  | 391   |
| Mid Lake CDA | 1997  | DBAC     | 5/16/97 | -    | 1.25       | 8        | 10.5      | 11.08                   | 8.92            | 7.02 | 30.7                 | 0.02 | 94.8                       | -                  | 389   |

|              |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|------|----|------|-------|-------|------|------|------|-------|---|-----|
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 7  | 12   | 11.29 | 8.45  | 7.01 | 33.4 | 0.02 | 95.6  | - | 389 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 6  | 13.5 | 11.32 | 7.81  | 6.99 | 34.7 | 0.02 | 94.3  | - | 390 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 5  | 15   | 11.28 | 7.47  | 6.98 | 34.9 | 0.02 | 93.2  | - | 389 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 4  | 16.5 | 11.25 | 6.66  | 6.96 | 37.2 | 0.02 | 91.1  | - | 390 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 3  | 18   | 11.12 | 6.08  | 6.94 | 40.1 | 0.03 | 88.7  | - | 391 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 2  | 19.5 | 11.15 | 5.69  | 6.93 | 45.9 | 0.03 | 88.1  | - | 391 |
| Mid Lake CDA | 1997 | DBAC | 5/16/97 | - | 1.25 | 1  | 20.9 | 10.91 | 5.6   | 6.92 | 47.2 | 0.03 | 85.9  | - | 390 |
|              |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 13 | 0.2  | 11.32 | 12.63 | 7.06 | 27.6 | 0.02 | 105.9 | - | 351 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 12 | 0.9  | 11.33 | 12.51 | 7.05 | 27.6 | 0.02 | 105.8 | - | 352 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 11 | 2.5  | 11.19 | 10.72 | 7    | 26.8 | 0.02 | 100.2 | - | 353 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 10 | 4    | 10.83 | 10.3  | 6.97 | 26.6 | 0.02 | 96    | - | 354 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 9  | 5.5  | 10.7  | 10.07 | 6.96 | 26.4 | 0.02 | 94.4  | - | 353 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 8  | 7    | 10.68 | 9.64  | 6.95 | 26.3 | 0.02 | 93.2  | - | 352 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 7  | 8.5  | 10.71 | 9.45  | 6.95 | 26.5 | 0.02 | 93    | - | 349 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 6  | 9.9  | 10.73 | 9.25  | 6.96 | 26.6 | 0.02 | 92.8  | - | 348 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 5  | 12.5 | 10.75 | 8.84  | 6.95 | 27   | 0.02 | 92    | - | 346 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 4  | 14.4 | 10.82 | 8.38  | 6.94 | 28.3 | 0.02 | 91.6  | - | 343 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 3  | 16.1 | 10.73 | 7.96  | 6.92 | 30.2 | 0.02 | 89.9  | - | 342 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 2  | 18.7 | 10.48 | 6.49  | 6.87 | 40.5 | 0.03 | 84.7  | - | 339 |
| Mid Lake CDA | 2197 | DBSS | 5/29/97 | - | 1.7  | 1  | 20.2 | 10.17 | 6.71  | 6.87 | 40.7 | 0.03 | 82.6  | - | 332 |
|              |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 14 | 0.4  | 10.81 | 15.76 | 7.1  | 30.6 | 0.02 | 108.8 | - | 411 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 13 | 1.3  | 10.86 | 15.58 | 7.11 | 30.7 | 0.02 | 108.9 | - | 411 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 12 | 2.9  | 10.9  | 13.8  | 7.1  | 29.8 | 0.02 | 105.2 | - | 411 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 11 | 4.3  | 10.86 | 13.09 | 7.05 | 29.4 | 0.02 | 103   | - | 414 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 10 | 5.9  | 11.2  | 11.54 | 7.04 | 28.9 | 0.02 | 102.6 | - | 415 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 9  | 7.4  | 11    | 10.45 | 7.02 | 29   | 0.02 | 98.2  | - | 416 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 8  | 9    | 10.92 | 10.12 | 7    | 28.6 | 0.02 | 96.8  | - | 417 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 7  | 10.5 | 10.84 | 9.71  | 6.99 | 28.6 | 0.02 | 95.2  | - | 417 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 6  | 11.9 | 10.62 | 8.94  | 6.96 | 31.6 | 0.02 | 91.4  | - | 419 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 5  | 13.4 | 10.6  | 8.42  | 6.95 | 34.4 | 0.02 | 90.1  | - | 420 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 4  | 15   | 10.23 | 7.81  | 6.92 | 37.7 | 0.02 | 85.7  | - | 421 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 3  | 16.4 | 10.36 | 7.05  | 6.92 | 41.9 | 0.03 | 85.2  | - | 421 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 2  | 17.9 | 10.31 | 6.82  | 6.92 | 43.2 | 0.03 | 84.3  | - | 425 |
| Mid Lake CDA | 2397 | DBAS | 6/11/97 | - | 2    | 1  | 19.5 | 10.02 | 6.99  | 6.93 | 43.2 | 0.03 | 82.3  | - | 424 |
|              |      |      |         |   |      |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 13 | 0.4  | 10.7  | 15.09 | 7.14 | 31.7 | 0.02 | 106.1 | - | 405 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 12 | 1.9  | 10.69 | 14.89 | 7.12 | 31.8 | 0.02 | 105.4 | - | 405 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 11 | 3.5  | 10.74 | 14.36 | 7.1  | 31.9 | 0.02 | 104.8 | - | 406 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 10 | 5    | 10.68 | 13.32 | 6.97 | 30.2 | 0.02 | 101.8 | - | 410 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 9  | 6.6  | 10.73 | 12.33 | 6.96 | 30.5 | 0.02 | 100   | - | 410 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 8  | 8.1  | 10.36 | 11.45 | 6.92 | 30.4 | 0.02 | 94.6  | - | 411 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3    | 7  | 9.6  | 10.26 | 10.59 | 6.89 | 29.6 | 0.02 | 91.8  | - | 412 |

|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 6  | 10.9 | 10.14 | 9.27  | 6.86 | 30.8 | 0.02 | 88    | - | 413 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 5  | 12.5 | 10.11 | 8.25  | 6.83 | 34.6 | 0.02 | 85.6  | - | 414 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 4  | 14   | 9.57  | 7.53  | 6.81 | 38   | 0.02 | 79.6  | - | 414 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 3  | 15.6 | 9.44  | 7.05  | 6.8  | 40.5 | 0.03 | 77.7  | - | 413 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 2  | 17   | 9.03  | 6.82  | 6.8  | 42.4 | 0.03 | 73.8  | - | 411 |
| Mid Lake CDA | 2597 | DBSS | 6/26/97 | - | 3   | 1  | 18.6 | 9.08  | 6.81  | 6.86 | 42.4 | 0.03 | 74.1  | - | 407 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 13 | 0.3  | 9.95  | 19.4  | 7.02 | 35.5 | 0.02 | ****  | - | 444 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 12 | 1.7  | 10.11 | 17.84 | 7    | 34.8 | 0.02 | ****  | - | 445 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 11 | 3.3  | 10.29 | 17.17 | 6.92 | 35   | 0.02 | ****  | - | 447 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 10 | 4.8  | 10.55 | 14.84 | 6.81 | 38.2 | 0.02 | ****  | - | 451 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 9  | 6.2  | 10.41 | 13.81 | 6.77 | 35.6 | 0.02 | 99.9  | - | 451 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 8  | 7.8  | 10.26 | 13.02 | 6.7  | 34.6 | 0.02 | 96.7  | - | 451 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 7  | 9.2  | 10.03 | 12.2  | 6.66 | 33.4 | 0.02 | 92.8  | - | 451 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 6  | 10.8 | 9.78  | 10.8  | 6.6  | 33.3 | 0.02 | 87.6  | - | 450 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 5  | 12.2 | 9.53  | 9.76  | 6.57 | 34.5 | 0.02 | 83.3  | - | 450 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 4  | 13.7 | 9.55  | 8.95  | 6.54 | 36.6 | 0.02 | 82    | - | 448 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 3  | 15.3 | 9.33  | 8.46  | 6.52 | 38.2 | 0.02 | 79.1  | - | 447 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 2  | 16.6 | 9.28  | 7.91  | 6.49 | 39.8 | 0.03 | 77.7  | - | 445 |
| Mid Lake CDA | 2797 | DBSS | 7/8/97  | - | 2.4 | 1  | 18.2 | 9.37  | 7.51  | 6.47 | 42   | 0.03 | 77.6  | - | 447 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 13 | 1.3  | 9.57  | 21.01 | 7.48 | 38.8 | 0.02 | ****  | - | 394 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 12 | 2.8  | 9.92  | 20.47 | 7.43 | 37.9 | 0.02 | ****  | - | 393 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 11 | 4.3  | 10.52 | 17.56 | 7.27 | 37.7 | 0.02 | ****  | - | 400 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 10 | 5.8  | 10.49 | 16.73 | 7.18 | 36.2 | 0.02 | ****  | - | 402 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 9  | 7.3  | 10.22 | 15.79 | 7.03 | 35.2 | 0.02 | ****  | - | 406 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 8  | 8.8  | 9.96  | 15.01 | 6.95 | 34.4 | 0.02 | 98.2  | - | 408 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 7  | 10.3 | 9.14  | 13.61 | 6.84 | 33.2 | 0.02 | 87.5  | - | 411 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 6  | 11.8 | 9.33  | 11.29 | 6.79 | 33.2 | 0.02 | 84.6  | - | 414 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 5  | 13.3 | 8.89  | 9.48  | 6.78 | 35.2 | 0.02 | 77.3  | - | 414 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 4  | 14.3 | 9.1   | 8.22  | 6.8  | 37.7 | 0.02 | 76.7  | - | 413 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 3  | 15.8 | 9.07  | 8.17  | 6.83 | 37.8 | 0.02 | 76.4  | - | 411 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 2  | 17.3 | 9.15  | 7.95  | 6.87 | 38.3 | 0.02 | 76.6  | - | 414 |
| Mid Lake CDA | 2997 | SSJD | 7/24/97 | - | 4.5 | 1  | 18.8 | 9.24  | 7.46  | 6.93 | 39.7 | 0.03 | 76.2  | - | 422 |
|              |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 14 | 0.4  | 9.23  | 24.49 | 7.8  | 41.3 | 0.03 | 109.2 | - | 345 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 13 | 1    | 9.29  | 23.73 | 7.7  | 41.5 | 0.03 | 108.4 | - | 348 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 12 | 2.6  | 9.34  | 22.74 | 7.49 | 43.4 | 0.03 | 106.9 | - | 354 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 11 | 4    | 9.57  | 21.92 | 7.41 | 49.1 | 0.03 | 107.8 | - | 357 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 10 | 5.5  | 9.89  | 20.76 | 7.43 | 41.8 | 0.03 | 108.9 | - | 354 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 9  | 7    | 10.15 | 19.78 | 7.29 | 39.8 | 0.03 | 109.7 | - | 356 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 8  | 8.6  | 10.13 | 17.86 | 7.05 | 36.9 | 0.02 | 105.4 | - | 361 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 7  | 10   | 9.33  | 15.01 | 6.96 | 34.5 | 0.02 | 91.3  | - | 364 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 6  | 11.5 | 8.92  | 12.66 | 6.9  | 34.4 | 0.02 | 82.9  | - | 364 |

|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 5  | 13.1 | 8.74 | 10.18 | 6.88 | 35.9 | 0.02 | 76.7  | - | 361 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 4  | 14.5 | 8.58 | 9.05  | 6.91 | 37.2 | 0.02 | 73.3  | - | 357 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 3  | 16.1 | 8.76 | 8.04  | 6.94 | 38.9 | 0.02 | 73    | - | 354 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 2  | 17.5 | 8.83 | 7.94  | 6.99 | 39   | 0.03 | 73.5  | - | 347 |
| Mid Lake CDA | 3197 | DBSS | 8/5/97  | - | 5.6 | 1  | 18.9 | 8.88 | 8.18  | 7.08 | 39   | 0.03 | 74    | - | 374 |
|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 13 | 0.5  | 8.7  | 22.93 | 7.65 | 45.4 | 0.03 | 100.6 | - | 361 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 12 | 2    | 8.72 | 22.56 | 7.6  | 45.2 | 0.03 | 100.1 | - | 362 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 11 | 3.5  | 8.77 | 22.29 | 7.56 | 45   | 0.03 | 100.2 | - | 362 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 10 | 5    | 8.81 | 22.12 | 7.44 | 45   | 0.03 | 100.4 | - | 365 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 9  | 6.5  | 8.87 | 21.44 | 7.28 | 49.4 | 0.03 | 99.7  | - | 370 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 8  | 8    | 9.05 | 20.09 | 7.19 | 42.6 | 0.03 | 99.1  | - | 373 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 7  | 9.5  | 9.2  | 18.61 | 7.05 | 38.7 | 0.02 | 97.7  | - | 374 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 6  | 11   | 8.12 | 14.03 | 6.91 | 35.4 | 0.02 | 78.3  | - | 379 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 5  | 12.5 | 7.96 | 10.62 | 6.9  | 36.8 | 0.02 | 71.1  | - | 380 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 4  | 14   | 7.79 | 9.22  | 6.93 | 38.5 | 0.02 | 67.3  | - | 378 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 3  | 15.5 | 7.71 | 8.49  | 6.97 | 39.3 | 0.03 | 65.4  | - | 376 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 2  | 17   | 7.29 | 8.15  | 7.08 | 40.1 | 0.03 | 61.3  | - | 370 |
| Mid Lake CDA | 3297 | DBSS | 8/13/97 | - | 6.5 | 1  | 18.7 | 8.21 | 7.69  | 6.92 | 41.3 | 0.03 | 68.3  | - | 393 |
|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 13 | 0.4  | 9.08 | 21.49 | 7.54 | 50.5 | 0.03 | 102.3 | - | 353 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 12 | 1.1  | 9.08 | 21.48 | 7.54 | 50.4 | 0.03 | 102.2 | - | 353 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 11 | 2.4  | 9.08 | 21.42 | 7.53 | 50.4 | 0.03 | 102.1 | - | 353 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 10 | 4.1  | 9.09 | 21.28 | 7.52 | 50.7 | 0.03 | 102   | - | 353 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 9  | 5.6  | 9.1  | 21.18 | 7.5  | 50.7 | 0.03 | 101.9 | - | 353 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 8  | 6.9  | 9.1  | 21.11 | 7.46 | 50.7 | 0.03 | 101.7 | - | 354 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 7  | 8.5  | 8.96 | 21.05 | 7.33 | 49.9 | 0.03 | 100.1 | - | 358 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 6  | 10   | 8.48 | 20.3  | 7.17 | 48.7 | 0.03 | 93.4  | - | 363 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 5  | 11.5 | 8.19 | 19.32 | 7.01 | 45.8 | 0.03 | 88.4  | - | 370 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 4  | 12.9 | 8.39 | 15.84 | 6.91 | 40.1 | 0.03 | 84.2  | - | 375 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 3  | 14.4 | 6.45 | 12.02 | 6.88 | 40.9 | 0.03 | 59.5  | - | 379 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 2  | 15.9 | 6.81 | 9.48  | 6.95 | 42.1 | 0.03 | 59.2  | - | 377 |
| Mid Lake CDA | 3497 | DBSS | 8/27/97 | - | 8   | 1  | 17.4 | 6.69 | 9.22  | 7.07 | 42.6 | 0.03 | 57.8  | - | 373 |
|              |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 11 | 0.2  | 8.99 | 16.75 | 7.02 | 49.9 | 0.03 | 93.5  | - | 384 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 10 | 1    | 8.9  | 16.75 | 6.94 | 49.9 | 0.03 | 92.6  | - | 386 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 9  | 3    | 8.75 | 16.43 | 6.79 | 49.7 | 0.03 | 90.4  | - | 392 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 8  | 5    | 8.08 | 10.87 | 6.71 | 43.5 | 0.03 | 73.4  | - | 398 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 7  | 7    | 8.27 | 9.37  | 6.69 | 42.7 | 0.03 | 72.8  | - | 399 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 6  | 9    | 8.19 | 8.88  | 6.7  | 43.4 | 0.03 | 71.3  | - | 399 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 5  | 10.9 | 9    | 7.66  | 6.71 | 43.9 | 0.03 | 76.1  | - | 398 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 4  | 13   | 8.96 | 7.58  | 6.74 | 44.1 | 0.03 | 75.6  | - | 396 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 3  | 15   | 9.05 | 7.43  | 6.75 | 44.2 | 0.03 | 76.2  | - | 395 |
| Mid Lake CDA | 3797 | ASRP | 9/17/97 | - | 5   | 2  | 16.9 | 9.45 | 7.12  | 6.78 | 44.6 | 0.03 | 78.8  | - | 393 |

|              |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|--------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Mid Lake CDA | 3797 | ASRP | 9/17/97  | - | 5   | 1  | 18.6 | 9.41  | 6.89  | 6.84 | 45.1 | 0.03 | 78   | -   | 390 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 10 | 0.5  | 9.59  | 16.46 | 7.17 | 50.6 | 0.03 | 99.2 | -   | 407 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 9  | 2.7  | 9.45  | 15.86 | 7.07 | 50.1 | 0.03 | 96.5 | -   | 411 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 8  | 4.6  | 9.24  | 15.24 | 6.92 | 50.3 | 0.03 | 93.1 | -   | 417 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 7  | 6.6  | 8.62  | 13.82 | 6.82 | 49.6 | 0.03 | 84.2 | -   | 421 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 6  | 8.6  | 8.64  | 12.23 | 6.78 | 46.7 | 0.03 | 81.4 | -   | 423 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 5  | 10.6 | 8.58  | 9.81  | 6.74 | 42.5 | 0.03 | 76.4 | -   | 425 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 4  | 12.6 | 8.65  | 8.46  | 6.74 | 42   | 0.03 | 74.5 | -   | 424 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 3  | 14.6 | 9.04  | 7.79  | 6.74 | 42   | 0.03 | 77.4 | -   | 424 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 2  | 16.6 | 8.92  | 7.46  | 6.81 | 42.2 | 0.03 | 75   | -   | 420 |
| Mid Lake CDA | 3997 | ASRA | 9/29/97  | - | 6.5 | 1  | 18.6 | 8.91  | 7.47  | 6.87 | 42.1 | 0.03 | 75   | -   | 417 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 10 | 0.2  | 10.5  | 11.58 | 7.09 | 49   | 0.03 | 96   | -   | 413 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 9  | 2    | 10.48 | 11.59 | 7.1  | 48.9 | 0.03 | 95.8 | -   | 412 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 8  | 4    | 10.47 | 11.59 | 7.06 | 49.1 | 0.03 | 95.7 | -   | 413 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 7  | 6    | 10.45 | 11.58 | 7.02 | 49.1 | 0.03 | 95.5 | -   | 414 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 6  | 8    | 10.43 | 11.59 | 7    | 49.2 | 0.03 | 95.3 | -   | 414 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 5  | 10   | 10.42 | 11.58 | 6.94 | 49.2 | 0.03 | 95.2 | -   | 415 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 4  | 12   | 10.4  | 11.56 | 6.88 | 49.2 | 0.03 | 95   | -   | 417 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 3  | 14   | 10.37 | 11.56 | 6.82 | 49.2 | 0.03 | 94.8 | -   | 418 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 2  | 16   | 10.21 | 11.51 | 6.69 | 49.3 | 0.03 | 93.2 | -   | 421 |
| Mid Lake CDA | 4297 | DBAS | 10/22/97 | - | 4.6 | 1  | 18   | 4.53  | 8.53  | 6.32 | 45.1 | 0.03 | 38.5 | -   | 434 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 13 | 0.5  | 10.64 | 10.18 | 7.08 | 53.4 | 0.03 | 93.8 | 104 | 396 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 12 | 2    | 10.49 | 9.51  | 7.02 | 53.4 | 0.03 | 91   | 56  | 398 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 11 | 3.5  | 10.25 | 9.4   | 6.97 | 53.7 | 0.03 | 88.7 | 55  | 401 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 10 | 5    | 10.05 | 9.32  | 6.94 | 53.7 | 0.03 | 86.7 | 49  | 401 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 9  | 6.5  | 9.89  | 9.24  | 6.9  | 53.4 | 0.03 | 85.2 | 120 | 402 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 8  | 8    | 8.86  | 9.01  | 6.85 | 52.9 | 0.03 | 75.9 | 119 | 403 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 7  | 9.5  | 9.57  | 8.71  | 6.85 | 54.6 | 0.03 | 81.4 | 102 | 403 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 6  | 11   | 9.28  | 8.6   | 6.84 | 53.2 | 0.03 | 78.8 | 58  | 403 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 5  | 12.5 | 9.13  | 8.5   | 6.79 | 52.6 | 0.03 | 77.3 | 102 | 405 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 4  | 14   | 9.01  | 8.43  | 6.75 | 52.1 | 0.03 | 76.2 | 227 | 406 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 3  | 15.5 | 8.76  | 7.68  | 6.72 | 47.9 | 0.03 | 72.6 | 51  | 407 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 2  | 17   | 8.81  | 7.38  | 6.72 | 46.3 | 0.03 | 72.6 | 58  | 408 |
| Mid Lake CDA | 4497 | DBAS | 11/4/97  | - | 4   | 1  | 18.4 | 8.78  | 7.27  | 6.75 | 46.2 | 0.03 | 72.1 | 713 | 407 |

| Location  | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-----------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Carey Bay | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 10       | 0.4       | 12.31                   | 6.22            | 7.43 | 38.4                 | 0.02 | 99.4                       | -                  | 374   |
| Carey Bay | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 9        | 1         | 12.3                    | 6.07            | 7.41 | 38.5                 | 0.02 | 99                         | -                  | 375   |
| Carey Bay | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 8        | 2         | 12.29                   | 5.77            | 7.41 | 38.2                 | 0.02 | 98.1                       | -                  | 375   |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 7  | 3.1  | 12.21 | 5.64  | 7.41 | 38.4 | 0.02 | 97.2  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 6  | 4.3  | 12.2  | 5.6   | 7.4  | 38.3 | 0.02 | 97    | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 5  | 5.5  | 12.14 | 5.57  | 7.39 | 38.2 | 0.02 | 96.6  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 4  | 6.8  | 12.15 | 5.56  | 7.39 | 38.3 | 0.02 | 96.5  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 3  | 8.1  | 12.12 | 5.54  | 7.38 | 38.4 | 0.02 | 96.2  | - | 375 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 2  | 9.3  | 12.12 | 5.49  | 7.38 | 38.4 | 0.02 | 96.1  | - | 374 |
| Carey Bay | 1597 | ASSS | 4/18/97 | - | 1.9 | 1  | 10.5 | 12.09 | 5.47  | 7.37 | 38.5 | 0.02 | 95.8  | - | 374 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 12 | 0.2  | 11.52 | 15.54 | 7.11 | 35.1 | 0.02 | 114.7 | - | 389 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 11 | 1    | 11.94 | 11.22 | 7.15 | 33.5 | 0.02 | 107.9 | - | 390 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 10 | 2    | 12.11 | 10.1  | 7.1  | 33.6 | 0.02 | 106.7 | - | 392 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 9  | 3.5  | 11.57 | 9.45  | 7.05 | 32.8 | 0.02 | 100.3 | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 8  | 5    | 11.39 | 9.02  | 7.02 | 33.9 | 0.02 | 97.7  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 7  | 6.5  | 11.37 | 8.88  | 7.01 | 34.2 | 0.02 | 97.2  | - | 396 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 6  | 8    | 11.34 | 8.69  | 7.01 | 34.6 | 0.02 | 96.6  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 5  | 9.5  | 11.36 | 8.33  | 7    | 34.4 | 0.02 | 95.8  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 4  | 11   | 11.32 | 8.07  | 6.99 | 34.8 | 0.02 | 95    | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 3  | 12.5 | 11.24 | 7.83  | 6.98 | 34.8 | 0.02 | 93.7  | - | 395 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 2  | 14   | 10.99 | 7.61  | 6.96 | 35.2 | 0.02 | 91    | - | 396 |
| Carey Bay | 1997 | DBAC | 5/16/97 | - | 1.6 | 1  | 15.3 | 10.38 | 6.86  | 6.95 | 37.3 | 0.02 | 84.5  | - | 396 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 14 | 0.3  | 10.67 | 14.15 | 6.97 | 28.5 | 0.02 | 103.3 | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 13 | 1    | 11.21 | 10.91 | 6.99 | 26.2 | 0.02 | 100.9 | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 12 | 2.1  | 10.88 | 10.38 | 6.97 | 26.3 | 0.02 | 96.7  | - | 367 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 11 | 3.1  | 10.63 | 10.2  | 6.94 | 26.3 | 0.02 | 94.1  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 10 | 4.1  | 10.64 | 9.86  | 6.91 | 26.2 | 0.02 | 93.3  | - | 369 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 9  | 5.1  | 10.43 | 9.79  | 6.92 | 26.4 | 0.02 | 91.3  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 8  | 6.2  | 10.43 | 9.74  | 6.93 | 26.5 | 0.02 | 91.3  | - | 368 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 7  | 7.2  | 10.61 | 9.64  | 6.92 | 26.3 | 0.02 | 92.6  | - | 367 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 6  | 8.2  | 10.61 | 9.53  | 6.94 | 26.4 | 0.02 | 92.4  | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 5  | 9.2  | 10.64 | 9.43  | 6.93 | 26.4 | 0.02 | 92.4  | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 4  | 10.2 | 10.64 | 9.3   | 6.93 | 26.5 | 0.02 | 92.1  | - | 366 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 3  | 11.2 | 10.61 | 9.12  | 6.94 | 26.7 | 0.02 | 91.5  | - | 365 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 2  | 12.3 | 10.56 | 8.86  | 6.93 | 27.5 | 0.02 | 90.3  | - | 365 |
| Carey Bay | 2197 | DBSS | 5/29/97 | - | 1.4 | 1  | 13.2 | 10.64 | 8.66  | 6.94 | 27.8 | 0.02 | 90.7  | - | 364 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 14 | 0.7  | 10.87 | 15.48 | 7.17 | 31.2 | 0.02 | 108.7 | - | 411 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 13 | 1.6  | 11.2  | 14.35 | 7.1  | 30.9 | 0.02 | 109.3 | - | 414 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 2.5  | 11.22 | 11.79 | 7.09 | 29.1 | 0.02 | 103.4 | - | 415 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 3.5  | 11.32 | 11.3  | 7.07 | 28.9 | 0.02 | 103.2 | - | 416 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 4.5  | 11.27 | 11.22 | 7.06 | 28.8 | 0.02 | 102.5 | - | 417 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 5.5  | 10.97 | 10.99 | 7.03 | 28.8 | 0.02 | 99.2  | - | 418 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 6.5  | 10.9  | 10.38 | 7.01 | 28.8 | 0.02 | 97.2  | - | 419 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 7.5  | 10.84 | 10.17 | 7.01 | 28.7 | 0.02 | 96.2  | - | 419 |

|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 8.5  | 10.69 | 9.91  | 6.98 | 28.9 | 0.02 | 94.2  | - | 420 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 9.5  | 10.63 | 9.45  | 6.98 | 29.4 | 0.02 | 92.7  | - | 420 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 10.5 | 10.61 | 9.22  | 6.97 | 30.2 | 0.02 | 92    | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 11.4 | 10.58 | 8.99  | 6.97 | 31.1 | 0.02 | 91.2  | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 12.5 | 10.56 | 8.84  | 6.96 | 31.9 | 0.02 | 90.6  | - | 421 |
| Carey Bay | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 13.4 | 10.14 | 8.11  | 6.95 | 36.1 | 0.02 | 85.6  | - | 423 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 12 | 0.3  | 10.81 | 14.98 | 7.2  | 30.4 | 0.02 | 106.8 | - | 427 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 0.9  | 10.81 | 14.93 | 7.17 | 30.4 | 0.02 | 106.8 | - | 429 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 1.8  | 10.85 | 14.49 | 7.11 | 30.2 | 0.02 | 106.1 | - | 432 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 2.8  | 10.91 | 13.73 | 7.12 | 29.9 | 0.02 | 104.9 | - | 431 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 3.8  | 10.94 | 13.11 | 7.1  | 29.8 | 0.02 | 103.8 | - | 432 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 4.8  | 10.91 | 12.73 | 7.06 | 29.6 | 0.02 | 102.6 | - | 434 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 5.7  | 10.84 | 12.33 | 6.99 | 29.8 | 0.02 | 101.1 | - | 436 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 6.8  | 10.61 | 11.74 | 6.94 | 29.5 | 0.02 | 97.5  | - | 438 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 7.7  | 10.42 | 10.66 | 6.92 | 29.3 | 0.02 | 93.4  | - | 440 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 3  | 8.7  | 10.01 | 10.1  | 6.89 | 29.9 | 0.02 | 88.6  | - | 441 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 2  | 9.7  | 9.88  | 9.97  | 6.91 | 30.4 | 0.02 | 87.1  | - | 440 |
| Carey Bay | 2597 | DBSS | 6/26/97 | - | 2.5 | 1  | 10.6 | 9.78  | 9.66  | 6.92 | 31.3 | 0.02 | 85.6  | - | 440 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 13 | 0.3  | 10.09 | 18.7  | 7.18 | 36.8 | 0.02 | ****  | - | 444 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 12 | 1.1  | 10.09 | 18.66 | 7.17 | 36.9 | 0.02 | ****  | - | 444 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 11 | 1.9  | 10.24 | 17.47 | 7.13 | 36.3 | 0.02 | ****  | - | 446 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 10 | 2.9  | 10.39 | 16.41 | 7.09 | 37.1 | 0.02 | ****  | - | 449 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 9  | 3.9  | 10.25 | 16.19 | 7.04 | 37.4 | 0.02 | ****  | - | 450 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 8  | 5    | 10.47 | 15.01 | 7.01 | 36.9 | 0.02 | ****  | - | 452 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 7  | 5.9  | 10.37 | 14.42 | 6.97 | 35.9 | 0.02 | ****  | - | 454 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 6  | 7.1  | 10.16 | 13.17 | 6.94 | 34.3 | 0.02 | 96.2  | - | 455 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 5  | 8    | 10    | 12.81 | 6.93 | 34.2 | 0.02 | 93.9  | - | 456 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 4  | 9    | 9.91  | 12.15 | 6.91 | 33.4 | 0.02 | 91.7  | - | 457 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 3  | 10   | 9.55  | 11.25 | 6.9  | 33   | 0.02 | 86.5  | - | 458 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 2  | 11   | 9.03  | 9.98  | 6.91 | 34.7 | 0.02 | 79.4  | - | 460 |
| Carey Bay | 2797 | DBSS | 7/8/97  | - | 2.9 | 1  | 11.9 | 9.2   | 9.77  | 6.97 | 35.1 | 0.02 | 80.4  | - | 457 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 14 | 0.3  | 9.94  | 20.53 | 7.53 | 40   | 0.03 | ****  | - | 408 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 13 | 0.8  | 9.99  | 20.55 | 7.51 | 40   | 0.03 | ****  | - | 408 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 12 | 1.8  | 10.43 | 19.63 | 7.5  | 39.7 | 0.03 | ****  | - | 410 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 11 | 2.8  | 10.53 | 18.87 | 7.48 | 38.4 | 0.02 | ****  | - | 410 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 10 | 3.7  | 10.45 | 18.25 | 7.34 | 38.9 | 0.02 | ****  | - | 414 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 9  | 4.8  | 10.36 | 17.29 | 7.24 | 37.6 | 0.02 | ****  | - | 418 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 8  | 5.8  | 10.3  | 16.48 | 7.19 | 36.7 | 0.02 | ****  | - | 419 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 7  | 6.7  | 10.29 | 16.07 | 7.16 | 36   | 0.02 | ****  | - | 419 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 6  | 7.9  | 9.77  | 15.61 | 7.07 | 35.2 | 0.02 | 97.7  | - | 422 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 5  | 8.4  | 10.09 | 15.67 | 7.02 | 35.3 | 0.02 | ****  | - | 422 |



|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-----------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 4  | 9.4  | 9.66  | 14.52 | 6.98 | 33.7 | 0.02 | 94.3  | - | 423 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 3  | 10.4 | 9.48  | 13.88 | 6.94 | 33.3 | 0.02 | 91.2  | - | 424 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 2  | 11.4 | 8.87  | 12.67 | 6.91 | 33.1 | 0.02 | 83    | - | 426 |
| Carey Bay | 2997 | SSJD | 7/24/97 | - | 5.5 | 1  | 12.4 | 7.87  | 10.18 | 6.93 | 36.1 | 0.02 | 69.6  | - | 427 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 12 | 0.3  | 9.47  | 23.01 | 7.6  | 43.4 | 0.03 | 109   | - | 364 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 11 | 1    | 9.51  | 22.72 | 7.59 | 43.5 | 0.03 | 108.8 | - | 365 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 10 | 2.1  | 9.61  | 22.15 | 7.53 | 43.8 | 0.03 | 108.7 | - | 366 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 9  | 3    | 9.66  | 21.88 | 7.5  | 42.6 | 0.03 | 108.8 | - | 367 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 8  | 3.9  | 9.81  | 21.53 | 7.47 | 43.8 | 0.03 | 109.7 | - | 368 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 7  | 4.9  | 9.89  | 21.11 | 7.42 | 43   | 0.03 | 109.5 | - | 369 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 6  | 6    | 9.91  | 20.84 | 7.34 | 41.5 | 0.03 | 109.4 | - | 371 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 5  | 6.9  | 9.66  | 19.4  | 7.27 | 39.5 | 0.03 | 103.6 | - | 374 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 4  | 7.9  | 10.34 | 18.36 | 7.26 | 37.2 | 0.02 | 108.6 | - | 373 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 3  | 9    | 9.68  | 17.1  | 7.14 | 36.6 | 0.02 | 99.1  | - | 377 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 2  | 10   | 8.87  | 15.07 | 7.06 | 35.5 | 0.02 | 86.9  | - | 380 |
| Carey Bay | 3197 | DBSS | 8/5/97  | - | 5.3 | 1  | 11   | 7.89  | 14.19 | 7.1  | 35.6 | 0.02 | 75.9  | - | 381 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 13 | 0.4  | 8.59  | 22.47 | 7.5  | 45.1 | 0.03 | 98.4  | - | 378 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 12 | 1.4  | 8.6   | 22.43 | 7.54 | 45.1 | 0.03 | 98.5  | - | 376 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 11 | 2.4  | 8.65  | 22.33 | 7.47 | 45.1 | 0.03 | 98.9  | - | 377 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 10 | 3.3  | 8.67  | 22.22 | 7.42 | 45.3 | 0.03 | 98.8  | - | 378 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 9  | 4.3  | 8.66  | 21.9  | 7.3  | 45.9 | 0.03 | 98.2  | - | 383 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 8  | 5.3  | 8.61  | 21.55 | 7.25 | 45.7 | 0.03 | 97    | - | 385 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 7  | 6.4  | 8.62  | 21.19 | 7.2  | 45.6 | 0.03 | 96.4  | - | 386 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 6  | 7.4  | 8.39  | 20.53 | 7.15 | 43   | 0.03 | 92.6  | - | 388 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 5  | 8.4  | 8.64  | 19.56 | 7.01 | 40.1 | 0.03 | 93.5  | - | 391 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 4  | 9.4  | 8.27  | 18.94 | 6.95 | 39.7 | 0.03 | 88.4  | - | 393 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 3  | 10.4 | 8.26  | 18.41 | 6.88 | 38.9 | 0.02 | 87.5  | - | 395 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 2  | 11.5 | 7.89  | 15.96 | 6.83 | 36.5 | 0.02 | 79.3  | - | 398 |
| Carey Bay | 3297 | DBSS | 8/14/97 | - | 8   | 1  | 12.4 | 6.74  | 13.07 | 6.82 | 37.1 | 0.02 | 62.7  | - | 401 |
|           |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 14 | 0.3  | 8.97  | 21.76 | 7.5  | 49.7 | 0.03 | 101.6 | - | 356 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 13 | 0.9  | 8.98  | 21.74 | 7.52 | 49.6 | 0.03 | 101.6 | - | 355 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 12 | 1.9  | 8.99  | 21.67 | 7.5  | 49.6 | 0.03 | 101.6 | - | 356 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 11 | 2.9  | 9     | 21.53 | 7.45 | 49.8 | 0.03 | 101.5 | - | 358 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 10 | 3.9  | 8.93  | 21.37 | 7.43 | 49.3 | 0.03 | 100.4 | - | 358 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 9  | 4.9  | 8.91  | 21.26 | 7.39 | 49.3 | 0.03 | 99.9  | - | 359 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 8  | 6.1  | 8.89  | 21.09 | 7.36 | 49.3 | 0.03 | 99.4  | - | 360 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 7  | 7    | 8.95  | 20.91 | 7.34 | 49.6 | 0.03 | 99.7  | - | 360 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 6  | 8    | 8.92  | 20.84 | 7.28 | 49.5 | 0.03 | 99.2  | - | 362 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 5  | 9    | 8.84  | 20.7  | 7.18 | 49.1 | 0.03 | 98.1  | - | 365 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 4  | 10   | 8.6   | 20.42 | 7.08 | 48.3 | 0.03 | 95    | - | 370 |
| Carey Bay | 3497 | DBSS | 8/27/97 | - | 8.2 | 3  | 11   | 8.16  | 16.21 | 6.99 | 40.9 | 0.03 | 82.6  | - | 376 |

|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
|-----------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-----|-----|
| Carey Bay | 3497 | DBSS | 8/27/97  | - | 8.2 | 2  | 11.9 | 8.02  | 15.24 | 7.04 | 40.2 | 0.03 | 79.5 | -   | 374 |
| Carey Bay | 3497 | DBSS | 8/27/97  | - | 8.2 | 1  | 12.9 | 8.03  | 14.64 | 7.15 | 40.2 | 0.03 | 78.6 | -   | 372 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 13 | 0.5  | 8.05  | 14.69 | 6.76 | 44.3 | 0.03 | 80.1 | -   | 408 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 12 | 1.5  | 7.89  | 14.03 | 6.72 | 43.7 | 0.03 | 77.4 | -   | 411 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 11 | 2.5  | 7.64  | 11.22 | 6.71 | 42.5 | 0.03 | 70.3 | -   | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 10 | 3.5  | 7.84  | 9.99  | 6.71 | 42.4 | 0.03 | 70.1 | -   | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 9  | 4.6  | 8.21  | 9.33  | 6.7  | 42.9 | 0.03 | 72.1 | -   | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 8  | 5.5  | 8.13  | 9.17  | 6.71 | 42.9 | 0.03 | 71.1 | -   | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 7  | 6.5  | 8.2   | 9.1   | 6.72 | 43.1 | 0.03 | 71.8 | -   | 415 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 6  | 7.5  | 8.27  | 8.99  | 6.73 | 43.2 | 0.03 | 72.2 | -   | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 5  | 8.5  | 8.31  | 8.89  | 6.75 | 43.3 | 0.03 | 72.3 | -   | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 4  | 9.3  | 8.34  | 8.78  | 6.75 | 43.3 | 0.03 | 72.3 | -   | 414 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 3  | 10.7 | 8.67  | 8.6   | 6.77 | 43.5 | 0.03 | 75.2 | -   | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 2  | 11.5 | 8.4   | 8.6   | 6.78 | 43.3 | 0.03 | 72.7 | -   | 413 |
| Carey Bay | 3797 | ASRP | 9/17/97  | - | 3.5 | 1  | 12.5 | 8.46  | 8.48  | 6.82 | 43.4 | 0.03 | 73.3 | -   | 412 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 13 | 0.2  | 9.16  | 16.28 | 7.09 | 50   | 0.03 | 94.3 | -   | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 12 | 1    | 9.09  | 15.89 | 7.08 | 49.9 | 0.03 | 92.9 | -   | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 11 | 2    | 9.02  | 15.58 | 7.07 | 49.9 | 0.03 | 91.6 | -   | 411 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 10 | 3    | 8.95  | 15.44 | 7.04 | 50   | 0.03 | 90.6 | -   | 412 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 9  | 4    | 8.83  | 15.34 | 6.99 | 49.9 | 0.03 | 89.2 | -   | 413 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 8  | 5    | 8.59  | 15.11 | 6.92 | 49.4 | 0.03 | 86.3 | -   | 415 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 7  | 6    | 7.93  | 14.51 | 6.8  | 47.5 | 0.03 | 78.7 | -   | 419 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 6  | 7    | 7.68  | 12.41 | 6.77 | 44.2 | 0.03 | 72.7 | -   | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 5  | 8    | 7.63  | 12.32 | 6.77 | 44.2 | 0.03 | 72.1 | -   | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 4  | 9    | 7.56  | 12.14 | 6.77 | 43.9 | 0.03 | 71.1 | -   | 420 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 3  | 10   | 7.3   | 10.38 | 6.75 | 42.4 | 0.03 | 65.9 | -   | 421 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 2  | 11   | 7.79  | 8.86  | 6.84 | 42.2 | 0.03 | 67.8 | -   | 419 |
| Carey Bay | 3997 | ASRA | 9/29/97  | - | 6.3 | 1  | 12   | 8.33  | 8.37  | 6.93 | 42.3 | 0.03 | 71.6 | -   | 416 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 10 | 0.3  | 10.25 | 11.54 | 7.25 | 49   | 0.03 | 93.6 | -   | 401 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 9  | 1.7  | 10.24 | 11.56 | 7.23 | 49.1 | 0.03 | 93.6 | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 8  | 2.7  | 10.24 | 11.54 | 7.24 | 49.2 | 0.03 | 93.5 | -   | 401 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 7  | 3.7  | 10.21 | 11.54 | 7.21 | 49.2 | 0.03 | 93.2 | -   | 403 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 6  | 4.7  | 10.18 | 11.54 | 7.23 | 49.1 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 5  | 5.7  | 10.19 | 11.53 | 7.22 | 49.2 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 4  | 6.7  | 10.19 | 11.53 | 7.22 | 49.3 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 3  | 7.7  | 10.19 | 11.53 | 7.22 | 49.3 | 0.03 | 93   | -   | 402 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 2  | 8.7  | 10.1  | 11.53 | 7.21 | 49.3 | 0.03 | 92.2 | -   | 403 |
| Carey Bay | 4297 | DBAS | 10/22/97 | - | 4.5 | 1  | 9.7  | 10.05 | 11.51 | 7.19 | 49.2 | 0.03 | 91.7 | -   | 404 |
|           |      |      |          |   |     |    |      |       |       |      |      |      |      |     |     |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 13 | 0.4  | 10.06 | 9.87  | 7.07 | 53.3 | 0.03 | 88   | 56  | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97  | - | 3.8 | 12 | 1.5  | 10.34 | 9.55  | 7.09 | 53.1 | 0.03 | 89.8 | 215 | 390 |

|           |      |      |         |   |     |    |      |       |      |      |      |      |      |     |     |
|-----------|------|------|---------|---|-----|----|------|-------|------|------|------|------|------|-----|-----|
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 11 | 2.5  | 10.46 | 9.5  | 7.12 | 52.9 | 0.03 | 90.7 | 54  | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 10 | 3.5  | 10.5  | 9.53 | 7.11 | 53   | 0.03 | 91.1 | 41  | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 9  | 4.5  | 10.5  | 9.45 | 7.12 | 52.9 | 0.03 | 90.9 | 56  | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 8  | 5.5  | 10.45 | 9.4  | 7.11 | 53   | 0.03 | 90.4 | 113 | 389 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 7  | 6.5  | 10.28 | 9.33 | 7.06 | 53   | 0.03 | 88.8 | 103 | 390 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 6  | 7.5  | 10.14 | 9.22 | 7.01 | 52.9 | 0.03 | 87.3 | 210 | 392 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 5  | 8.5  | 9.65  | 8.89 | 6.93 | 53.1 | 0.03 | 82.5 | 101 | 393 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 4  | 9.5  | 9.51  | 8.81 | 6.89 | 52.9 | 0.03 | 81.1 | 54  | 394 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 3  | 10.6 | 9.08  | 8.58 | 6.86 | 52.4 | 0.03 | 77   | 37  | 395 |
| Carey Bay | 4497 | DBAS | 11/4/97 | - | 3.8 | 2  | 11.5 | 8.98  | 8.47 | 6.87 | 52.1 | 0.03 | 76   | 111 | 394 |

| Location       | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|----------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 13       | 0.4       | 12.35                   | 6.31            | 7.45 | 36.4                 | 0.02 | 99.9                       | -                  | 372   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 12       | 1.9       | 12.27                   | 6.23            | 7.43 | 36.4                 | 0.02 | 99.2                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 11       | 3.2       | 12.29                   | 6.2             | 7.43 | 36.6                 | 0.02 | 99.2                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 10       | 4.5       | 12.27                   | 6.13            | 7.43 | 36.7                 | 0.02 | 98.7                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 9        | 5.7       | 12.23                   | 6.05            | 7.42 | 37                   | 0.02 | 98.4                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 8        | 7         | 12.26                   | 5.97            | 7.41 | 37.4                 | 0.02 | 98.4                       | -                  | 374   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 7        | 8.2       | 12.24                   | 5.97            | 7.4  | 37.3                 | 0.02 | 98.2                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 6        | 9.4       | 12.26                   | 5.9             | 7.4  | 37.6                 | 0.02 | 98.2                       | -                  | 373   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 5        | 10.4      | 12.24                   | 5.82            | 7.39 | 37.8                 | 0.02 | 97.9                       | -                  | 374   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 4        | 11.8      | 12.21                   | 5.62            | 7.38 | 38.2                 | 0.02 | 97.1                       | -                  | 374   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 3        | 13.1      | 12.2                    | 5.57            | 7.37 | 38.3                 | 0.02 | 96.9                       | -                  | 374   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 2        | 14.4      | 12.14                   | 5.51            | 7.36 | 38.2                 | 0.02 | 96.3                       | -                  | 374   |
| Conkling Point | 1597  | ASSS     | 4/18/97 | -    | 1.1        | 1        | 15.9      | 12.16                   | 5.29            | 7.36 | 38.5                 | 0.02 | 95.9                       | -                  | 374   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 12       | 0.2       | 10.99                   | 11.35           | 6.99 | 28                   | 0.02 | 99.6                       | -                  | 367   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 11       | 2.1       | 11.1                    | 9.61            | 6.98 | 27.8                 | 0.02 | 96.7                       | -                  | 369   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 10       | 3.9       | 11.09                   | 9.46            | 7    | 27.8                 | 0.02 | 96.2                       | -                  | 368   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 9        | 5.5       | 11.08                   | 9.42            | 6.98 | 27.7                 | 0.02 | 96                         | -                  | 369   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 8        | 7         | 11.05                   | 9.35            | 6.97 | 27.9                 | 0.02 | 95.6                       | -                  | 369   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 7        | 8.6       | 11.04                   | 9.28            | 6.98 | 28.2                 | 0.02 | 95.3                       | -                  | 368   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 6        | 10.1      | 11.03                   | 9.28            | 6.97 | 28.2                 | 0.02 | 95.3                       | -                  | 368   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 5        | 11.6      | 11.01                   | 9.15            | 6.97 | 28.6                 | 0.02 | 94.7                       | -                  | 368   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 4        | 13        | 11.03                   | 8.35            | 6.95 | 31.4                 | 0.02 | 93.1                       | -                  | 369   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 3        | 14.5      | 11.15                   | 7.45            | 6.93 | 34.8                 | 0.02 | 92.1                       | -                  | 369   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 2        | 16        | 10.71                   | 6.08            | 6.89 | 41                   | 0.03 | 85.6                       | -                  | 372   |
| Conkling Point | 1997  | DBAC     | 5/16/97 | -    | 0.8        | 1        | 17.5      | 10.51                   | 5.9             | 6.89 | 44.8                 | 0.03 | 83.4                       | -                  | 370   |
| Conkling Point | 2197  | DBSS     | 5/29/97 | -    | 1.8        | 12       | 0.2       | 11.09                   | 11.41           | 7.01 | 26.9                 | 0.02 | 101                        | -                  | 362   |
| Conkling Point | 2197  | DBSS     | 5/29/97 | -    | 1.8        | 11       | 1.4       | 11.11                   | 10.79           | 7.02 | 26.9                 | 0.02 | 99.7                       | -                  | 363   |
| Conkling Point | 2197  | DBSS     | 5/29/97 | -    | 1.8        | 10       | 2.9       | 11.16                   | 10.36           | 7    | 26.8                 | 0.02 | 99                         | -                  | 364   |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 9  | 4.5  | 11.12 | 10.22 | 6.98 | 26.8 | 0.02 | 98.4  | - | 365 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 8  | 6    | 11.12 | 10.05 | 6.96 | 26.9 | 0.02 | 98    | - | 365 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 7  | 7.4  | 11.05 | 9.74  | 6.94 | 26.8 | 0.02 | 96.6  | - | 364 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 6  | 8.9  | 11.01 | 9.61  | 6.93 | 26.9 | 0.02 | 96    | - | 364 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 5  | 10.4 | 10.97 | 9.43  | 6.92 | 26.9 | 0.02 | 95.3  | - | 363 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 4  | 12.1 | 10.89 | 8.96  | 6.91 | 27.2 | 0.02 | 93.5  | - | 362 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 3  | 13.4 | 10.75 | 8.71  | 6.9  | 27.8 | 0.02 | 91.8  | - | 362 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 2  | 15   | 10.75 | 7.97  | 6.88 | 31.3 | 0.02 | 90.1  | - | 361 |
| Conkling Point | 2197 | DBSS | 5/29/97 | - | 1.8 | 1  | 16.8 | 10.29 | 7.01  | 6.86 | 36   | 0.02 | 84.3  | - | 360 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 12 | 0.2  | 10.9  | 15.61 | 7.09 | 30.8 | 0.02 | 109.4 | - | 395 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 11 | 1    | 10.7  | 14.08 | 7.08 | 30.1 | 0.02 | 103.8 | - | 396 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 2.5  | 10.95 | 12.4  | 7.06 | 29.4 | 0.02 | 102.2 | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 4    | 10.98 | 11.91 | 7.04 | 29.1 | 0.02 | 101.5 | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 5.3  | 10.87 | 11.77 | 6.99 | 29.2 | 0.02 | 100.1 | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 6.9  | 10.85 | 10.68 | 6.99 | 29.6 | 0.02 | 97.3  | - | 400 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 8.3  | 10.82 | 9.94  | 6.98 | 29.1 | 0.02 | 95.5  | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 9.9  | 10.78 | 9.78  | 6.97 | 28.9 | 0.02 | 94.7  | - | 399 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 11.4 | 10.73 | 9.33  | 6.96 | 29.8 | 0.02 | 93.3  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 13   | 10.53 | 8.72  | 6.94 | 32.2 | 0.02 | 90.2  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 14.5 | 10.48 | 8.47  | 6.93 | 33.3 | 0.02 | 89.3  | - | 398 |
| Conkling Point | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 16.2 | 10.17 | 7.86  | 6.93 | 37.1 | 0.02 | 85.3  | - | 395 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 16 | 0.4  | 10.79 | 15.18 | 7.23 | 30.6 | 0.02 | 107.1 | - | 421 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 15 | 1    | 10.79 | 15.14 | 7.23 | 30.6 | 0.02 | 107.1 | - | 422 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 14 | 2.2  | 10.8  | 14.94 | 7.21 | 30.7 | 0.02 | 106.7 | - | 424 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 13 | 3.3  | 10.82 | 14.66 | 7.15 | 31   | 0.02 | 106.2 | - | 426 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 12 | 4.2  | 10.75 | 14.46 | 7.11 | 30.9 | 0.02 | 105   | - | 427 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 5.2  | 10.87 | 13.07 | 7.06 | 31.2 | 0.02 | 103   | - | 430 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 6.2  | 10.75 | 12.66 | 7.02 | 30.8 | 0.02 | 101   | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 7.2  | 10.78 | 12.17 | 7    | 30.6 | 0.02 | 100.1 | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 8.2  | 10.73 | 12    | 6.95 | 30.8 | 0.02 | 99.3  | - | 431 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 9.3  | 10.02 | 10.77 | 6.9  | 29.5 | 0.02 | 90    | - | 434 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 10.2 | 10.21 | 9.2   | 6.83 | 31.2 | 0.02 | 88.4  | - | 435 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 11.3 | 9.97  | 8.66  | 6.8  | 32.7 | 0.02 | 85.2  | - | 435 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 12.4 | 9.36  | 8.27  | 6.78 | 35.8 | 0.02 | 79.3  | - | 436 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 3  | 13.3 | 9.52  | 7.91  | 6.8  | 37.2 | 0.02 | 79.9  | - | 434 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 2  | 14.4 | 9.05  | 7.56  | 6.8  | 38.9 | 0.02 | 75.3  | - | 434 |
| Conkling Point | 2597 | DBSS | 6/26/97 | - | 2.5 | 1  | 15.3 | 8.74  | 7.43  | 6.82 | 40.2 | 0.03 | 72.5  | - | 433 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 16 | 0.3  | 10.17 | 18.54 | 7.16 | 36.5 | 0.02 | ****  | - | 434 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 15 | 1.1  | 10.15 | 18    | 7.13 | 35.6 | 0.02 | ****  | - | 436 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 14 | 2    | 10.31 | 17.25 | 7.11 | 36.3 | 0.02 | ****  | - | 437 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 13 | 3    | 10.36 | 16.56 | 7.09 | 36.9 | 0.02 | ****  | - | 438 |

|                |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 12 | 3.9  | 10.45 | 16.12 | 7.06 | 36.7 | 0.02 | ****  | - | 439 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 11 | 5    | 10.41 | 15.54 | 7.02 | 36.3 | 0.02 | ****  | - | 440 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 10 | 6    | 10.41 | 14.52 | 7    | 35.3 | 0.02 | ****  | - | 441 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 9  | 7.1  | 10.1  | 13.28 | 6.92 | 33.6 | 0.02 | 95.9  | - | 444 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 8  | 8    | 10.03 | 12.71 | 6.89 | 33.8 | 0.02 | 93.9  | - | 445 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 7  | 9    | 9.76  | 12.11 | 6.87 | 33.3 | 0.02 | 90.2  | - | 446 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 6  | 10   | 9.26  | 11.34 | 6.84 | 33.2 | 0.02 | 84    | - | 447 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 5  | 11   | 8.94  | 10.69 | 6.84 | 33.3 | 0.02 | 79.9  | - | 447 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 4  | 11.9 | 8.68  | 9.59  | 6.84 | 35   | 0.02 | 75.5  | - | 448 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 3  | 12.8 | 8.73  | 8.72  | 6.86 | 36.7 | 0.02 | 74.5  | - | 447 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 2  | 13.9 | 8.74  | 8.45  | 6.88 | 37.7 | 0.02 | 73.9  | - | 448 |
| Conkling Point | 2797 | DBSS | 7/8/97  | - | 2.8 | 1  | 14.9 | 8.73  | 8.22  | 6.92 | 38.6 | 0.02 | 73.6  | - | 446 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 10 | 0.4  | 9.56  | 21.37 | 7.46 | 39.7 | 0.03 | ****  | - | 408 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 9  | 2    | 9.59  | 20.79 | 7.37 | 37.6 | 0.02 | ****  | - | 411 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 8  | 3.5  | 10.4  | 18.18 | 7.26 | 38.4 | 0.02 | ****  | - | 417 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 7  | 5    | 10.43 | 17.3  | 7.13 | 38   | 0.02 | ****  | - | 420 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 6  | 6.5  | 10    | 16.31 | 7.05 | 35.9 | 0.02 | ****  | - | 422 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 5  | 8    | 10.02 | 15.41 | 6.93 | 34.8 | 0.02 | 99.8  | - | 427 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 4  | 9.5  | 9.46  | 12.97 | 6.83 | 32.6 | 0.02 | 89.2  | - | 431 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 3  | 11   | 9.31  | 12.3  | 6.77 | 32   | 0.02 | 86.4  | - | 433 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 2  | 12.5 | 8.63  | 9.87  | 6.74 | 33.2 | 0.02 | 75.8  | - | 435 |
| Conkling Point | 2997 | SSJD | 7/24/97 | - | 4.5 | 1  | 14   | 8.17  | 8.72  | 6.74 | 36.4 | 0.02 | 69.7  | - | 435 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 15 | 0.3  | 9.57  | 24    | 8.18 | 41.5 | 0.03 | 112.2 | - | 329 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 14 | 1.8  | 9.56  | 22.25 | 7.5  | 43.2 | 0.03 | 108.4 | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 13 | 2.7  | 9.49  | 22.04 | 7.46 | 42.3 | 0.03 | 107.2 | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 12 | 3.6  | 9.48  | 21.7  | 7.39 | 41.4 | 0.03 | 106.4 | - | 352 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 11 | 4.7  | 9.88  | 21.04 | 7.33 | 43.5 | 0.03 | 109.5 | - | 352 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 10 | 5.6  | 9.4   | 20.23 | 7.19 | 40.7 | 0.03 | 102.5 | - | 356 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 9  | 6.7  | 9.32  | 19.21 | 7.13 | 39.5 | 0.03 | 99.6  | - | 356 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 8  | 7.7  | 9.69  | 18.15 | 7.13 | 37.4 | 0.02 | 101.4 | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 7  | 8.7  | 10.04 | 17.18 | 7.09 | 36.3 | 0.02 | 103   | - | 353 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 6  | 9.7  | 9.28  | 15.82 | 6.99 | 35.4 | 0.02 | 92.4  | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 5  | 10.7 | 8.73  | 14.39 | 6.93 | 34.7 | 0.02 | 84.3  | - | 354 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 4  | 11.8 | 8.22  | 12.28 | 6.91 | 33.7 | 0.02 | 75.9  | - | 350 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 3  | 12.6 | 8.21  | 11.23 | 6.93 | 34.5 | 0.02 | 73.8  | - | 346 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 2  | 13.7 | 8     | 10.16 | 6.95 | 35.9 | 0.02 | 70.2  | - | 342 |
| Conkling Point | 3197 | DBSS | 8/5/97  | - | 4.1 | 1  | 14.8 | 7.22  | 9.44  | 7.02 | 38.4 | 0.02 | 62.2  | - | 388 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 11 | 0.3  | 8.64  | 22.75 | 7.98 | 44.7 | 0.03 | 99.5  | - | 368 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 10 | 2.1  | 8.73  | 22.56 | 8.11 | 44.7 | 0.03 | 100.3 | - | 361 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 9  | 3.6  | 8.81  | 22.33 | 7.87 | 44.3 | 0.03 | 100.7 | - | 363 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 8  | 5    | 8.57  | 21.86 | 7.31 | 45.2 | 0.03 | 97.1  | - | 384 |

|                |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|----------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 7  | 6.5  | 8.56 | 21.26 | 7.14 | 46.4 | 0.03 | 95.9  | - | 389 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 6  | 8    | 8.56 | 20.65 | 7.06 | 43.6 | 0.03 | 94.7  | - | 391 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 5  | 9.6  | 8.58 | 18.75 | 6.94 | 39   | 0.03 | 91.4  | - | 396 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 4  | 11   | 7.69 | 15.02 | 6.9  | 35.7 | 0.02 | 75.8  | - | 399 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 3  | 12.6 | 6.19 | 15.07 | 6.87 | 36.5 | 0.02 | 77    | - | 399 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 2  | 14   | 6.57 | 9.2   | 6.95 | 39.1 | 0.03 | 56.7  | - | 401 |
| Conkling Point | 3297 | DBSS | 8/14/97 | - | 5   | 1  | 15.6 | 6.79 | 8.9   | 7.08 | 39.3 | 0.03 | 58.2  | - | 398 |
|                |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 11 | 0.4  | 9.02 | 21.92 | 7.53 | 50   | 0.03 | 102.3 | - | 346 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 10 | 2    | 9.03 | 21.7  | 7.51 | 50   | 0.03 | 102.1 | - | 346 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 9  | 3.6  | 9    | 21.56 | 7.45 | 50   | 0.03 | 101.6 | - | 347 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 8  | 5    | 9.01 | 21.44 | 7.4  | 50.3 | 0.03 | 101.4 | - | 349 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 7  | 6.6  | 8.84 | 21.14 | 7.33 | 49.2 | 0.03 | 98.9  | - | 350 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 6  | 8    | 8.73 | 21    | 7.17 | 49.5 | 0.03 | 97.4  | - | 355 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 5  | 9.5  | 8.19 | 20.49 | 6.96 | 49.5 | 0.03 | 90.5  | - | 363 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 4  | 11   | 7.84 | 17.3  | 6.89 | 42.7 | 0.03 | 81.2  | - | 368 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 3  | 12.6 | 8.37 | 15.61 | 6.88 | 40.2 | 0.03 | 83.6  | - | 368 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 2  | 14.1 | 7.65 | 12.49 | 6.86 | 39.9 | 0.03 | 71.3  | - | 369 |
| Conkling Point | 3497 | DBSS | 8/27/97 | - | 7.2 | 1  | 15.8 | 6.43 | 9.48  | 6.87 | 42.8 | 0.03 | 56    | - | 369 |
|                |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 16 | 0.2  | 8.46 | 14.83 | 6.87 | 47.6 | 0.03 | 84.4  | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 15 | 1    | 8.42 | 14.86 | 6.86 | 47.6 | 0.03 | 84.1  | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 14 | 2    | 8.3  | 14.58 | 6.79 | 46.9 | 0.03 | 82.4  | - | 395 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 13 | 3    | 7.76 | 13.75 | 6.68 | 45.7 | 0.03 | 76.9  | - | 399 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 12 | 4    | 7.6  | 10.69 | 6.62 | 42.6 | 0.03 | 69.1  | - | 404 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 11 | 5    | 7.76 | 9.81  | 6.62 | 42.7 | 0.03 | 69.1  | - | 404 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 10 | 6    | 7.91 | 9.5   | 6.63 | 42.7 | 0.03 | 69.9  | - | 403 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 9  | 7    | 7.78 | 9.43  | 6.63 | 43.1 | 0.03 | 68.6  | - | 402 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 8  | 8    | 7.61 | 9.37  | 6.63 | 42.9 | 0.03 | 67.2  | - | 402 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 7  | 9    | 7.5  | 9.06  | 6.62 | 43   | 0.03 | 65.7  | - | 401 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 6  | 10   | 7.18 | 8.74  | 6.63 | 43.3 | 0.03 | 62.5  | - | 400 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 5  | 11   | 7.35 | 8.68  | 6.66 | 43.4 | 0.03 | 63.7  | - | 398 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 4  | 12   | 8.42 | 8.29  | 6.71 | 43.8 | 0.03 | 72.1  | - | 395 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 3  | 13   | 8.59 | 8.2   | 6.74 | 43.8 | 0.03 | 74.2  | - | 392 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 2  | 14.1 | 8.63 | 8.19  | 6.76 | 43.8 | 0.03 | 73.9  | - | 391 |
| Conkling Point | 3797 | ASRP | 9/17/97 | - | 3   | 1  | 15.2 | 8.7  | 8.19  | 6.79 | 43.8 | 0.03 | 74.5  | - | 389 |
|                |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 10 | 0.2  | 9.53 | 16.26 | 7.21 | 50.8 | 0.03 | 98.2  | - | 388 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 9  | 1.5  | 9.3  | 15.79 | 7.13 | 50.3 | 0.03 | 94.8  | - | 391 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 8  | 3    | 9.24 | 15.61 | 7.06 | 50.2 | 0.03 | 93.9  | - | 392 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 7  | 4.5  | 8.78 | 14.69 | 6.94 | 48.9 | 0.03 | 87.4  | - | 395 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 6  | 6    | 8.32 | 13.67 | 6.83 | 47.1 | 0.03 | 81    | - | 397 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 5  | 7.5  | 7.78 | 12.38 | 6.77 | 44.3 | 0.03 | 73.6  | - | 399 |
| Conkling Point | 3997 | ASRA | 9/29/97 | - | 6.2 | 4  | 9    | 7.84 | 10.9  | 6.76 | 44   | 0.03 | 71.7  | - | 398 |

|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
|----------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-------|-----|
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 3  | 10.5 | 7.38  | 9.64  | 6.77 | 42.2 | 0.03 | 65.1 | -     | 397 |
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 2  | 12   | 7.86  | 8.4   | 6.82 | 42.3 | 0.03 | 67.6 | -     | 395 |
| Conkling Point | 3997 | ASRA | 9/29/97  | - | 6.2 | 1  | 13.4 | 8.18  | 8.22  | 6.92 | 42.4 | 0.03 | 70.1 | -     | 389 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 16 | 0.2  | 10.3  | 11.54 | 7.28 | 48   | 0.03 | 94   | -     | 378 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 15 | 1    | 10.28 | 11.56 | 7.31 | 49   | 0.03 | 93.9 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 14 | 2    | 10.27 | 11.57 | 7.29 | 49.1 | 0.03 | 93.8 | -     | 377 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 13 | 3    | 10.26 | 11.54 | 7.31 | 49.2 | 0.03 | 93.7 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 12 | 4    | 10.28 | 11.54 | 7.29 | 49.3 | 0.03 | 93.8 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 11 | 5    | 10.25 | 11.54 | 7.29 | 49.1 | 0.03 | 93.6 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 10 | 6    | 10.28 | 11.54 | 7.29 | 49.2 | 0.03 | 93.8 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 9  | 7    | 10.26 | 11.54 | 7.28 | 49.3 | 0.03 | 93.7 | -     | 376 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 8  | 8    | 10.25 | 11.54 | 7.29 | 49.3 | 0.03 | 93.6 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 7  | 9    | 10.23 | 11.53 | 7.27 | 49.1 | 0.03 | 93.4 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 6  | 10   | 10.24 | 11.53 | 7.27 | 49.1 | 0.03 | 93.5 | -     | 375 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 5  | 11   | 10.25 | 11.51 | 7.31 | 49.3 | 0.03 | 93.5 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 4  | 12   | 10.26 | 11.5  | 7.3  | 49.3 | 0.03 | 93.4 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 3  | 13   | 10.25 | 11.43 | 7.3  | 49.3 | 0.03 | 93.4 | -     | 372 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 2  | 14   | 10.63 | 11.15 | 7.31 | 49.5 | 0.03 | 96.2 | -     | 371 |
| Conkling Point | 4297 | DBAS | 10/22/97 | - | 4.1 | 1  | 14.9 | 10.61 | 10.86 | 7.32 | 50   | 0.03 | 95.3 | -     | 370 |
|                |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 11 | 0.3  | 10.59 | 7.04  | 9.79 | 53   | 0.03 | 93.2 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 10 | 1.5  | 10.54 | 7.05  | 9.74 | 53   | 0.03 | 92.5 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 9  | 3    | 10.52 | 7.02  | 9.6  | 53.1 | 0.03 | 92   | ***** | 442 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 8  | 4.5  | 10.75 | 6.99  | 9.24 | 52.9 | 0.03 | 93.3 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 7  | 6    | 10.97 | 6.93  | 8.74 | 53.7 | 0.03 | 94   | ***** | 440 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 6  | 7.5  | 9.25  | 6.79  | 8.48 | 52.9 | 0.03 | 78.8 | ***** | 442 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 5  | 9    | 9.27  | 6.77  | 8.48 | 52.8 | 0.03 | 78.9 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 4  | 10.5 | 9.32  | 6.75  | 8.43 | 52.5 | 0.03 | 79.3 | ***** | 439 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 3  | 12   | 8.81  | 6.66  | 7.73 | 48.8 | 0.03 | 73.7 | ***** | 441 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 2  | 13.6 | 8.64  | 6.66  | 7.41 | 47   | 0.03 | 71.6 | ***** | 439 |
| Conkling Point | 4497 | DBAS | 11/3/97  | - | 3.1 | 1  | 14.9 | 8.65  | 6.71  | 7.37 | 46.7 | 0.03 | 71.7 | ***** | 441 |

| Location    | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Hidden Lake | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 5        | 0.2       | 12.49                   | 6.73            | 7.46 | 38.6                 | 0.02 | *****                      | -                  | 374   |
| Hidden Lake | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 4        | 2.3       | 12.51                   | 6.31            | 7.45 | 38.4                 | 0.02 | *****                      | -                  | 375   |
| Hidden Lake | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 3        | 3.7       | 12.54                   | 6.28            | 7.45 | 38.4                 | 0.02 | *****                      | -                  | 375   |
| Hidden Lake | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 2        | 5.3       | 12.47                   | 6.26            | 7.43 | 38.5                 | 0.02 | *****                      | -                  | 375   |
| Hidden Lake | 1597  | ASSS     | 4/18/97 | -    | 1.9        | 1        | 6.9       | 12.38                   | 6.23            | 7.42 | 38.3                 | 0.02 | 99.9                       | -                  | 375   |
|             |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Hidden Lake | 1997  | DBAC     | 5/16/97 | -    | 0.7        | 11       | 0.2       | 11.1                    | 11.97           | 7    | 29.9                 | 0.02 | 102.2                      | -                  | 371   |
| Hidden Lake | 1997  | DBAC     | 5/16/97 | -    | 0.7        | 10       | 1.4       | 11                      | 10.64           | 6.97 | 28.7                 | 0.02 | 98.1                       | -                  | 376   |

|             |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 9  | 2.4  | 11.06 | 9.38  | 6.97 | 28.8 | 0.02 | 95.8  | - | 377 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 8  | 3.4  | 11.03 | 9.3   | 6.97 | 29   | 0.02 | 95.3  | - | 377 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 7  | 4.5  | 11.03 | 9.27  | 6.96 | 29.3 | 0.02 | 95.3  | - | 377 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 6  | 5.4  | 11.01 | 9.25  | 6.98 | 29.4 | 0.02 | 95    | - | 375 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 5  | 6.4  | 10.95 | 9.17  | 6.96 | 30.2 | 0.02 | 94.4  | - | 376 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 4  | 7.5  | 10.89 | 9.1   | 6.98 | 29.1 | 0.02 | 93.6  | - | 375 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 3  | 8.5  | 10.87 | 9.07  | 6.98 | 28.8 | 0.02 | 93.4  | - | 374 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 2  | 9.3  | 10.86 | 9.04  | 6.99 | 28.9 | 0.02 | 93.3  | - | 373 |
| Hidden Lake | 1997 | DBAC | 5/16/97 | - | 0.7 | 1  | 10.4 | 10.86 | 8.97  | 6.99 | 30   | 0.02 | 93.1  | - | 373 |
|             |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 10 | 0.7  | 10.91 | 12.61 | 6.99 | 26.6 | 0.02 | 101.9 | - | 361 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 9  | 1.8  | 11.25 | 9.84  | 7    | 26.9 | 0.02 | 98.6  | - | 361 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 8  | 2.8  | 11.18 | 9.53  | 6.99 | 26.7 | 0.02 | 97.3  | - | 361 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 7  | 3.7  | 11.14 | 9.3   | 6.97 | 26.6 | 0.02 | 96.5  | - | 362 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 6  | 4.7  | 11.13 | 9.19  | 6.95 | 26.6 | 0.02 | 96.1  | - | 362 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 5  | 5.7  | 11.03 | 9.12  | 6.95 | 26.4 | 0.02 | 95.3  | - | 361 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 4  | 6.8  | 10.74 | 8.45  | 6.92 | 26.4 | 0.02 | 91.1  | - | 361 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 3  | 7.7  | 10.76 | 8.38  | 6.92 | 26.3 | 0.02 | 91.1  | - | 360 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 2  | 8.8  | 10.62 | 8.33  | 6.91 | 26.5 | 0.02 | 89.8  | - | 359 |
| Hidden Lake | 2197 | DBSS | 5/29/97 | - | 1.7 | 1  | 9.8  | 10    | 8.3   | 6.91 | 27   | 0.02 | 84.5  | - | 367 |
|             |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 10 | 0.5  | 11.03 | 13.17 | 7.09 | 29.3 | 0.02 | 104.9 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 9  | 1.6  | 11.24 | 12.28 | 7.11 | 29.1 | 0.02 | 104.7 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 8  | 2.6  | 11.15 | 11.98 | 7.1  | 29   | 0.02 | 103.2 | - | 405 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 7  | 3.6  | 11.13 | 11.91 | 7.06 | 29   | 0.02 | 102.9 | - | 407 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 6  | 4.3  | 11.1  | 10.59 | 7.02 | 28.5 | 0.02 | 99.5  | - | 409 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 5  | 5.3  | 10.97 | 10.02 | 7.01 | 28.2 | 0.02 | 97.1  | - | 410 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 4  | 6.3  | 10.96 | 9.61  | 6.99 | 27.7 | 0.02 | 96    | - | 410 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 3  | 7.2  | 10.72 | 9.35  | 6.98 | 27.7 | 0.02 | 93.3  | - | 411 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 2  | 8.1  | 10.16 | 9.13  | 6.98 | 28.1 | 0.02 | 88    | - | 411 |
| Hidden Lake | 2397 | DBAS | 6/11/97 | - | 2   | 1  | 8.2  | 10.15 | 9.12  | 6.98 | 28.1 | 0.02 | 87.8  | - | 411 |
|             |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 11 | 0.4  | 10.75 | 13.72 | 7.16 | 30.6 | 0.02 | 103.3 | - | 406 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 10 | 0.3  | 10.75 | 13.77 | 7.15 | 30.6 | 0.02 | 103.5 | - | 406 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 9  | 1.2  | 10.78 | 13.3  | 7.14 | 30.4 | 0.02 | 102.7 | - | 409 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 8  | 2.4  | 10.8  | 12.94 | 7.13 | 30.5 | 0.02 | 102   | - | 409 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 7  | 3.4  | 10.75 | 12.74 | 7.11 | 30.2 | 0.02 | 101   | - | 410 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 6  | 4.5  | 10.52 | 12.44 | 7.06 | 30.3 | 0.02 | 98.3  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 5  | 4.5  | 10.54 | 12.36 | 7.02 | 30.3 | 0.02 | 98.3  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 4  | 5.5  | 10.38 | 11.89 | 6.98 | 30   | 0.02 | 95.8  | - | 411 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 3  | 6.5  | 7.98  | 11.15 | 6.83 | 29.8 | 0.02 | 72.3  | - | 416 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 2  | 7.5  | 6.41  | 10.43 | 6.86 | 30.3 | 0.02 | 57.2  | - | 418 |
| Hidden Lake | 2597 | DBSS | 6/26/97 | - | 2.4 | 1  | 7.4  | 6.45  | 10.38 | 6.88 | 30.3 | 0.02 | 57.4  | - | 419 |
|             |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| Hidden Lake | 2797 | DBSS | 7/8/97  | - | 2.4 | 12 | 0.3  | 10.08 | 18.56 | 7.14 | 34.7 | 0.02 | ***** | - | 392 |





|             |      |      |          |   |     |    |     |       |       |      |      |      |       |       |     |
|-------------|------|------|----------|---|-----|----|-----|-------|-------|------|------|------|-------|-------|-----|
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 10 | 0.5 | 9.39  | 22.02 | 7.9  | 44.9 | 0.03 | 106.8 | -     | 235 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 9  | 0.3 | 9.39  | 22.02 | 7.9  | 44.9 | 0.03 | 106.8 | -     | 235 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 8  | 1.6 | 9.41  | 21.92 | 7.88 | 45   | 0.03 | 106.9 | -     | 231 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 7  | 2.4 | 9.43  | 21.88 | 7.83 | 45   | 0.03 | 107   | -     | 223 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 6  | 3.5 | 9.45  | 21.85 | 7.74 | 45.1 | 0.03 | 107.2 | -     | 219 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 5  | 4.6 | 9.33  | 21.58 | 7.44 | 44.9 | 0.03 | 105.3 | -     | 218 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 4  | 5.4 | 10.33 | 19.92 | 7.05 | 42.8 | 0.03 | 112.9 | -     | 228 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 3  | 6.5 | 3.7   | 17.44 | 6.73 | 47.1 | 0.03 | 38.4  | -     | 208 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 2  | 7.5 | 0.3   | 14.92 | 6.67 | 57.7 | 0.04 | 2.9   | -     | 147 |
| Hidden Lake | 3497 | DBSS | 8/27/97  | - | 5.7 | 1  | 8.4 | 0.4   | 13.88 | 6.6  | 79.9 | 0.05 | 3.9   | -     | 139 |
|             |      |      |          |   |     |    |     |       |       |      |      |      |       |       |     |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 9  | 0.2 | 8.74  | 18.01 | 7.31 | 47.1 | 0.03 | 92.1  | -     | 296 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 8  | 1.2 | 8.75  | 18    | 7.3  | 47.2 | 0.03 | 92.3  | -     | 295 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 7  | 2.2 | 8.68  | 17.9  | 7.27 | 47.1 | 0.03 | 91.3  | -     | 294 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 6  | 3.2 | 8.57  | 17.84 | 7.27 | 47.1 | 0.03 | 90    | -     | 290 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 5  | 4.2 | 8.63  | 17.79 | 7.27 | 47.1 | 0.03 | 90.4  | -     | 286 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 4  | 5.2 | 8.75  | 17.73 | 7.24 | 47.1 | 0.03 | 91.7  | -     | 281 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 3  | 6.2 | 8     | 17.44 | 7.12 | 49.2 | 0.03 | 83.4  | -     | 278 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 2  | 7.2 | 7.77  | 17.12 | 7.07 | 50.9 | 0.03 | 80.4  | -     | 264 |
| Hidden Lake | 3797 | DBAS | 9/16/97  | - | 2.5 | 1  | 8.2 | 7.96  | 17.05 | 7.05 | 51   | 0.03 | 82.3  | -     | 245 |
|             |      |      |          |   |     |    |     |       |       |      |      |      |       |       |     |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 10 | 0.2 | 9.65  | 17.34 | 7.45 | 48.2 | 0.03 | 101.6 | -     | 350 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 9  | 1.6 | 9.68  | 16.21 | 7.44 | 48.1 | 0.03 | 99.6  | -     | 349 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 8  | 2.4 | 9.67  | 15.91 | 7.44 | 48.2 | 0.03 | 98.8  | -     | 348 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 7  | 3.2 | 9.61  | 15.86 | 7.42 | 48.2 | 0.03 | 98.1  | -     | 347 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 6  | 4   | 9.5   | 15.82 | 7.4  | 48.2 | 0.03 | 97.1  | -     | 346 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 5  | 4.7 | 9.4   | 15.77 | 7.37 | 48.3 | 0.03 | 95.7  | -     | 344 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 4  | 5.6 | 9.36  | 15.67 | 7.34 | 48.1 | 0.03 | 95.2  | -     | 340 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 3  | 6.4 | 8.83  | 15.47 | 7.33 | 48.7 | 0.03 | 89.4  | -     | 333 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 2  | 7.2 | 9.16  | 15.34 | 7.35 | 48.9 | 0.03 | 92.4  | -     | 328 |
| Hidden Lake | 3997 | ASRA | 9/29/97  | - | 3.5 | 1  | 8   | 9.44  | 15.32 | 7.36 | 48.6 | 0.03 | 95.2  | -     | 319 |
|             |      |      |          |   |     |    |     |       |       |      |      |      |       |       |     |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 9  | 0.3 | 10.76 | 11.63 | 7.44 | 54.2 | 0.03 | 97.8  | -     | 350 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 8  | 1   | 10.76 | 11.43 | 7.46 | 54   | 0.03 | 97.5  | -     | 349 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 7  | 2   | 10.76 | 11.25 | 7.42 | 54.2 | 0.03 | 97.1  | -     | 351 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 6  | 3   | 10.64 | 11.12 | 7.39 | 54.3 | 0.03 | 95.7  | -     | 351 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 5  | 4   | 10.71 | 11.05 | 7.38 | 54.1 | 0.03 | 96.2  | -     | 350 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 4  | 5   | 10.72 | 11.02 | 7.37 | 54.2 | 0.03 | 96.2  | -     | 349 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 3  | 6   | 10.75 | 11    | 7.31 | 54.3 | 0.03 | 96.5  | -     | 351 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 2  | 7   | 10.77 | 11    | 7.28 | 54.2 | 0.03 | 96.6  | -     | 351 |
| Hidden Lake | 4297 | DBAS | 10/20/97 | - | 2.5 | 1  | 8   | 10.89 | 10.86 | 7.25 | 54.3 | 0.03 | 97.4  | -     | 350 |
|             |      |      |          |   |     |    |     |       |       |      |      |      |       |       |     |
| Hidden Lake | 4497 | DBAS | 11/3/97  | - | 2.4 | 9  | 0.2 | 11.64 | 7.48  | 8.83 | 53.2 | 0.03 | 100   | ***** | 425 |
| Hidden Lake | 4497 | DBAS | 11/3/97  | - | 2.4 | 8  | 1   | 11.57 | 7.44  | 8.73 | 53.2 | 0.03 | 99.1  | ***** | 427 |
| Hidden Lake | 4497 | DBAS | 11/3/97  | - | 2.4 | 7  | 2   | 11.5  | 7.44  | 8.65 | 53.1 | 0.03 | 98.3  | ***** | 426 |

|             |      |      |         |   |     |   |     |       |      |      |      |      |      |       |     |
|-------------|------|------|---------|---|-----|---|-----|-------|------|------|------|------|------|-------|-----|
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 6 | 3   | 11.5  | 7.41 | 8.61 | 53.3 | 0.03 | 98.2 | ***** | 425 |
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 5 | 4   | 11.48 | 7.43 | 8.61 | 53.2 | 0.03 | 98.1 | ***** | 422 |
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 4 | 5   | 11.38 | 7.37 | 8.6  | 53.2 | 0.03 | 97.2 | ***** | 423 |
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 3 | 6   | 11.17 | 7.36 | 8.45 | 53.5 | 0.03 | 95   | ***** | 422 |
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 2 | 7   | 11.03 | 7.35 | 8.32 | 53.6 | 0.03 | 93.6 | ***** | 422 |
| Hidden Lake | 4497 | DBAS | 11/3/97 | - | 2.4 | 1 | 7.8 | 11.02 | 7.36 | 8.32 | 53.7 | 0.03 | 93.3 | ***** | 433 |

| Location   | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 12       | 0.2       | 11.03                   | 10.79           | 7    | 26.7                 | 0.02 | 99                         | -                  | 372   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 11       | 0.2       | 11.11                   | 10.63           | 7.03 | 27                   | 0.02 | 99                         | -                  | 372   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 10       | 0.2       | 11.09                   | 10.63           | 7.05 | 27                   | 0.02 | 98.9                       | -                  | 371   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 9        | 1.7       | 11.09                   | 10.33           | 7.04 | 26.9                 | 0.02 | 98.3                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 8        | 1.7       | 11.06                   | 10.35           | 7.03 | 26.8                 | 0.02 | 98                         | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 7        | 1.7       | 11.07                   | 10.36           | 7.02 | 26.9                 | 0.02 | 98.2                       | -                  | 374   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 6        | 2.8       | 11.15                   | 8.61            | 7.06 | 26.5                 | 0.02 | 94.7                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 5        | 2.8       | 11.13                   | 8.6             | 7.06 | 26.5                 | 0.02 | 94.5                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 4        | 2.8       | 11.1                    | 8.59            | 7.06 | 26.5                 | 0.02 | 94.2                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 3        | 3.9       | 11.17                   | 8.33            | 7.06 | 26.4                 | 0.02 | 94.3                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 2        | 3.9       | 11.2                    | 8.32            | 7.06 | 26.3                 | 0.02 | 94.5                       | -                  | 373   |
| Round Lake | 1597  | DBAC     | 5/16/97 | -    | 0.4        | 1        | 3.9       | 11.24                   | 8.3             | 7.06 | 26.6                 | 0.02 | 94.8                       | -                  | 373   |
|            |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 12       | 0.3       | 11.02                   | 10.86           | 7.11 | 26.5                 | 0.02 | 99                         | -                  | 365   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 11       | 0.3       | 11.03                   | 10.92           | 7.1  | 26.4                 | 0.02 | 99.3                       | -                  | 366   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 10       | 0.2       | 11.01                   | 11              | 7.1  | 26.2                 | 0.02 | 99.3                       | -                  | 366   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 9        | 1.2       | 11.39                   | 9.71            | 7.11 | 26.6                 | 0.02 | 99.7                       | -                  | 366   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 8        | 1.2       | 11.37                   | 9.69            | 7.13 | 26.7                 | 0.02 | 99.4                       | -                  | 366   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 7        | 1.2       | 11.35                   | 9.73            | 7.12 | 26.7                 | 0.02 | 99.4                       | -                  | 366   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 6        | 2.4       | 11.44                   | 9.5             | 7.09 | 26.7                 | 0.02 | 99.5                       | -                  | 368   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 5        | 2.4       | 11.44                   | 9.51            | 7.1  | 26.3                 | 0.02 | 99.4                       | -                  | 368   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 4        | 2.4       | 11.5                    | 9.45            | 7.1  | 26.7                 | 0.02 | 99.8                       | -                  | 367   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 3        | 3.3       | 11.53                   | 9.2             | 7.09 | 26.7                 | 0.02 | 99.6                       | -                  | 368   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 2        | 3.3       | 11.54                   | 9.22            | 7.08 | 26.4                 | 0.02 | 99.7                       | -                  | 369   |
| Round Lake | 2197  | DBSS     | 5/29/97 | -    | 1.4        | 1        | 3.3       | 11.54                   | 9.23            | 7.09 | 26.4                 | 0.02 | 99.8                       | -                  | 368   |
|            |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 10       | 0.3       | 10.67                   | 13.47           | 7.14 | 29.4                 | 0.02 | 102.1                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 9        | 0.3       | 10.65                   | 13.45           | 7.15 | 29.6                 | 0.02 | 101.9                      | -                  | 397   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 8        | 0.3       | 10.67                   | 13.45           | 7.15 | 29.5                 | 0.02 | 102.1                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 7        | 1         | 11.08                   | 11.69           | 7.19 | 29                   | 0.02 | 101.9                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 6        | 1         | 11.02                   | 11.81           | 7.19 | 29.1                 | 0.02 | 101.6                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 5        | 1         | 10.99                   | 12              | 7.19 | 29                   | 0.02 | 101.8                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 4        | 0.9       | 10.85                   | 12.21           | 7.19 | 29                   | 0.02 | 100.9                      | -                  | 398   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 3        | 1.5       | 11.29                   | 11.28           | 7.21 | 28.6                 | 0.02 | 102.8                      | -                  | 401   |
| Round Lake | 2397  | DBAS     | 6/11/97 | -    | 1.8        | 2        | 1.5       | 11.25                   | 11.38           | 7.22 | 28.7                 | 0.02 | 102.7                      | -                  | 400   |

|            |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Round Lake | 2397 | DBAS | 6/11/97 | - | 1.8 | 1  | 1.6 | 11.27 | 11.43 | 7.21 | 28.6 | 0.02 | 103   | - | 401 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 10 | 0.3 | 11.34 | 11.69 | 7.21 | 32.4 | 0.02 | 104.1 | - | 398 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 9  | 0.3 | 11.37 | 11.7  | 7.21 | 31.9 | 0.02 | 104.3 | - | 398 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 8  | 0.3 | 11.26 | 11.95 | 7.2  | 32.1 | 0.02 | 104.2 | - | 399 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 7  | 0.9 | 11.34 | 11.21 | 7.18 | 32.1 | 0.02 | 103   | - | 401 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 6  | 0.9 | 11.25 | 11.23 | 7.19 | 32   | 0.02 | 102.2 | - | 400 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 5  | 1   | 11.3  | 11.02 | 7.2  | 31.9 | 0.02 | 102.2 | - | 401 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 4  | 1.4 | 11.28 | 10.98 | 7.19 | 31.7 | 0.02 | 101.9 | - | 401 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 3  | 1.4 | 11.32 | 10.92 | 7.2  | 31.9 | 0.02 | 102.1 | - | 402 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 2  | 1.5 | 11.31 | 10.95 | 7.2  | 31.9 | 0.02 | 102.1 | - | 402 |
| Round Lake | 2597 | DBSS | 6/26/97 | - | 1.5 | 1  | 1.6 | 11.32 | 10.95 | 7.19 | 31.8 | 0.02 | 102.1 | - | 403 |
| Round Lake | 2797 | DBSS | 7/8/97  | - | 0.6 | 2  | 0.2 | 9.73  | 16.76 | 7.33 | 35.3 | 0.02 | 99.6  | - | 399 |
| Round Lake | 2797 | DBSS | 7/8/97  | - | 0.6 | 1  | 0.6 | 10.33 | 16.11 | 7.36 | 35.1 | 0.02 | ***** | - | 397 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 10 | 0.3 | 9.48  | 19.4  | 7.27 | 37.9 | 0.02 | ***** | - | 405 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 9  | 0.3 | 9.47  | 19.26 | 7.26 | 37.9 | 0.02 | ***** | - | 405 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 8  | 0.3 | 9.49  | 19.28 | 7.25 | 38   | 0.02 | ***** | - | 406 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 7  | 0.9 | 9.37  | 18.66 | 7.25 | 37.8 | 0.02 | 99.8  | - | 408 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 6  | 0.9 | 9.31  | 18.63 | 7.26 | 37.8 | 0.02 | 99    | - | 408 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 5  | 0.9 | 9.36  | 18.66 | 7.25 | 37.8 | 0.02 | 99.6  | - | 408 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 4  | 0.9 | 9.37  | 18.7  | 7.25 | 37.9 | 0.02 | 99.6  | - | 408 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 3  | 1.7 | 9.47  | 18.51 | 7.32 | 37.8 | 0.02 | ***** | - | 406 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 2  | 1.7 | 9.48  | 18.51 | 7.34 | 37.9 | 0.02 | ***** | - | 408 |
| Round Lake | 2997 | SSJD | 7/24/97 | - | 1.7 | 1  | 1.7 | 9.56  | 18.51 | 7.38 | 37.9 | 0.02 | ***** | - | 411 |
| Round Lake | 3197 | DBSS | 8/5/97  | - | 1.7 | 2  | 0.3 | 9.22  | 22.77 | 7.55 | 44.2 | 0.03 | 105.6 | - | 334 |
| Round Lake | 3197 | DBSS | 8/5/97  | - | 1.7 | 1  | 1.5 | 9.56  | 21.9  | 7.68 | 44.2 | 0.03 | 107.6 | - | 329 |
| Round Lake | 3297 | DBSS | 8/14/97 | - | 1.4 | 2  | 0.3 | 9.3   | 23.91 | 7.92 | 48.7 | 0.03 | 109.5 | - | 292 |
| Round Lake | 3297 | DBSS | 8/14/97 | - | 1.4 | 1  | 1.3 | 9.75  | 23.42 | 8.12 | 49.1 | 0.03 | 113.8 | - | 280 |
| Round Lake | 3497 | DBSS | 8/27/97 | - | 2.1 | 3  | 0.3 | 9.01  | 21.21 | 7.35 | 55.7 | 0.04 | 100.9 | - | 294 |
| Round Lake | 3497 | DBSS | 8/27/97 | - | 2.1 | 2  | 1   | 8.93  | 20.84 | 7.35 | 55.5 | 0.04 | 99.5  | - | 292 |
| Round Lake | 3497 | DBSS | 8/27/97 | - | 2.1 | 1  | 1.9 | 9.11  | 20.72 | 7.38 | 55.6 | 0.04 | 101.1 | - | 287 |
| Round Lake | 3797 | DBAS | 9/16/97 | - | 1.3 | 2  | 0.2 | 8.62  | 17.39 | 7.29 | 55.6 | 0.04 | 89.7  | - | 286 |
| Round Lake | 3797 | DBAS | 9/16/97 | - | 1.3 | 1  | 1   | 8.87  | 17.39 | 7.31 | 55.5 | 0.04 | 92.2  | - | 285 |
| Round Lake | 3997 | ASRA | 9/29/97 | - | 1.5 | 8  | 0.3 | 11.35 | 15.24 | 7.44 | 51.9 | 0.03 | 114.5 | - | 377 |
| Round Lake | 3997 | ASRA | 9/29/97 | - | 1.5 | 7  | 0.3 | 11.35 | 15.25 | 7.45 | 51.9 | 0.03 | 114.3 | - | 376 |
| Round Lake | 3997 | ASRA | 9/29/97 | - | 1.5 | 6  | 0.3 | 11.37 | 15.27 | 7.45 | 51.9 | 0.03 | 114.6 | - | 376 |
| Round Lake | 3997 | ASRA | 9/29/97 | - | 1.5 | 5  | 0.3 | 11.36 | 15.29 | 7.46 | 51.9 | 0.03 | 114.6 | - | 376 |

|            |      |      |          |   |     |   |     |       |       |      |      |      |       |       |     |
|------------|------|------|----------|---|-----|---|-----|-------|-------|------|------|------|-------|-------|-----|
| Round Lake | 3997 | ASRA | 9/29/97  | - | 1.5 | 4 | 1.5 | 11.38 | 14.79 | 7.49 | 51.9 | 0.03 | 113.5 | -     | 375 |
| Round Lake | 3997 | ASRA | 9/29/97  | - | 1.5 | 3 | 1.5 | 11.28 | 14.69 | 7.5  | 52.1 | 0.03 | 112.3 | -     | 374 |
| Round Lake | 3997 | ASRA | 9/29/97  | - | 1.5 | 2 | 1.5 | 10.91 | 14.71 | 7.5  | 51.9 | 0.03 | 108.6 | -     | 374 |
| Round Lake | 3997 | ASRA | 9/29/97  | - | 1.5 | 1 | 1.5 | 11.51 | 14.67 | 7.51 | 51.9 | 0.03 | 115.9 | -     | 373 |
|            |      |      |          |   |     |   |     |       |       |      |      |      |       |       |     |
| Round Lake | 4297 | DBAS | 10/20/97 | - | 1.2 | 3 | 0.3 | 10.65 | 9.38  | 6.97 | 57.2 | 0.04 | 91.9  | -     | 381 |
| Round Lake | 4297 | DBAS | 10/20/97 | - | 1.2 | 2 | 0.8 | 10.68 | 9.35  | 6.96 | 57.1 | 0.04 | 92.2  | -     | 382 |
| Round Lake | 4297 | DBAS | 10/20/97 | - | 1.2 | 1 | 1.3 | 10.87 | 9.19  | 6.93 | 57.1 | 0.04 | 93.4  | -     | 384 |
|            |      |      |          |   |     |   |     |       |       |      |      |      |       |       |     |
| Round Lake | 4497 | DBAS | 11/3/97  | - | 1   | 2 | 0.3 | 11.93 | 7.13  | 6.68 | 38.7 | 0.02 | 97.1  | ***** | 429 |
| Round Lake | 4497 | DBAS | 11/3/97  | - | 1   | 1 | 1.2 | 11.99 | 7.11  | 6.68 | 38.5 | 0.02 | 97.6  | ***** | 431 |

| Location          | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Chatcolet Lake DP | 1597  | ASSS     | 4/18/97 | -    | 1.5        | 3        | 0.5       | 12.66                   | 6.96            | 7.41 | 38.3                 | 0.02 | *****                      | -                  | 380   |
| Chatcolet Lake DP | 1597  | ASSS     | 4/18/97 | -    | 1.5        | 2        | 5.4       | 12.37                   | 5.98            | 7.41 | 36.3                 | 0.02 | 99.3                       | -                  | 381   |
| Chatcolet Lake DP | 1597  | ASSS     | 4/18/97 | -    | 1.5        | 1        | 10.4      | 12.45                   | 5.7             | 7.41 | 33.5                 | 0.02 | 99                         | -                  | 381   |
|                   |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 14       | 0.2       | 11.31                   | 14.46           | 7.12 | 33.7                 | 0.02 | 110                        | -                  | 368   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 13       | 0.8       | 11.32                   | 13.83           | 7.08 | 33.3                 | 0.02 | 108.6                      | -                  | 371   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 12       | 1.8       | 11.46                   | 12.2            | 7.03 | 32.6                 | 0.02 | 106.3                      | -                  | 373   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 11       | 2.7       | 11.37                   | 11.66           | 7.01 | 32                   | 0.02 | 103.9                      | -                  | 374   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 10       | 3.7       | 11                      | 10.94           | 6.95 | 30.5                 | 0.02 | 98.7                       | -                  | 376   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 9        | 4.7       | 10.91                   | 9.76            | 6.93 | 28.8                 | 0.02 | 95.4                       | -                  | 378   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 8        | 5.6       | 10.84                   | 9.3             | 6.9  | 10.84                | **** | 5.6                        | -                  | 376   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 7        | 6.7       | 10.87                   | 9.09            | 6.95 | 27.9                 | 0.02 | 93.4                       | -                  | 376   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 6        | 7.7       | 10.88                   | 9.02            | 6.95 | 27.9                 | 0.02 | 93.4                       | -                  | 376   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 5        | 8.7       | 10.89                   | 8.6             | 6.94 | 30.2                 | 0.02 | 92.5                       | -                  | 377   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 4        | 9.7       | 10.86                   | 8.5             | 6.94 | 30.3                 | 0.02 | 92.1                       | -                  | 377   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 3        | 10.6      | 10.75                   | 8.32            | 6.94 | 31                   | 0.02 | 90.7                       | -                  | 376   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 2        | 11.6      | 10.03                   | 7.54            | 6.92 | 35.2                 | 0.02 | 83.2                       | -                  | 378   |
| Chatcolet Lake DP | 1997  | DBSS     | 5/16/97 | -    | 1.4        | 1        | 12.7      | 10.04                   | 7.48            | 6.92 | 35.4                 | 0.02 | 83                         | -                  | 377   |
|                   |       |          |         |      |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 11       | 0.3       | 10.62                   | 13.44           | 6.98 | 27.4                 | 0.02 | 101.2                      | -                  | 360   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 10       | 2.3       | 10.9                    | 9.6             | 6.97 | 26.8                 | 0.02 | 95                         | -                  | 363   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 9        | 3.4       | 10.8                    | 8.6             | 6.98 | 26.7                 | 0.02 | 91.9                       | -                  | 362   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 8        | 4.5       | 10.85                   | 8.55            | 6.96 | 26.6                 | 0.02 | 92.2                       | -                  | 363   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 7        | 6.5       | 10.88                   | 8.35            | 6.96 | 26.5                 | 0.02 | 92.1                       | -                  | 362   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 6        | 7.4       | 10.88                   | 8.29            | 6.95 | 26.4                 | 0.02 | 91.9                       | -                  | 362   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 5        | 8.4       | 11.07                   | 8.11            | 6.95 | 26.5                 | 0.02 | 93.1                       | -                  | 361   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 4        | 9.3       | 11.05                   | 7.88            | 6.95 | 26.4                 | 0.02 | 92.4                       | -                  | 360   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 3        | 10.4      | 11.03                   | 7.86            | 6.93 | 26.4                 | 0.02 | 92.2                       | -                  | 360   |
| Chatcolet Lake DP | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 2        | 11.4      | 11                      | 7.78            | 6.92 | 26.4                 | 0.02 | 91.8                       | -                  | 359   |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Chatcolet Lake DP | 2197 | DBSS | 5/29/97 | - | 1.6 | 1  | 12.3 | 10.3  | 7.73  | 6.91 | 27.3 | 0.02 | 85.8  | - | 359 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 12 | 0.1  | 11.24 | 12.99 | 7.2  | 29.1 | 0.02 | 106.4 | - | 417 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 11 | 1    | 11.23 | 11.66 | 7.23 | 28.3 | 0.02 | 103.2 | - | 417 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 10 | 1.9  | 11.08 | 10.38 | 7.21 | 27.5 | 0.02 | 98.8  | - | 419 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 9  | 3    | 11.09 | 10    | 7.19 | 27.5 | 0.02 | 98    | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 8  | 3.9  | 11.16 | 9.82  | 7.19 | 27.2 | 0.02 | 98.1  | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 7  | 4.9  | 11.2  | 9.72  | 7.19 | 27.1 | 0.02 | 98.2  | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 6  | 6    | 11.07 | 9.67  | 7.2  | 27.1 | 0.02 | 97    | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 5  | 7    | 11.15 | 9.15  | 7.21 | 26.9 | 0.02 | 96.5  | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 4  | 8    | 11.07 | 9.02  | 7.21 | 27   | 0.02 | 95.5  | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 3  | 8.9  | 10.94 | 8.99  | 7.23 | 27.1 | 0.02 | 94.3  | - | 421 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 2  | 9.8  | 10.84 | 9.04  | 7.26 | 27.3 | 0.02 | 93.5  | - | 419 |
| Chatcolet Lake DP | 2397 | DBAS | 6/13/97 | - | 2.1 | 1  | 10.9 | 10.85 | 8.99  | 7.29 | 27.3 | 0.02 | 93.4  | - | 417 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 11 | 0.3  | 10.94 | 14.59 | 7.21 | 31.5 | 0.02 | 107.3 | - | 373 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 10 | 1.4  | 10.94 | 14.45 | 7.17 | 31.5 | 0.02 | 106.9 | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 9  | 2.5  | 11.03 | 13.45 | 7.17 | 31.5 | 0.02 | 105.5 | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 8  | 3.6  | 10.97 | 13.25 | 7.15 | 31.4 | 0.02 | 104.4 | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 7  | 4.6  | 10.99 | 12.95 | 7.12 | 31.2 | 0.02 | 103.8 | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 6  | 5.6  | 10.93 | 12.74 | 7.09 | 31.5 | 0.02 | 102.8 | - | 374 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 5  | 6.6  | 10.97 | 12.1  | 7.04 | 31.6 | 0.02 | 101.8 | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 4  | 7.7  | 10.35 | 10.92 | 7    | 30.9 | 0.02 | 93.4  | - | 375 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 3  | 8.7  | 9.68  | 10.68 | 6.98 | 30.2 | 0.02 | 86.8  | - | 374 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 2  | 9.7  | 9.38  | 10.12 | 6.99 | 30.4 | 0.02 | 83    | - | 372 |
| Chatcolet Lake DP | 2597 | DBSS | 6/26/97 | - | 2.5 | 1  | 10.6 | 9.44  | 10.1  | 7.02 | 30.6 | 0.02 | 83.5  | - | 370 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 12 | 0.3  | 10.25 | 17.45 | 7.19 | 35.2 | 0.02 | ****  | - | 411 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 11 | 0.9  | 10.23 | 17.39 | 7.17 | 35.2 | 0.02 | ****  | - | 411 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 10 | 2.1  | 10.2  | 17.17 | 7.11 | 35.1 | 0.02 | ****  | - | 413 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 9  | 3.1  | 9.99  | 16.11 | 7.05 | 35.1 | 0.02 | ****  | - | 416 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 8  | 4    | 9.88  | 15.35 | 7.02 | 34.9 | 0.02 | 98.1  | - | 417 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 7  | 5    | 9.83  | 14.71 | 6.97 | 33.8 | 0.02 | 96.2  | - | 418 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 6  | 6    | 9.84  | 13.13 | 6.94 | 33.3 | 0.02 | 93    | - | 420 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 5  | 7    | 9.5   | 12.69 | 6.91 | 33.1 | 0.02 | 88.9  | - | 421 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 4  | 8.1  | 8.93  | 12.39 | 6.89 | 33.3 | 0.02 | 83    | - | 421 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 3  | 9    | 9.12  | 12.25 | 6.89 | 33.2 | 0.02 | 84.5  | - | 421 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 2  | 10.1 | 8.4   | 11.77 | 6.9  | 33.3 | 0.02 | 77    | - | 418 |
| Chatcolet Lake DP | 2797 | DBSS | 7/8/97  | - | 1.8 | 1  | 11   | 8.23  | 11.64 | 7    | 33.7 | 0.02 | 75.2  | - | 423 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 12 | 0.6  | 9.86  | 22.15 | 7.99 | 36.4 | 0.02 | ****  | - | 378 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 11 | 1.8  | 9.85  | 22.15 | 7.96 | 36.4 | 0.02 | ****  | - | 378 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 10 | 2.9  | 9.97  | 21.67 | 7.85 | 36.8 | 0.02 | ****  | - | 379 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 9  | 3.9  | 9.81  | 19.33 | 7.18 | 35.8 | 0.02 | ****  | - | 402 |

|                   |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|-------------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 8  | 4.9  | 10.24 | 16.32 | 7.08 | 33.2 | 0.02 | ****  | - | 406 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 7  | 5.9  | 9.79  | 14.49 | 6.93 | 32.4 | 0.02 | 95.4  | - | 410 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 6  | 6.9  | 8.29  | 13.58 | 6.84 | 32.1 | 0.02 | 79.2  | - | 413 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 5  | 7.9  | 7.71  | 13.1  | 6.82 | 32.2 | 0.02 | 72.9  | - | 413 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 4  | 8.9  | 6.88  | 12.72 | 6.82 | 32.4 | 0.02 | 64.5  | - | 411 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 3  | 10   | 5.03  | 12.03 | 6.87 | 34.1 | 0.02 | 46.5  | - | 408 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 2  | 10.9 | 5.19  | 12.08 | 6.97 | 34.1 | 0.02 | 47.9  | - | 402 |
| Chatcolet Lake DP | 2997 | SSJD | 7/24/97 | - | 3.7 | 1  | 0    | 0     | 0     | 0    | 0    | 0    | 0     | - | 0   |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 13 | 0.3  | 11.59 | 24.51 | 9.28 | 42.3 | 0.03 | 137.1 | - | 289 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 12 | 1.7  | 11.55 | 24.37 | 9.22 | 42.2 | 0.03 | 136.3 | - | 290 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 11 | 2.7  | 10.64 | 23.4  | 8.86 | 40.4 | 0.03 | 123.3 | - | 299 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 10 | 3.6  | 10.13 | 22.02 | 7.49 | 39.2 | 0.03 | 114.4 | - | 331 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 9  | 4.6  | 9.8   | 18.9  | 7.05 | 35.8 | 0.02 | 104.2 | - | 349 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 8  | 5.6  | 9.77  | 16.32 | 6.97 | 33   | 0.02 | 98.4  | - | 351 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 7  | 6.5  | 8.09  | 14.44 | 6.84 | 33   | 0.02 | 78.2  | - | 353 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 6  | 7.5  | 6.42  | 13.55 | 6.79 | 33.4 | 0.02 | 60.9  | - | 353 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 5  | 8.4  | 5.49  | 13.15 | 6.79 | 33.6 | 0.02 | 51.6  | - | 351 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 4  | 9.5  | 3.45  | 12.3  | 6.81 | 36.1 | 0.02 | 31.8  | - | 347 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 3  | 10.5 | 3.05  | 12.11 | 6.89 | 36.7 | 0.02 | 28    | - | 342 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 2  | 11.5 | 2.57  | 12.07 | 7.03 | 38.4 | 0.02 | 23.5  | - | 335 |
| Chatcolet Lake DP | 3197 | DBSS | 8/5/97  | - | 3   | 1  | 12.4 | 2.5   | 12.13 | 7.18 | 39.3 | 0.03 | 23    | - | 343 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 10 | 0.8  | 9.41  | 22.99 | 8.77 | 42.9 | 0.03 | 108.9 | - | 339 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 9  | 1.8  | 9.41  | 22.58 | 8.71 | 42.7 | 0.03 | 108.2 | - | 341 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 8  | 2.8  | 9.27  | 21.79 | 8.4  | 42   | 0.03 | 104.7 | - | 348 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 7  | 3.8  | 8.8   | 21    | 7.58 | 39.8 | 0.03 | 98    | - | 371 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 6  | 4.8  | 8.7   | 20.35 | 7.29 | 39.1 | 0.03 | 95.7  | - | 379 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 5  | 5.8  | 8.91  | 18.18 | 6.95 | 35.4 | 0.02 | 93.9  | - | 392 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 4  | 6.8  | 6.33  | 15.19 | 6.8  | 34.4 | 0.02 | 62.6  | - | 399 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 3  | 7.9  | 5.85  | 14.18 | 6.8  | 34.9 | 0.02 | 56.6  | - | 399 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 2  | 8.8  | 4.53  | 13.48 | 6.82 | 35.8 | 0.02 | 43.2  | - | 399 |
| Chatcolet Lake DP | 3297 | DBSS | 8/13/97 | - | 3.5 | 1  | 9.8  | 2.8   | 13.1  | 6.87 | 39.2 | 0.03 | 26.5  | - | 398 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 12 | 0.4  | 9.51  | 21.99 | 8.48 | 47.3 | 0.03 | 108.2 | - | 271 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 11 | 0.7  | 9.54  | 21.97 | 8.49 | 47.4 | 0.03 | 108.4 | - | 267 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 10 | 1.7  | 9.55  | 21.93 | 8.47 | 47.4 | 0.03 | 108.5 | - | 267 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 9  | 2.7  | 9.52  | 21.65 | 8.38 | 47.3 | 0.03 | 107.5 | - | 267 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 8  | 3.7  | 9.54  | 21.51 | 8.34 | 48.8 | 0.03 | 107.5 | - | 265 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 7  | 4.7  | 9.29  | 21.19 | 7.74 | 45.6 | 0.03 | 104.1 | - | 276 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 6  | 5.7  | 8.58  | 20.37 | 7.14 | 44.4 | 0.03 | 94.7  | - | 297 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 5  | 6.7  | 5.59  | 17.27 | 6.77 | 40.6 | 0.03 | 57.9  | - | 313 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 4  | 7.7  | 4.07  | 16.17 | 6.73 | 40.6 | 0.03 | 41.1  | - | 311 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97 | - | 4.4 | 3  | 8.8  | 1.34  | 13.79 | 6.75 | 42.7 | 0.03 | 12.9  | - | 305 |

|                   |      |      |          |   |     |    |      |       |       |      |      |      |       |       |     |
|-------------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|-------|-------|-----|
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97  | - | 4.4 | 2  | 9.8  | 0.64  | 13.33 | 6.84 | 46.9 | 0.03 | 6.1   | -     | 296 |
| Chatcolet Lake DP | 3497 | DBSS | 8/27/97  | - | 4.4 | 1  | 10.7 | 0.52  | 13.09 | 6.98 | 49   | 0.03 | 4.9   | -     | 284 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 11 | 0.3  | 9.12  | 17.71 | 7.49 | 46.4 | 0.03 | 95.6  | -     | 251 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 10 | 1.7  | 9.11  | 17.71 | 7.48 | 46.4 | 0.03 | 95.4  | -     | 248 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 9  | 2.7  | 9.05  | 17.69 | 7.45 | 46.4 | 0.03 | 94.8  | -     | 246 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 8  | 3.7  | 9.02  | 17.67 | 7.42 | 46.4 | 0.03 | 94.4  | -     | 244 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 7  | 4.7  | 8.93  | 17.66 | 7.39 | 46.3 | 0.03 | 93.5  | -     | 241 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 6  | 5.7  | 9.01  | 17.56 | 7.31 | 46.1 | 0.03 | 93.8  | -     | 231 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 5  | 6.7  | 8.67  | 17.49 | 7.23 | 45.9 | 0.03 | 90.4  | -     | 219 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 4  | 7.7  | 8.58  | 17.49 | 7.15 | 45.8 | 0.03 | 89.5  | -     | 205 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 3  | 8.7  | 2.12  | 16.21 | 6.79 | 46.4 | 0.03 | 21    | -     | 156 |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 2  | 9.7  | 0.27  | 12.49 | 6.72 | 74.3 | 0.05 | 2.4   | -     | 34  |
| Chatcolet Lake DP | 3797 | DBAS | 9/16/97  | - | 2.2 | 1  | 10.7 | 0.35  | 12.36 | 6.69 | 77.1 | 0.05 | 3.3   | -     | 42  |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 11 | 0.6  | 9.55  | 16.12 | 7.43 | 48.5 | 0.03 | 98    | -     | 362 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 10 | 1.6  | 9.51  | 15.72 | 7.41 | 48.3 | 0.03 | 96.8  | -     | 362 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 9  | 2.6  | 9.25  | 15.67 | 7.37 | 48.3 | 0.03 | 94.1  | -     | 363 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 8  | 3.6  | 9.15  | 15.54 | 7.35 | 48.4 | 0.03 | 92.8  | -     | 362 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 7  | 4.6  | 8.95  | 15.47 | 7.31 | 48.4 | 0.03 | 90.7  | -     | 362 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 6  | 5.6  | 8.67  | 15.37 | 7.3  | 48.7 | 0.03 | 87.6  | -     | 362 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 5  | 6.6  | 8.63  | 15.32 | 7.3  | 48.7 | 0.03 | 87.1  | -     | 361 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 4  | 7.6  | 8.53  | 15.27 | 7.3  | 48.8 | 0.03 | 86.2  | -     | 359 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 3  | 8.6  | 8.41  | 15.22 | 7.3  | 48.9 | 0.03 | 84.7  | -     | 358 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 2  | 9.6  | 8.06  | 15.06 | 7.29 | 49.4 | 0.03 | 80.8  | -     | 356 |
| Chatcolet Lake DP | 3997 | ASRA | 9/29/97  | - | 3.3 | 1  | 10.6 | 7.99  | 14.99 | 7.32 | 50.4 | 0.03 | 80    | -     | 353 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 11 | 0.3  | 11.17 | 11.63 | 7.54 | 54.9 | 0.04 | 101.7 | -     | 372 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 10 | 1    | 11.1  | 11.56 | 7.5  | 55   | 0.04 | 100.9 | -     | 373 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 9  | 2    | 11.08 | 11.25 | 7.5  | 54.7 | 0.04 | 100   | -     | 373 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 8  | 3    | 11.05 | 11.23 | 7.45 | 54.9 | 0.04 | 99.6  | -     | 375 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 7  | 4    | 10.97 | 11.17 | 7.4  | 54.6 | 0.03 | 98.8  | -     | 376 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 6  | 5    | 10.89 | 11.13 | 7.32 | 54.5 | 0.03 | 98    | -     | 378 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 5  | 6    | 10.81 | 11.07 | 7.24 | 54.5 | 0.03 | 97.1  | -     | 381 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 4  | 7    | 10.67 | 11.04 | 7.15 | 54.6 | 0.03 | 95.8  | -     | 384 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 3  | 8    | 10.53 | 11    | 7.08 | 54.6 | 0.03 | 94.5  | -     | 386 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 2  | 9    | 9.17  | 10.74 | 6.92 | 56.4 | 0.04 | 81.8  | -     | 392 |
| Chatcolet Lake DP | 4297 | DBAS | 10/20/97 | - | 2.8 | 1  | 10   | 8.97  | 10.64 | 6.88 | 56.8 | 0.04 | 79.8  | -     | 393 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97  | - | 2   | 11 | 0.2  | 11.7  | 7.53  | 8.55 | 53.1 | 0.03 | 99.8  | ***** | 418 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97  | - | 2   | 10 | 1    | 11.7  | 7.5   | 8.55 | 53.3 | 0.03 | 99.8  | ***** | 419 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97  | - | 2   | 9  | 2    | 11.69 | 7.45  | 8.5  | 53.2 | 0.03 | 99.5  | ***** | 421 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97  | - | 2   | 8  | 3    | 11.69 | 7.43  | 8.4  | 53.2 | 0.03 | 99.3  | ***** | 419 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97  | - | 2   | 7  | 4    | 11.69 | 7.39  | 8.38 | 53.1 | 0.03 | 99.3  | ***** | 420 |



|                   |      |      |         |   |   |   |     |       |      |      |      |      |      |      |     |
|-------------------|------|------|---------|---|---|---|-----|-------|------|------|------|------|------|------|-----|
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 6 | 5   | 11.7  | 7.38 | 8.37 | 53.1 | 0.03 | 99.4 | **** | 417 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 5 | 6   | 11.72 | 7.37 | 8.33 | 53.1 | 0.03 | 99.4 | **** | 416 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 4 | 7   | 11.77 | 7.31 | 8.19 | 52.5 | 0.03 | 99.5 | **** | 417 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 3 | 8   | 11.73 | 7.29 | 7.97 | 52   | 0.03 | 98.6 | **** | 416 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 2 | 9   | 11.46 | 7.25 | 7.55 | 52.2 | 0.03 | 95.4 | **** | 415 |
| Chatcolet Lake DP | 4497 | DBAS | 11/3/97 | - | 2 | 1 | 9.8 | 11.55 | 7.26 | 6.91 | 47.7 | 0.03 | 94.6 | **** | 414 |

| Location          | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Chatcolet Lake SH | 1597  | ASSS     | 4/18/97 | -    | 0.5        | 2        | 0.2       | 12.75                   | 5.46            | 7.43 | 32.4                 | 0.02 | ****                       | -                  | 325   |
| Chatcolet Lake SH | 1597  | ASSS     | 4/18/97 | -    | 0.5        | 1        | 0.7       | 13.38                   | 5.47            | 7.41 | 32.6                 | 0.02 | ****                       | -                  | 320   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 13       | 0.3       | 11.04                   | 10.48           | 6.99 | 27.2                 | 0.02 | 98.1                       | -                  | 369   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 12       | 0.3       | 11.05                   | 10.49           | 7    | 27.2                 | 0.02 | 98.2                       | -                  | 369   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 11       | 0.3       | 11.04                   | 10.44           | 7.01 | 27.3                 | 0.02 | 98                         | -                  | 369   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 10       | 0.3       | 11.03                   | 10.46           | 7.01 | 27.2                 | 0.02 | 98                         | -                  | 369   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 9        | 1.3       | 11.02                   | 9.97            | 6.97 | 27                   | 0.02 | 95.7                       | -                  | 374   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 8        | 1.3       | 11                      | 9.89            | 6.96 | 27.1                 | 0.02 | 96.4                       | -                  | 374   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 7        | 1.3       | 11.01                   | 9.51            | 6.99 | 27                   | 0.02 | 95.6                       | -                  | 373   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 6        | 1.3       | 10.97                   | 10.15           | 6.95 | 27.4                 | 0.02 | 96.7                       | -                  | 375   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 5        | 3.1       | 11.21                   | 8.56            | 7    | 26.4                 | 0.02 | 95.1                       | -                  | 375   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 4        | 3.1       | 11.21                   | 8.58            | 6.99 | 26.3                 | 0.02 | 95.2                       | -                  | 375   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 3        | 3.1       | 11.22                   | 8.55            | 6.99 | 26.3                 | 0.02 | 95.2                       | -                  | 376   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 2        | 3.1       | 11.22                   | 8.59            | 7    | 26.3                 | 0.02 | 95.3                       | -                  | 375   |
| Chatcolet Lake SH | 1997  | DBSS     | 5/16/97 | -    | 0.3        | 1        | 3.1       | 11.21                   | 8.56            | 7.01 | 26.4                 | 0.02 | 95.1                       | -                  | 375   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 10       | 0.2       | 10.8                    | 11.38           | 7.04 | 26.8                 | 0.02 | 98.2                       | -                  | 368   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 9        | 0.2       | 10.85                   | 11.72           | 7.1  | 18.3                 | 0.01 | 99.5                       | -                  | 363   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 8        | 0.2       | 10.87                   | 11.59           | 7.1  | 15.3                 | 0.01 | 99.3                       | -                  | 364   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 7        | 1.4       | 11.32                   | 9.74            | 7.07 | 26.7                 | 0.02 | 99                         | -                  | 368   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 6        | 1.4       | 11.32                   | 9.69            | 7.07 | 26.7                 | 0.02 | 99.1                       | -                  | 368   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 5        | 1.4       | 11.35                   | 9.68            | 7.08 | 26.6                 | 0.02 | 99.4                       | -                  | 368   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 4        | 2.3       | 11.51                   | 9.32            | 7.08 | 26.5                 | 0.02 | 99.7                       | -                  | 369   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 3        | 2.3       | 11.53                   | 9.32            | 7.08 | 26.5                 | 0.02 | 99.9                       | -                  | 369   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 2        | 2.3       | 11.52                   | 9.33            | 7.07 | 26.5                 | 0.02 | 100                        | -                  | 369   |
| Chatcolet Lake SH | 2197  | DBSS     | 5/29/97 | -    | 1.6        | 1        | 2.4       | 11.57                   | 9.35            | 7.08 | 26.6                 | 0.02 | 100.3                      | -                  | 368   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 10       | 0.4       | 10.75                   | 11.8            | 7.09 | 28.7                 | 0.02 | 99.1                       | -                  | 394   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 9        | 0.4       | 10.7                    | 11.74           | 7.1  | 28.8                 | 0.02 | 98.5                       | -                  | 395   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 8        | 0.3       | 10.71                   | 11.86           | 7.11 | 28.7                 | 0.02 | 98.8                       | -                  | 394   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 7        | 0.9       | 10.95                   | 11.15           | 7.1  | 28.7                 | 0.02 | 99.4                       | -                  | 396   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 6        | 0.9       | 10.93                   | 11.31           | 7.11 | 28.5                 | 0.02 | 99.6                       | -                  | 396   |
| Chatcolet Lake SH | 2397  | DBAS     | 6/11/97 | -    | 1.6        | 5        | 0.9       | 10.94                   | 11.18           | 7.12 | 28.3                 | 0.02 | 99.4                       | -                  | 395   |

|                   |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
|-------------------|------|------|----------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Chatcolet Lake SH | 2397 | DBAS | 6/11/97  | - | 1.6 | 4  | 0.9 | 10.97 | 11.1  | 7.11 | 28.4 | 0.02 | 99.5  | - | 396 |
| Chatcolet Lake SH | 2397 | DBAS | 6/11/97  | - | 1.6 | 3  | 1.5 | 11.07 | 10.98 | 7.11 | 28.3 | 0.02 | 100.1 | - | 398 |
| Chatcolet Lake SH | 2397 | DBAS | 6/11/97  | - | 1.6 | 2  | 1.5 | 11.12 | 10.89 | 7.11 | 28.4 | 0.02 | 100.4 | - | 398 |
| Chatcolet Lake SH | 2397 | DBAS | 6/11/97  | - | 1.6 | 1  | 1.5 | 11.16 | 10.88 | 7.12 | 28.3 | 0.02 | 100.7 | - | 398 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 10 | 0.2 | 11.42 | 11.88 | 7.25 | 31.8 | 0.02 | 105.4 | - | 388 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 9  | 0.2 | 11.38 | 11.97 | 6.9  | 0.7  | 0    | 105.1 | - | 405 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 8  | 0.2 | 11.43 | 11.9  | 7.25 | 31.8 | 0.02 | 105.4 | - | 389 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 7  | 0.2 | 11.44 | 11.88 | 7.26 | 31.8 | 0.02 | 105.6 | - | 389 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 6  | 0.2 | 11.44 | 11.92 | 7.26 | 31.8 | 0.02 | 105.5 | - | 389 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 5  | 0.7 | 11.39 | 11.87 | 7.24 | 31.9 | 0.02 | 105   | - | 392 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 4  | 0.7 | 11.41 | 11.85 | 7.27 | 31.8 | 0.02 | 105.2 | - | 391 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 3  | 0.7 | 11.43 | 11.84 | 7.27 | 31.8 | 0.02 | 105.3 | - | 391 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 2  | 0.7 | 11.4  | 11.87 | 7.27 | 31.8 | 0.02 | 105.2 | - | 391 |
| Chatcolet Lake SH | 2597 | DBAC | 6/26/97  | - | 0.7 | 1  | 0.7 | 11.44 | 11.87 | 7.27 | 31.8 | 0.02 | 105.5 | - | 391 |
| Chatcolet Lake SH | 2797 | DBSS | 7/8/97   | - | 0.7 | 2  | 0.2 | 10.12 | 16.22 | 7.28 | 35.2 | 0.02 | ****  | - | 406 |
| Chatcolet Lake SH | 2797 | DBSS | 7/8/97   | - | 0.7 | 1  | 0.8 | 10.17 | 16.19 | 7.29 | 35.2 | 0.02 | ****  | - | 409 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 10 | 0.3 | 11.24 | 22.77 | 9.09 | 38.9 | 0.02 | ****  | - | 356 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 9  | 0.3 | 11.33 | 22.93 | 9.1  | 39   | 0.02 | ****  | - | 357 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 8  | 0.3 | 11.39 | 23.04 | 9.12 | 39   | 0.03 | ****  | - | 356 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 7  | 0.3 | 11.5  | 22.99 | 9.1  | 39   | 0.03 | ****  | - | 357 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 6  | 0.2 | 11.19 | 22.95 | 9.03 | 39   | 0.02 | ****  | - | 361 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 5  | 1   | 11.36 | 20.13 | 8.43 | 38.2 | 0.02 | ****  | - | 382 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 4  | 1   | 11.45 | 20.09 | 8.47 | 38   | 0.02 | ****  | - | 379 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 3  | 1   | 11.51 | 20.11 | 8.49 | 38.1 | 0.02 | ****  | - | 378 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 2  | 0   | 11.47 | 20.18 | 8.45 | 38.1 | 0.02 | ****  | - | 379 |
| Chatcolet Lake SH | 2997 | SSJD | 7/24/97  | - | 0.8 | 1  | 1   | 11.55 | 20.09 | 8.47 | 38   | 0.02 | ****  | - | 378 |
| Chatcolet Lake SH | 3197 | DBSS | 8/5/97   | - | 1   | 2  | 0.3 | 12.26 | 25.72 | 9.28 | 46   | 0.03 | 148.1 | - | 295 |
| Chatcolet Lake SH | 3197 | DBSS | 8/5/97   | - | 1   | 1  | 0.9 | 12.94 | 22.95 | 9.04 | 46.6 | 0.03 | 148.9 | - | 304 |
| Chatcolet Lake SH | 3297 | DBSS | 8/14/97  | - | 1.1 | 2  | 0.3 | 11.62 | 24.37 | 9.54 | 53.3 | 0.03 | 138   | - | 279 |
| Chatcolet Lake SH | 3297 | DBSS | 8/14/97  | - | 1.1 | 1  | 1.1 | 11.73 | 24.29 | 9.63 | 52.4 | 0.03 | 139.1 | - | 276 |
| Chatcolet Lake SH | 3497 | DBSS | 8/27/97  | - | 0.8 | 1  | 0.8 | 11.39 | 21.49 | 8.62 | 56.9 | 0.04 | 128.4 | - | 277 |
| Chatcolet Lake SH | 3797 | DBAS | 9/16/97  | - | 1   | 1  | 0.8 | 10.41 | 15.56 | 7.79 | 54.3 | 0.03 | 104.2 | - | 256 |
| Chatcolet Lake SH | 3997 | ASRA | 9/29/97  | - | 0.4 | 1  | 0.4 | 12.26 | 16.46 | 8.31 | 53.6 | 0.03 | 126.8 | - | 362 |
| Chatcolet Lake SH | 4297 | DBAS | 10/20/97 | - | 1   | 2  | 0.3 | 10.98 | 10.61 | 7.11 | 53.4 | 0.03 | 97.6  | - | 372 |
| Chatcolet Lake SH | 4297 | DBAS | 10/20/97 | - | 1   | 1  | 0.7 | 10.16 | 9.02  | 7    | 57.2 | 0.04 | 87    | - | 379 |

|                   |      |      |         |   |     |   |     |       |      |      |      |      |      |       |     |
|-------------------|------|------|---------|---|-----|---|-----|-------|------|------|------|------|------|-------|-----|
| Chatcolet Lake SH | 4497 | DBAS | 11/3/97 | - | 0.9 | 2 | 0.2 | 11.72 | 7.28 | 6.97 | 42.6 | 0.03 | 96.2 | ***** | 418 |
| Chatcolet Lake SH | 4497 | DBAS | 11/3/97 | - | 0.9 | 1 | 0.8 | 11.86 | 7.26 | 6.96 | 43   | 0.03 | 97.3 | ***** | 424 |

| Location     | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (us/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Benewah Lake | 1597  | ASSS     | 4/18/97 | -    | 0.8        | 3        | 0.2       | 11.66                   | 9.27            | 7.36 | 28.1                 | 0.02 | *****                      | -                  | 363   |
| Benewah Lake | 1597  | ASSS     | 4/18/97 | -    | 0.8        | 2        | 1.6       | 11.46                   | 8.92            | 7.32 | 27.9                 | 0.02 | 99                         | -                  | 367   |
| Benewah Lake | 1597  | ASSS     | 4/18/97 | -    | 0.8        | 1        | 4.3       | 11.42                   | 7.02            | 7.36 | 24.8                 | 0.02 | *****                      | -                  | 366   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 10       | 0.3       | 10.12                   | 16.39           | 6.97 | 31.6                 | 0.02 | 102.6                      | -                  | 369   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 9        | 1.1       | 9.81                    | 15.1            | 6.95 | 31.2                 | 0.02 | 96.8                       | -                  | 371   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 8        | 1.8       | 11.11                   | 8.71            | 7    | 26.6                 | 0.02 | 94.8                       | -                  | 372   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 7        | 2.5       | 11.1                    | 8.53            | 7.02 | 26.6                 | 0.02 | 94.1                       | -                  | 372   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 6        | 3.3       | 11.07                   | 8.4             | 7.01 | 26.5                 | 0.02 | 93.6                       | -                  | 372   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 5        | 3.9       | 11.04                   | 8.38            | 7.01 | 26.7                 | 0.02 | 93.3                       | -                  | 372   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 4        | 4.7       | 11                      | 8.36            | 7.02 | 26.7                 | 0.02 | 92.9                       | -                  | 372   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 3        | 5.6       | 10.94                   | 8.36            | 7.03 | 27                   | 0.02 | 92.5                       | -                  | 371   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 2        | 6.2       | 10.8                    | 8.35            | 7.02 | 27.2                 | 0.02 | 91                         | -                  | 371   |
| Benewah Lake | 1997  | DBSS     | 5/16/97 | -    | 1.1        | 1        | 7         | 10.47                   | 8.25            | 7.04 | 27.7                 | 0.02 | 88.2                       | -                  | 371   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 12       | 0.8       | 10.09                   | 13.9            | 6.97 | 28                   | 0.02 | 97.1                       | -                  | 351   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 11       | 0.8       | 10.05                   | 13.88           | 6.97 | 27.6                 | 0.02 | 96.7                       | -                  | 350   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 10       | 1.8       | 10.85                   | 9.35            | 6.97 | 27                   | 0.02 | 94.1                       | -                  | 351   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 9        | 1.8       | 11.02                   | 9.17            | 6.96 | 27                   | 0.02 | 95.1                       | -                  | 351   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 8        | 2.8       | 11.1                    | 8.71            | 6.94 | 26.7                 | 0.02 | 94.8                       | -                  | 351   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 7        | 2.8       | 11.08                   | 8.65            | 6.96 | 26.7                 | 0.02 | 94.4                       | -                  | 350   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 6        | 3.9       | 10.82                   | 8.37            | 6.92 | 26.8                 | 0.02 | 91.6                       | -                  | 350   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 5        | 3.9       | 10.83                   | 8.35            | 6.93 | 26.6                 | 0.02 | 91.6                       | -                  | 349   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 4        | 4.9       | 10.54                   | 7.89            | 6.92 | 26.5                 | 0.02 | 88.2                       | -                  | 346   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 3        | 4.9       | 10.59                   | 7.89            | 6.92 | 26.7                 | 0.02 | 88.6                       | -                  | 347   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 2        | 5.9       | 10.3                    | 7.66            | 6.92 | 26.6                 | 0.02 | 85.7                       | -                  | 352   |
| Benewah Lake | 2197  | DBSS     | 5/29/97 | -    | 2          | 1        | 5.9       | 10.35                   | 7.65            | 6.92 | 26.5                 | 0.02 | 86.1                       | -                  | 350   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 10       | 0.4       | 9.39                    | 17.73           | 6.99 | 32.6                 | 0.02 | 98.5                       | -                  | 399   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 9        | 0.4       | 9.43                    | 17.83           | 7    | 32.5                 | 0.02 | 99.1                       | -                  | 399   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 8        | 0.3       | 9.43                    | 17.9            | 7    | 32.4                 | 0.02 | 99.2                       | -                  | 399   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 7        | 0.9       | 9.82                    | 15.89           | 7.01 | 32.7                 | 0.02 | 99.1                       | -                  | 399   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 6        | 1.9       | 10.25                   | 14.21           | 7    | 30.3                 | 0.02 | 99.8                       | -                  | 400   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 5        | 2.8       | 11.19                   | 11.04           | 6.95 | 28.6                 | 0.02 | 101.4                      | -                  | 402   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 4        | 3.9       | 9.41                    | 8.78            | 6.86 | 28.8                 | 0.02 | 80.9                       | -                  | 407   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 3        | 4.8       | 8.36                    | 8.6             | 6.85 | 30.2                 | 0.02 | 71.5                       | -                  | 407   |
| Benewah Lake | 2397  | DBAS     | 6/11/97 | -    | 2.5        | 2        | 4.8       | 8.38                    | 8.6             | 6.86 | 29.9                 | 0.02 | 71.6                       | -                  | 407   |

|              |      |      |         |   |     |    |     |       |       |      |      |      |       |   |     |
|--------------|------|------|---------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Benewah Lake | 2397 | DBAS | 6/11/97 | - | 2.5 | 1  | 4.8 | 8.41  | 8.56  | 6.87 | 29.7 | 0.02 | 71.8  | - | 406 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 12 | 0.3 | 10.31 | 18.63 | 7.36 | 33.9 | 0.02 | 110   | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 11 | 0.3 | 10.3  | 18.65 | 7.36 | 33.9 | 0.02 | 110   | - | 383 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 10 | 0.9 | 10.32 | 18.57 | 7.34 | 33.9 | 0.02 | 110   | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 9  | 0.9 | 10.31 | 18.59 | 7.32 | 33.9 | 0.02 | 109.9 | - | 384 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 8  | 1.9 | 10.39 | 17.97 | 7.22 | 34   | 0.02 | 109.4 | - | 386 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 7  | 1.9 | 10.33 | 18.02 | 7.21 | 33.8 | 0.02 | 108.9 | - | 386 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 6  | 2.8 | 11.07 | 13.35 | 7.02 | 31.1 | 0.02 | 105.6 | - | 391 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 5  | 2.8 | 11.04 | 13.35 | 7    | 30.9 | 0.02 | 105.3 | - | 391 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 4  | 3.7 | 8.49  | 10.92 | 6.78 | 30.5 | 0.02 | 76.6  | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 3  | 3.7 | 8.43  | 10.9  | 6.78 | 30.3 | 0.02 | 76    | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 2  | 4.7 | 5.4   | 9.84  | 6.77 | 33.7 | 0.02 | 47.5  | - | 398 |
| Benewah Lake | 2597 | DBSS | 6/26/97 | - | 3.1 | 1  | 4.7 | 5.43  | 9.84  | 6.78 | 33.7 | 0.02 | 47.7  | - | 398 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 6  | 0.3 | 10.59 | 21.76 | 8.31 | 35.9 | 0.02 | ****  | - | 370 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 5  | 0.8 | 10.62 | 21.56 | 7.85 | 36   | 0.02 | ****  | - | 377 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 4  | 1.8 | 11.49 | 17.39 | 7.29 | 35.3 | 0.02 | ****  | - | 398 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 3  | 2.7 | 11.51 | 14.86 | 7.09 | 33.8 | 0.02 | ****  | - | 408 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 2  | 3.6 | 9.61  | 13.13 | 6.92 | 33.1 | 0.02 | 90.9  | - | 417 |
| Benewah Lake | 2797 | DBSS | 7/8/97  | - | 2.3 | 1  | 4.6 | 5.87  | 11.66 | 6.9  | 36   | 0.02 | 53.6  | - | 421 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 10 | 0.4 | 10.59 | 23.51 | 9.06 | 37.8 | 0.02 | ****  | - | 356 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 9  | 0   | 10.66 | 23.29 | 9.02 | 37.8 | 0.02 | ****  | - | 358 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 8  | 1   | 10.58 | 23.37 | 9.03 | 37.8 | 0.02 | ****  | - | 358 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 7  | 1.9 | 10.62 | 23.2  | 8.9  | 37.5 | 0.02 | ****  | - | 361 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 6  | 2   | 10.67 | 23.01 | 8.72 | 37.6 | 0.02 | ****  | - | 366 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 5  | 3   | 11.46 | 18.22 | 7.41 | 34   | 0.02 | ****  | - | 407 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 4  | 3   | 11.38 | 18.08 | 7.31 | 33.6 | 0.02 | ****  | - | 409 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 3  | 4   | 10.54 | 15.01 | 7.05 | 33   | 0.02 | ****  | - | 421 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 2  | 4   | 9.11  | 14.54 | 6.99 | 34.9 | 0.02 | 88.9  | - | 426 |
| Benewah Lake | 2997 | SSJD | 7/24/97 | - | 3   | 1  | 4.8 | 5.06  | 13.37 | 6.97 | 40.5 | 0.03 | 48.2  | - | 437 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 5  | 0.3 | 11.3  | 26.45 | 9.38 | 42.4 | 0.03 | 138.4 | - | 259 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 4  | 1.6 | 11.69 | 24.11 | 9.41 | 42.5 | 0.03 | 137.3 | - | 256 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 3  | 2.8 | 10.7  | 22.54 | 8.41 | 39.8 | 0.03 | 122   | - | 276 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 2  | 3.6 | 10.2  | 20.44 | 7.28 | 37.3 | 0.02 | 111.7 | - | 309 |
| Benewah Lake | 3197 | DBSS | 8/5/97  | - | 2.7 | 1  | 4.7 | 2.78  | 15.59 | 7.06 | 47.9 | 0.03 | 27.2  | - | 337 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 9  | 0.4 | 8.91  | 23.87 | 8.9  | 42   | 0.03 | 104.8 | - | 260 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 8  | 0.9 | 8.91  | 23.86 | 8.9  | 42   | 0.03 | 104.8 | - | 256 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 7  | 1.4 | 8.89  | 23.86 | 8.91 | 42   | 0.03 | 104.6 | - | 251 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 6  | 1.9 | 8.89  | 23.76 | 8.83 | 42.1 | 0.03 | 104.4 | - | 250 |
| Benewah Lake | 3297 | DBSS | 8/14/97 | - | 4.3 | 5  | 2.3 | 8.88  | 23.75 | 8.74 | 42   | 0.03 | 104.3 | - | 248 |

|              |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
|--------------|------|------|----------|---|-----|----|-----|-------|-------|------|------|------|-------|---|-----|
| Benewah Lake | 3297 | DBSS | 8/14/97  | - | 4.3 | 4  | 2.9 | 7.94  | 22.77 | 8.28 | 41.6 | 0.03 | 91.6  | - | 254 |
| Benewah Lake | 3297 | DBSS | 8/14/97  | - | 4.3 | 3  | 3.4 | 8.65  | 20.97 | 7.3  | 41.2 | 0.03 | 96.3  | - | 277 |
| Benewah Lake | 3297 | DBSS | 8/14/97  | - | 4.3 | 2  | 3.9 | 8.53  | 20.2  | 7.12 | 41.6 | 0.03 | 93.4  | - | 277 |
| Benewah Lake | 3297 | DBSS | 8/14/97  | - | 4.3 | 1  | 4.4 | 2.15  | 18.22 | 6.93 | 49.6 | 0.03 | 22.7  | - | 293 |
|              |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 10 | 0.3 | 9.39  | 22.33 | 8.27 | 46.4 | 0.03 | 107.6 | - | 271 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 9  | 0.8 | 9.41  | 22.18 | 8.36 | 46.5 | 0.03 | 107.5 | - | 264 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 8  | 1.3 | 9.42  | 22.08 | 8.35 | 46.6 | 0.03 | 107.3 | - | 263 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 7  | 1.8 | 9.46  | 21.85 | 8.37 | 46.6 | 0.03 | 107.3 | - | 260 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 6  | 2.3 | 9.64  | 21.58 | 8.49 | 46.6 | 0.03 | 108.8 | - | 253 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 5  | 2.8 | 9.6   | 21.49 | 8.45 | 46.6 | 0.03 | 108.1 | - | 250 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 4  | 3.3 | 9.59  | 21.44 | 8.39 | 46.7 | 0.03 | 108   | - | 248 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 3  | 3.7 | 9.2   | 21.32 | 8.03 | 46.6 | 0.03 | 103.3 | - | 248 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 2  | 4.3 | 3.62  | 19.73 | 6.74 | 54.2 | 0.03 | 39.4  | - | 292 |
| Benewah Lake | 3497 | DBSS | 8/27/97  | - | 3.8 | 1  | 4.8 | 1.44  | 18.97 | 6.69 | 57.1 | 0.04 | 15.5  | - | 280 |
|              |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 10 | 0.2 | 8.58  | 17.07 | 7.22 | 46.1 | 0.03 | 88.5  | - | 308 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 9  | 0.5 | 8.57  | 17.07 | 7.22 | 46   | 0.03 | 88.6  | - | 308 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 8  | 1   | 8.56  | 17.05 | 7.22 | 46.1 | 0.03 | 88.4  | - | 307 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 7  | 1.5 | 8.53  | 17.07 | 7.23 | 46   | 0.03 | 88.3  | - | 306 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 6  | 2   | 8.57  | 17.07 | 7.23 | 46   | 0.03 | 88.6  | - | 304 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 5  | 2.5 | 8.64  | 17.05 | 7.23 | 46.1 | 0.03 | 89.3  | - | 302 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 4  | 3   | 8.7   | 17.05 | 7.23 | 46   | 0.03 | 90    | - | 300 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 3  | 3.5 | 8.73  | 17.03 | 7.23 | 46.1 | 0.03 | 90.2  | - | 297 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 2  | 4   | 8.71  | 17.03 | 7.24 | 46   | 0.03 | 90    | - | 296 |
| Benewah Lake | 3797 | DBAS | 9/16/97  | - | 1.5 | 1  | 4.5 | 8.73  | 17.03 | 7.23 | 46   | 0.03 | 90.2  | - | 300 |
|              |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 8  | 0.3 | 9.77  | 16.41 | 7.41 | 45.9 | 0.03 | 100.9 | - | 382 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 7  | 1   | 9.9   | 15.76 | 7.41 | 46   | 0.03 | 100.8 | - | 382 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 6  | 1.5 | 9.87  | 15.4  | 7.39 | 45.8 | 0.03 | 99.8  | - | 383 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 5  | 2   | 9.78  | 15.3  | 7.37 | 45.8 | 0.03 | 98.7  | - | 383 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 4  | 2.5 | 9.49  | 15.19 | 7.33 | 45.8 | 0.03 | 95.5  | - | 385 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 3  | 3   | 9.42  | 15.1  | 7.32 | 45.8 | 0.03 | 94.6  | - | 384 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 2  | 3.5 | 9.16  | 15.04 | 7.3  | 45.8 | 0.03 | 91.9  | - | 385 |
| Benewah Lake | 3997 | ASRA | 9/29/97  | - | 2.2 | 1  | 4   | 9.09  | 15.04 | 7.31 | 45.9 | 0.03 | 91.2  | - | 384 |
|              |      |      |          |   |     |    |     |       |       |      |      |      |       |   |     |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 9  | 0.3 | 10.61 | 10.45 | 7.15 | 49.1 | 0.03 | 94    | - | 369 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 8  | 0.6 | 10.61 | 10.48 | 7.12 | 49.3 | 0.03 | 94    | - | 372 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 7  | 1.1 | 10.63 | 10.48 | 7.11 | 49.3 | 0.03 | 94.2  | - | 372 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 6  | 1.6 | 10.63 | 10.43 | 7.11 | 49.4 | 0.03 | 94.1  | - | 372 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 5  | 2.1 | 10.6  | 10.45 | 7.08 | 49.3 | 0.03 | 93.8  | - | 372 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 4  | 2.6 | 10.55 | 10.41 | 7.04 | 49.3 | 0.03 | 93.3  | - | 373 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 3  | 3.1 | 10.17 | 10.41 | 7.02 | 49.5 | 0.03 | 90    | - | 373 |
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 2  | 3.6 | 9.84  | 10.4  | 6.97 | 49.8 | 0.03 | 87    | - | 375 |

|              |      |      |          |   |     |   |     |      |       |      |      |      |      |   |     |
|--------------|------|------|----------|---|-----|---|-----|------|-------|------|------|------|------|---|-----|
| Benewah Lake | 4297 | DBAS | 10/20/97 | - | 2.2 | 1 | 4.1 | 9.71 | 10.38 | 6.86 | 49.8 | 0.03 | 85.8 | - | 378 |
|--------------|------|------|----------|---|-----|---|-----|------|-------|------|------|------|------|---|-----|

| Location      | Phase | Samplers | Date    | Time | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|---------------|-------|----------|---------|------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| St. Joe River | 1597  | AS SS    | 4/18/97 | -    | 1.1        | 2        | 0.2       | 12.89                   |                 | 7.41 | 32.4                 | 0.02 | 102.6                      | -                  | -     |
| St. Joe River | 1597  | AS SS    | 4/18/97 | -    | 1.1        | 1        | 5         | 13.21                   | 5.54            | 7.41 | 32.3                 | 0.02 | *****                      | -                  | 356   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 12       | 0.2       | 11.4                    | 8.54            | 6.99 | 25.9                 | 0.02 | 96.7                       | -                  | 356   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 11       | 1         | 11.41                   | 8.46            | 6.96 | 25.9                 | 0.02 | 96.6                       | -                  | 364   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 10       | 1.9       | 11.44                   | 8.36            | 6.97 | 25.9                 | 0.02 | 96.6                       | -                  | 365   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 9        | 3         | 11.44                   | 8.36            | 6.96 | 25.9                 | 0.02 | 96.4                       | -                  | 365   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 8        | 4.1       | 11.44                   | 8.36            | 6.96 | 25.9                 | 0.02 | 96.5                       | -                  | 366   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 7        | 5         | 11.44                   | 8.35            | 6.95 | 25.9                 | 0.02 | 96.6                       | -                  | 367   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 6        | 6         | 11.46                   | 8.35            | 6.96 | 26                   | 0.02 | 96.7                       | -                  | 367   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 5        | 6.9       | 11.43                   | 8.38            | 6.96 | 25.9                 | 0.02 | 96.5                       | -                  | 368   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 4        | 8         | 11.43                   | 8.38            | 6.96 | 26                   | 0.02 | 96.6                       | -                  | 369   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 3        | 9.2       | 11.39                   | 8.4             | 6.97 | 26                   | 0.02 | 96.3                       | -                  | 369   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 2        | 10        | 11.44                   | 8.4             | 6.99 | 25.9                 | 0.02 | 96.7                       | -                  | 371   |
| St. Joe River | 1997  | DBSS     | 5/16/97 | -    | 0.4        | 1        | 11.2      | 11.45                   | 8.4             | 6.99 | 25.9                 | 0.02 | 96.8                       | -                  | 373   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 10       | 0.3       | 11.53                   | 8.97            | 7.08 | 26.3                 | 0.02 | 99.1                       | -                  | 363   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 9        | 0.3       | 11.53                   | 8.99            | 7.09 | 26.4                 | 0.02 | 99.1                       | -                  | 362   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 8        | 0.3       | 11.51                   | 8.99            | 7.08 | 26.6                 | 0.02 | 98.9                       | -                  | 363   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 7        | 0.3       | 11.53                   | 8.97            | 7.09 | 26.1                 | 0.02 | 99.1                       | -                  | 363   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 6        | 0.3       | 11.53                   | 8.99            | 7.09 | 26.4                 | 0.02 | 99.1                       | -                  | 363   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 5        | 9.1       | 11.53                   | 8.97            | 7.06 | 26.5                 | 0.02 | 98.7                       | -                  | 366   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 4        | 10        | 11.51                   | 8.97            | 7.06 | 26.5                 | 0.02 | 98.9                       | -                  | 366   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 3        | 10.1      | 11.51                   | 8.99            | 7.07 | 26.7                 | 0.02 | 98.8                       | -                  | 366   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 2        | 10.5      | 11.5                    | 8.97            | 7.06 | 26.4                 | 0.02 | 98.9                       | -                  | 366   |
| St. Joe River | 2197  | DBSS     | 5/29/97 | -    | 1.1        | 1        | 10.6      | 11.49                   | 8.97            | 7.06 | 26.6                 | 0.02 | 99.1                       | -                  | 366   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 10       | 0.8       | 11.35                   | 10.28           | 7.13 | 28.1                 | 0.02 | 100.9                      | -                  | 394   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 9        | 0.8       | 11.33                   | 10.3            | 7.13 | 28.1                 | 0.02 | 100.9                      | -                  | 394   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 8        | 0.9       | 11.33                   | 10.3            | 7.13 | 28.1                 | 0.02 | 100.9                      | -                  | 394   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 7        | 0.9       | 11.33                   | 10.3            | 7.13 | 28.1                 | 0.02 | 100.9                      | -                  | 394   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 6        | 0.8       | 11.35                   | 10.3            | 7.12 | 28.1                 | 0.02 | 101                        | -                  | 395   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 5        | 0.9       | 11.35                   | 10.3            | 7.13 | 28.1                 | 0.02 | 101                        | -                  | 395   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 4        | 0.9       | 11.36                   | 10.3            | 7.13 | 28.1                 | 0.02 | 101.1                      | -                  | 395   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 3        | 0.9       | 11.37                   | 10.28           | 7.13 | 28.1                 | 0.02 | 101.1                      | -                  | 395   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 2        | 1.1       | 11.37                   | 10.28           | 7.14 | 28.1                 | 0.02 | 101.2                      | -                  | 395   |
| St. Joe River | 2397  | DBAS     | 6/11/97 | -    | 1.7        | 1        | 0.9       | 11.38                   | 10.3            | 7.11 | 28                   | 0.02 | 101.3                      | -                  | 398   |
| St. Joe River | 2597  | DBSS     | 6/26/97 | -    | 3.1        | 11       | 0.2       | 11.21                   | 10.68           | 7.18 | 32.1                 | 0.02 | 100.6                      | -                  | 399   |

|               |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
|---------------|------|------|---------|---|-----|----|------|-------|-------|------|------|------|-------|---|-----|
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 10 | 1    | 11.17 | 10.66 | 7.17 | 32.1 | 0.02 | 100.1 | - | 400 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 9  | 1.9  | 11.2  | 10.64 | 7.17 | 32.1 | 0.02 | 100.4 | - | 402 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 8  | 2.9  | 11.17 | 10.64 | 7.17 | 32.1 | 0.02 | 100.2 | - | 402 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 7  | 4    | 11.19 | 10.64 | 7.17 | 32   | 0.02 | 100.3 | - | 402 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 6  | 4.9  | 11.17 | 10.64 | 7.17 | 32.1 | 0.02 | 100.2 | - | 402 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 5  | 6.1  | 11.19 | 10.64 | 7.17 | 32   | 0.02 | 100.3 | - | 402 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 4  | 7.1  | 11.18 | 10.61 | 7.18 | 32   | 0.02 | 100.1 | - | 401 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 3  | 8.1  | 11.18 | 10.62 | 7.17 | 32.1 | 0.02 | 100   | - | 401 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 2  | 9    | 11.18 | 10.62 | 7.18 | 32   | 0.02 | 100.2 | - | 400 |
| St. Joe River | 2597 | DBSS | 6/26/97 | - | 3.1 | 1  | 11.2 | 11.2  | 10.67 | 7.19 | 32   | 0.02 | 100.4 | - | 400 |
|               |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 13 | 0.4  | 9.9   | 15.06 | 7.14 | 34.6 | 0.02 | 97.6  | - | 414 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 12 | 1.3  | 9.89  | 15.04 | 7.16 | 34.6 | 0.02 | 97.5  | - | 414 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 11 | 2.4  | 9.9   | 15.07 | 7.15 | 34.6 | 0.02 | 97.6  | - | 415 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 10 | 3.5  | 9.92  | 15.02 | 7.15 | 34.6 | 0.02 | 97.7  | - | 415 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 9  | 4.4  | 9.9   | 15.09 | 7.15 | 34.6 | 0.02 | 97.8  | - | 415 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 8  | 5.4  | 9.9   | 15.06 | 7.15 | 34.6 | 0.02 | 97.7  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 7  | 6.4  | 9.89  | 15.02 | 7.15 | 34.6 | 0.02 | 97.5  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 6  | 7.4  | 9.88  | 15.02 | 7.16 | 34.6 | 0.02 | 97.4  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 5  | 8.3  | 9.9   | 15.02 | 7.16 | 34.6 | 0.02 | 97.6  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 4  | 9.4  | 9.89  | 15.06 | 7.16 | 34.6 | 0.02 | 97.5  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 3  | 10.4 | 9.91  | 15.06 | 7.18 | 34.6 | 0.02 | 97.7  | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 2  | 11.4 | 9.93  | 15.06 | 7.19 | 34.6 | 0.02 | 98    | - | 416 |
| St. Joe River | 2797 | DBSS | 7/8/97  | - | 2.5 | 1  | 12.3 | 10.07 | 15.07 | 7.22 | 34.6 | 0.02 | 99.3  | - | 416 |
|               |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 11 | 0.3  | 9.37  | 18.94 | 7.2  | 38.3 | 0.02 | ****  | - | 430 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 10 | 1.3  | 9.42  | 18.47 | 7.2  | 38.2 | 0.02 | 99.9  | - | 431 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 9  | 2.3  | 9.4   | 18.22 | 7.21 | 38.1 | 0.02 | 99.2  | - | 432 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 8  | 3.3  | 9.42  | 18.15 | 7.21 | 38.2 | 0.02 | 99.2  | - | 432 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 7  | 4.3  | 9.43  | 18.22 | 7.21 | 38.2 | 0.02 | 99.5  | - | 433 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 6  | 5.3  | 9.41  | 18.22 | 7.2  | 38.1 | 0.02 | 99.3  | - | 434 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 5  | 6.3  | 9.4   | 18.1  | 7.21 | 38.1 | 0.02 | 99    | - | 433 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 4  | 7.3  | 9.41  | 18.22 | 7.21 | 38.1 | 0.02 | 99.3  | - | 434 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 3  | 8.3  | 9.4   | 18.18 | 7.22 | 38.1 | 0.02 | 99.1  | - | 434 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 2  | 9.3  | 9.43  | 18.1  | 7.25 | 38.1 | 0.02 | 99.3  | - | 434 |
| St. Joe River | 2997 | SSJD | 7/24/97 | - | 3   | 1  | 10.3 | 9.47  | 18.08 | 7.29 | 38.1 | 0.02 | 99.7  | - | 434 |
|               |      |      |         |   |     |    |      |       |       |      |      |      |       |   |     |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 13 | 0.3  | 9.07  | 23.01 | 7.3  | 44.8 | 0.03 | 104.4 | - | 351 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 12 | 1.5  | 9.11  | 21.37 | 7.26 | 44.5 | 0.03 | 101.5 | - | 354 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 11 | 2.4  | 9     | 21.28 | 7.25 | 44.7 | 0.03 | 100.2 | - | 355 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 10 | 3.4  | 8.91  | 20.55 | 7.22 | 44.3 | 0.03 | 97.8  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 9  | 4.5  | 8.64  | 19.94 | 7.19 | 44   | 0.03 | 93.7  | - | 357 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 8  | 5.5  | 8.62  | 19.71 | 7.19 | 44.1 | 0.03 | 93.1  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 7  | 6.5  | 8.54  | 19.57 | 7.19 | 43.7 | 0.03 | 92    | - | 356 |

|               |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
|---------------|------|------|---------|---|-----|----|------|------|-------|------|------|------|-------|---|-----|
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 6  | 7.5  | 8.51 | 19.45 | 7.19 | 43.7 | 0.03 | 91.3  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 5  | 8.5  | 8.46 | 19.33 | 7.19 | 43.8 | 0.03 | 90.6  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 4  | 9.5  | 8.44 | 19.26 | 7.19 | 43.6 | 0.03 | 90.3  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 3  | 10.5 | 8.43 | 19.25 | 7.19 | 43.7 | 0.03 | 90.2  | - | 356 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 2  | 11.5 | 8.4  | 19.25 | 7.19 | 43.7 | 0.03 | 89.8  | - | 355 |
| St. Joe River | 3197 | DBSS | 8/5/97  | - | 3   | 1  | 12.1 | 8.4  | 19.3  | 7.2  | 43.6 | 0.03 | 89.9  | - | 354 |
|               |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 14 | 0.2  | 8.47 | 20.77 | 7.23 | 55.6 | 0.04 | 94.3  | - | 317 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 13 | 0.8  | 8.49 | 20.27 | 7.25 | 55.8 | 0.04 | 93.4  | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 12 | 1.7  | 8.47 | 20.2  | 7.24 | 55.7 | 0.04 | 93.1  | - | 317 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 11 | 2.7  | 8.49 | 20.2  | 7.25 | 55.6 | 0.04 | 93.2  | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 10 | 3.7  | 8.49 | 20.2  | 7.24 | 55.7 | 0.04 | 93.3  | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 9  | 4.6  | 8.48 | 20.15 | 7.24 | 55.6 | 0.04 | 93    | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 8  | 5.6  | 8.46 | 20.09 | 7.24 | 55.6 | 0.04 | 92.9  | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 7  | 6.4  | 8.41 | 20.08 | 7.23 | 55.5 | 0.04 | 92.3  | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 6  | 7.5  | 8.4  | 20.02 | 7.23 | 55.6 | 0.04 | 92    | - | 316 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 5  | 8.5  | 8.39 | 20.02 | 7.24 | 55.5 | 0.04 | 91.9  | - | 315 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 4  | 9.6  | 8.38 | 20.01 | 7.24 | 55.6 | 0.04 | 91.7  | - | 315 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 3  | 10.4 | 8.38 | 20.01 | 7.25 | 55.5 | 0.04 | 91.7  | - | 314 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 2  | 11.5 | 8.41 | 20.01 | 7.26 | 55.7 | 0.04 | 92    | - | 313 |
| St. Joe River | 3497 | DBSS | 8/27/97 | - | 4   | 1  | 12.7 | 8.43 | 20.04 | 7.27 | 55.6 | 0.04 | 92.3  | - | 313 |
|               |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 13 | 0.2  | 8.51 | 17.45 | 7.2  | 54.9 | 0.04 | 88.7  | - | 308 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 12 | 0.9  | 8.48 | 17.51 | 7.21 | 55.2 | 0.04 | 88.5  | - | 307 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 11 | 1.9  | 8.47 | 17.47 | 7.2  | 55.1 | 0.04 | 88.3  | - | 308 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 10 | 2.9  | 8.47 | 17.45 | 7.19 | 55.2 | 0.04 | 88.3  | - | 307 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 9  | 3.9  | 8.48 | 17.44 | 7.21 | 55.1 | 0.04 | 88.3  | - | 305 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 8  | 4.9  | 8.49 | 17.43 | 7.2  | 55.1 | 0.04 | 88.5  | - | 305 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 7  | 5.9  | 8.51 | 17.44 | 7.2  | 55.3 | 0.04 | 88.6  | - | 305 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 6  | 6.9  | 8.49 | 17.42 | 7.2  | 55.2 | 0.04 | 88.6  | - | 303 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 5  | 7.9  | 8.48 | 17.42 | 7.21 | 55.2 | 0.04 | 88.3  | - | 303 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 4  | 8.9  | 8.51 | 17.42 | 7.21 | 55.3 | 0.04 | 88.6  | - | 302 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 3  | 9.9  | 8.51 | 17.42 | 7.21 | 55.2 | 0.04 | 88.6  | - | 301 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 2  | 10.9 | 8.53 | 17.44 | 7.22 | 55.2 | 0.04 | 88.8  | - | 300 |
| St. Joe River | 3797 | DBAS | 9/16/97 | - | 3.5 | 1  | 11.9 | 8.54 | 17.41 | 7.23 | 55.2 | 0.04 | 89    | - | 299 |
|               |      |      |         |   |     |    |      |      |       |      |      |      |       |   |     |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 11 | 0.2  | 9.83 | 16.29 | 7.24 | 52.9 | 0.03 | 101.3 | - | 397 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 10 | 1.6  | 9.77 | 14.01 | 7.22 | 52.5 | 0.03 | 95.8  | - | 400 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 9  | 2.6  | 9.71 | 13.91 | 7.22 | 52.4 | 0.03 | 95    | - | 400 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 8  | 3.6  | 9.72 | 13.88 | 7.23 | 52.5 | 0.03 | 95.1  | - | 400 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 7  | 4.6  | 9.66 | 13.86 | 7.23 | 52.4 | 0.03 | 94.4  | - | 399 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 6  | 5.6  | 9.47 | 13.86 | 7.24 | 52.5 | 0.03 | 92.6  | - | 399 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 5  | 6.6  | 9.44 | 13.86 | 7.25 | 52.3 | 0.03 | 92.3  | - | 399 |
| St. Joe River | 3997 | ASRA | 9/29/97 | - | 2.7 | 4  | 7.6  | 9.46 | 13.83 | 7.25 | 52.3 | 0.03 | 92.3  | - | 398 |



|               |      |      |          |   |     |    |      |       |       |      |      |      |      |       |     |
|---------------|------|------|----------|---|-----|----|------|-------|-------|------|------|------|------|-------|-----|
| St. Joe River | 3997 | ASRA | 9/29/97  | - | 2.7 | 3  | 8.6  | 9.43  | 13.88 | 7.27 | 52.2 | 0.03 | 92.1 | -     | 398 |
| St. Joe River | 3997 | ASRA | 9/29/97  | - | 2.7 | 2  | 9.6  | 9.55  | 13.83 | 7.27 | 52.3 | 0.03 | 93.3 | -     | 398 |
| St. Joe River | 3997 | ASRA | 9/29/97  | - | 2.7 | 1  | 10.6 | 9.5   | 13.83 | 7.29 | 52.3 | 0.03 | 92.7 | -     | 397 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 11 | 0.3  | 10.77 | 9.06  | 6.97 | 56.7 | 0.04 | 92.3 | -     | 399 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 10 | 1.3  | 10.75 | 9.07  | 6.95 | 56.7 | 0.04 | 92.1 | -     | 402 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 9  | 2.3  | 10.76 | 9.06  | 6.93 | 56.7 | 0.04 | 92.1 | -     | 403 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 8  | 3.3  | 10.76 | 9.06  | 6.91 | 56.7 | 0.04 | 92.1 | -     | 404 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 7  | 4.3  | 10.74 | 9.06  | 6.91 | 56.6 | 0.04 | 92   | -     | 405 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 6  | 5.3  | 10.76 | 9.06  | 6.88 | 56.6 | 0.04 | 92.1 | -     | 406 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 5  | 6.3  | 10.76 | 9.04  | 6.87 | 56.7 | 0.04 | 92.2 | -     | 406 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 4  | 7.3  | 10.78 | 9.02  | 6.85 | 56.6 | 0.04 | 92.3 | -     | 407 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 3  | 8.3  | 10.78 | 9.04  | 6.83 | 56.6 | 0.04 | 92.3 | -     | 408 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 2  | 9.3  | 10.8  | 9.02  | 6.81 | 56.5 | 0.04 | 92.4 | -     | 408 |
| St. Joe River | 4297 | DBAS | 10/20/97 | - | 2.1 | 1  | 10.3 | 10.83 | 9.04  | 6.76 | 56.7 | 0.04 | 92.8 | -     | 411 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 11 | 0.4  | 11.9  | 7.03  | 6.63 | 39.7 | 0.03 | 96.8 | ***** | 436 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 10 | 1.8  | 11.93 | 7.02  | 6.61 | 39.9 | 0.03 | 96.9 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 9  | 2.8  | 11.9  | 7.03  | 6.63 | 39.8 | 0.03 | 96.8 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 8  | 3.8  | 11.92 | 7.03  | 6.61 | 39.8 | 0.03 | 96.9 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 7  | 4.8  | 11.89 | 7.04  | 6.61 | 39.9 | 0.03 | 96.6 | ***** | 437 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 6  | 5.8  | 11.93 | 7.06  | 6.61 | 39.8 | 0.03 | 96.9 | ***** | 436 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 5  | 6.8  | 11.93 | 7.06  | 6.61 | 39.9 | 0.03 | 96.9 | ***** | 436 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 4  | 7.8  | 11.95 | 7.07  | 6.61 | 40.2 | 0.03 | 97.1 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 3  | 8.8  | 11.96 | 7.08  | 6.61 | 39.9 | 0.03 | 97.3 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 2  | 9.8  | 12    | 7.08  | 6.61 | 39.9 | 0.03 | 97.6 | ***** | 438 |
| St. Joe River | 4497 | DBAS | 11/3/97  | - | 1   | 1  | 10.7 | 11.99 | 7.09  | 6.63 | 39.9 | 0.03 | 97.5 | ***** | 441 |

Appendix A.2 Vertical hydrolab profiles for all thirteen station on Coeur d'Alene Lake, 1998.

| Location     | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 13       | 0.3       | 13.29                   | 4.16            | 6.94 | 49.5                 | 0.03 | 101                        | 59                 | 405   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 12       | 1.2       | 13.33                   | 4.03            | 6.98 | 49.5                 | 0.03 | 101                        | 43                 | 403   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 11       | 2.2       | 13.39                   | 3.91            | 6.98 | 49.6                 | 0.03 | 101.1                      | 113                | 403   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 10       | 3.2       | 13.38                   | 3.91            | 6.91 | 49.7                 | 0.03 | 101                        | 101                | 407   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 9        | 4.2       | 13.43                   | 3.88            | 6.91 | 49.8                 | 0.03 | 101.3                      | 149                | 406   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 8        | 5.2       | 13.5                    | 3.88            | 6.9  | 49.8                 | 0.03 | 101.8                      | 59                 | 405   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 7        | 6.2       | 13.52                   | 3.91            | 6.86 | 49.8                 | 0.03 | 102.1                      | 131                | 406   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 6        | 7.2       | 13.53                   | 3.94            | 6.81 | 49.7                 | 0.03 | 102.2                      | 51                 | 406   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 5        | 8.2       | 13.54                   | 3.98            | 6.82 | 49.9                 | 0.03 | 102.4                      | 154                | 405   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 4        | 9.2       | 13.5                    | 3.96            | 6.72 | 49.6                 | 0.03 | 102.1                      | 103                | 407   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 3        | 10.2      | 13.49                   | 4.01            | 6.69 | 49.8                 | 0.03 | 102.1                      | 158                | 407   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 2        | 11.2      | 13.39                   | 4.26            | 6.58 | 49.7                 | 0.03 | 102                        | 201                | 408   |
| Rockford Bay | 1098  | DBRP     | 3/13/98 | 10:09:11 | 3          | 1        | 12.2      | 13.08                   | 4.54            | 6.42 | 50.9                 | 0.03 | 100.4                      | 215                | 409   |
|              |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 14       | 0.3       | 12.86                   | 5.87            | 7.09 | 51.3                 | 0.03 | 103                        | 124                | 441   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 13       | 1.3       | 12.9                    | 5.77            | 7.05 | 51.5                 | 0.03 | 103.1                      | 128                | 444   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 12       | 2.3       | 12.88                   | 5.6             | 7.06 | 51.2                 | 0.03 | 102.5                      | 53                 | 444   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 11       | 3.3       | 12.82                   | 5.36            | 6.97 | 51                   | 0.03 | 101.4                      | 121                | 448   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 10       | 4.3       | 12.79                   | 5.41            | 6.95 | 51.2                 | 0.03 | 101.3                      | 111                | 449   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 9        | 5.2       | 12.76                   | 5.14            | 6.96 | 51.1                 | 0.03 | 100.3                      | 141                | 447   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 8        | 6.2       | 12.75                   | 5.03            | 6.97 | 51.7                 | 0.03 | 100                        | 114                | 446   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 7        | 7.1       | 12.75                   | 5.03            | 6.92 | 51.5                 | 0.03 | 100                        | 107                | 448   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 6        | 8.1       | 12.71                   | 4.99            | 6.92 | 51.7                 | 0.03 | 99.5                       | 120                | 447   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 5        | 9.2       | 12.68                   | 4.96            | 6.91 | 51.8                 | 0.03 | 99.2                       | 104                | 447   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 4        | 10.2      | 12.49                   | 4.88            | 6.87 | 52.2                 | 0.03 | 97.6                       | 137                | 447   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 3        | 11.3      | 12.43                   | 4.79            | 6.85 | 52.2                 | 0.03 | 96.9                       | 110                | 447   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 2        | 12.3      | 12.3                    | 4.81            | 6.86 | 52.5                 | 0.03 | 95.9                       | 242                | 446   |
| Rockford Bay | 1498  | DBAS     | 4/8/98  | 9:40:15  | 2.9        | 1        | 13.2      | 11.97                   | 4.96            | 6.79 | 53.9                 | 0.03 | 93.7                       | 433                | 447   |
|              |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 13       | 0.1       | 11.7                    | 15              | 7    | 47.4                 | 0.03 | 114.2                      | 129                | 390   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 12       | 1.6       | 12.33                   | 10.3            | 7.17 | 45.9                 | 0.03 | 108.3                      | 40                 | 383   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 11       | 2.6       | 12.33                   | 9.86            | 7.15 | 46.5                 | 0.03 | 107.1                      | 34                 | 383   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 10       | 3.6       | 12.33                   | 9.2             | 7.11 | 47.1                 | 0.03 | 105.4                      | 40                 | 385   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 9        | 4.5       | 12.33                   | 8.79            | 7.05 | 47.1                 | 0.03 | 104.4                      | 56                 | 387   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 8        | 5.6       | 12.36                   | 8.3             | 7.04 | 47.5                 | 0.03 | 103.4                      | 34                 | 387   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 7        | 6.6       | 12.36                   | 8.25            | 7.03 | 47.7                 | 0.03 | 103.3                      | 38                 | 386   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 6        | 7.4       | 12.36                   | 7.53            | 6.98 | 47.9                 | 0.03 | 101.4                      | 57                 | 388   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 5        | 8.5       | 12.36                   | 7.4             | 6.99 | 48.3                 | 0.03 | 101.1                      | 44                 | 387   |
| Rockford Bay | 1698  | DBASRP   | 4/23/98 | 13:24:29 | 5          | 4        | 9.4       | 12.36                   | 6.94            | 6.91 | 48.4                 | 0.03 | 100                        | 59                 | 391   |

|              |      |        |         |          |     |    |      |       |      |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|------|------|------|------|-------|-----|-----|
| Rockford Bay | 1698 | DBASRP | 4/23/98 | 13:24:29 | 5   | 3  | 10.6 | 12.36 | 6.82 | 6.94 | 48.6 | 0.03 | 99.7  | 56  | 388 |
| Rockford Bay | 1698 | DBASRP | 4/23/98 | 13:24:29 | 5   | 2  | 11.6 | 12.29 | 6.69 | 6.89 | 48.4 | 0.03 | 98.8  | 132 | 390 |
| Rockford Bay | 1698 | DBASRP | 4/23/98 | 13:24:29 | 5   | 1  | 12.5 | 12.4  | 6.44 | 6.87 | 48.8 | 0.03 | 99.1  | 157 | 390 |
|              |      |        |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 14 | 0.3  | 11.57 | 10.9 | 7.25 | 42.3 | 0.03 | 105.6 | 119 | 399 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 13 | 1.6  | 11.6  | 10.8 | 7.23 | 42.2 | 0.03 | 105.8 | 101 | 401 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 12 | 2.6  | 11.6  | 10.7 | 7.21 | 42.2 | 0.03 | 105.5 | 135 | 401 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 11 | 3.6  | 11.66 | 10.5 | 7.22 | 42.5 | 0.03 | 105.5 | 327 | 399 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 10 | 4.6  | 11.61 | 10.3 | 7.23 | 41.9 | 0.03 | 104.5 | 118 | 395 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 9  | 5.6  | 11.65 | 10.2 | 7.12 | 43.1 | 0.03 | 104.8 | 120 | 398 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 8  | 6.6  | 11.49 | 10   | 7.08 | 43.2 | 0.03 | 102.8 | 301 | 398 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 7  | 7.6  | 11.43 | 9.74 | 7.08 | 42.8 | 0.03 | 101.6 | 112 | 394 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 6  | 8.6  | 11.49 | 9.51 | 7.02 | 43.4 | 0.03 | 101.6 | 133 | 395 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 5  | 9.6  | 11.39 | 9.02 | 6.96 | 44.2 | 0.03 | 99.5  | 147 | 395 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 4  | 10.6 | 11.39 | 8.92 | 7    | 44.4 | 0.03 | 99.3  | 235 | 390 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 3  | 11.6 | 11.42 | 8.6  | 6.89 | 44.8 | 0.03 | 98.8  | 118 | 391 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 2  | 12.6 | 11.42 | 8.07 | 6.88 | 45.6 | 0.03 | 97.5  | 57  | 388 |
| Rockford Bay | 2198 | ASJLDT | 5/26/98 | 11:37:29 | 3.6 | 1  | 13.6 | 11.4  | 7.86 | 6.89 | 46.2 | 0.03 | 96.8  | 318 | 386 |
|              |      |        |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 14 | 0.3  | 10.87 | 16.9 | 7.63 | 44.2 | 0.03 | 112.2 | 54  | 442 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 13 | 1.4  | 10.89 | 16.7 | 7.62 | 44.2 | 0.03 | 112.1 | 52  | 445 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 12 | 2.4  | 10.92 | 16.4 | 7.61 | 44.3 | 0.03 | 111.7 | 56  | 447 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 11 | 3.4  | 11    | 16.1 | 7.6  | 44.1 | 0.03 | 111.8 | 53  | 450 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 10 | 4.4  | 10.98 | 16   | 7.57 | 43.6 | 0.03 | 111.2 | 125 | 451 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 9  | 5.4  | 11.04 | 15.8 | 7.55 | 43.7 | 0.03 | 111.5 | 43  | 454 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 8  | 6.4  | 11.1  | 15.7 | 7.54 | 44.3 | 0.03 | 111.9 | 128 | 455 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 7  | 7.4  | 11.11 | 15.6 | 7.51 | 44.3 | 0.03 | 111.7 | 111 | 458 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 6  | 8.4  | 11.2  | 15.4 | 7.5  | 44.1 | 0.03 | 112   | 52  | 460 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 5  | 9.4  | 11.3  | 15.2 | 7.49 | 44.3 | 0.03 | 112.7 | 57  | 462 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 4  | 10.4 | 11.47 | 14.7 | 7.46 | 44.8 | 0.03 | 113.2 | 46  | 464 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 3  | 11.4 | 11.53 | 13.7 | 7.37 | 44.3 | 0.03 | 111.3 | 101 | 468 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 2  | 12.4 | 11.13 | 12.9 | 7.16 | 44.1 | 0.03 | 105.5 | 116 | 475 |
| Rockford Bay | 2398 | DTJLAS | 6/8/98  | 10:30    | 4   | 1  | 13.4 | 11    | 11.9 | 7.02 | 43.5 | 0.03 | 101.9 | 231 | 483 |
|              |      |        |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 14 | 0.2  | 10.9  | 17   | 7.89 | 42.7 | 0.03 | 113.5 | 32  | 437 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 13 | 1.6  | 11.09 | 16.1 | 7.79 | 42.5 | 0.03 | 113.3 | 41  | 443 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 12 | 2.6  | 11.1  | 15.8 | 7.76 | 42.5 | 0.03 | 112.7 | 42  | 445 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 11 | 3.6  | 11.02 | 15.7 | 7.66 | 42.6 | 0.03 | 111.7 | 52  | 448 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 10 | 4.6  | 11    | 15.6 | 7.63 | 42.2 | 0.03 | 111.1 | 105 | 451 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 9  | 5.6  | 10.99 | 15.5 | 7.53 | 42.3 | 0.03 | 110.8 | 44  | 455 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 8  | 6.6  | 10.61 | 15.4 | 7.33 | 43.2 | 0.03 | 106.8 | 49  | 459 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 7  | 7.6  | 10.76 | 15.2 | 7.35 | 42.7 | 0.03 | 107.8 | 105 | 461 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 6  | 8.6  | 10.61 | 14.5 | 7.15 | 42.3 | 0.03 | 104.6 | 38  | 467 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 5  | 9.6  | 10.41 | 13.1 | 7.04 | 42.2 | 0.03 | 99.6  | 59  | 470 |

|              |      |        |         |          |     |    |      |       |      |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|------|------|------|------|-------|-----|-----|
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 4  | 10.6 | 10.21 | 12.4 | 6.96 | 42.4 | 0.03 | 96.1  | 58  | 472 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 3  | 11.6 | 10.19 | 11.3 | 6.91 | 41.7 | 0.03 | 93.6  | 124 | 474 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 2  | 12.6 | 9.92  | 10.7 | 6.84 | 42   | 0.03 | 89.8  | 133 | 477 |
| Rockford Bay | 2598 | DTASBH | 6/24/98 | 11:01    | 3.3 | 1  | 13.6 | 9.85  | 9.92 | 6.78 | 42.3 | 0.03 | 87.5  | 619 | 480 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 14 | 0.4  | 10.13 | 20.7 | 8.12 | 44.5 | 0.03 | 112.8 | 133 | 359 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 13 | 1.2  | 10.2  | 20.3 | 8.16 | 44.4 | 0.03 | 112.8 | 134 | 359 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 12 | 2.2  | 10.27 | 20.2 | 8.19 | 44.4 | 0.03 | 113.1 | 131 | 359 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 11 | 3.2  | 10.37 | 20   | 8.22 | 44.4 | 0.03 | 113.9 | 123 | 359 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 10 | 4.2  | 10.38 | 20   | 8.22 | 44.5 | 0.03 | 114   | 125 | 360 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 9  | 5.2  | 10.52 | 19.7 | 8.27 | 44   | 0.03 | 114.9 | 111 | 360 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 8  | 6.2  | 10.64 | 19.4 | 8.31 | 43.7 | 0.03 | 115.5 | 135 | 360 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 7  | 7.2  | 10.73 | 19.1 | 8.31 | 43.6 | 0.03 | 115.7 | 136 | 361 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 6  | 8.2  | 10.81 | 18.5 | 8.18 | 43.4 | 0.03 | 115.2 | 130 | 366 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 5  | 9.2  | 10.77 | 17   | 7.88 | 43.3 | 0.03 | 111.1 | 109 | 375 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 4  | 10.2 | 9.98  | 14.8 | 7.21 | 42.2 | 0.03 | 98.3  | 136 | 386 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 3  | 11.2 | 9.47  | 13.2 | 6.98 | 41.6 | 0.03 | 90    | 159 | 389 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 2  | 12.2 | 9.27  | 12.4 | 6.89 | 41.6 | 0.03 | 86.5  | 19  | 392 |
| Rockford Bay | 2798 | ASDNBH | 7/8/98  | 9:03:16  | 3.3 | 1  | 13.2 | 9     | 12.2 | 6.85 | 41.6 | 0.03 | 83.6  | 412 | 394 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 13 | 0.7  | 9.39  | 22   | 7.84 | 50.7 | 0.03 | 107   | 115 | 333 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 12 | 1.7  | 9.46  | 21.6 | 7.84 | 50.3 | 0.03 | 107   | 101 | 332 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 11 | 2.7  | 9.55  | 21.3 | 7.84 | 50.3 | 0.03 | 107.3 | 101 | 332 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 10 | 3.7  | 9.59  | 21.1 | 7.84 | 49.8 | 0.03 | 107.4 | 104 | 332 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 9  | 4.7  | 9.7   | 20.6 | 7.83 | 49.6 | 0.03 | 107.6 | 112 | 332 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 8  | 5.7  | 9.78  | 20.3 | 7.85 | 49.3 | 0.03 | 107.8 | 110 | 330 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 7  | 6.7  | 9.89  | 19.9 | 7.85 | 49.2 | 0.03 | 108.2 | 58  | 330 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 6  | 7.7  | 9.91  | 19.7 | 7.81 | 49.1 | 0.03 | 108   | 58  | 329 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 5  | 8.7  | 9.94  | 19.6 | 7.76 | 48.9 | 0.03 | 108   | 104 | 329 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 4  | 9.7  | 10.18 | 18.5 | 7.72 | 48.2 | 0.03 | 108.4 | 122 | 329 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 3  | 10.7 | 10.29 | 17.2 | 7.54 | 47.5 | 0.03 | 106.7 | 118 | 330 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 2  | 11.7 | 10.08 | 14.8 | 7.34 | 46.3 | 0.03 | 99.1  | 133 | 329 |
| Rockford Bay | 2998 | ASBH   | 7/20/98 | 10:54:42 | 6.5 | 1  | 12.7 | 9.9   | 13.4 | 7.16 | 45.7 | 0.03 | 94.6  | 207 | 326 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 14 | 0.3  | 8.77  | 24   | 7.81 | 55   | 0.04 | 103.8 | 43  | 334 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 13 | 1.7  | 8.79  | 23.9 | 7.81 | 55   | 0.04 | 103.7 | 46  | 333 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 12 | 2.8  | 8.82  | 23.8 | 7.82 | 54.8 | 0.04 | 103.9 | 37  | 333 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 11 | 3.9  | 8.85  | 23.7 | 7.81 | 54.9 | 0.04 | 104.1 | 51  | 333 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 10 | 4.9  | 8.87  | 23.6 | 7.8  | 54.9 | 0.04 | 104.2 | 37  | 333 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 9  | 5.8  | 8.86  | 23.5 | 7.78 | 54.7 | 0.04 | 103.8 | 101 | 333 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 8  | 6.9  | 8.87  | 23.4 | 7.77 | 54.7 | 0.04 | 103.8 | 102 | 332 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 7  | 7.9  | 8.87  | 23.4 | 7.68 | 54.6 | 0.04 | 103.7 | 117 | 332 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 6  | 8.8  | 8.49  | 23.1 | 7.45 | 54.3 | 0.03 | 98.8  | 113 | 335 |
| Rockford Bay | 3298 | DBAS   | 8/10/98 | 10:22:34 | 9.9 | 5  | 9.7  | 9.36  | 20.9 | 7.48 | 50.8 | 0.03 | 104.4 | 159 | 333 |

|              |      |      |         |          |     |    |      |       |      |      |      |      |       |     |     |
|--------------|------|------|---------|----------|-----|----|------|-------|------|------|------|------|-------|-----|-----|
| Rockford Bay | 3298 | DBAS | 8/10/98 | 10:22:34 | 9.9 | 4  | 10.9 | 10.52 | 16.9 | 7.39 | 47.9 | 0.03 | 108.3 | 124 | 330 |
| Rockford Bay | 3298 | DBAS | 8/10/98 | 10:22:34 | 9.9 | 3  | 11.7 | 9.81  | 15.3 | 7.19 | 47.1 | 0.03 | 97.5  | 143 | 328 |
| Rockford Bay | 3298 | DBAS | 8/10/98 | 10:22:34 | 9.9 | 2  | 12.7 | 9.66  | 13.8 | 7.06 | 46.8 | 0.03 | 93    | 117 | 316 |
| Rockford Bay | 3298 | DBAS | 8/10/98 | 10:22:34 | 9.9 | 1  | 13.7 | 8.63  | 12.3 | 7.03 | 46.8 | 0.03 | 80.3  | 131 | 376 |
|              |      |      |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 18 | 0.4  | 9.09  | 21.9 | 7.74 | 54.6 | 0.03 | 103.6 | 51  | 347 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 17 | 1.5  | 9.07  | 21.8 | 7.78 | 54.5 | 0.03 | 103.2 | 123 | 348 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 16 | 2.2  | 9.06  | 21.7 | 7.78 | 54.8 | 0.04 | 103   | 111 | 348 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 15 | 3    | 9.12  | 21.4 | 7.77 | 54.8 | 0.04 | 103   | 59  | 350 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 14 | 3.7  | 9.12  | 21.4 | 7.77 | 54.6 | 0.04 | 102.9 | 125 | 350 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 13 | 4.5  | 9.13  | 21.3 | 7.73 | 54.8 | 0.04 | 102.9 | 52  | 352 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 12 | 5.2  | 9.13  | 21.3 | 7.71 | 54.7 | 0.04 | 102.8 | 102 | 352 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 11 | 6    | 9.1   | 21.2 | 7.67 | 54.7 | 0.04 | 102.4 | 117 | 354 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 10 | 6.7  | 9.08  | 21.2 | 7.64 | 54.8 | 0.04 | 102.2 | 103 | 355 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 9  | 7.5  | 9.04  | 21.2 | 7.56 | 54.9 | 0.04 | 101.6 | 103 | 357 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 8  | 8.2  | 9.02  | 21.1 | 7.49 | 55   | 0.04 | 101.2 | 113 | 359 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 7  | 9    | 8.92  | 20.5 | 7.34 | 54   | 0.03 | 99    | 109 | 364 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 6  | 9.7  | 9.51  | 19.3 | 7.22 | 51.9 | 0.03 | 103.1 | 47  | 368 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 5  | 10.5 | 10.19 | 17.3 | 7.13 | 47.9 | 0.03 | 105.9 | 133 | 370 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 4  | 11.2 | 9.73  | 15.4 | 6.95 | 46.9 | 0.03 | 97.3  | 110 | 375 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 3  | 12   | 8.71  | 12.9 | 6.76 | 46.7 | 0.03 | 82.4  | 50  | 378 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 2  | 12.7 | 8.56  | 11.5 | 6.73 | 47   | 0.03 | 78.4  | 131 | 379 |
| Rockford Bay | 3498 | DTAS | 8/24/98 | 15:12:19 | 10  | 1  | 13.5 | 8.74  | 11.5 | 6.77 | 47.1 | 0.03 | 80    | 239 | 377 |
|              |      |      |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 14 | 0.3  | 8.93  | 21.9 | 7.83 | 53.6 | 0.03 | 102.1 | 58  | 330 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 13 | 1.6  | 8.95  | 21.8 | 7.79 | 53.4 | 0.03 | 102   | 57  | 333 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 12 | 2.6  | 8.98  | 21.5 | 7.78 | 53.5 | 0.03 | 101.8 | 108 | 335 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 11 | 3.6  | 8.99  | 21.4 | 7.76 | 53.4 | 0.03 | 101.8 | 109 | 336 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 10 | 4.6  | 8.99  | 21.4 | 7.73 | 53.6 | 0.03 | 101.7 | 55  | 337 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 9  | 5.6  | 8.98  | 21.3 | 7.7  | 53.4 | 0.03 | 101.5 | 58  | 339 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 8  | 6.6  | 8.99  | 21.3 | 7.69 | 53.5 | 0.03 | 101.6 | 131 | 339 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 7  | 7.6  | 8.98  | 21.3 | 7.66 | 53.2 | 0.03 | 101.4 | 58  | 341 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 6  | 8.6  | 8.91  | 21.2 | 7.56 | 53.8 | 0.03 | 100.4 | 103 | 345 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 5  | 9.6  | 8.8   | 21.1 | 7.46 | 53.5 | 0.03 | 99    | 123 | 349 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 4  | 10.6 | 8.88  | 20.4 | 7.34 | 53   | 0.03 | 98.6  | 132 | 353 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 3  | 11.6 | 8.79  | 19   | 7.01 | 50   | 0.03 | 94.8  | 54  | 365 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 2  | 12.6 | 8.5   | 14.1 | 6.74 | 46.1 | 0.03 | 82.7  | 38  | 370 |
| Rockford Bay | 3598 | DTAS | 9/2/98  | 10:48:53 | 8.2 | 1  | 13.6 | 7.11  | 12.3 | 6.64 | 47.2 | 0.03 | 66.5  | 715 | 372 |
|              |      |      |         |          |     |    |      |       |      |      |      |      |       |     |     |
| Rockford Bay | 3998 | ASKB | 9/30/98 | 10:40:23 | 7.9 | 14 | 0.2  | 9.21  | 18   | 7.66 | 55.4 | 0.04 | 97.2  | 52  | 361 |
| Rockford Bay | 3998 | ASKB | 9/30/98 | 10:40:23 | 7.9 | 13 | 1.3  | 9.21  | 18   | 7.61 | 55.4 | 0.04 | 97.2  | 142 | 365 |
| Rockford Bay | 3998 | ASKB | 9/30/98 | 10:40:23 | 7.9 | 12 | 2.3  | 9.18  | 17.9 | 7.59 | 55.6 | 0.04 | 96.7  | 52  | 366 |
| Rockford Bay | 3998 | ASKB | 9/30/98 | 10:40:23 | 7.9 | 11 | 3.2  | 9.15  | 17.9 | 7.57 | 55.3 | 0.04 | 96.3  | 110 | 367 |
| Rockford Bay | 3998 | ASKB | 9/30/98 | 10:40:23 | 7.9 | 10 | 4.3  | 9.14  | 17.9 | 7.55 | 55.2 | 0.04 | 96.1  | 136 | 369 |

|              |      |        |          |          |     |    |      |      |      |      |      |      |      |     |     |
|--------------|------|--------|----------|----------|-----|----|------|------|------|------|------|------|------|-----|-----|
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 9  | 5.3  | 9.09 | 17.8 | 7.52 | 55.4 | 0.04 | 95.6 | 116 | 371 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 8  | 6.3  | 9.07 | 17.8 | 7.49 | 55.7 | 0.04 | 95.3 | 213 | 372 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 7  | 7.2  | 9.04 | 17.8 | 0    | 55.4 | 0.04 | 94.9 | 122 | 372 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 6  | 8.3  | 9.06 | 17.8 | 7.48 | 55.7 | 0.04 | 95.1 | 210 | 374 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 5  | 9.3  | 9.03 | 17.7 | 7.48 | 55.3 | 0.04 | 94.7 | 332 | 374 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 4  | 10.3 | 9.03 | 17.7 | 7.45 | 55.6 | 0.04 | 94.7 | 234 | 376 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 3  | 11.3 | 8.97 | 17.7 | 7.43 | 55.7 | 0.04 | 94.1 | 346 | 378 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 2  | 12.3 | 8.93 | 17.7 | 7.33 | 55.5 | 0.04 | 93.6 | 203 | 382 |
| Rockford Bay | 3998 | ASKB   | 9/30/98  | 10:40:23 | 7.9 | 1  | 13.3 | 7.81 | 17.3 | 7.11 | 55.5 | 0.04 | 81.1 | 850 | 389 |
|              |      |        |          |          |     |    |      |      |      |      |      |      |      |     |     |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 14 | 0.4  | 9.39 | 12.9 | 7.37 | 52   | 0.03 | 87.7 | 39  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 13 | 0.9  | 9.35 | 12.9 | 7.37 | 52   | 0.03 | 87.4 | 37  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 12 | 1.9  | 9.36 | 12.9 | 7.39 | 52   | 0.03 | 87.5 | 26  | 354 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 11 | 2.9  | 9.32 | 12.9 | 7.34 | 52.1 | 0.03 | 87.1 | 50  | 356 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 10 | 3.9  | 9.35 | 12.9 | 7.36 | 52.1 | 0.03 | 87.4 | 39  | 356 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 9  | 4.9  | 9.33 | 12.9 | 7.34 | 52   | 0.03 | 87.2 | 25  | 356 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 8  | 5.9  | 9.33 | 12.9 | 7.34 | 52.1 | 0.03 | 87.1 | 37  | 356 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 7  | 6.9  | 9.35 | 12.9 | 7.34 | 51.8 | 0.03 | 87.4 | 33  | 356 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 6  | 7.9  | 9.35 | 12.9 | 7.35 | 52.2 | 0.03 | 87.4 | 33  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 5  | 8.9  | 9.36 | 12.9 | 7.34 | 51.8 | 0.03 | 87.5 | 37  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 4  | 9.9  | 9.35 | 12.8 | 7.34 | 51.8 | 0.03 | 87.3 | 42  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 3  | 10.9 | 9.37 | 12.8 | 7.32 | 51.9 | 0.03 | 87.3 | 37  | 355 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 2  | 11.9 | 9.4  | 12.7 | 7.32 | 51.7 | 0.03 | 87.6 | 49  | 354 |
| Rockford Bay | 4298 | DTASKB | 10/20/98 | 10:13:09 | 8.2 | 1  | 12.9 | 9.4  | 12.6 | 7.29 | 52.1 | 0.03 | 87.4 | 138 | 363 |
|              |      |        |          |          |     |    |      |      |      |      |      |      |      |     |     |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 17 | 0.5  | 8.58 | 10   | 7.26 | 52.7 | 0.03 | 75.5 | 30  | 385 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 16 | 1.7  | 8.59 | 10   | 7.24 | 52.7 | 0.03 | 75.6 | 35  | 387 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 15 | 2.7  | 8.58 | 10   | 7.22 | 52.6 | 0.03 | 75.5 | 32  | 388 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 14 | 2.7  | 8.57 | 0    | 7.23 | 52.7 | 0.03 | 82.5 | 6   | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 13 | 3.6  | 8.56 | 10   | 7.24 | 52.7 | 0.03 | 75.4 | 32  | 387 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 12 | 4.6  | 8.56 | 10   | 7.22 | 52.6 | 0.03 | 75.4 | 37  | 388 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 11 | 5.7  | 8.58 | 10   | 7.2  | 52.7 | 0.03 | 75.5 | 37  | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 10 | 6.6  | 8.58 | 10   | 7.21 | 52.7 | 0.03 | 96.8 | 5   | 389 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 9  | 6.7  | 8.56 | 0    | 7.2  | 52.6 | 0.03 | 94.9 | 8   | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 8  | 7.7  | 8.56 | 10   | 7.19 | 52.9 | 0.03 | 75.3 | 34  | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 7  | 8.7  | 8.56 | 10   | 7.2  | 52.3 | 0.03 | 75.3 | 41  | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 6  | 9.7  | 8.57 | 9.99 | 0    | 52.9 | 0.03 | 75.4 | 52  | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 5  | 10.6 | 8.59 | 9.92 | 7.2  | 52.8 | 0.03 | 75.4 | 107 | 390 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 3  | 11.7 | 8.51 | 9.82 | 7.16 | 52.5 | 0.03 | 74.6 | 101 | 391 |
| Rockford Bay | 4598 | KB AS  | 11/12/98 | 10:46:32 | 9.1 | 2  | 12.7 | 8.49 | 9.78 | 7.14 | 52.8 | 0.03 | 74.3 | 59  | 392 |

| Location          | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 15       | 0.3       | 13.5                    | 4.31            | 7.03 | 53.4                 | 0.03 | 103.3                      | 36                 | 394   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 14       | 1         | 13.6                    | 4.09            | 7.09 | 53.3                 | 0.03 | 102.8                      | 53                 | 393   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 13       | 2         | 13.5                    | 4.08            | 7.06 | 53.4                 | 0.03 | 102.6                      | 46                 | 395   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 12       | 3         | 13.5                    | 4.04            | 7.08 | 53.2                 | 0.03 | 102.5                      | 55                 | 393   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 11       | 4         | 13.5                    | 4.06            | 7.07 | 53.5                 | 0.03 | 102.3                      | 208                | 393   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 10       | 5         | 13.5                    | 4.03            | 7.01 | 53.4                 | 0.03 | 102.2                      | 107                | 395   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 9        | 6         | 13.5                    | 4.01            | 6.97 | 53.7                 | 0.03 | 102.2                      | 114                | 395   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 8        | 7         | 13.5                    | 4.03            | 6.99 | 53.9                 | 0.03 | 102.1                      | 39                 | 394   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 7        | 8         | 13.4                    | 4.06            | 6.94 | 53.9                 | 0.03 | 101.9                      | 110                | 396   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 6        | 9         | 13.5                    | 4.04            | 6.97 | 54.3                 | 0.03 | 102                        | 136                | 393   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 5        | 10        | 13.5                    | 4.08            | 6.93 | 54.3                 | 0.03 | 102.2                      | 102                | 393   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 4        | 11        | 13.5                    | 4.08            | 6.91 | 54.5                 | 0.03 | 102.2                      | 116                | 392   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 3        | 12        | 13.5                    | 4.11            | 6.85 | 54.5                 | 0.03 | 102.2                      | 118                | 394   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 2        | 13        | 13.5                    | 4.24            | 6.79 | 55.2                 | 0.04 | 102.6                      | 120                | 395   |
| Windy Bay Shallow | 1098  | DBRP     | 3/13/98 | 10:42:36 | 2          | 1        | 14        | 13.6                    | 4.29            | 6.74 | 55.4                 | 0.04 | 103.5                      | 147                | 394   |
|                   |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 16       | 0.3       | 12.6                    | 5.49            | 6.97 | 52.4                 | 0.03 | 100.2                      | 109                | 437   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 15       | 1         | 12.6                    | 5.5             | 7.01 | 52.4                 | 0.03 | 100.2                      | 108                | 436   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 14       | 2         | 12.6                    | 5.41            | 7.05 | 52.5                 | 0.03 | 99.7                       | 54                 | 435   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 13       | 3         | 12.6                    | 5.29            | 7.08 | 52.4                 | 0.03 | 99.5                       | 109                | 434   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 12       | 4         | 12.6                    | 5.34            | 7    | 52.4                 | 0.03 | 99.6                       | 142                | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 11       | 4.9       | 12.7                    | 5.26            | 7.03 | 52.7                 | 0.03 | 99.9                       | 41                 | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 10       | 5.9       | 12.7                    | 5.17            | 7.02 | 52.7                 | 0.03 | 99.8                       | 100                | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 9        | 7         | 12.7                    | 5.12            | 6.99 | 52.7                 | 0.03 | 99.6                       | 50                 | 439   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 8        | 8         | 12.7                    | 5.09            | 7    | 52.7                 | 0.03 | 99.6                       | 134                | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 7        | 9         | 12.7                    | 5.11            | 7    | 52.8                 | 0.03 | 99.4                       | 138                | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 6        | 10        | 12.7                    | 5.11            | 7.02 | 52.9                 | 0.03 | 99.6                       | 59                 | 437   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 5        | 11        | 12.7                    | 5.09            | 6.95 | 52.9                 | 0.03 | 99.8                       | 108                | 440   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 4        | 12        | 12.7                    | 5.06            | 6.95 | 53                   | 0.03 | 99.7                       | 157                | 440   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 3        | 13        | 12.7                    | 5.06            | 6.95 | 53                   | 0.03 | 99.8                       | 216                | 439   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 2        | 14        | 12.8                    | 5.08            | 6.96 | 52.9                 | 0.03 | 100.2                      | 140                | 438   |
| Windy Bay Shallow | 1498  | DBAS     | 4/8/98  | 10:45:10 | 2.8        | 1        | 15        | 12.8                    | 5.19            | 6.92 | 52.8                 | 0.03 | 100.6                      | 250                | 439   |
|                   |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 11       | 0.3       | 12.2                    | 6.05            | 7.08 | 49                   | 0.03 | 99.9                       | 103                | 418   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 10       | 1.7       | 12.2                    | 6.05            | 7.12 | 49                   | 0.03 | 99.4                       | 59                 | 422   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 9        | 2.9       | 12.2                    | 6.02            | 7.05 | 49                   | 0.03 | 99.8                       | 114                | 425   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 8        | 4.6       | 12.2                    | 6.03            | 6.96 | 49.2                 | 0.03 | 99.4                       | 48                 | 431   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 7        | 6.1       | 12.2                    | 5.79            | 7.01 | 49.3                 | 0.03 | 98.7                       | 124                | 428   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 6        | 7.7       | 12.2                    | 5.77            | 6.95 | 49.2                 | 0.03 | 98.7                       | 107                | 431   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 5        | 9         | 12.1                    | 5.82            | 6.95 | 49.2                 | 0.03 | 98.5                       | 41                 | 431   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 4        | 10.4      | 12.1                    | 5.67            | 6.99 | 49.4                 | 0.03 | 97.6                       | 134                | 428   |
| Windy Bay Shallow | 1698  | DBASRP   | 4/24/98 | 9:31:58  | 3.5        | 3        | 11.9      | 12.2                    | 4.88            | 6.88 | 50.5                 | 0.03 | 96.3                       | 130                | 433   |

|                   |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-------------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Windy Bay Shallow | 1698 | DBASRP | 4/24/98 | 9:31:58  | 3.5 | 2  | 14   | 12.2 | 5.56  | 6.91 | 49.6 | 0.03 | 98.4  | 56  | 431 |
| Windy Bay Shallow | 1698 | DBASRP | 4/24/98 | 9:31:58  | 3.5 | 1  | 15.7 | 12.8 | 5.14  | 6.93 | 50.1 | 0.03 | 102   | 116 | 430 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 13 | 0.3  | 11.7 | 10.51 | 7.15 | 42.2 | 0.03 | 106.2 | 52  | 396 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 12 | 1.5  | 11.7 | 10.48 | 7.19 | 42.2 | 0.03 | 106.1 | 108 | 394 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 11 | 3    | 11.7 | 10.45 | 7.14 | 42.2 | 0.03 | 105.8 | 128 | 395 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 10 | 4.5  | 11.7 | 10.09 | 7.04 | 42.9 | 0.03 | 104.9 | 141 | 397 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 9  | 6    | 11.6 | 9.91  | 7.05 | 43   | 0.03 | 103.5 | 153 | 392 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 8  | 7.5  | 11.6 | 9.73  | 7    | 43.1 | 0.03 | 102.7 | 204 | 391 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 6  | 9    | 11.5 | 9.58  | 6.94 | 43.3 | 0.03 | 101.9 | 258 | 390 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 5  | 10.5 | 11.5 | 9.5   | 6.93 | 43.5 | 0.03 | 101.7 | 245 | 384 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 4  | 12   | 11.5 | 9.4   | 6.88 | 43.7 | 0.03 | 101.4 | 125 | 378 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 3  | 13.5 | 11.5 | 9.22  | 6.77 | 44.1 | 0.03 | 100.9 | 147 | 379 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 2  | 15   | 11.5 | 9.09  | 6.77 | 44.1 | 0.03 | 100.4 | 156 | 370 |
| Windy Bay Shallow | 2198 | ASJLDT | 5/26/98 | 10:33:29 | 3.4 | 1  | 16.2 | 11   | 8.81  | 6.73 | 44.7 | 0.03 | 95.9  | 604 | 383 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 11 | 0.3  | 11   | 17.22 | 7.82 | 43.8 | 0.03 | 114.8 | 28  | 420 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 10 | 2    | 11   | 16.69 | 7.79 | 43.7 | 0.03 | 113   | 47  | 423 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 9  | 3.5  | 11   | 16.36 | 7.77 | 43.8 | 0.03 | 112.9 | 37  | 424 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 8  | 5    | 11   | 16.21 | 7.75 | 43.8 | 0.03 | 112.5 | 116 | 426 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 7  | 6.5  | 11.2 | 15.71 | 7.71 | 43.6 | 0.03 | 112.4 | 140 | 428 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 6  | 8    | 11.3 | 15.39 | 7.69 | 43.5 | 0.03 | 112.7 | 143 | 430 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 5  | 9.5  | 11.3 | 15.01 | 7.61 | 43.4 | 0.03 | 112   | 129 | 433 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 4  | 11   | 11.3 | 13.86 | 7.45 | 43   | 0.03 | 109.7 | 50  | 437 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 3  | 12.5 | 11.3 | 13.33 | 7.38 | 43.2 | 0.03 | 107.9 | 114 | 440 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 2  | 14   | 11   | 12.89 | 7.27 | 43.3 | 0.03 | 103.9 | 128 | 443 |
| Windy Bay Shallow | 2398 | DTJLAS | 6/8/98  | 10:36:03 | 4   | 1  | 15.5 | 10.8 | 12.36 | 7.21 | 43.5 | 0.03 | 101.1 | 240 | 445 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 16 | 0.3  | 10.9 | 16.73 | 8.09 | 43.1 | 0.03 | 112.5 | 104 | 399 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 15 | 1.6  | 10.9 | 16.72 | 8.07 | 43.2 | 0.03 | 112.6 | 102 | 402 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 14 | 2.6  | 10.9 | 16.68 | 8.04 | 43   | 0.03 | 112.4 | 59  | 404 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 13 | 3.6  | 10.9 | 16.63 | 8.02 | 43   | 0.03 | 112.7 | 32  | 405 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 12 | 4.6  | 10.9 | 16.55 | 7.95 | 43   | 0.03 | 112.5 | 56  | 408 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 11 | 5.6  | 11   | 15.79 | 7.78 | 42.9 | 0.03 | 111.2 | 54  | 411 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 10 | 6.6  | 10.9 |       | 7.65 | 42.1 | 0.03 | 109.3 | 134 | 415 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 9  | 7.6  | 10.8 | 15.23 | 7.58 | 42.2 | 0.03 | 108.6 | 101 | 417 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 8  | 8.6  | 10.8 | 15.04 | 7.48 | 42   | 0.03 | 107.9 | 101 | 419 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 7  | 9.6  | 10.4 | 14.03 | 7.28 | 41.8 | 0.03 | 101.5 | 121 | 424 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 6  | 10.6 | 10.3 | 13.55 | 7.17 | 42   | 0.03 | 99.7  | 137 | 427 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 5  | 11.6 | 9.78 | 9.92  | 6.97 | 42.4 | 0.03 | 86.9  | 155 | 431 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 4  | 12.6 | 9.97 | 12.04 | 6.97 | 41.7 | 0.03 | 93.1  | 122 | 432 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 3  | 13.6 | 9.71 | 9.06  | 6.87 | 43.6 | 0.03 | 84.5  | 120 | 434 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 2  | 14.6 | 9.7  | 9.1   | 6.85 | 43.4 | 0.03 | 84.5  | 206 | 436 |
| Windy Bay Shallow | 2598 | DTASBH | 6/24/98 | 11:03:05 | 3.1 | 1  | 15.6 | 9.7  | 9.02  | 6.86 | 43.5 | 0.03 | 84.4  | 256 | 436 |



|                   |      |        |         |         |     |    |      |      |       |      |      |      |       |     |     |
|-------------------|------|--------|---------|---------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
|                   |      |        |         |         |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 16 | 0.4  | 9.83 | 21.33 | 7.89 | 46.5 | 0.03 | 110.8 | 134 | 323 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 15 | 1.1  | 10.1 | 20.86 | 8    | 46.2 | 0.03 | 112.9 | 103 | 318 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 14 | 2.1  | 10.1 | 20.67 | 8.04 | 46.3 | 0.03 | 112.2 | 50  | 318 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 13 | 3.1  | 10.2 | 20.35 | 8.05 | 46   | 0.03 | 112.9 | 140 | 317 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 12 | 4.1  | 10.5 | 19.59 | 8.17 | 45.9 | 0.03 | 113.9 | 114 | 314 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 11 | 5.1  | 10.8 | 18.63 | 8.23 | 44.6 | 0.03 | 115   | 107 | 313 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 10 | 6.1  | 10.9 | 17.86 | 8.27 | 44.4 | 0.03 | 115   | 57  | 313 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 9  | 7.1  | 10.5 | 17.47 | 8.21 | 44   | 0.03 | 109.6 | 130 | 313 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 8  | 8.1  | 10.8 | 17.1  | 8.05 | 44   | 0.03 | 112.2 | 122 | 316 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 7  | 9.1  | 10.7 | 16.66 | 7.84 | 43.8 | 0.03 | 109.9 | 118 | 319 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 6  | 10.1 | 10.5 | 16.39 | 7.63 | 43.6 | 0.03 | 107.2 | 126 | 322 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 5  | 11.1 | 9.94 | 15.74 | 7.29 | 43.2 | 0.03 | 100   | 146 | 326 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 4  | 12.1 | 9.73 | 15.22 | 7.16 | 43   | 0.03 | 96.7  | 112 | 326 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 3  | 13.1 | 9.1  | 13.53 | 6.96 | 42   | 0.03 | 87.3  | 109 | 326 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 2  | 14.1 | 9.03 | 12.89 | 6.93 | 41.4 | 0.03 | 85.3  | 153 | 323 |
| Windy Bay Shallow | 2798 | ASDNBH | 7/8/98  | 9:04:55 | 3.1 | 1  | 15.1 | 8.55 | 12.82 | 6.93 | 41.5 | 0.03 | 80.6  | 301 | 316 |
|                   |      |        |         |         |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 15 | 0.4  | 9.71 | 20.91 | 7.89 | 50.9 | 0.03 | 108.4 | 116 | 399 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 14 | 1.4  | 9.73 | 20.9  | 7.92 | 50.8 | 0.03 | 108.5 | 114 | 400 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 13 | 2.4  | 9.81 | 20.49 | 7.94 | 50.7 | 0.03 | 108.6 | 23  | 401 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 12 | 3.4  | 9.93 | 20.13 | 7.97 | 50.3 | 0.03 | 109.2 | 33  | 402 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 11 | 4.4  | 10.1 | 19.76 | 8.09 | 50.1 | 0.03 | 110.4 | 112 | 402 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 10 | 5.4  | 10.3 | 19.18 | 7.94 | 49.5 | 0.03 | 110.5 | 25  | 407 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 9  | 6.4  | 10.3 | 18.8  | 7.8  | 49   | 0.03 | 110.7 | 36  | 412 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 8  | 7.4  | 10.2 | 18.39 | 7.63 | 48.7 | 0.03 | 108.3 | 112 | 418 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 7  | 8.4  | 10.4 | 17.96 | 7.58 | 48.2 | 0.03 | 109.3 | 117 | 420 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 6  | 9.4  | 10.6 | 17.34 | 7.58 | 47.6 | 0.03 | 109.7 | 111 | 422 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 5  | 10.4 | 10.5 | 17.07 | 7.49 | 47.8 | 0.03 | 108.6 | 112 | 426 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 4  | 11.4 | 10.5 | 16.78 | 7.31 | 47.6 | 0.03 | 107.3 | 55  | 430 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 3  | 12.4 | 9.92 | 16.17 | 7.08 | 47   | 0.03 | 100.5 | 100 | 435 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 2  | 13.4 | 8.85 | 14.16 | 6.8  | 46.8 | 0.03 | 85.9  | 117 | 440 |
| Windy Bay Shallow | 2998 | ASBH   | 7/20/98 | 10:30   | 6   | 1  | 14.4 | 7.95 | 11.31 | 6.61 | 46.4 | 0.03 | 72.4  | 238 | 444 |
|                   |      |        |         |         |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 17 | 0.4  | 8.85 | 23.46 | 7.74 | 55.6 | 0.04 | 103.6 | 21  | 383 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 16 | 0.9  | 8.85 | 23.46 | 7.75 | 55.5 | 0.04 | 103.6 | 39  | 383 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 15 | 1.9  | 8.86 | 23.42 | 7.74 | 55.6 | 0.04 | 103.7 | 54  | 385 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 14 | 2.9  | 8.86 | 23.4  | 7.7  | 55.5 | 0.04 | 103.6 | 35  | 387 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 13 | 3.9  | 8.85 | 23.37 | 7.67 | 55.5 | 0.04 | 103.5 | 104 | 391 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 12 | 4.9  | 8.85 | 23.37 | 7.66 | 55.5 | 0.04 | 103.4 | 105 | 392 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 11 | 5.9  | 8.85 | 23.33 | 7.64 | 55.4 | 0.04 | 103.4 | 114 | 395 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 10 | 6.9  | 8.85 | 23.31 | 7.6  | 55.4 | 0.04 | 103.3 | 57  | 398 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 9  | 7.9  | 8.91 | 22.92 | 7.5  | 54.7 | 0.04 | 103.3 | 104 | 403 |
| Windy Bay Shallow | 3298 | DBAS   | 8/10/98 | 9:29:19 | 7.9 | 8  | 8.9  | 9.09 | 22.02 | 7.39 | 53.6 | 0.03 | 103.5 | 122 | 407 |

|                   |      |      |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-------------------|------|------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 7  | 9.9  | 9.28 | 20.27 | 7.23 | 50.9 | 0.03 | 102.2 | 122 | 414 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 6  | 10.9 | 9.69 | 16.98 | 7.06 | 48.1 | 0.03 | 99.9  | 101 | 420 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 5  | 11.9 | 9.83 | 14.71 | 6.95 | 47.1 | 0.03 | 96.5  | 102 | 424 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 4  | 12.9 | 8.9  | 13.37 | 6.78 | 47   | 0.03 | 84.8  | 115 | 430 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 3  | 13.9 | 8.24 | 12.62 | 6.65 | 47.2 | 0.03 | 77.2  | 101 | 434 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 2  | 14.9 | 8.24 | 11.49 | 6.6  | 46.9 | 0.03 | 75.2  | 59  | 435 |
| Windy Bay Shallow | 3298 | DBAS | 8/10/98 | 9:29:19  | 7.9 | 1  | 15.9 | 7.59 | 10.59 | 6.53 | 48   | 0.03 | 67.8  | 226 | 438 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 22 | 0.3  | 9.11 | 21.99 | 7.78 | 55.7 | 0.04 | 104   | 123 | 345 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 21 | 1.5  | 9.09 | 21.79 | 7.81 | 55.5 | 0.04 | 103.4 | 131 | 346 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 20 | 2.2  | 9.19 | 21.3  | 7.83 | 55.5 | 0.04 | 103.6 | 157 | 346 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 19 | 3    | 9.2  | 21.18 | 7.82 | 55.6 | 0.04 | 103.4 | 119 | 347 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 18 | 3.7  | 9.2  | 21.09 | 7.81 | 55.6 | 0.04 | 103.2 | 150 | 349 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 17 | 4.5  | 9.19 | 21.04 | 7.78 | 55.6 | 0.04 | 103.1 | 48  | 350 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 16 | 5.2  | 9.18 | 20.98 | 7.76 | 55.7 | 0.04 | 102.8 | 111 | 351 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 15 | 6    | 9.15 | 20.97 | 7.72 | 55.3 | 0.04 | 102.4 | 122 | 352 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 14 | 6.7  | 9.13 | 20.95 | 7.72 | 55.1 | 0.04 | 102.2 | 100 | 351 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 13 | 7.5  | 9.11 | 20.9  | 7.68 | 55.6 | 0.04 | 101.8 | 207 | 352 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 12 | 8.2  | 9.12 | 20.86 | 7.66 | 55.3 | 0.04 | 101.9 | 136 | 353 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 11 | 9    | 9.06 | 20.84 | 7.59 | 55.1 | 0.04 | 101.2 | 57  | 356 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 10 | 9.7  | 8.91 | 20.62 | 7.46 | 55.3 | 0.04 | 99.1  | 113 | 359 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 8  | 10.5 | 8.49 | 20.15 | 7.29 | 54.9 | 0.04 | 93.5  | 154 | 363 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 7  | 11.2 | 9.51 | 17.23 | 7.08 | 49.4 | 0.03 | 98.8  | 147 | 369 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 6  | 12   | 9.7  | 15.97 | 7.07 | 47.5 | 0.03 | 98.1  | 140 | 369 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 5  | 12.7 | 9.79 | 15.57 | 7.07 | 47.1 | 0.03 | 98.1  | 240 | 368 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 4  | 13.5 | 9.85 | 15.01 | 7.03 | 46.6 | 0.03 | 97.6  | 150 | 368 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 3  | 14.2 | 8.71 | 14.41 | 6.88 | 47.6 | 0.03 | 85.1  | 126 | 370 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 2  | 15   | 8.17 | 13.71 | 6.85 | 47.9 | 0.03 | 78.7  | 221 | 369 |
| Windy Bay Shallow | 3498 | DTAS | 8/24/98 | 12:44:32 | 9   | 1  | 15.7 | 8.39 | 13.46 | 6.92 | 46.9 | 0.03 | 80.3  | 116 | 365 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 17 | 0.3  | 9.07 | 21.51 | 7.79 | 55.7 | 0.04 | 102.9 | 28  | 336 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 16 | 1.2  | 9.07 | 21.48 | 7.79 | 55.7 | 0.04 | 102.8 | 50  | 336 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 15 | 2.2  | 9.07 | 21.39 | 7.83 | 55.6 | 0.04 | 102.7 | 48  | 334 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 14 | 3.2  | 9.08 | 21.37 | 7.78 | 55.5 | 0.04 | 102.7 | 113 | 337 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 13 | 4.2  | 9.08 | 21.35 | 7.8  | 55.6 | 0.04 | 102.7 | 104 | 336 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 12 | 5.2  | 9.09 | 21.33 | 7.76 | 55.6 | 0.04 | 102.8 | 103 | 339 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 11 | 6.2  | 9.08 | 21.33 | 7.75 | 55.6 | 0.04 | 102.7 | 119 | 339 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 10 | 7.2  | 9.1  | 21.3  | 7.7  | 55.4 | 0.04 | 102.8 | 121 | 343 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 9  | 8.2  | 9.1  | 21.29 | 7.64 | 55.7 | 0.04 | 102.8 | 117 | 346 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 8  | 9.2  | 9.07 | 21.17 | 7.49 | 55.5 | 0.04 | 102.2 | 108 | 353 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 7  | 10.2 | 8.53 | 19.93 | 7.17 | 53.5 | 0.03 | 93.8  | 114 | 363 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 6  | 11.2 | 9.17 | 17.05 | 6.92 | 46.9 | 0.03 | 95.1  | 100 | 371 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 5  | 12.2 | 8.61 | 14.96 | 6.81 | 46.5 | 0.03 | 85.4  | 114 | 373 |
| Windy Bay Shallow | 3598 | DTAS | 9/2/98  | 13:02:24 | 8.2 | 4  | 13.2 | 8.52 | 13.46 | 6.73 | 45.8 | 0.03 | 81.8  | 122 | 376 |

|                   |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
|-------------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| Windy Bay Shallow | 3598 | DTAS   | 9/2/98   | 13:02:24 | 8.2 | 3  | 14.2 | 7.79 | 12.25 | 6.62 | 45.9 | 0.03 | 72.8 | 141 | 379 |
| Windy Bay Shallow | 3598 | DTAS   | 9/2/98   | 13:02:24 | 8.2 | 2  | 15.2 | 7.39 | 11.21 | 6.55 | 46.3 | 0.03 | 67.4 | 130 | 382 |
| Windy Bay Shallow | 3598 | DTAS   | 9/2/98   | 13:02:24 | 8.2 | 1  | 16.2 | 6.13 | 11.11 | 6.49 | 46.5 | 0.03 | 55.8 | 856 | 383 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 16 | 0.2  | 9.32 | 18    | 7.7  | 54.9 | 0.04 | 98.2 | 151 | 348 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 15 | 1.1  | 9.27 | 18.1  | 7.75 | 57.7 | 0.04 | 98   | 49  | 349 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 14 | 2.1  | 9.27 | 18.03 | 7.73 | 57.8 | 0.04 | 97.8 | 47  | 350 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 13 | 3.1  | 9.26 | 17.98 | 7.69 | 58.1 | 0.04 | 97.7 | 52  | 353 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 12 | 4.1  | 9.28 | 17.95 | 7.69 | 58.1 | 0.04 | 97.8 | 35  | 352 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 11 | 5.1  | 9.26 | 17.95 | 7.67 | 58.1 | 0.04 | 97.6 | 53  | 354 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 10 | 6.1  | 9.26 | 17.95 | 7.66 | 57.7 | 0.04 | 97.5 | 33  | 354 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 9  | 7.1  | 9.26 | 17.93 | 7.66 | 57.8 | 0.04 | 97.5 | 53  | 354 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 8  | 8.1  | 9.26 | 17.93 | 7.65 | 58.2 | 0.04 | 97.5 | 103 | 354 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 7  | 9.1  | 9.25 | 17.93 | 7.63 | 57.7 | 0.04 | 97.4 | 38  | 356 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 6  | 10.1 | 9.24 | 17.91 | 7.63 | 57.6 | 0.04 | 97.3 | 50  | 356 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 5  | 11.1 | 9.23 | 17.91 | 7.61 | 57.9 | 0.04 | 97.2 | 129 | 356 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 4  | 12.1 | 9.23 | 17.9  | 7.57 | 57.9 | 0.04 | 97.1 | 114 | 358 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 3  | 13.1 | 9.18 | 17.86 | 7.5  | 58.2 | 0.04 | 96.5 | 108 | 359 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 2  | 14.1 | 9.15 | 17.74 | 7.42 | 58.8 | 0.04 | 96   | 102 | 363 |
| Windy Bay Shallow | 3998 | ASKB   | 9/30/98  | 11:21:17 | 8.5 | 1  | 15.1 | 8.18 | 16.38 | 7.19 | 56.7 | 0.04 | 83.4 | 17  | 368 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 16 | 0.4  | 9.33 | 13.06 | 7.41 | 53.4 | 0.03 | 87.5 | 29  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 15 | 1    | 9.31 | 13.04 | 7.4  | 53.4 | 0.03 | 87.4 | 36  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 14 | 2    | 9.34 | 13.02 | 7.44 | 53.4 | 0.03 | 87.6 | 39  | 346 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 13 | 3    | 9.36 | 13.02 | 7.41 | 53.3 | 0.03 | 87.8 | 39  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 12 | 4    | 9.33 | 13.01 | 7.4  | 53.4 | 0.03 | 87.5 | 26  | 349 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 11 | 5    | 9.36 | 12.99 | 7.42 | 53.4 | 0.03 | 87.7 | 21  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 10 | 6    | 9.34 | 12.99 | 7.41 | 53.2 | 0.03 | 87.5 | 51  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 9  | 7    | 9.35 | 12.97 | 7.4  | 53.2 | 0.03 | 87.5 | 48  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 8  | 8    | 9.34 | 12.96 | 7.39 | 53.2 | 0.03 | 87.5 | 34  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 7  | 9    | 9.34 | 12.96 | 7.39 | 53.1 | 0.03 | 87.5 | 38  | 348 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 6  | 10   | 9.34 | 12.94 | 7.4  | 53.5 | 0.03 | 87.4 | 34  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 5  | 11   | 9.35 | 12.94 | 7.39 | 53.2 | 0.03 | 87.5 | 37  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 4  | 12   | 9.34 | 12.94 | 7.39 | 53.5 | 0.03 | 87.4 | 44  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 3  | 13   | 9.35 | 12.94 | 7.38 | 53.1 | 0.03 | 87.5 | 39  | 347 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 2  | 14   | 9.34 | 12.94 | 7.38 | 53.4 | 0.03 | 87.4 | 114 | 346 |
| Windy Bay Shallow | 4298 | DTASKB | 10/20/98 | 10:40:14 | 7.1 | 1  | 15   | 9.33 | 12.92 | 7.37 | 53.5 | 0.03 | 87.3 | 311 | 351 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 15 | 0.5  | 8.61 | 10.05 | 7.31 | 53.3 | 0.03 | 75.9 | 52  | 375 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 14 | 1.5  | 8.6  | 10.07 | 7.33 | 53.4 | 0.03 | 75.8 | 42  | 375 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 13 | 2.5  | 8.6  | 10.07 | 7.32 | 53.3 | 0.03 | 75.8 | 37  | 375 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 12 | 3.5  | 8.59 | 10.07 | 7.3  | 53.2 | 0.03 | 75.7 | 111 | 376 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 11 | 4.5  | 8.6  | 10.07 | 7.31 | 53.2 | 0.03 | 75.8 | 46  | 375 |
| Windy Bay Shallow | 4598 | KB AS  | 11/12/98 | 11:22:39 | 8   | 10 | 5.5  | 8.59 | 10.07 | 7.3  | 53.2 | 0.03 | 75.7 | 50  | 376 |

|                   |      |       |          |          |   |   |      |      |       |      |      |      |      |      |     |
|-------------------|------|-------|----------|----------|---|---|------|------|-------|------|------|------|------|------|-----|
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 9 | 6.5  | 8.59 | 10.07 | 7.29 | 53.1 | 0.03 | 75.7 | 49   | 376 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 8 | 7.6  | 8.59 | 10.07 | 7.27 | 53   | 0.03 | 75.7 | 41   | 378 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 7 | 8.5  | 8.59 | 10.05 | 7.29 | 53.1 | 0.03 | 75.7 | 36   | 377 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 6 | 9.5  | 8.58 | 10.04 | 7.3  | 53.3 | 0.03 | 75.6 | 36   | 376 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 5 | 10.5 | 8.58 | 10.02 | 7.29 | 53.5 | 0.03 | 75.5 | 36   | 377 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 4 | 11.5 | 8.58 | 10.02 | 7.29 | 53.1 | 0.03 | 75.5 | 37   | 377 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 3 | 12.5 | 8.57 | 10.02 | 7.3  | 53.3 | 0.03 | 75.4 | 105  | 377 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 2 | 13.4 | 8.56 | 9.99  | 7.29 | 53.2 | 0.03 | 75.3 | 35   | 377 |
| Windy Bay Shallow | 4598 | KB AS | 11/12/98 | 11:22:39 | 8 | 1 | 14.5 | 8.55 | 9.73  | 7.31 | 53.8 | 0.03 | 74.7 | 1114 | 377 |

| Location       | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|----------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 16       | 0.3       | 13.02                   | 3.89            | 7    | 52.1                 | 0.03 | 98.2                       | 55                 | 394   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 15       | 1.2       | 13.11                   | 3.74            | 7.05 | 52                   | 0.03 | 98.5                       | 215                | 396   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 14       | 3.2       | 13.19                   | 3.73            | 7.04 | 52                   | 0.03 | 99.1                       | 131                | 398   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 13       | 5.2       | 13.18                   | 3.73            | 7.02 | 52.1                 | 0.03 | 99                         | 213                | 398   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 12       | 7.2       | 13.12                   | 3.69            | 7.01 | 52.2                 | 0.03 | 98.5                       | 139                | 398   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 11       | 9.2       | 13.14                   | 3.69            | 7.03 | 52.1                 | 0.03 | 98.6                       | 147                | 395   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 10       | 11.2      | 13.12                   | 3.74            | 7.02 | 52.2                 | 0.03 | 98.7                       | 205                | 395   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 9        | 13.2      | 13.14                   | 3.69            | 6.98 | 52.2                 | 0.03 | 98.6                       | 335                | 397   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 8        | 15.2      | 13.13                   | 3.71            | 6.99 | 52.1                 | 0.03 | 98.6                       | 224                | 394   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 7        | 17.2      | 13.16                   | 3.71            | 7    | 52.1                 | 0.03 | 98.8                       | 133                | 392   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 6        | 19.2      | 13.12                   | 3.69            | 6.95 | 52.1                 | 0.03 | 98.5                       | 129                | 393   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 5        | 21.2      | 13.1                    | 3.73            | 6.93 | 52.1                 | 0.03 | 98.5                       | 136                | 393   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 4        | 23.2      | 13.1                    | 3.78            | 6.93 | 52.2                 | 0.03 | 98.6                       | 224                | 391   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 3        | 25.2      | 13.15                   | 3.79            | 6.92 | 52.1                 | 0.03 | 99                         | 142                | 389   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 2        | 27.2      | 13.17                   | 3.86            | 6.91 | 52.1                 | 0.03 | 99.3                       | 139                | 388   |
| Windy Bay Deep | 1098  | DBRP     | 3/13/98 | 11:09:06 | 2.5        | 1        | 29.2      | 13.29                   | 3.83            | 6.87 | 52.3                 | 0.03 | 100.1                      | 137                | 388   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 15       | 0.5       | 12.87                   | 5.77            | 7.06 | 51.5                 | 0.03 | 102.9                      | 201                | 421   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 14       | 2.5       | 12.92                   | 5.64            | 7.01 | 51.5                 | 0.03 | 102.9                      | 246                | 425   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 13       | 4.6       | 12.82                   | 5.49            | 7.05 | 51.6                 | 0.03 | 101.7                      | 200                | 423   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 12       | 6.5       | 12.76                   | 5.36            | 7.08 | 51.6                 | 0.03 | 100.9                      | 102                | 420   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 11       | 8.6       | 12.68                   | 5.29            | 7.03 | 51.5                 | 0.03 | 100                        | 103                | 422   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 10       | 10.5      | 12.68                   | 5.08            | 6.99 | 51.6                 | 0.03 | 99.5                       | 115                | 423   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 9        | 12.5      | 12.63                   | 5.04            | 7    | 51.7                 | 0.03 | 99.1                       | 217                | 422   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 8        | 14.6      | 12.67                   | 5.01            | 7.02 | 52                   | 0.03 | 99.3                       | 206                | 420   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 7        | 16.6      | 12.65                   | 4.91            | 6.97 | 52.1                 | 0.03 | 98.8                       | 100                | 421   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 6        | 18.5      | 12.65                   | 4.76            | 6.96 | 52.1                 | 0.03 | 98.5                       | 117                | 421   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 5        | 20.6      | 12.64                   | 4.76            | 6.99 | 52.1                 | 0.03 | 98.4                       | 133                | 418   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 4        | 22.5      | 12.64                   | 4.71            | 6.99 | 52.4                 | 0.03 | 98.3                       | 116                | 417   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 3        | 24.5      | 12.6                    | 4.71            | 6.96 | 52.3                 | 0.03 | 98                         | 110                | 417   |
| Windy Bay Deep | 1498  | DBAS     | 4/8/98  | 11:26:34 | 3          | 2        | 26.5      | 12.46                   | 4.69            | 6.96 | 52.8                 | 0.03 | 96.8                       | 141                | 415   |

|                |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|----------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Windy Bay Deep | 1498 | DBAS   | 4/8/98  | 11:26:34 | 3   | 1  | 28.5 | 12.46 | 4.69  | 6.96 | 53   | 0.03 | 96.8  | 217 | 413 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 15 | 0.3  | 12.4  | 6.79  | 6.97 | 48.5 | 0.03 | 103.1 | 43  | 444 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 14 | 1.4  | 12.3  | 6.59  | 7.03 | 48.4 | 0.03 | 101.8 | 58  | 441 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 13 | 3.4  | 12.22 | 6.81  | 6.99 | 48.4 | 0.03 | 101.7 | 100 | 443 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 12 | 5.5  | 12.27 | 6.63  | 6.98 | 48.4 | 0.03 | 101.6 | 49  | 444 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 11 | 7.7  | 12.3  | 6.56  | 6.95 | 48.5 | 0.03 | 101.7 | 125 | 445 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 10 | 9.6  | 12.27 | 6.33  | 6.95 | 48.4 | 0.03 | 100.8 | 138 | 444 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 9  | 11.8 | 12.21 | 6.13  | 6.92 | 48.5 | 0.03 | 99.9  | 150 | 445 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 8  | 13.8 | 12.12 | 5.85  | 6.89 | 48.9 | 0.03 | 98.5  | 122 | 446 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 7  | 15.8 | 11.96 | 5.34  | 6.8  | 50   | 0.03 | 95.9  | 110 | 449 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 6  | 17.7 | 11.96 | 5.17  | 6.8  | 50.1 | 0.03 | 95.4  | 107 | 449 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 5  | 19.6 | 11.98 | 4.88  | 6.8  | 50.1 | 0.03 | 94.9  | 133 | 449 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 4  | 21.6 | 11.98 | 4.79  | 6.77 | 50.2 | 0.03 | 94.6  | 139 | 450 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 3  | 23.7 | 11.95 | 4.78  | 6.77 | 50.1 | 0.03 | 94.4  | 240 | 450 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 2  | 25.7 | 12.02 | 4.78  | 6.75 | 50.2 | 0.03 | 94.9  | 109 | 450 |
| Windy Bay Deep | 1698 | DBASRP | 4/23/98 | 8:48:30  | 3.1 | 1  | 27.7 | 12.02 | 4.78  | 6.71 | 50.1 | 0.03 | 94.9  | 346 | 451 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 16 | 0.3  | 11.7  | 10.53 | 7.21 | 42   | 0.03 | 105.9 | 53  | 412 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 15 | 2.3  | 11.67 | 10.48 | 7.2  | 42.1 | 0.03 | 105.6 | 105 | 411 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 14 | 4.3  | 11.64 | 10.43 | 7.17 | 42.1 | 0.03 | 105.2 | 133 | 412 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 13 | 6.3  | 11.62 | 10.35 | 7.16 | 42.1 | 0.03 | 104.7 | 201 | 411 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 12 | 8.3  | 11.41 | 9.76  | 7.1  | 43   | 0.03 | 101.4 | 9   | 410 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 11 | 10.3 | 11.44 | 9.4   | 7.03 | 43.5 | 0.03 | 100.8 | 153 | 411 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 10 | 12.3 | 11.4  | 8.68  | 6.9  | 44.7 | 0.03 | 98.7  | 227 | 416 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 9  | 14.3 | 11.4  | 8.61  | 6.92 | 44.8 | 0.03 | 98.6  | 244 | 413 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 8  | 16.3 | 11.12 |       | 6.85 | 45.9 | 0.03 | 94.6  | 153 | 412 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 7  | 18.3 | 11.35 | 7.14  | 6.79 | 47.3 | 0.03 | 94.6  | 137 | 415 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 6  | 20.3 | 11.28 | 6.94  | 6.82 | 47.7 | 0.03 | 93.6  | 124 | 411 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 5  | 22.3 | 11.29 | 6.69  | 6.81 | 48   | 0.03 | 93.1  | 119 | 411 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 4  | 24.3 | 11.19 | 6.61  | 6.72 | 48.1 | 0.03 | 92.1  | 159 | 414 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 3  | 26.3 | 11.05 | 6.23  | 6.7  | 48.8 | 0.03 | 90.1  | 118 | 412 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 2  | 28.3 | 11.05 | 6.15  | 6.65 | 49   | 0.03 | 89.9  | 427 | 414 |
| Windy Bay Deep | 2198 | JLDTAS | 5/26/98 | 11:37    | 3.4 | 1  | 30.3 | 11.12 | 6.18  | 6.65 | 48.9 | 0.03 | 90.6  | 227 | 411 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 11 | 0.4  | 10.91 | 17.52 | 7.78 | 43.9 | 0.03 | 113.5 | 52  | 399 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 10 | 2.6  | 11.07 | 16.07 | 7.73 | 43.4 | 0.03 | 112.4 | 54  | 403 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 9  | 5.6  | 11.06 | 15.49 | 7.66 | 43.5 | 0.03 | 110.9 | 100 | 406 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 8  | 8.6  | 11.12 | 15.28 | 7.62 | 43.5 | 0.03 | 111.1 | 105 | 407 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 7  | 11.6 | 11.55 | 13.87 | 7.62 | 43.1 | 0.03 | 111.8 | 128 | 408 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 6  | 14.6 | 11.15 | 12.04 | 7.28 | 43.4 | 0.03 | 103.6 | 203 | 416 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 5  | 17.6 | 10.83 | 9.74  | 7.1  | 44.3 | 0.03 | 95.3  | 204 | 420 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 4  | 20.6 | 10.7  | 8.61  | 7.04 | 45   | 0.03 | 91.6  | 151 | 421 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 3  | 23.6 | 10.64 | 8.22  | 7.02 | 46.6 | 0.03 | 90.3  | 117 | 421 |

|                |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|----------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 2  | 26.6 | 10.59 | 7.63  | 6.99 | 48.2 | 0.03 | 88.5  | 115 | 421 |
| Windy Bay Deep | 2398 | DTJLAS | 6/8/98  | 11:07:23 | 4.2 | 1  | 29.6 | 10.11 | 8.1   | 7    | 47.8 | 0.03 | 85.6  | 216 | 432 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 12 | 0.5  | 10.5  | 18.17 | 8.02 | 43.4 | 0.03 | 112   | 126 | 391 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 11 | 1.5  | 10.55 | 18.05 | 7.99 | 43.3 | 0.03 | 112.2 | 115 | 395 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 10 | 4.5  | 11.1  | 15.86 | 7.88 | 42.4 | 0.03 | 112.8 | 124 | 400 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 9  | 7.5  | 10.78 | 14.43 | 7.46 | 41.8 | 0.03 | 106.2 | 212 | 409 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 8  | 10.5 | 10.49 | 12.94 | 7.21 | 41.3 | 0.03 | 100   | 143 | 415 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 7  | 13.5 | 10.22 | 10.81 | 7.04 | 41.1 | 0.03 | 92.7  | 244 | 418 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 6  | 16.5 | 9.88  | 8.37  | 6.9  | 43.5 | 0.03 | 84.6  | 212 | 421 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 5  | 19.5 | 10.12 | 7.43  | 6.89 | 45.7 | 0.03 | 84.7  | 53  | 422 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 4  | 22.5 | 10.18 | 7.15  | 6.89 | 46.2 | 0.03 | 84.6  | 46  | 422 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 3  | 25.5 | 10.25 | 6.89  | 6.9  | 46.7 | 0.03 | 84.6  | 206 | 422 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 2  | 28.5 | 10.24 | 6.71  | 6.9  | 47.2 | 0.03 | 84.1  | 207 | 421 |
| Windy Bay Deep | 2598 | DTASBH | 6/24/98 | 11:43:39 | 4.5 | 1  | 31.5 | 9.98  | 6.71  | 6.91 | 47.4 | 0.03 | 82    | 455 | 421 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 16 | 0.5  | 9.56  | 21.94 | 7.83 | 47.3 | 0.03 | 109   | 105 | 330 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 15 | 2.2  | 9.66  | 21.06 | 7.91 | 47.1 | 0.03 | 108.3 | 142 | 326 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 14 | 4.2  | 10.07 | 20.47 | 7.97 | 46.3 | 0.03 | 111.6 | 216 | 326 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 13 | 6.2  | 10.77 | 18.6  | 8.23 | 44.8 | 0.03 | 115   | 205 | 323 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 12 | 8.2  | 10.98 | 17.36 | 8.23 | 44.2 | 0.03 | 114.2 | 138 | 325 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 11 | 10.2 | 10.37 | 15.91 | 7.57 | 43.2 | 0.03 | 104.6 | 225 | 338 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 10 | 12.2 | 9.1   | 13.02 | 6.97 | 41.7 | 0.03 | 86.2  | 155 | 348 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 9  | 14.2 | 9.06  | 10.18 | 6.83 | 41.1 | 0.03 | 80.4  | 135 | 349 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 8  | 16.2 | 9.07  | 9.32  | 6.81 | 41.8 | 0.03 | 78.8  | 142 | 348 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 7  | 18.2 | 9     | 8.47  | 6.78 | 43.2 | 0.03 | 76.6  | 133 | 347 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 6  | 20.2 | 9.06  | 8.25  | 6.78 | 43.5 | 0.03 | 76.8  | 208 | 345 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 5  | 22.2 | 9.35  | 7.7   | 6.79 | 44.7 | 0.03 | 78.1  | 131 | 343 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 4  | 24.2 | 9.67  | 6.96  | 6.78 | 46   | 0.03 | 79.4  | 104 | 342 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 3  | 26.2 | 9.38  | 6.66  | 6.76 | 47   | 0.03 | 76.4  | 128 | 340 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 2  | 28.2 | 9.36  | 6.48  | 6.76 | 47.4 | 0.03 | 75.9  | 144 | 339 |
| Windy Bay Deep | 2798 | ASDNBH | 7/8/98  | 9:49:41  | 4.5 | 1  | 30.2 | 9.38  | 6.48  | 6.77 | 47.4 | 0.03 | 76.1  | 416 | 336 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 17 | 0.4  | 9.39  | 21.93 | 7.86 | 51.8 | 0.03 | 106.9 | 50  | 372 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 16 | 1    | 9.45  | 21.62 | 7.88 | 51.5 | 0.03 | 107   | 110 | 373 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 15 | 3    | 9.75  | 20.66 | 7.9  | 50.2 | 0.03 | 108.4 | 107 | 373 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 14 | 5    | 9.71  | 20.05 | 7.89 | 49.6 | 0.03 | 106.6 | 113 | 375 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 13 | 7    | 10.12 | 19.22 | 7.77 | 49.2 | 0.03 | 109.3 | 101 | 378 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 12 | 9    | 10.44 | 16.9  | 7.54 | 47.6 | 0.03 | 107.5 | 112 | 384 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 11 | 11   | 10.19 | 15.36 | 7.23 | 46.2 | 0.03 | 101.5 | 103 | 390 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 10 | 13   | 9.24  | 13.45 | 6.94 | 45.5 | 0.03 | 88.3  | 221 | 394 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 9  | 15   | 8.53  | 10.76 | 6.76 | 45.2 | 0.03 | 76.6  | 50  | 397 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 8  | 17   | 8.54  | 9.56  | 6.71 | 46   | 0.03 | 74.6  | 55  | 397 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 7  | 19   | 8.85  | 9.06  | 6.72 | 46.4 | 0.03 | 76.3  | 104 | 396 |

|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|----------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 6  | 21   | 8.48 | 8.81  | 6.72 | 47.2 | 0.03 | 72.7  | 117 | 395 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 5  | 23   | 8.89 | 8.22  | 6.75 | 47.9 | 0.03 | 75.2  | 105 | 394 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 4  | 25   | 9.29 | 7.43  | 6.77 | 49.7 | 0.03 | 77    | 51  | 394 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 3  | 27   | 9.27 | 6.9   | 6.77 | 50.7 | 0.03 | 75.8  | 122 | 393 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 2  | 29   | 9.35 | 6.67  | 6.79 | 51   | 0.03 | 76    | 204 | 392 |
| Windy Bay Deep | 2998 | ASBH   | 7/20/98 | 10:06:45 | 8.2 | 1  | 31   | 8.76 | 7.22  | 6.83 | 51.3 | 0.03 | 72.2  | 315 | 393 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 16 | 0.4  | 8.78 | 23.55 | 7.81 | 57.5 | 0.04 | 103   | 42  | 362 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 15 | 1.8  | 8.8  | 23.5  | 7.8  | 57.5 | 0.04 | 103.1 | 46  | 364 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 14 | 3.9  | 8.86 | 23.27 | 7.81 | 58.2 | 0.04 | 103.4 | 121 | 365 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 13 | 5.8  | 8.89 | 23.17 | 7.76 | 57.9 | 0.04 | 103.5 | 116 | 367 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 12 | 7.8  | 8.87 | 23.12 | 7.67 | 57.7 | 0.04 | 103.1 | 128 | 371 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 11 | 9.9  | 9.98 | 19.39 | 7.36 | 49.6 | 0.03 | 108   | 108 | 379 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 10 | 11.8 | 9.98 | 14.05 | 6.96 | 46.8 | 0.03 | 96.6  | 100 | 389 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 9  | 13.8 | 9.01 | 12.02 | 6.8  | 46.3 | 0.03 | 83.2  | 128 | 391 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 8  | 15.8 | 8.34 | 10.49 | 6.7  | 46.6 | 0.03 | 74.4  | 110 | 393 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 7  | 17.8 | 7.77 | 9.1   | 6.63 | 48.1 | 0.03 | 67    | 104 | 394 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 6  | 19.8 | 8.22 | 8.17  | 6.65 | 49.3 | 0.03 | 69.4  | 114 | 393 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 5  | 21.8 | 8.76 | 7.61  | 6.69 | 49.8 | 0.03 | 72.9  | 117 | 391 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 4  | 23.8 | 8.89 | 7.07  | 6.69 | 51.4 | 0.03 | 73    | 123 | 391 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 3  | 25.8 | 8.69 | 6.79  | 6.69 | 51.8 | 0.03 | 70.8  | 125 | 389 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 2  | 27.8 | 8.21 | 6.64  | 6.68 | 52.4 | 0.03 | 66.7  | 101 | 387 |
| Windy Bay Deep | 3298 | DBAS   | 8/10/98 | 9:53:02  | 8.8 | 1  | 29.8 | 7.72 | 7.04  | 6.7  | 53.6 | 0.03 | 63.3  | 252 | 385 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 16 | 0.9  | 9.13 | 21.8  | 7.8  | 53.4 | 0.03 | 104.1 | 116 | 355 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 15 | 2.9  | 9.21 | 21.41 | 7.8  | 53.6 | 0.03 | 104.2 | 55  | 359 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 14 | 4.9  | 9.18 | 21.38 | 7.76 | 53.6 | 0.03 | 103.8 | 46  | 360 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 13 | 6.9  | 9.19 | 21.31 | 7.69 | 53.4 | 0.03 | 103.7 | 48  | 364 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 12 | 8.9  | 9.18 | 21.22 | 7.59 | 52.8 | 0.03 | 103.4 | 37  | 368 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 11 | 10.9 | 9.04 | 20.76 | 7.45 | 52.7 | 0.03 | 101   | 200 | 373 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 10 | 12.9 | 9.49 | 15.33 | 6.93 | 45.1 | 0.03 | 94.8  | 142 | 387 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 9  | 14.9 | 8.71 | 12.04 | 6.74 | 44.2 | 0.03 | 80.9  | 126 | 392 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 8  | 16.9 | 7.48 | 9.99  | 6.59 | 45.9 | 0.03 | 66.3  | 113 | 395 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 7  | 18.9 | 7.5  | 9.15  | 6.57 | 45.8 | 0.03 | 65.1  | 110 | 396 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 6  | 20.9 | 7.64 | 8.42  | 6.57 | 46.6 | 0.03 | 65.2  | 126 | 396 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 5  | 22.9 | 8.13 | 7.97  | 6.58 | 46.9 | 0.03 | 68.6  | 228 | 395 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 4  | 24.9 | 8.54 | 7.28  | 6.59 | 48.1 | 0.03 | 70.8  | 301 | 395 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 3  | 26.9 | 8.47 | 7.02  | 6.57 | 48.5 | 0.03 | 69.8  | 226 | 395 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 2  | 28.9 | 8.27 | 6.71  | 6.54 | 49.4 | 0.03 | 67.6  | 142 | 396 |
| Windy Bay Deep | 3498 | ASDNGL | 8/24/98 | 13:50:30 | 9   | 1  | 30.9 | 7.93 | 7.15  | 6.54 | 48.7 | 0.03 | 65.5  | 820 | 401 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98  | 10:03:06 | 8.7 | 16 | 0.3  | 9.07 | 21.96 | 7.89 | 56.1 | 0.04 | 103.7 | 31  | 336 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98  | 10:03:06 | 8.7 | 15 | 2.2  | 9.05 | 21.55 | 7.79 | 55.8 | 0.04 | 102.7 | 44  | 342 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98  | 10:03:06 | 8.7 | 14 | 4.2  | 9.08 | 21.41 | 7.79 | 55.9 | 0.04 | 102.8 | 51  | 343 |

|                |      |        |          |          |     |    |      |      |       |      |      |      |       |     |     |
|----------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 13 | 6.2  | 9.08 | 21.36 | 7.74 | 55.9 | 0.04 | 102.7 | 54  | 347 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 12 | 8.2  | 9.04 | 21.25 | 7.6  | 55.8 | 0.04 | 102   | 110 | 353 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 11 | 10.2 | 8.76 | 20.59 | 7.22 | 55   | 0.04 | 97.6  | 54  | 366 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 10 | 12.2 | 9.24 | 13.72 | 6.81 | 46   | 0.03 | 89.2  | 115 | 376 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 9  | 14.2 | 7.86 | 11.48 | 6.65 | 45.5 | 0.03 | 72.1  | 111 | 380 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 8  | 16.2 | 7.26 | 9.94  | 6.56 | 46.6 | 0.03 | 64.2  | 110 | 382 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 7  | 18.2 | 7.4  | 8.96  | 6.57 | 46.9 | 0.03 | 64    | 122 | 381 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 6  | 20.2 | 7.67 | 8.13  | 6.56 | 48.2 | 0.03 | 65    | 109 | 381 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 5  | 22.2 | 8.12 | 7.66  | 6.58 | 49.1 | 0.03 | 68    | 103 | 381 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 4  | 24.2 | 8.23 | 7.35  | 6.57 | 49.3 | 0.03 | 68.4  | 53  | 381 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 3  | 26.2 | 8.31 | 7.05  | 6.57 | 50   | 0.03 | 68.6  | 158 | 381 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 2  | 28.2 | 7.92 | 6.76  | 6.54 | 51   | 0.03 | 64.9  | 122 | 382 |
| Windy Bay Deep | 3598 | DTAS   | 9/2/98   | 10:03:06 | 8.7 | 1  | 30.2 | 7.39 | 7.08  | 6.54 | 50.5 | 0.03 | 61    | 606 | 382 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 16 | 0.5  | 9.27 | 18.28 | 7.7  | 57.5 | 0.04 | 98.3  | 111 | 347 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 15 | 2.9  | 9.29 | 18    | 7.69 | 57.3 | 0.04 | 98    | 121 | 348 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 14 | 4.9  | 9.29 | 17.97 | 7.66 | 57.5 | 0.04 | 98    | 130 | 351 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 13 | 6.9  | 9.27 | 17.95 | 7.63 | 57.8 | 0.04 | 97.7  | 109 | 352 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 12 | 8.9  | 9.27 | 17.88 | 7.58 | 57.8 | 0.04 | 97.5  | 121 | 354 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 11 | 10.9 | 9.25 | 17.87 | 7.52 | 58.2 | 0.04 | 97.3  | 152 | 356 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 10 | 12.9 | 9.12 | 17.85 | 7.33 | 57.6 | 0.04 | 95.9  | 140 | 363 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 9  | 14.9 | 6.72 | 12.6  | 6.64 | 47.7 | 0.03 | 63.1  | 122 | 377 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 8  | 16.9 | 6.29 | 9.92  | 6.6  | 48.5 | 0.03 | 55.5  | 141 | 377 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 7  | 18.9 | 6.58 | 9.01  | 6.59 | 48.8 | 0.03 | 56.8  | 157 | 377 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 6  | 20.9 | 7.02 | 8.48  | 6.62 | 49.2 | 0.03 | 59.8  | 153 | 375 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 5  | 22.9 | 7.6  | 7.88  | 6.64 | 49.7 | 0.03 | 63.8  | 142 | 374 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 4  | 24.9 | 7.87 | 7.63  | 6.66 | 49.8 | 0.03 | 65.6  | 143 | 373 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 3  | 26.9 | 7.92 | 6.92  | 6.65 | 51   | 0.03 | 64.9  | 135 | 373 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 2  | 28.8 | 7.26 | 6.73  | 6.63 | 51.8 | 0.03 | 59.2  | 124 | 372 |
| Windy Bay Deep | 3998 | ASKB   | 9/30/98  | 12:07:00 | 10  | 1  | 30.9 | 6.89 | 6.76  | 6.63 | 52   | 0.03 | 56.3  | 549 | 374 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 16 | 0.4  | 9.2  | 13.14 | 7.36 | 53.3 | 0.03 | 86.5  | 41  | 353 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 15 | 2.3  | 9.18 | 13.12 | 7.36 | 53.3 | 0.03 | 86.3  | 31  | 354 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 14 | 4.3  | 9.18 | 13.06 | 7.33 | 53.1 | 0.03 | 86.2  | 30  | 356 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 13 | 6.3  | 9.18 | 13.06 | 7.31 | 53.3 | 0.03 | 86.2  | 35  | 357 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 12 | 8.3  | 9.19 | 13.04 | 7.31 | 53.1 | 0.03 | 86.2  | 34  | 357 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 11 | 10.3 | 9.17 | 13.02 | 7.3  | 53.3 | 0.03 | 86    | 46  | 358 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 10 | 12.3 | 9.17 | 13.02 | 7.26 | 52.9 | 0.03 | 86    | 30  | 359 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 9  | 14.3 | 9.15 | 13.02 | 7.24 | 52.9 | 0.03 | 85.8  | 32  | 360 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 8  | 16.3 | 9.14 | 12.99 | 7.2  | 53.1 | 0.03 | 85.6  | 34  | 361 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 7  | 18.3 | 9.03 | 13.01 | 7.1  | 52.9 | 0.03 | 84.6  | 54  | 364 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 6  | 20.3 | 6.75 | 11.54 | 6.67 | 49.1 | 0.03 | 61.2  | 40  | 370 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 5  | 22.3 | 6.13 | 9.69  | 6.6  | 47.2 | 0.03 | 53.2  | 113 | 370 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 4  | 24.3 | 6.89 | 8.2   | 6.62 | 46.8 | 0.03 | 57.7  | 53  | 369 |



|                |      |        |          |          |     |    |      |      |       |      |      |      |      |      |     |
|----------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|------|-----|
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 3  | 26.3 | 7.02 | 7.63  | 6.62 | 47.3 | 0.03 | 57.9 | 53   | 369 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 2  | 28.3 | 7.14 | 7.24  | 6.63 | 47.7 | 0.03 | 58.3 | 50   | 367 |
| Windy Bay Deep | 4298 | DTASKB | 10/20/98 | 11:14:13 | 9   | 1  | 30.3 | 7.05 | 7.04  | 6.64 | 48.2 | 0.03 | 57.4 | 113  | 367 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 16 | 0.4  | 8.51 | 10.2  | 7.28 | 53.2 | 0.03 | 75.2 | 37   | 377 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 15 | 1.4  | 8.5  | 10.2  | 7.28 | 53.1 | 0.03 | 75.1 | 53   | 377 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 14 | 3.5  | 8.47 | 10.22 | 7.23 | 53.3 | 0.03 | 74.9 | 108  | 379 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 13 | 5.5  | 8.49 | 10.2  | 7.23 | 52.9 | 0.03 | 75   | 114  | 379 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 12 | 7.5  | 8.47 | 10.22 | 7.18 | 52.9 | 0.03 | 74.9 | 104  | 381 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 11 | 9.5  | 8.46 | 10.22 | 7.14 | 53.2 | 0.03 | 74.8 | 46   | 382 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 10 | 11.5 | 8.46 | 10.22 | 7.13 | 53.1 | 0.03 | 74.8 | 50   | 382 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 9  | 13.5 | 8.45 | 10.23 | 7.08 | 53.1 | 0.03 | 74.8 | 104  | 383 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 8  | 15.5 | 8.4  | 10.2  | 6.98 | 53.8 | 0.03 | 74.3 | 145  | 385 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 7  | 17.4 | 6.04 | 8.55  | 6.6  | 50.9 | 0.03 | 51.3 | 147  | 391 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 6  | 19.5 | 6.07 | 7.4   | 6.56 | 48.4 | 0.03 | 50.1 | 115  | 391 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 5  | 21.5 | 5.94 | 7.04  | 6.55 | 48.4 | 0.03 | 48.6 | 104  | 392 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 4  | 23.5 | 5.71 | 6.99  | 6.54 | 49   | 0.03 | 46.7 | 51   | 391 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 3  | 25.5 | 5.68 | 6.97  | 6.55 | 49   | 0.03 | 46.4 | 32   | 391 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 2  | 27.7 | 5.74 | 6.92  | 6.55 | 48.9 | 0.03 | 46.9 | 112  | 391 |
| Windy Bay Deep | 4598 | KB AS  | 11/12/98 | 11:44:00 | 8.6 | 1  | 29.5 | 5.69 | 6.92  | 6.56 | 48.8 | 0.03 | 46.5 | 1828 | 390 |

| Location  | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-----------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 18       | 0.4       | 13.07                   | 4.76            | 6.9  | 58.3                 | 0.04 | 100.9                      | 59                 | 395   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 17       | 1         | 13.08                   | 4.72            | 6.9  | 58.3                 | 0.04 | 100.9                      | 51                 | 396   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 16       | 1.5       | 13.08                   | 4.66            | 6.9  | 57.8                 | 0.04 | 100.7                      | 50                 | 397   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 15       | 2         | 13.09                   | 4.54            | 6.91 | 57                   | 0.04 | 100.5                      | 42                 | 396   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 14       | 2.5       | 13.08                   | 4.49            | 6.9  | 58.2                 | 0.04 | 100.3                      | 50                 | 396   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 13       | 3         | 13.09                   | 4.41            | 6.93 | 57.5                 | 0.04 | 100.1                      | 40                 | 394   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 12       | 3.5       | 13.09                   | 4.44            | 6.94 | 57.4                 | 0.04 | 100.3                      | 57                 | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 11       | 4         | 13.08                   | 4.44            | 6.94 | 58                   | 0.04 | 100.2                      | 104                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 10       | 4.5       | 13.11                   | 4.44            | 6.92 | 57.9                 | 0.04 | 100.4                      | 122                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 9        | 5         | 13.14                   | 4.51            | 6.9  | 60.9                 | 0.04 | 100.8                      | 105                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 8        | 5.5       | 13.14                   | 4.47            | 6.87 | 61.4                 | 0.04 | 100.7                      | 103                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 7        | 6         | 13.14                   | 4.38            | 6.88 | 61.5                 | 0.04 | 100.4                      | 44                 | 392   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 6        | 6.5       | 13.09                   | 4.43            | 6.86 | 61.7                 | 0.04 | 100.2                      | 121                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 5        | 7         | 13.14                   | 4.47            | 6.82 | 61.7                 | 0.04 | 100.7                      | 102                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 4        | 7.5       | 13.15                   | 4.44            | 6.83 | 62                   | 0.04 | 100.7                      | 108                | 392   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 3        | 8         | 13.2                    | 4.46            | 6.77 | 62.1                 | 0.04 | 101.1                      | 52                 | 394   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 2        | 8.5       | 13.22                   | 4.49            | 6.77 | 62.1                 | 0.04 | 101.4                      | 102                | 393   |
| CDA River | 1098  | DBRP     | 3/13/98 | 11:54:53 | 2          | 1        | 9         | 13.25                   | 4.39            | 6.76 | 63.3                 | 0.04 | 101.3                      | 151                | 392   |

|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 10 | 0.3  | 12.28 | 7.14  | 7.04 | 43.5 | 0.03 | 101.6 | 110 | 411 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 9  | 1.6  | 12.23 | 7.04  | 7.01 | 43.6 | 0.03 | 100.9 | 139 | 414 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 8  | 2.6  | 12.26 | 6.96  | 6.99 | 43.6 | 0.03 | 100.9 | 111 | 415 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 7  | 3.6  | 12.26 | 7.04  | 7.04 | 43.5 | 0.03 | 101.2 | 118 | 413 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 6  | 4.6  | 12.26 | 6.96  | 6.93 | 43.6 | 0.03 | 100.9 | 140 | 418 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 5  | 5.6  | 12.23 | 6.96  | 6.99 | 43.7 | 0.03 | 100.7 | 110 | 415 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 4  | 6.6  | 12.23 | 6.87  | 6.98 | 43.9 | 0.03 | 100.5 | 109 | 415 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 3  | 7.6  | 12.23 | 6.86  | 6.95 | 44   | 0.03 | 100.4 | 56  | 415 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 2  | 8.6  | 12.23 | 6.84  | 6.95 | 43.9 | 0.03 | 100.4 | 100 | 415 |
| CDA River | 1498 | DBAS   | 4/8/98  | 13:19:28 | 3.2 | 1  | 9.6  | 12.24 | 6.78  | 6.93 | 44.4 | 0.03 | 100.3 | 407 | 415 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 11 | 0.1  | 11.53 | 12.99 | 7.01 | 47.1 | 0.03 | 107.6 | 57  | 376 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 10 | 1.3  | 11.68 | 10.12 | 7.04 | 49.1 | 0.03 | 102.1 | 53  | 373 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 9  | 2.3  | 11.7  | 9.56  | 6.98 | 50.3 | 0.03 | 100.9 | 42  | 374 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 8  | 3.3  | 11.72 | 9.49  | 6.91 | 50.3 | 0.03 | 100.9 | 34  | 377 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 7  | 4.3  | 11.65 | 9.35  | 6.89 | 50.2 | 0.03 | 99.9  | 53  | 377 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 6  | 5.3  | 11.54 | 8.28  | 6.82 | 49.8 | 0.03 | 96.5  | 51  | 378 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 5  | 6.2  | 11.5  | 7.46  | 6.8  | 48.4 | 0.03 | 94.3  | 108 | 378 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 4  | 7.2  | 11.39 | 7.02  | 6.75 | 48.5 | 0.03 | 92.3  | 107 | 378 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 3  | 8.2  | 11.3  | 6.59  | 6.77 | 47.9 | 0.03 | 90.6  | 59  | 375 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 2  | 9.3  | 11.35 | 6.08  | 6.73 | 48.5 | 0.03 | 89.8  | 109 | 375 |
| CDA River | 1698 | DBASRP | 4/23/98 | 12:46:10 | 2.3 | 1  | 10.3 | 11.42 | 5.67  | 6.7  | 49.4 | 0.03 | 89.4  | 207 | 374 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 11 | 0.3  | 11.07 | 11.13 | 7.03 | 40.5 | 0.03 | 101.7 | 42  | 395 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 10 | 1.8  | 11.08 | 11.15 | 6.98 | 40.5 | 0.03 | 101.8 | 49  | 397 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 9  | 2.8  | 11.05 | 11.12 | 6.98 | 40.4 | 0.03 | 101.5 | 103 | 396 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 8  | 3.8  | 11.04 | 11.05 | 7.06 | 39.9 | 0.03 | 101.2 | 112 | 389 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 7  | 4.8  | 11.01 | 11    | 7    | 40   | 0.03 | 100.9 | 119 | 390 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 6  | 5.8  | 10.99 | 10.94 | 7.02 | 40.2 | 0.03 | 100.5 | 55  | 386 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 5  | 6.8  | 10.98 | 10.68 | 6.92 | 42.7 | 0.03 | 99.7  | 120 | 390 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 4  | 7.8  | 11.03 | 10.58 | 6.89 | 44   | 0.03 | 100   | 128 | 389 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 3  | 8.8  | 11.09 | 10.12 | 6.85 | 49.1 | 0.03 | 99.4  | 57  | 387 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 2  | 9.8  | 10.99 | 9.66  | 6.76 | 49.5 | 0.03 | 97.5  | 118 | 389 |
| CDA River | 2198 | ASJLDT | 5/26/98 | 12:34:22 | 3.5 | 1  | 10.8 | 11.14 | 8.47  | 6.78 | 44.5 | 0.03 | 96    | 200 | 387 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 12 | 0.3  | 10.65 | 17.18 | 7.58 | 42.6 | 0.03 | 110.7 | 129 | 399 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 11 | 1.1  | 10.75 | 16.19 | 7.55 | 42.9 | 0.03 | 109.5 | 118 | 403 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 10 | 2.1  | 10.85 | 15.56 | 7.54 | 42.9 | 0.03 | 109   | 113 | 405 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 9  | 3.1  | 10.92 | 14.96 | 7.52 | 42.2 | 0.03 | 108.3 | 56  | 407 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 8  | 4.1  | 10.85 | 14.86 | 7.49 | 43.4 | 0.03 | 107.4 | 150 | 408 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 7  | 5.1  | 10.84 | 14.56 | 7.38 | 49.4 | 0.03 | 106.6 | 119 | 412 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 6  | 6.1  | 10.6  | 13.98 | 7.25 | 53.8 | 0.03 | 102.9 | 115 | 416 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 5  | 7.1  | 10.64 | 13.63 | 7.21 | 55.1 | 0.04 | 102.5 | 49  | 417 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 4  | 8.1  | 10.64 | 13.4  | 7.22 | 53.1 | 0.03 | 102   | 115 | 417 |

|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 3  | 9.1  | 10.68 | 12.94 | 7.25 | 44.1 | 0.03 | 101.3 | 119 | 415 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 2  | 10.1 | 10.76 | 12.43 | 7.24 | 41.4 | 0.03 | 100.8 | 132 | 415 |
| CDA River | 2398 | DTJLAS | 6/8/98  | 11:39:46 | 3.6 | 1  | 11.1 | 10.57 | 11.58 | 7.2  | 40.8 | 0.03 | 97.2  | 244 | 416 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 12 | 0.6  | 9.81  | 19.83 | 7.59 | 45.2 | 0.03 | 107.9 | 46  | 385 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 11 | 1.9  | 9.98  | 18.92 | 7.6  | 47.1 | 0.03 | 107.7 | 46  | 387 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 10 | 2.9  | 10.01 | 18.85 | 7.59 | 47.2 | 0.03 | 107.9 | 33  | 388 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 9  | 3.9  | 10    | 18.85 | 7.56 | 47.3 | 0.03 | 107.8 | 105 | 390 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 8  | 4.9  | 10.58 | 17.54 | 7.46 | 62   | 0.04 | 111   | 53  | 395 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 7  | 5.9  | 10.4  | 16.69 | 7.36 | 50.3 | 0.03 | 107.2 | 50  | 395 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 6  | 6.9  | 10.44 | 15.79 | 7.33 | 45.2 | 0.03 | 105.6 | 50  | 396 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 5  | 7.9  | 10.5  | 15.54 | 7.28 | 48   | 0.03 | 105.6 | 112 | 397 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 4  | 8.9  | 10.49 | 14.86 | 7.21 | 47.1 | 0.03 | 104   | 54  | 399 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 3  | 9.9  | 10.46 | 14.74 | 7.15 | 48.1 | 0.03 | 103.4 | 42  | 400 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 2  | 10.9 | 10.32 | 14.16 | 7.08 | 48.3 | 0.03 | 100.8 | 126 | 401 |
| CDA River | 2598 | DTASJL | 6/23/98 | 13:30:13 | 5.4 | 1  | 11.9 | 9.96  | 12.44 | 6.94 | 42.5 | 0.03 | 93.6  | 29  | 407 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 11 | 0.4  | 9.23  | 22.99 | 7.87 | 50.4 | 0.03 | 107.4 | 126 | 324 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 10 | 1.4  | 9.26  | 22.92 | 7.96 | 51.2 | 0.03 | 107.5 | 112 | 319 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 9  | 2.4  | 9.26  | 22.45 | 7.89 | 52.2 | 0.03 | 106.6 | 133 | 321 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 8  | 3.4  | 9.59  | 21.86 | 7.9  | 56.9 | 0.04 | 109.1 | 129 | 322 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 7  | 4.4  | 9.52  | 21.39 | 7.81 | 48.4 | 0.03 | 107.4 | 100 | 322 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 6  | 5.4  | 10.51 | 19.06 | 7.87 | 54.2 | 0.03 | 113.3 | 208 | 322 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 5  | 6.4  | 10.68 | 17.95 | 7.92 | 47.6 | 0.03 | 112.5 | 201 | 320 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 4  | 7.4  | 10.64 | 16.93 | 7.72 | 45.5 | 0.03 | 109.8 | 112 | 323 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 3  | 8.4  | 10.45 | 16.19 | 7.54 | 44.1 | 0.03 | 106.2 | 138 | 325 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 2  | 9.4  | 10.16 | 15.96 | 7.41 | 43.8 | 0.03 | 102.7 | 153 | 326 |
| CDA River | 2798 | ASDNBH | 7/8/98  | 10:29:22 | 5.4 | 1  | 10.4 | 9.81  | 16.09 | 7.3  | 43.3 | 0.03 | 99.4  | 550 | 327 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 12 | 0.3  | 9.01  | 23.86 | 7.85 | 54.8 | 0.04 | 106.4 | 47  | 333 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 11 | 1.3  | 9.05  | 23.6  | 7.87 | 55   | 0.04 | 106.3 | 108 | 332 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 10 | 2.3  | 9.13  | 23.37 | 7.89 | 59.5 | 0.04 | 106.8 | 51  | 333 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 9  | 3.3  | 9.15  | 23.29 | 7.88 | 63.1 | 0.04 | 106.9 | 56  | 333 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 8  | 4.3  | 9.05  | 23.08 | 7.81 | 63.3 | 0.04 | 105.2 | 57  | 336 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 7  | 5.3  | 8.94  | 22.31 | 7.59 | 65.3 | 0.04 | 102.5 | 34  | 339 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 6  | 6.3  | 10.21 | 19.61 | 7.9  | 49.5 | 0.03 | 111.1 | 141 | 335 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 5  | 7.3  | 10.35 | 18.9  | 7.87 | 48.8 | 0.03 | 111   | 102 | 334 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 4  | 8.3  | 10.54 | 18.25 | 7.76 | 48.3 | 0.03 | 111.6 | 108 | 338 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 3  | 9.3  | 9.66  | 16.91 | 7.21 | 48.7 | 0.03 | 99.4  | 110 | 346 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 2  | 10.3 | 9.16  | 14.81 | 7.06 | 46.3 | 0.03 | 90.2  | 138 | 347 |
| CDA River | 2998 | ASBH   | 7/20/98 | 10:55:07 | 4.6 | 1  | 11.3 | 8.12  | 13.91 | 6.95 | 46.4 | 0.03 | 78.4  | 244 | 347 |
|           |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 13 | 0.3  | 8.62  | 24.59 | 7.71 | 64.6 | 0.04 | 103.1 | 47  | 343 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 12 | 1    | 8.67  | 24.09 | 7.73 | 64.6 | 0.04 | 102.7 | 46  | 342 |

|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 11 | 1.9  | 8.7  | 23.84 | 7.71 | 65.5 | 0.04 | 102.6 | 27  | 343 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 10 | 2.8  | 8.7  | 23.82 | 7.71 | 65.9 | 0.04 | 102.5 | 44  | 343 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 9  | 3.7  | 8.71 | 23.78 | 7.71 | 65.5 | 0.04 | 102.6 | 102 | 343 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 8  | 4.8  | 8.72 | 23.76 | 7.7  | 66.2 | 0.04 | 102.7 | 37  | 344 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 7  | 5.7  | 8.73 | 23.75 | 7.67 | 67.3 | 0.04 | 102.8 | 36  | 345 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 6  | 6.7  | 8.73 | 23.75 | 7.64 | 68.5 | 0.04 | 102.8 | 47  | 346 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 5  | 7.7  | 8.17 | 23.53 | 7.44 | 71   | 0.05 | 95.8  | 125 | 351 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 4  | 8.7  | 9.57 | 19.38 | 7.33 | 50.7 | 0.03 | 103.6 | 42  | 353 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 3  | 9.8  | 9.75 | 18.95 | 7.33 | 50   | 0.03 | 104.6 | 50  | 353 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 2  | 10.8 | 9.69 | 17.86 | 7.28 | 49   | 0.03 | 101.6 | 132 | 353 |
| CDA River | 3298 | DBAS   | 8/10/98 | 10:52:11 | 6.6 | 1  | 11.8 | 9.14 | 16.86 | 7.21 | 48.7 | 0.03 | 93.9  | 206 | 354 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 13 | 0.2  | 9.13 | 22.93 | 7.93 | 57.7 | 0.04 | 106.4 | 34  | 331 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 12 | 1    | 9.17 | 22.29 | 7.96 | 57.5 | 0.04 | 105.5 | 42  | 330 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 11 | 2    | 9.26 | 21.83 | 7.95 | 57.7 | 0.04 | 105.6 | 51  | 332 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 10 | 3    | 9.24 | 21.76 | 7.91 | 58.6 | 0.04 | 105.2 | 107 | 333 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 9  | 4    | 9.24 | 21.7  | 7.9  | 59.2 | 0.04 | 105.1 | 116 | 333 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 8  | 5    | 9.24 | 21.65 | 7.87 | 59.8 | 0.04 | 105   | 118 | 333 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 7  | 6    | 9.23 | 21.6  | 7.85 | 59.2 | 0.04 | 104.8 | 108 | 334 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 6  | 7    | 9.19 | 21.48 | 7.73 | 62.3 | 0.04 | 104.1 | 55  | 338 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 5  | 8    | 9.01 | 21.39 | 7.59 | 70   | 0.04 | 101.9 | 29  | 342 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 4  | 9    | 8.9  | 21.26 | 7.5  | 80.4 | 0.05 | 100.4 | 107 | 345 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 3  | 10   | 8.8  | 21.09 | 7.33 | 85.9 | 0.06 | 98.9  | 118 | 350 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 2  | 11   | 8.74 | 16.59 | 6.82 | 47   | 0.03 | 89.8  | 114 | 357 |
| CDA River | 3498 | ASDNGL | 8/25/98 | 11:44:46 | 7   | 1  | 11.6 | 7.89 | 16.04 | 6.78 | 46.2 | 0.03 | 80    | 737 | 358 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 12 | 0.3  | 9.2  | 22.43 | 7.96 | 59.9 | 0.04 | 106.2 | 101 | 323 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 11 | 1.3  | 9.23 | 21.99 | 7.9  | 61   | 0.04 | 105.6 | 47  | 327 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 10 | 2.3  | 9.05 | 21.55 | 7.75 | 69.5 | 0.04 | 102.7 | 47  | 333 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 9  | 3.3  | 8.95 | 21.46 | 7.72 | 73.7 | 0.05 | 101.5 | 112 | 334 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 8  | 4.3  | 9.13 | 21.39 | 7.82 | 64.4 | 0.04 | 103.4 | 47  | 330 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 7  | 5.3  | 9.21 | 21.37 | 7.87 | 60.3 | 0.04 | 104.2 | 51  | 329 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 6  | 6.3  | 9.2  | 21.32 | 7.84 | 59.6 | 0.04 | 104   | 122 | 330 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 5  | 7.3  | 9.23 | 21.3  | 7.81 | 59.2 | 0.04 | 104.3 | 130 | 331 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 4  | 8.3  | 9.15 | 21.25 | 7.62 | 59.6 | 0.04 | 103.2 | 51  | 338 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 3  | 9.3  | 9.06 | 21.21 | 7.41 | 60.4 | 0.04 | 102.2 | 100 | 345 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 2  | 10.3 | 7.94 | 19.4  | 6.96 | 56.1 | 0.04 | 86.4  | 118 | 356 |
| CDA River | 3598 | DTAS   | 9/2/98  | 11:23:48 | 7.1 | 1  | 11.3 | 7.74 | 17.32 | 6.87 | 48.7 | 0.03 | 80.8  | 759 | 356 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| CDA River | 3998 | ASKB   | 9/30/98 | 12:55:00 | 7.4 | 11 | 0.3  | 9.24 | 18.73 | 7.76 | 62.9 | 0.04 | 98.9  | 52  | 335 |
| CDA River | 3998 | ASKB   | 9/30/98 | 12:55:00 | 7.4 | 10 | 1.9  | 9.28 | 18.08 | 7.77 | 63   | 0.04 | 98    | 104 | 336 |
| CDA River | 3998 | ASKB   | 9/30/98 | 12:55:00 | 7.4 | 9  | 2.9  | 9.29 | 18.05 | 7.75 | 63.5 | 0.04 | 98.1  | 133 | 338 |
| CDA River | 3998 | ASKB   | 9/30/98 | 12:55:00 | 7.4 | 8  | 3.9  | 9.32 | 18.03 | 7.75 | 65.3 | 0.04 | 98.4  | 108 | 338 |
| CDA River | 3998 | ASKB   | 9/30/98 | 12:55:00 | 7.4 | 7  | 4.9  | 9.34 | 17.98 | 7.73 | 65.1 | 0.04 | 98.5  | 119 | 339 |

|           |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
|-----------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 6  | 5.9  | 9.33 | 17.9  | 7.7  | 66   | 0.04 | 98.2 | 123 | 340 |
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 5  | 7    | 9.32 | 17.83 | 7.63 | 72.4 | 0.05 | 98   | 107 | 343 |
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 4  | 7.9  | 9.26 | 17.62 | 7.57 | 74.8 | 0.05 | 96.9 | 104 | 345 |
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 3  | 8.9  | 9.23 | 17.52 | 7.53 | 78.2 | 0.05 | 96.4 | 140 | 345 |
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 2  | 9.9  | 8.96 | 17.37 | 7.45 | 87.7 | 0.06 | 93.2 | 135 | 347 |
| CDA River | 3998 | ASKB   | 9/30/98  | 12:55:00 | 7.4 | 1  | 10.9 | 8.85 | 17.32 | 7.39 | 89.7 | 0.06 | 92.1 | 818 | 353 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 12 | 0.4  | 9.37 | 12.45 | 7.36 | 58.5 | 0.04 | 86.8 | 34  | 357 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 11 | 0.8  | 9.38 | 12.43 | 7.38 | 58.5 | 0.04 | 86.8 | 30  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 10 | 1.8  | 9.36 | 12.43 | 7.38 | 58.6 | 0.04 | 86.6 | 44  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 9  | 2.8  | 9.36 | 12.41 | 7.34 | 58.6 | 0.04 | 86.5 | 39  | 358 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 8  | 3.8  | 9.34 | 12.4  | 7.34 | 58.3 | 0.04 | 86.4 | 30  | 357 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 7  | 4.8  | 9.26 | 12.37 | 7.32 | 58.3 | 0.04 | 85.6 | 46  | 357 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 6  | 5.8  | 9.23 | 12.3  | 7.31 | 59   | 0.04 | 85.1 | 47  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 5  | 6.8  | 9.23 | 12.2  | 7.3  | 59.2 | 0.04 | 85   | 31  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 4  | 7.8  | 9.36 | 11.84 | 7.3  | 61.3 | 0.04 | 85.4 | 42  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 3  | 8.8  | 9.51 | 11.4  | 7.27 | 66.5 | 0.04 | 86   | 30  | 356 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 2  | 9.8  | 9.63 | 11.17 | 7.25 | 68.4 | 0.04 | 86.5 | 26  | 354 |
| CDA River | 4298 | DTASKB | 10/20/98 | 9:51:12  | 7.1 | 1  | 10.8 | 9.73 | 10.63 | 7.22 | 74.3 | 0.05 | 86.3 | 218 | 353 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 12 | 0.3  | 8.53 | 9.55  | 7.37 | 57.2 | 0.04 | 74.3 | 30  | 340 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 11 | 1    | 8.51 | 9.55  | 7.33 | 57.3 | 0.04 | 74.1 | 43  | 342 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 10 | 2    | 8.52 | 9.55  | 7.36 | 57.4 | 0.04 | 74.2 | 38  | 340 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 9  | 3    | 8.53 | 9.55  | 7.33 | 57.7 | 0.04 | 74.3 | 42  | 341 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 8  | 4    | 8.52 | 9.55  | 7.31 | 57.6 | 0.04 | 74.2 | 46  | 342 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 7  | 5    | 8.51 | 9.53  | 7.3  | 57.8 | 0.04 | 74.1 | 46  | 343 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 6  | 6    | 8.55 | 9.43  | 7.3  | 58.6 | 0.04 | 74.2 | 42  | 342 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 5  | 7    | 8.53 | 9.46  | 7.28 | 58.2 | 0.04 | 74.1 | 112 | 343 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 4  | 8    | 8.65 | 9.07  | 7.3  | 63.1 | 0.04 | 74.4 | 119 | 340 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 3  | 9    | 8.74 | 8.68  | 7.28 | 66.5 | 0.04 | 74.5 | 101 | 339 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 2  | 9.9  | 9.21 | 6.91  | 7.22 | 88.4 | 0.06 | 75.2 | 307 | 339 |
| CDA River | 4598 | KB AS  | 11/12/98 | 12:34:00 | 5.5 | 1  | 11   | 9.23 | 6.99  | 7.21 | 90   | 0.06 | 81.1 | 9   | 342 |

| Location     | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 18       | 0.5       | 12.84                   | 4.14            | 6.84 | 44.3                 | 0.01 | 149.5                      | 4                  | 375   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 17       | 1.6       | 12.87                   | 3.98            | 7.01 | 44.3                 | 0.03 | 97.3                       | 115                | 392   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 16       | 2.6       | 12.79                   | 3.79            | 6.96 | 44.3                 | 0.03 | 96.3                       | 116                | 395   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 15       | 3.6       | 12.79                   | 3.84            | 7    | 44.1                 | 0.03 | 96.4                       | 108                | 392   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 14       | 4.6       | 12.83                   | 3.89            | 6.89 | 44.3                 | 0.03 | 96.8                       | 121                | 398   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 13       | 5.6       | 12.79                   | 3.86            | 6.96 | 44.3                 | 0.03 | 96.5                       | 53                 | 392   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 12       | 6.6       | 12.78                   | 3.76            | 6.92 | 44.5                 | 0.03 | 96.1                       | 102                | 395   |
| Mid Lake CDA | 1098  | DBRP     | 3/13/98 | 12:25:26 | 1.9        | 11       | 7.6       | 12.78                   | 3.76            | 6.99 | 44.4                 | 0.03 | 96.1                       | 126                | 390   |

|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 10 | 8.6  | 12.8  | 3.76  | 7    | 44.5 | 0.03 | 96.2  | 50  | 389 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 9  | 9.6  | 12.77 | 3.78  | 6.96 | 44.3 | 0.03 | 96.1  | 111 | 391 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 8  | 10.6 | 12.77 | 3.79  | 6.94 | 44.4 | 0.03 | 96.2  | 114 | 391 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 7  | 11.6 | 12.76 | 3.79  | 6.96 | 44.5 | 0.03 | 96.1  | 159 | 389 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 6  | 12.6 | 12.79 | 3.78  | 6.95 | 44.6 | 0.03 | 96.2  | 120 | 389 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 5  | 13.6 | 12.79 | 3.79  | 6.88 | 44.6 | 0.03 | 96.3  | 107 | 391 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 4  | 14.6 | 12.79 | 3.78  | 6.9  | 44.6 | 0.03 | 96.2  | 109 | 390 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 3  | 15.7 | 12.74 |       | 6.9  | 44.5 | 0.03 | 96    | 146 | 388 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 2  | 16.6 | 12.73 | 3.83  | 6.87 | 44.6 | 0.03 | 95.9  | 135 | 388 |
| Mid Lake CDA | 1098 | DBRP   | 3/13/98 | 12:25:26 | 1.9 | 1  | 17.6 | 12.77 | 3.83  | 6.88 | 44.7 | 0.03 | 96.2  | 227 | 387 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 19 | 0.5  | 12.45 | 8.56  | 7.14 | 39.4 | 0.03 | 106.7 | 111 | 410 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 18 | 1.5  | 12.41 | 8.15  | 7.08 | 39.1 | 0.03 | 105.2 | 133 | 415 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 17 | 2.5  | 12.6  | 7.6   | 7.08 | 39.2 | 0.03 | 105.4 | 126 | 416 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 16 | 3.5  | 12.56 | 7.12  | 7.08 | 39   | 0.02 | 103.8 | 39  | 415 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 15 | 4.6  | 12.48 | 7.02  | 7.02 | 39   | 0.02 | 102.9 | 110 | 417 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 14 | 5.5  | 12.31 | 6.99  | 7.06 | 39.1 | 0.03 | 101.5 | 141 | 415 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 13 | 6.4  | 12.3  | 6.91  | 7.02 | 39.2 | 0.03 | 101.2 | 112 | 416 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 12 | 7.4  | 12.21 | 6.66  | 6.96 | 39.3 | 0.03 | 99.8  | 103 | 418 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 11 | 8.6  | 12.24 | 6.61  | 6.92 | 39.4 | 0.03 | 99.9  | 109 | 419 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 10 | 9.5  | 12.21 | 6.53  | 6.94 | 39.5 | 0.03 | 99.5  | 113 | 417 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 9  | 10.5 | 12.05 | 6.3   | 6.94 | 39.9 | 0.03 | 97.6  | 121 | 416 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 8  | 11.6 | 12.03 | 6.23  | 6.9  | 39.9 | 0.03 | 97.3  | 135 | 417 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 7  | 12.5 | 12.02 | 5.97  | 6.88 | 39.8 | 0.03 | 96.6  | 121 | 417 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 6  | 13.6 | 12.02 | 5.8   | 6.83 | 40.9 | 0.03 | 96.1  | 225 | 419 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 5  | 14.5 | 12.08 | 5.64  | 6.9  | 42.7 | 0.03 | 96.2  | 119 | 415 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 4  | 15.5 | 12    | 5.59  | 6.87 | 43   | 0.03 | 95.4  | 202 | 415 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 3  | 16.5 | 11.95 | 5.57  | 6.87 | 43.2 | 0.03 | 95    | 103 | 414 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 2  | 17.5 | 11.51 | 5.46  | 6.84 | 43.9 | 0.03 | 91.2  | 129 | 414 |
| Mid Lake CDA | 1498 | DBAS   | 4/8/98  | 13:44:51 | 2.2 | 1  | 18.5 | 11.51 | 5.46  | 6.82 | 44   | 0.03 | 91.2  | 358 | 414 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 13 | 0.3  | 12.01 | 11.69 | 7.24 | 42.7 | 0.03 | 110.3 | 52  | 396 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 12 | 1.5  | 12.01 | 9.63  | 7.19 | 42.2 | 0.03 | 105.1 | 113 | 402 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 11 | 3    | 12.12 | 8.58  | 7.13 | 42.1 | 0.03 | 103.4 | 109 | 405 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 10 | 4.4  | 12    | 7.32  | 7.04 | 43.1 | 0.03 | 99.3  | 55  | 405 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 9  | 6    | 11.88 | 6.69  | 6.95 | 45   | 0.03 | 96.8  | 103 | 409 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 8  | 7.4  | 11.88 | 6.53  | 6.93 | 46.2 | 0.03 | 96.4  | 133 | 409 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 7  | 9    | 11.9  | 6.36  | 6.88 | 47.1 | 0.03 | 96.2  | 112 | 410 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 6  | 10.5 | 11.87 | 6.31  | 6.88 | 47.2 | 0.03 | 95.8  | 154 | 409 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 5  | 12   | 11.87 | 6.2   | 6.91 | 47.7 | 0.03 | 95.5  | 130 | 406 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 4  | 13.5 | 11.82 | 6.15  | 6.91 | 48.2 | 0.03 | 95    | 134 | 405 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 3  | 15   | 11.72 | 6     | 6.82 | 47.8 | 0.03 | 93.8  | 124 | 407 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 2  | 16.6 | 11.62 | 5.42  | 6.87 | 49.9 | 0.03 | 91.7  | 144 | 402 |
| Mid Lake CDA | 1698 | DBASRP | 4/21/98 | 14:41:39 | 2.7 | 1  | 18.2 | 11.21 | 5.36  | 6.83 | 51   | 0.03 | 88.3  | 235 | 404 |

|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 13 | 0.3  | 11.17 | 11.95 | 7.17 | 35   | 0.02 | 104.5 | 49  | 382 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 12 | 2    | 11.16 | 11.89 | 7.17 | 35.5 | 0.02 | 104.3 | 54  | 382 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 11 | 3.5  | 11.13 | 11.43 | 7.06 | 36.8 | 0.02 | 102.9 | 40  | 384 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 10 | 5    | 11.12 | 11.13 | 7.08 | 34.5 | 0.02 | 102.1 | 50  | 381 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 9  | 6.5  | 11.12 | 11.1  | 7.07 | 34.2 | 0.02 | 102.1 | 129 | 379 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 8  | 8    | 11.15 | 10.97 | 7.04 | 33.5 | 0.02 | 102   | 113 | 378 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 7  | 9.5  | 11.05 | 9.76  | 6.96 | 35.5 | 0.02 | 98.2  | 102 | 378 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 6  | 11   | 11.18 | 8.96  | 6.86 | 40.7 | 0.03 | 97.5  | 119 | 381 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 5  | 12.5 | 11.22 | 8.45  | 6.88 | 43.3 | 0.03 | 96.7  | 111 | 377 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 4  | 14   | 11.19 | 7.92  | 6.8  | 44.8 | 0.03 | 95.1  | 140 | 378 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 3  | 15.5 | 11.09 | 7.32  | 6.77 | 46.5 | 0.03 | 92.9  | 152 | 375 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 2  | 17   | 11.05 | 7.1   | 6.74 | 47.1 | 0.03 | 92.1  | 145 | 369 |
| Mid Lake CDA | 2198 | ASJLDT | 5/26/98 | 12:58:43 | 3.4 | 1  | 18.8 | 10.3  | 6.26  | 6.61 | 49   | 0.03 | 84    | 156 | 366 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 11 | 0.4  | 10.86 | 16.72 | 7.61 | 38.6 | 0.02 | 111.9 | 49  | 375 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 10 | 1.5  | 11.02 | 15.73 | 7.6  | 38.7 | 0.02 | 111.1 | 110 | 376 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 9  | 3    | 10.94 | 15.39 | 7.51 | 39.2 | 0.03 | 109.5 | 103 | 379 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 8  | 5    | 10.96 | 13.09 | 7.32 | 37.7 | 0.02 | 104.3 | 110 | 382 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 7  | 7    | 10.97 | 12.5  | 7.26 | 41.8 | 0.03 | 103   | 114 | 385 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 6  | 9    | 10.97 | 12.28 | 7.2  | 41.9 | 0.03 | 102.5 | 33  | 384 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 5  | 11   | 10.88 | 11.5  | 7.17 | 36.7 | 0.02 | 99.8  | 121 | 383 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 4  | 13   | 10.87 | 10.77 | 7.11 | 35.9 | 0.02 | 98.1  | 50  | 383 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 3  | 15   | 10.54 | 8.92  | 7.02 | 39.9 | 0.03 | 91    | 118 | 385 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 2  | 17   | 10.01 | 7.69  | 6.92 | 46.5 | 0.03 | 83.9  | 125 | 386 |
| Mid Lake CDA | 2398 | DTJLAS | 6/8/98  | 12:11:07 | 2.9 | 1  | 19   | 9.63  | 8.04  | 6.92 | 47   | 0.03 | 81.3  | 236 | 383 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 11 | 0.2  | 9.91  | 19.71 | 7.53 | 45.7 | 0.03 | 108.7 | 49  | 417 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 10 | 1    | 9.97  | 19.06 | 7.54 | 46   | 0.03 | 107.9 | 102 | 423 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 9  | 3    | 10.62 | 17.15 | 7.49 | 48   | 0.03 | 110.5 | 53  | 428 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 8  | 5    | 10.88 | 15.76 | 7.44 | 52.7 | 0.03 | 110   | 48  | 431 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 7  | 7    | 10.84 | 14.86 | 7.35 | 46.3 | 0.03 | 107.5 | 118 | 434 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 6  | 9    | 10.84 | 14.47 | 7.24 | 46.1 | 0.03 | 106.6 | 47  | 438 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 5  | 11   | 10.88 | 13.81 | 7.16 | 45.8 | 0.03 | 105.4 | 102 | 440 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 4  | 13   | 10.72 | 12.31 | 6.97 | 43.8 | 0.03 | 100.5 | 100 | 446 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 3  | 15   | 9.99  | 10.51 | 6.79 | 43   | 0.03 | 89.7  | 130 | 450 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 2  | 17   | 9.14  | 8.46  | 6.65 | 46.4 | 0.03 | 78.2  | 131 | 454 |
| Mid Lake CDA | 2598 | DTJLAS | 6/23/98 | 14:06:28 | 5   | 1  | 19   | 8.81  | 8.02  | 6.59 | 47.9 | 0.03 | 74.5  | 152 | 457 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 20 | 0.4  | 9.49  | 24.64 | 8.03 | 50.7 | 0.03 | 113.5 | 100 | 337 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 19 | 1    | 9.48  | 23.4  | 7.92 | 50.5 | 0.03 | 110.8 | 203 | 343 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 18 | 2    | 10.36 | 19.56 | 7.89 | 50.4 | 0.03 | 112.5 | 126 | 345 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 17 | 3    | 10.67 | 18.5  | 8.03 | 50.7 | 0.03 | 113.4 | 114 | 343 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 16 | 4    | 11.01 | 17.51 | 8.34 | 49.4 | 0.03 | 114.6 | 149 | 338 |

|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 15 | 5    | 11    | 16.75 | 8.11 | 49.3 | 0.03 | 112.8 | 122 | 343 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 14 | 6    | 10.95 | 16.58 | 8.08 | 48.3 | 0.03 | 111.9 | 124 | 344 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 13 | 7    | 10.92 | 16.53 | 8.07 | 47.8 | 0.03 | 111.5 | 115 | 346 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 12 | 8    | 10.76 | 16.27 | 7.8  | 47.5 | 0.03 | 109.2 | 112 | 352 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 11 | 9    | 10.42 | 15.61 | 7.48 | 46.7 | 0.03 | 104.3 | 105 | 356 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 10 | 10   | 10.23 | 14.91 | 7.27 | 46.6 | 0.03 | 100.9 | 117 | 359 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 9  | 11   | 10.21 | 14.72 | 7.24 | 46.6 | 0.03 | 100.3 | 223 | 359 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 8  | 12   | 9.93  | 14.03 | 7.08 | 45.5 | 0.03 | 96    | 136 | 361 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 7  | 13.1 | 9.53  | 12.84 | 6.93 | 44.3 | 0.03 | 89.8  | 122 | 362 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 6  | 14   | 9.24  | 12.07 | 6.89 | 43.2 | 0.03 | 85.5  | 53  | 362 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 5  | 15   | 8.96  | 10.85 | 6.77 | 43.8 | 0.03 | 80.6  | 113 | 362 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 4  | 16   | 8.85  | 10.46 | 6.75 | 43.9 | 0.03 | 78.9  | 123 | 362 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 3  | 17   | 8.66  | 9.72  | 6.72 | 45.4 | 0.03 | 75.9  | 118 | 361 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 2  | 18   | 7.82  | 8.76  | 6.64 | 47.2 | 0.03 | 66.9  | 145 | 360 |
| Mid Lake CDA | 2798 | DTASBH | 7/7/98  | 13:23:07 | 5.6 | 1  | 19   | 7.73  | 8.71  | 6.66 | 47.5 | 0.03 | 66.1  | 315 | 357 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 20 | 0.4  | 9.47  | 23.5  | 8.11 | 46.3 | 0.03 | 111.2 | 39  | 325 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 19 | 1    | 9.41  | 22.88 | 7.97 | 46.3 | 0.03 | 109.3 | 102 | 330 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 18 | 2    | 9.42  | 21.74 | 7.83 | 46.3 | 0.03 | 107   | 59  | 332 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 17 | 3    | 10.25 | 19.88 | 8.08 | 47.1 | 0.03 | 112.3 | 101 | 329 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 16 | 4    | 10.8  | 18.07 | 8.37 | 44.4 | 0.03 | 114.1 | 51  | 325 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 15 | 5    | 10.93 | 17.13 | 8.38 | 43.1 | 0.03 | 113.2 | 110 | 325 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 14 | 6    | 10.95 | 17.08 | 8.29 | 43.9 | 0.03 | 113.3 | 121 | 328 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 13 | 7    | 10.9  | 16.95 | 8.16 | 44.4 | 0.03 | 112.4 | 128 | 331 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 12 | 8    | 10.7  | 16.63 | 8.01 | 43.8 | 0.03 | 109.7 | 111 | 335 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 11 | 9    | 10.58 | 16.23 | 7.71 | 43.6 | 0.03 | 107.6 | 56  | 341 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 10 | 10   | 10.47 | 15.96 | 7.6  | 43.1 | 0.03 | 105.8 | 116 | 343 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 9  | 11   | 10.32 | 15.42 | 7.42 | 43   | 0.03 | 103.1 | 122 | 346 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 8  | 12   | 10.18 | 15.12 | 7.3  | 42.5 | 0.03 | 101   | 116 | 348 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 7  | 13   | 9.77  | 14.1  | 7.11 | 41.5 | 0.03 | 94.8  | 152 | 351 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 6  | 14   | 9.44  | 13.12 | 6.98 | 40.4 | 0.03 | 89.7  | 205 | 352 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 5  | 15   | 8.86  | 11.99 | 6.85 | 39.2 | 0.03 | 82    | 217 | 353 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 4  | 16   | 8.41  | 11.25 | 6.77 | 39.2 | 0.03 | 76.5  | 154 | 353 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 3  | 17   | 7.92  | 10.48 | 6.71 | 39.7 | 0.03 | 70.8  | 215 | 352 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 2  | 18   | 7.29  | 8.63  | 6.62 | 43.9 | 0.03 | 62.4  | 146 | 352 |
| Mid Lake CDA | 2798 | ASDNBH | 7/8/98  | 11:24:33 | 5.6 | 1  | 19   | 7.25  | 8.59  | 6.63 | 43.9 | 0.03 | 62    | 555 | 351 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 19 | 0.3  | 9.18  | 24.55 | 7.93 | 54.1 | 0.03 | 109.8 | 56  | 334 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 18 | 1.8  | 9.26  | 23.15 | 7.9  | 53.6 | 0.03 | 107.8 | 134 | 334 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 17 | 2.8  | 9.35  | 22.64 | 7.9  | 53.7 | 0.03 | 107.9 | 123 | 335 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 16 | 3.8  | 9.68  | 22.09 | 7.97 | 53.3 | 0.03 | 110.5 | 219 | 335 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 15 | 4.8  | 9.98  | 20.97 | 8.02 | 51.4 | 0.03 | 111.5 | 125 | 336 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 14 | 5.8  | 10.2  | 19.95 | 8.01 | 50.1 | 0.03 | 111.7 | 125 | 337 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 13 | 6.8  | 10.52 | 18.74 | 7.85 | 49   | 0.03 | 112.5 | 155 | 342 |



|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|------|-------|-------|------|------|------|-------|-----|-----|
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 12 | 7.8  | 10.08 | 16.31 | 7.27 | 47.5 | 0.03 | 102.5 | 135 | 354 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 11 | 8.8  | 9.14  | 13.96 | 6.94 | 45.8 | 0.03 | 88.3  | 121 | 359 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 10 | 9.8  | 8.59  | 12.87 | 6.8  | 45   | 0.03 | 81    | 111 | 361 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 9  | 10.8 | 8.6   | 11.7  | 6.76 | 44.9 | 0.03 | 79    | 139 | 360 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 8  | 11.8 | 8.44  | 11.21 | 6.73 | 45.2 | 0.03 | 76.6  | 151 | 360 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 7  | 12.8 | 8.37  | 10.84 | 6.71 | 45.4 | 0.03 | 75.4  | 157 | 359 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 6  | 13.8 | 8.3   | 10.05 | 6.67 | 46   | 0.03 | 73.4  | 155 | 358 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 5  | 14.8 | 7.79  | 9.28  | 6.63 | 47.2 | 0.03 | 67.6  | 128 | 357 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 4  | 15.8 | 7.75  | 8.97  | 6.63 | 47.3 | 0.03 | 66.7  | 32  | 356 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 3  | 16.8 | 8.21  | 8.74  | 6.68 | 47.9 | 0.03 | 70.3  | 55  | 351 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 2  | 17.8 | 8.42  | 8.72  | 6.71 | 47.8 | 0.03 | 72    | 22  | 349 |
| Mid Lake CDA | 2998 | ASBH   | 7/20/98 | 11:13:27 | 8.7 | 1  | 18.8 | 8.49  | 8.66  | 6.77 | 47.9 | 0.03 | 72.6  | 222 | 343 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 13 | 0.3  | 8.5   | 24.59 | 7.73 | 59   | 0.04 | 101.6 | 48  | 343 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 12 | 1.9  | 8.56  | 23.87 | 7.73 | 58.8 | 0.04 | 101   | 39  | 343 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 11 | 3.4  | 8.59  | 23.64 | 7.71 | 59.1 | 0.04 | 100.9 | 49  | 343 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 10 | 4.9  | 8.7   | 23.49 | 7.7  | 58.8 | 0.04 | 102   | 110 | 344 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 9  | 6.6  | 8.7   | 23.4  | 7.62 | 59.2 | 0.04 | 101.8 | 104 | 347 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 8  | 7.8  | 9.57  | 20.65 | 7.39 | 52.7 | 0.03 | 106.2 | 54  | 355 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 7  | 9.8  | 10.06 | 19.06 | 7.33 | 49.8 | 0.03 | 108.1 | 46  | 356 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 6  | 11.1 | 9.85  | 17.61 | 7.12 | 48.5 | 0.03 | 102.8 | 25  | 361 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 5  | 12.4 | 9.14  | 15.99 | 6.95 | 47.8 | 0.03 | 92.2  | 103 | 365 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 4  | 14.1 | 7.96  | 12.74 | 6.76 | 47   | 0.03 | 74.8  | 108 | 368 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 3  | 15.4 | 7.03  | 11.16 | 6.69 | 47.7 | 0.03 | 63.7  | 46  | 368 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 2  | 17.1 | 6.06  | 9.58  | 6.66 | 49.1 | 0.03 | 52.9  | 56  | 368 |
| Mid Lake CDA | 3298 | DBAS   | 8/10/98 | 11:21:08 | 9.8 | 1  | 18.8 | 4.84  | 10.07 | 6.71 | 50.9 | 0.03 | 42.7  | 130 | 369 |
|              |      |        |         |          |     |    |      |       |       |      |      |      |       |     |     |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 20 | 0.3  | 9.16  | 22.36 | 7.92 | 56   | 0.04 | 105.6 | 105 | 334 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 19 | 1    | 9.25  | 21.95 | 7.98 | 55.3 | 0.04 | 105.7 | 107 | 334 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 18 | 2    | 9.25  | 21.78 | 7.96 | 55.3 | 0.04 | 105.4 | 113 | 337 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 17 | 3    | 9.25  | 21.54 | 7.92 | 55.2 | 0.04 | 104.9 | 100 | 336 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 16 | 4    | 9.24  | 21.46 | 7.87 | 55.3 | 0.04 | 104.6 | 102 | 338 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 15 | 5    | 9.25  | 21.39 | 7.82 | 55.3 | 0.04 | 104.6 | 44  | 339 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 14 | 6    | 9.19  | 21.33 | 7.79 | 55.4 | 0.04 | 103.8 | 56  | 340 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 13 | 7    | 9.14  | 21.4  | 7.73 | 55.1 | 0.04 | 103.4 | 51  | 343 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 12 | 8    | 9.13  | 21.32 | 7.58 | 55.3 | 0.04 | 103.1 | 28  | 349 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 11 | 9    | 9.04  | 20.84 | 7.34 | 53.5 | 0.03 | 101.1 | 113 | 355 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 10 | 10   | 9.31  | 19.37 | 7.11 | 49.2 | 0.03 | 101.2 | 115 | 361 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 9  | 11   | 8.07  | 16.51 | 6.82 | 45.9 | 0.03 | 82.7  | 107 | 368 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 8  | 12   | 7.71  | 14.82 | 6.73 | 45.6 | 0.03 | 76.2  | 152 | 370 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 7  | 13   | 6.98  | 13.2  | 6.63 | 44.8 | 0.03 | 66.6  | 224 | 372 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 6  | 14   | 6.93  | 11.79 | 6.54 | 45.6 | 0.03 | 64    | 30  | 374 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 5  | 15   | 6.46  | 10.72 | 6.5  | 46   | 0.03 | 58.2  | 37  | 375 |
| Mid Lake CDA | 3498 | ASDNGL | 8/25/98 | 12:13:57 | 7.5 | 4  | 16   | 6.34  | 9.9   | 6.48 | 46.5 | 0.03 | 56.1  | 123 | 375 |



|              |      |        |          |          |     |    |      |      |       |      |      |      |      |     |      |
|--------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|------|
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 19 | 0.4  | 9.11 | 12.65 | 7.32 | 57.5 | 0.04 | 84.7 | 59  | 385  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 18 | 1.3  | 9.12 | 12.66 | 7.31 | 57.7 | 0.04 | 84.8 | 42  | 387  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 17 | 2.3  | 9.11 | 12.66 | 7.32 | 57.7 | 0.04 | 84.7 | 119 | 387  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 16 | 3.3  | 9.1  | 12.66 | 7.29 | 57.7 | 0.04 | 84.6 | 129 | 390  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 15 | 4.3  | 9.11 | 12.63 | 7.32 | 57.4 | 0.04 | 84.7 | 109 | 388  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 14 | 5.3  | 9.1  | 12.63 | 7.31 | 57.7 | 0.04 | 84.6 | 53  | 389  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 13 | 6.3  | 9.11 | 12.6  | 7.3  | 57.7 | 0.04 | 84.6 | 117 | 389  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 12 | 7.3  | 9.13 | 12.56 | 7.3  | 57   | 0.04 | 84.7 | 59  | 390  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 11 | 8.3  | 9.11 | 12.58 | 7.3  | 57.3 | 0.04 | 84.5 | 131 | 390  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 10 | 9.3  | 9.13 | 12.55 | 7.3  | 56.8 | 0.04 | 84.7 | 47  | 390  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 9  | 10.3 | 9.13 | 12.55 | 7.29 | 57.3 | 0.04 | 84.7 | 57  | 391  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 8  | 11.3 | 9.07 | 12.55 | 7.27 | 57.3 | 0.04 | 84.1 | 51  | 392  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 7  | 12.3 | 9.03 | 12.5  | 7.26 | 56.6 | 0.04 | 83.6 | 124 | 392  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 6  | 13.3 | 8.98 | 12.46 | 7.22 | 56.7 | 0.04 | 83.1 | 104 | 393  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 5  | 14.3 | 8.76 | 11.81 | 7.17 | 56.7 | 0.04 | 79.9 | 151 | 395  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 4  | 15.3 | 7.23 | 11.2  | 6.91 | 55.1 | 0.04 | 65.1 | 215 | 399  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 3  | 16.3 | 4.21 | 10.12 | 6.56 | 50.9 | 0.03 | 36.9 | 152 | 409  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 2  | 17.3 | 5.14 | 8.73  | 6.48 | 49.2 | 0.03 | 43.6 | 315 | 411  |
| Mid Lake CDA | 4298 | DTASKB | 10/20/98 | 9:38:00  | 7.5 | 1  | 18.3 | 5.42 | 8.09  | 6.5  | 49.1 | 0.03 | 45.3 | 357 | 410  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 19 | 0.4  | 8.58 | 8.78  | 7.31 | 54.7 | 0.04 | 74.6 | 31  | 324  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 18 | 0.9  | 8.58 | 8.79  | 7.33 | 54.8 | 0.04 | 74.6 | 38  | 323  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 17 | 1.9  | 8.58 | 8.79  | 7.31 | 54.8 | 8.57 | 325  | 47  | 11.3 |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 16 | 2.9  | 8.55 | 8.78  | 7.28 | 54.8 | 0.04 | 74.3 | 111 | 325  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 15 | 3.9  | 8.57 | 8.78  | 7.27 | 55   | 0    | 8.56 | 58  | 4    |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 14 | 4.9  | 8.52 | 8.78  | 7.23 | 55.2 | 0.04 | 74   | 41  | 327  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 13 | 5.9  | 8.52 | 8.79  | 7.22 | 54.7 | 0.04 | 74.1 | 40  | 327  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 12 | 6.9  | 8.48 | 8.78  | 7.19 | 55.3 | 0.04 | 73.7 | 43  | 327  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 11 | 7.9  | 8.3  | 8.74  | 7.14 | 55.3 | 0.04 | 72.1 | 53  | 329  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 10 | 8.9  | 8.19 | 8.71  | 7.1  | 55.4 | 0.04 | 71   | 43  | 329  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 9  | 9.9  | 8.19 | 8.69  | 7.09 | 55.7 | 0.04 | 71.1 | 55  | 329  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 8  | 10.9 | 8.04 | 8.69  | 7.05 | 55.7 | 0.04 | 69.7 | 39  | 329  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 7  | 11.9 | 7.87 | 8.66  | 6.99 | 56   | 0.04 | 69.1 | 25  | 331  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 6  | 12.9 | 7.49 | 8.48  | 6.89 | 55.6 | 0.04 | 64.6 | 42  | 331  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 5  | 13.9 | 7.14 | 8.3   | 6.83 | 54.7 | 0.04 | 61.3 | 42  | 331  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 4  | 14.9 | 7.11 | 8.2   | 6.82 | 54.2 | 0.03 | 60.9 | 54  | 331  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 3  | 16   | 6.95 | 8.06  | 6.77 | 53.8 | 0.03 | 59.4 | 50  | 330  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 2  | 16.9 | 6.37 | 7.7   | 6.72 | 50.9 | 0.03 | 53.9 | 59  | 330  |
| Mid Lake CDA | 4698 | KBAS   | 11/17/98 | 13:20:00 | 5.3 | 1  | 17.9 | 6.07 | 7.45  | 6.71 | 49.5 | 0.03 | 51   | 505 | 328  |

| Location  | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-----------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Carey Bay | 1098  | DBRP     | 3/13/98 | 13:01:20 | 1.8        | 12       | 0.4       | 12.8                    | 5.07            | 6.92 | 44.8                 | 0.03 | 99.4                       | 104                | 391   |

|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 11 | 0.8  | 12.7 | 4.74  | 6.9  | 44.6 | 0.03 | 98.3  | 37  | 397 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 10 | 1.8  | 12.8 | 4.01  | 6.99 | 44.8 | 0.03 | 96.6  | 116 | 392 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 9  | 2.8  | 12.8 | 4.19  | 6.93 | 44.6 | 0.03 | 97.2  | 43  | 395 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 8  | 3.8  | 12.8 | 4.14  | 6.96 | 44.6 | 0.03 | 96.9  | 57  | 393 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 7  | 4.8  | 12.8 | 3.98  | 6.95 | 44.7 | 0.03 | 96.6  | 56  | 393 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 6  | 5.8  | 12.7 | 3.99  | 6.99 | 44.7 | 0.03 | 96.4  | 108 | 391 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 5  | 6.8  | 12.8 | 4.01  | 6.96 | 44.7 | 0.03 | 96.8  | 110 | 392 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 4  | 7.8  | 12.8 | 4.06  | 6.92 | 44.7 | 0.03 | 96.8  | 146 | 393 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 3  | 8.8  | 12.8 | 3.98  | 6.86 | 44.6 | 0.03 | 96.5  | 32  | 394 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 2  | 9.8  | 12.7 | 4.04  | 6.84 | 46.3 | 0.03 | 96.2  | 121 | 393 |
| Carey Bay | 1098 | DBRP   | 3/13/98 | 13:01:20 | 1.8 | 1  | 10.8 | 12.4 | 3.98  | 6.75 | 48.1 | 0.03 | 93.4  | 144 | 395 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 13 | 0.3  | 12.4 | 7.7   | 7.04 | 40.3 | 0.03 | 103.8 | 113 | 411 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 12 | 1.2  | 12.3 | 7.7   | 7.02 | 40.4 | 0.03 | 103.4 | 126 | 414 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 11 | 2.3  | 12.3 | 7.71  | 7.03 | 40.3 | 0.03 | 103.3 | 104 | 414 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 10 | 3.3  | 12.3 | 7.63  | 7    | 40   | 0.03 | 103   | 121 | 415 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 9  | 4.3  | 12.3 | 7.05  | 7.06 | 40.2 | 0.03 | 101.4 | 140 | 411 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 8  | 5.4  | 12.2 | 6.61  | 6.93 | 40.6 | 0.03 | 99.6  | 55  | 417 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 7  | 6.4  | 12.2 | 6.5   | 6.87 | 41   | 0.03 | 99.1  | 107 | 419 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 6  | 7.4  | 12.1 | 6.45  | 6.91 | 40.9 | 0.03 | 98.4  | 116 | 416 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 5  | 8.4  | 11.9 | 6.13  | 6.88 | 42.2 | 0.03 | 95.8  | 139 | 416 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 4  | 9.4  | 11.8 | 5.97  | 6.84 | 42.7 | 0.03 | 94.4  | 115 | 417 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 3  | 10.4 | 11.8 | 5.85  | 6.8  | 42.9 | 0.03 | 94.1  | 125 | 418 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 2  | 11.4 | 11.8 | 5.75  | 6.84 | 43   | 0.03 | 94.1  | 46  | 415 |
| Carey Bay | 1498 | DBAS   | 4/8/98  | 14:24:27 | 2.2 | 1  | 12.5 | 11.7 | 5.75  | 6.79 | 43.2 | 0.03 | 93.2  | 203 | 418 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 13 | 0.3  | 11.7 | 13.63 | 7.2  | 43.7 | 0.03 | 111.9 | 120 | 391 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 12 | 1.6  | 12.3 | 9.86  | 7.24 | 43.2 | 0.03 | 107.9 | 108 | 392 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 11 | 2.6  | 12.4 | 8.22  | 7.24 | 42.8 | 0.03 | 104.8 | 105 | 392 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 10 | 3.5  | 12.4 | 7.78  | 7.13 | 42.2 | 0.03 | 103.5 | 50  | 396 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 9  | 4.5  | 12.2 | 7.42  | 7.15 | 42.4 | 0.03 | 101.3 | 115 | 393 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 8  | 5.6  | 12   | 7.14  | 7    | 42.9 | 0.03 | 98.8  | 142 | 398 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 7  | 6.6  | 11.9 | 6.97  | 7.03 | 43.4 | 0.03 | 97.4  | 154 | 394 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 6  | 7.7  | 11.8 | 6.79  | 6.95 | 43.9 | 0.03 | 96.5  | 118 | 395 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 5  | 8.6  | 11.7 | 6.54  | 6.96 | 44.5 | 0.03 | 95    | 210 | 393 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 4  | 9.7  | 11.7 | 6.13  | 6.94 | 45.7 | 0.03 | 93.7  | 116 | 391 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 3  | 10.6 | 11.7 | 6.08  | 6.92 | 46.2 | 0.03 | 93.6  | 131 | 390 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 2  | 11.5 | 11.6 | 5.9   | 6.85 | 47   | 0.03 | 92.9  | 124 | 391 |
| Carey Bay | 1698 | DBASRP | 4/21/98 | 15:07:34 | 2.8 | 1  | 12.6 | 11.6 | 5.87  | 6.93 | 47.4 | 0.03 | 92.2  | 217 | 385 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 10 | 0.3  | 11.2 | 11.87 | 7.19 | 33.7 | 0.02 | 104.6 | 46  | 380 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 9  | 1    | 11.2 | 11.86 | 7.18 | 33.7 | 0.02 | 104.3 | 124 | 380 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 8  | 2.5  | 11.2 | 11.86 | 7.13 | 33.7 | 0.02 | 104.4 | 55  | 381 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 7  | 4    | 11.2 | 11.82 | 7.15 | 33.9 | 0.02 | 104.5 | 48  | 379 |

|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 6  | 5.5  | 11.2 | 11.66 | 7.14 | 33.8 | 0.02 | 104.1 | 49  | 376 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 5  | 7    | 11.2 | 11.02 | 7.07 | 34.1 | 0.02 | 102.7 | 114 | 377 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 4  | 8.5  | 11.3 | 10.82 | 7.06 | 34.5 | 0.02 | 102.8 | 117 | 375 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 3  | 10   | 11.2 | 9.99  | 6.99 | 35.9 | 0.02 | 100.4 | 40  | 374 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 2  | 11.5 | 11.3 | 9.56  | 6.9  | 37.9 | 0.02 | 99.6  | 40  | 377 |
| Carey Bay | 2198 | ASJLDT | 5/26/98 | 13:24:59 | 3.9 | 1  | 13   | 11.2 | 9.3   | 6.9  | 38.8 | 0.02 | 98.8  | 202 | 375 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 13 | 0.3  | 10.9 | 16.57 | 7.64 | 38.5 | 0.02 | 112.2 | 50  | 375 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 12 | 1.3  | 10.9 | 16.25 | 7.62 | 38.5 | 0.02 | 111   | 52  | 378 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 11 | 2.3  | 11.1 | 14.81 | 7.6  | 38.5 | 0.02 | 109.8 | 56  | 379 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 10 | 3.3  | 11.1 | 14.54 | 7.55 | 38.4 | 0.02 | 109.4 | 43  | 381 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 9  | 4.3  | 11.2 | 14.21 | 7.49 | 38.7 | 0.02 | 108.7 | 56  | 383 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 8  | 5.3  | 11.1 | 13.56 | 7.4  | 38.6 | 0.02 | 106.8 | 55  | 385 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 7  | 6.3  | 11.1 | 13.14 | 7.34 | 38.6 | 0.02 | 105.5 | 56  | 386 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 6  | 7.3  | 10.9 | 12.55 | 7.25 | 38.2 | 0.02 | 102   | 41  | 387 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 5  | 8.3  | 10.8 | 12.45 | 7.23 | 37.9 | 0.02 | 101.6 | 57  | 387 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 4  | 9.3  | 10.8 | 12.14 | 7.2  | 37.2 | 0.02 | 100.3 | 109 | 388 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 3  | 10.3 | 10.2 | 11.58 | 7.11 | 37.6 | 0.02 | 93.9  | 58  | 389 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 2  | 11.3 | 10.3 | 10.86 | 7.08 | 37.8 | 0.02 | 93.1  | 58  | 389 |
| Carey Bay | 2398 | DTJLAS | 6/8/98  | 12:38:40 | 3.1 | 1  | 12.3 | 10.2 | 10.43 | 7.06 | 38.6 | 0.02 | 91.1  | 143 | 391 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 14 | 0.2  | 10.1 | 19.38 | 7.55 | 42.7 | 0.03 | 110   | 32  | 355 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 13 | 1.6  | 10.1 | 19.16 | 7.54 | 42.8 | 0.03 | 109.2 | 107 | 356 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 12 | 2.6  | 10.6 | 16.73 | 7.58 | 43.5 | 0.03 | 109.4 | 59  | 358 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 11 | 3.6  | 10.7 | 16.61 | 7.59 | 43.6 | 0.03 | 109.7 | 116 | 357 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 10 | 4.6  | 10.8 | 16.39 | 7.6  | 45.6 | 0.03 | 110.8 | 46  | 357 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 9  | 5.6  | 10.9 | 16.11 | 7.6  | 46.3 | 0.03 | 111   | 108 | 357 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 8  | 6.6  | 10.8 | 15.86 | 7.49 | 46.7 | 0.03 | 109   | 52  | 359 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 7  | 7.6  | 10.8 | 15.54 | 7.45 | 43.6 | 0.03 | 108.1 | 137 | 359 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 6  | 8.6  | 10.9 | 14.57 | 7.38 | 42.6 | 0.03 | 107.1 | 104 | 359 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 5  | 9.6  | 11   | 13.89 | 7.33 | 42.2 | 0.03 | 106.1 | 115 | 360 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 4  | 10.6 | 10.9 | 13.17 | 7.23 | 40.3 | 0.03 | 103.7 | 114 | 361 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 3  | 11.6 | 10.5 | 12.79 | 7.15 | 39.7 | 0.03 | 99.6  | 118 | 361 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 2  | 12.6 | 10.6 | 12.16 | 7.13 | 39.7 | 0.03 | 98.7  | 155 | 359 |
| Carey Bay | 2598 | DTASBH | 6/22/98 | 13:41:37 | 4.3 | 1  | 13.6 | 10   | 11.29 | 7.05 | 38.9 | 0.02 | 91.7  | 129 | 358 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 13 | 0.3  | 9.02 | 24.79 | 7.68 | 51.4 | 0.03 | 108.2 | 200 | 338 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 12 | 1.7  | 10.3 | 20.18 | 7.78 | 51.1 | 0.03 | 113.6 | 103 | 340 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 11 | 2.7  | 10.6 | 18.99 | 7.84 | 50.5 | 0.03 | 114   | 103 | 339 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 10 | 3.7  | 10.7 | 17.88 | 7.8  | 50.5 | 0.03 | 112.3 | 126 | 341 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 9  | 4.7  | 11   | 17.3  | 7.96 | 50.2 | 0.03 | 113.5 | 114 | 339 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 8  | 5.7  | 10.9 | 17.1  | 7.81 | 50.3 | 0.03 | 112.3 | 125 | 342 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 7  | 6.7  | 10.4 | 16.41 | 7.54 | 48.2 | 0.03 | 105.6 | 123 | 346 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 6  | 7.7  | 10.6 | 16.02 | 7.53 | 47.7 | 0.03 | 107.1 | 138 | 346 |

|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-----------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 5  | 8.7  | 10.4 | 15.67 | 7.4  | 47.2 | 0.03 | 104   | 113 | 347 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 4  | 9.7  | 10.3 | 15.49 | 7.37 | 47.2 | 0.03 | 103.2 | 119 | 347 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 3  | 10.7 | 10.2 | 15.32 | 7.29 | 47   | 0.03 | 101.6 | 143 | 348 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 2  | 11.7 | 9.97 |       | 7.21 | 46.2 | 0.03 | 98.2  | 246 | 348 |
| Carey Bay | 2798 | DTASBH | 7/7/98  | 12:50:28 | 5.1 | 1  | 12.7 | 9.41 | 14.14 | 7.05 | 45.9 | 0.03 | 91.2  | 218 | 348 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 14 | 0.5  | 8.93 | 23.64 | 7.79 | 51.7 | 0.03 | 105   | 120 | 395 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 13 | 1.5  | 9.24 | 22.7  | 7.86 | 51.2 | 0.03 | 106.7 | 124 | 398 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 12 | 2.5  | 9.42 | 22.29 | 7.95 | 50.9 | 0.03 | 108   | 102 | 398 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 11 | 3.5  | 10.2 | 20.83 | 8.11 | 48.7 | 0.03 | 113.2 | 152 | 398 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 10 | 4.5  | 10.4 | 19.82 | 8.02 | 47.5 | 0.03 | 113.4 | 110 | 403 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 9  | 5.5  | 10.3 | 19.75 | 7.91 | 47.7 | 0.03 | 112.1 | 135 | 406 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 8  | 6.5  | 10.6 | 19.06 | 7.93 | 46.7 | 0.03 | 113.7 | 51  | 408 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 7  | 7.5  | 10.5 | 18.58 | 7.7  | 46.3 | 0.03 | 111.4 | 204 | 416 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 6  | 8.5  | 10.1 | 17.69 | 7.32 | 45.7 | 0.03 | 105.5 | 44  | 427 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 5  | 9.5  | 9.84 | 16.27 | 7.09 | 45.4 | 0.03 | 100   | 131 | 431 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 4  | 10.5 | 9.27 | 15.45 | 6.94 | 45.1 | 0.03 | 92.6  | 136 | 434 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 3  | 11.5 | 9.16 | 13.88 | 6.83 | 43.5 | 0.03 | 88.4  | 59  | 438 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 2  | 12.5 | 8.32 | 12.74 | 6.65 | 42.9 | 0.03 | 78.2  | 40  | 445 |
| Carey Bay | 2998 | ASBH   | 7/20/98 | 11:26:44 | 8.6 | 1  | 13.5 | 7.51 | 12.15 | 6.59 | 42.7 | 0.03 | 69.7  | 252 | 447 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 14 | 0.6  | 8.54 | 24.09 | 7.66 | 58.5 | 0.04 | 101.1 | 52  | 346 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 13 | 1.7  | 8.59 | 23.62 | 7.63 | 58.3 | 0.04 | 100.9 | 49  | 347 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 12 | 2.7  | 8.67 | 23.42 | 7.62 | 57.6 | 0.04 | 101.4 | 40  | 347 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 11 | 3.6  | 8.81 | 23.13 | 7.63 | 57.7 | 0.04 | 102.5 | 124 | 348 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 10 | 4.5  | 8.84 | 22.99 | 7.59 | 57   | 0.04 | 102.6 | 51  | 350 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 9  | 5.6  | 9.06 | 22.22 | 7.52 | 55.6 | 0.04 | 103.7 | 54  | 353 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 8  | 6.7  | 9.38 | 21.16 | 7.46 | 53.2 | 0.03 | 105.1 | 58  | 355 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 7  | 7.8  | 9.63 | 20.32 | 7.4  | 50.8 | 0.03 | 106.2 | 39  | 356 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 6  | 8.8  | 9.61 | 19.8  | 7.37 | 50.7 | 0.03 | 104.9 | 56  | 357 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 5  | 9.7  | 9.69 | 19.5  | 7.31 | 50.4 | 0.03 | 105.1 | 104 | 358 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 4  | 10.8 | 9.59 | 18.65 | 7.18 | 49.5 | 0.03 | 102.2 | 114 | 361 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 3  | 11.8 | 8.84 | 17.25 | 6.96 | 49.2 | 0.03 | 91.6  | 41  | 366 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 2  | 12.7 | 7.87 | 15.27 | 6.84 | 48.4 | 0.03 | 78.2  | 144 | 368 |
| Carey Bay | 3298 | DBAS   | 8/10/98 | 11:43:00 | 9.1 | 1  | 13.7 | 6.62 | 13.55 | 6.77 | 48.8 | 0.03 | 63.3  | 23  | 369 |
|           |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 14 | 0.2  | 9.14 | 22.49 | 7.89 | 55.5 | 0.04 | 105.6 | 101 | 332 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 13 | 1.2  | 9.16 | 22.49 | 7.93 | 55.5 | 0.04 | 105.8 | 101 | 329 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 12 | 2.2  | 9.19 | 22.4  | 7.93 | 55.6 | 0.04 | 105.9 | 40  | 329 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 11 | 3.2  | 9.28 | 21.85 | 7.94 | 55.4 | 0.04 | 105.9 | 56  | 329 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 10 | 4.2  | 9.2  | 21.72 | 7.89 | 55.1 | 0.04 | 104.7 | 52  | 331 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 9  | 5.2  | 9.25 | 21.55 | 7.82 | 55.3 | 0.04 | 105   | 59  | 333 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 8  | 6.2  | 8.89 | 21.49 | 7.61 | 55.3 | 0.04 | 100.7 | 56  | 339 |
| Carey Bay | 3498 | ASDNGL | 8/25/98 | 13:20:35 | 6.8 | 7  | 7.2  | 9.04 | 21.37 | 7.62 | 55.3 | 0.04 | 102.2 | 130 | 339 |

|           |      |        |          |          |     |    |      |      |       |      |      |      |       |     |     |
|-----------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 6  | 8.2  | 8.65 | 21.26 | 7.33 | 55.2 | 0.04 | 97.6  | 115 | 346 |
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 5  | 9.2  | 8.43 | 19.87 | 7.06 | 50.7 | 0.03 | 92.6  | 108 | 352 |
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 4  | 10.2 | 8.05 | 18.94 | 6.96 | 48.9 | 0.03 | 86.8  | 210 | 354 |
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 3  | 11.2 | 7.39 | 16.69 | 6.77 | 46.5 | 0.03 | 76    | 125 | 358 |
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 2  | 12.2 | 5.61 | 14.03 | 6.59 | 47.5 | 0.03 | 54.5  | 128 | 361 |
| Carey Bay | 3498 | ASDNGL | 8/25/98  | 13:20:35 | 6.8 | 1  | 13.2 | 4.67 | 12.41 | 6.54 | 48.5 | 0.03 | 43.8  | 330 | 362 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 14 | 0.3  | 8.96 | 22.9  | 7.78 | 58.7 | 0.04 | 104.4 | 50  | 300 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 13 | 1.3  | 8.98 | 22.43 | 7.87 | 58.6 | 0.04 | 103.8 | 38  | 296 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 12 | 2.3  | 9.23 | 21.48 | 7.93 | 58.7 | 0.04 | 104.8 | 30  | 295 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 11 | 3.3  | 9.21 | 21.26 | 7.92 | 58.4 | 0.04 | 104.1 | 133 | 296 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 10 | 4.3  | 9.23 | 21.14 | 7.86 | 58.4 | 0.04 | 104.1 | 107 | 297 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 9  | 5.3  | 9.25 | 21.05 | 7.83 | 58.4 | 0.04 | 104.2 | 118 | 297 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 8  | 6.3  | 9.19 | 21.07 | 7.73 | 58.3 | 0.04 | 103.5 | 147 | 300 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 7  | 7.3  | 9.06 | 20.97 | 7.6  | 58   | 0.04 | 101.8 | 323 | 303 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 6  | 8.3  | 8.78 | 20.27 | 7.17 | 55.7 | 0.04 | 97.3  | 57  | 319 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 5  | 9.3  | 7.54 | 19.09 | 6.9  | 52.3 | 0.03 | 81.7  | 120 | 325 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 4  | 10.3 | 7.08 | 16.44 | 6.76 | 47.4 | 0.03 | 72.6  | 111 | 328 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 3  | 11.3 | 6.29 | 15.04 | 6.65 | 48.5 | 0.03 | 62.6  | 115 | 328 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 2  | 12.3 | 4.6  | 11.46 | 0    | 49.9 | 0.03 | 42.2  | 204 | 329 |
| Carey Bay | 3598 | DTAS   | 9/3/98   | 14:16:15 | 6.8 | 1  | 13.3 | 4.88 | 11.2  | 6.57 | 49.5 | 0.03 | 44.6  | 159 | 317 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 13 | 0.3  | 9.04 | 19.23 | 7.66 | 61.1 | 0.04 | 97.8  | 56  | 336 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 12 | 1.9  | 9.14 | 17.95 | 7.62 | 61   | 0.04 | 96.3  | 44  | 339 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 11 | 2.9  | 9.1  | 17.76 | 7.6  | 61.1 | 0.04 | 95.5  | 53  | 340 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 10 | 3.9  | 9.1  | 17.76 | 7.55 | 61.2 | 0.04 | 95.5  | 111 | 342 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 9  | 4.9  | 8.89 | 17.61 | 7.47 | 60.8 | 0.04 | 93    | 43  | 344 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 8  | 5.9  | 8.82 | 17.57 | 7.43 | 60.7 | 0.04 | 92.2  | 28  | 346 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 7  | 6.9  | 8.81 | 17.57 | 7.42 | 60.9 | 0.04 | 92.1  | 104 | 345 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 6  | 8    | 8.8  | 17.56 | 7.39 | 61.3 | 0.04 | 92    | 53  | 347 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 5  | 8.9  | 8.78 | 17.56 | 7.35 | 60.6 | 0.04 | 91.8  | 112 | 348 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 4  | 9.9  | 8.49 | 17.51 | 7.23 | 60.7 | 0.04 | 88.6  | 34  | 351 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 3  | 11   | 8.13 | 17.37 | 7.17 | 61   | 0.04 | 84.7  | 154 | 352 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 2  | 11.9 | 7.35 | 16.91 | 6.98 | 60   | 0.04 | 75.8  | 128 | 356 |
| Carey Bay | 3998 | ASKB   | 9/30/98  | 14:15:00 | 6   | 1  | 12.9 | 5.88 | 16.12 | 6.86 | 57.2 | 0.04 | 59.7  | 636 | 358 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |       |     |     |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 13 | 0.4  | 9.13 | 13.25 | 7.35 | 58.7 | 0.04 | 86.2  | 30  | 334 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 12 | 1.4  | 9.04 | 12.99 | 7.36 | 58.6 | 0.04 | 84.8  | 245 | 334 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 11 | 2.4  | 9    | 12.78 | 7.35 | 58.6 | 0.04 | 84.1  | 133 | 333 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 10 | 3.4  | 9.02 | 12.71 | 7.35 | 58.4 | 0.04 | 84    | 46  | 332 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 9  | 4.4  | 9    | 12.68 | 7.32 | 58.7 | 0.04 | 83.8  | 49  | 333 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 8  | 5.4  | 8.98 | 12.68 | 7.31 | 58.6 | 0.04 | 83.6  | 39  | 333 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 7  | 6.4  | 8.99 | 12.68 | 7.3  | 58.6 | 0.04 | 83.7  | 42  | 333 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 6  | 7.4  | 8.98 | 12.66 | 7.29 | 58.7 | 0.04 | 83.6  | 35  | 333 |

|           |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
|-----------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 5  | 8.4  | 8.93 | 12.66 | 7.29 | 58.6 | 0.04 | 83.2 | 109 | 332 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 4  | 9.4  | 8.94 | 12.65 | 7.29 | 58.2 | 0.04 | 83.2 | 59  | 330 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 3  | 10.4 | 8.92 | 12.65 | 7.3  | 58.8 | 0.04 | 83   | 103 | 328 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 2  | 11.4 | 8.98 | 12.6  | 7.3  | 58.4 | 0.04 | 83.5 | 56  | 327 |
| Carey Bay | 4298 | DTASJB | 10/19/98 | 12:57:51 | 6.2 | 1  | 12.4 | 8.96 | 12.56 | 7.3  | 58.7 | 0.04 | 83.2 | 624 | 329 |
|           |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 12 | 0.6  | 8.25 | 8.47  | 7.21 | 55   | 0.04 | 71.1 | 26  | 320 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 11 | 1.7  | 8.23 | 8.45  | 7.19 | 54.9 | 0.04 | 71   | 37  | 321 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 10 | 2.7  | 8.2  | 8.46  | 7.14 | 54.9 | 0.04 | 70.7 | 35  | 323 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 9  | 3.7  | 8.19 | 8.47  | 7.14 | 55   | 0.04 | 70.6 | 37  | 323 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 8  | 4.7  | 8.2  | 8.47  | 7.14 | 54.9 | 0.04 | 70.7 | 46  | 322 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 7  | 5.7  | 8.19 | 8.47  | 0    | 55.3 | 0.04 | 70.6 | 115 | 322 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 6  | 6.7  | 8.13 | 8.45  | 7.11 | 54.9 | 0.04 | 70.1 | 103 | 322 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 5  | 7.7  | 8.17 | 8.45  | 7.1  | 55.4 | 0.04 | 70.5 | 51  | 322 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 4  | 8.7  | 8.15 | 8.45  | 7.08 | 54.9 | 0.04 | 70.3 | 130 | 321 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 3  | 9.7  | 7.79 | 8.3   | 7.01 | 54.5 | 0.03 | 66.9 | 57  | 320 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 2  | 10.8 | 7.61 | 8.2   | 6.99 | 54.3 | 0.03 | 65.2 | 105 | 318 |
| Carey Bay | 4698 | KBAS   | 11/17/98 | 12:50:00 | 4.8 | 1  | 11.7 | 7.65 | 8.17  | 7.02 | 55.1 | 0.04 | 65.5 | 659 | 320 |

| Location       | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|----------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 15       | 0.5       | 12.6                    | 7.37            | 7    | 41.8                 | 0.03 | 105.6                      | 44                 | 428   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 14       | 1.5       | 12.6                    | 6.79            | 6.98 | 39.5                 | 0.03 | 103.7                      | 56                 | 431   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 13       | 2.5       | 12.6                    | 6.41            | 6.99 | 38.7                 | 0.02 | 102.9                      | 101                | 431   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 12       | 3.5       | 12.6                    | 6.36            | 6.97 | 38.9                 | 0.02 | 102.8                      | 32                 | 432   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 11       | 4.5       | 12.6                    | 6.31            | 6.91 | 38.8                 | 0.02 | 102.6                      | 32                 | 435   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 10       | 5.5       | 12.6                    | 6.26            | 6.91 | 38.9                 | 0.02 | 102.4                      | 39                 | 435   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 9        | 6.5       | 12.6                    | 6.25            | 6.87 | 39                   | 0.02 | 102.3                      | 31                 | 438   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 8        | 7.5       | 12.5                    | 6.22            | 6.9  | 39.1                 | 0.03 | 101.4                      | 103                | 436   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 7        | 8.5       | 12.5                    | 6.17            | 6.83 | 39                   | 0.02 | 101.6                      | 43                 | 439   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 6        | 9.5       | 12.4                    | 6.07            | 6.83 | 39.1                 | 0.03 | 100.6                      | 46                 | 439   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 5        | 10.5      | 12.5                    | 6.05            | 6.77 | 39                   | 0.03 | 100.6                      | 25                 | 442   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 4        | 11.5      | 12.4                    | 5.85            | 6.75 | 39.3                 | 0.03 | 99.7                       | 48                 | 443   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 3        | 12.5      | 12.3                    | 5.77            | 6.73 | 40                   | 0.03 | 98.6                       | 34                 | 445   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 2        | 13.5      | 12.2                    | 5.77            | 6.7  | 41.1                 | 0.03 | 97.7                       | 44                 | 447   |
| Conkling Point | 1398  | DBRP     | 4/3/98  | 10:42:23 | 3.1        | 1        | 14.5      | 11.8                    | 5.8             | 6.67 | 43.4                 | 0.03 | 95                         | 228                | 449   |
|                |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 16       | 0.3       | 11.9                    | 11.12           | 7.08 | 42.8                 | 0.03 | 107.7                      | 100                | 402   |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 15       | 0.8       | 11.8                    | 10.68           | 7.13 | 42.6                 | 0.03 | 105.9                      | 132                | 403   |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 14       | 1.7       | 12                      | 8.65            | 7.09 | 42.2                 | 0.03 | 103                        | 121                | 404   |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 13       | 2.8       | 11.9                    | 7.88            | 7.08 | 42.4                 | 0.03 | 100.1                      | 105                | 407   |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 12       | 3.7       | 11.9                    | 7.35            | 6.99 | 42.7                 | 0.03 | 98.5                       | 56                 | 410   |
| Conkling Point | 1698  | DBASRP   | 4/21/98 | 14:06:07 | 2.6        | 11       | 4.8       | 11.8                    | 7.14            | 6.9  | 42.8                 | 0.03 | 97.2                       | 125                | 414   |



|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|----------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 10 | 5.8  | 11.8 | 7.01  | 6.95 | 42.8 | 0.03 | 97    | 145 | 410 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 9  | 6.8  | 11.7 | 6.71  | 6.96 | 43.2 | 0.03 | 95.4  | 142 | 407 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 8  | 7.8  | 11.7 | 6.64  | 6.89 | 43.5 | 0.03 | 95.5  | 140 | 409 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 7  | 8.8  | 11.7 | 6.51  | 6.85 | 44.2 | 0.03 | 94.9  | 118 | 410 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 6  | 9.8  | 11.7 | 6.38  | 6.9  | 44.6 | 0.03 | 94.6  | 142 | 405 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 5  | 10.8 | 11.7 | 6.17  | 6.85 | 45.5 | 0.03 | 94    | 145 | 406 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 4  | 11.8 | 11.7 | 5.92  | 6.83 | 46.8 | 0.03 | 93.1  | 104 | 404 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 3  | 12.7 | 11.3 | 5.89  | 6.77 | 47.4 | 0.03 | 90.3  | 117 | 405 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 2  | 13.7 | 11.1 | 5.8   | 6.72 | 47.9 | 0.03 | 88.5  | 113 | 406 |
| Conkling Point | 1698 | DBASRP | 4/21/98 | 14:06:07 | 2.6 | 1  | 14.8 | 11   | 5.59  | 6.72 | 49.3 | 0.03 | 87.3  | 325 | 406 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 11 | 0.3  | 11.2 | 11.87 | 7.08 | 33.4 | 0.02 | 104.3 | 23  | 385 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 10 | 1.5  | 11.2 | 11.89 | 7.15 | 33.5 | 0.02 | 104.3 | 32  | 379 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 9  | 3    | 11.2 | 11.45 | 7.09 | 33.4 | 0.02 | 103.5 | 44  | 379 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 8  | 4.5  | 11.2 | 11.1  | 7.15 | 33.3 | 0.02 | 103.2 | 38  | 374 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 7  | 6    | 11.3 | 10.99 | 7.07 | 33.2 | 0.02 | 103   | 46  | 377 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 6  | 7.5  | 11.3 | 10.74 | 6.99 | 33.1 | 0.02 | 102.6 | 33  | 378 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 5  | 9    | 11.2 | 10.25 | 6.99 | 32.6 | 0.02 | 100.7 | 40  | 376 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 4  | 10.5 | 11.2 | 9.45  | 6.9  | 32.5 | 0.02 | 98.6  | 44  | 378 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 3  | 12   | 11.2 | 9.25  | 6.89 | 32   | 0.02 | 98.4  | 104 | 377 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 2  | 13.5 | 11.3 | 9.24  | 6.93 | 32   | 0.02 | 99    | 107 | 373 |
| Conkling Point | 2198 | ASJLDT | 5/26/98 | 13:45:03 | 3.8 | 1  | 15.2 | 11.2 | 9.15  | 6.9  | 32.8 | 0.02 | 98.5  | 153 | 376 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 11 | 0.4  | 10.7 | 15.62 | 7.52 | 38.9 | 0.02 | 108   | 28  | 362 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 10 | 1.5  | 10.7 | 15.51 | 7.5  | 38.9 | 0.02 | 107.3 | 101 | 363 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 9  | 3    | 10.6 | 14.84 | 7.43 | 38.5 | 0.02 | 105   | 125 | 365 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 8  | 4.5  | 10.8 | 13.06 | 7.3  | 38   | 0.02 | 102.6 | 109 | 367 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 7  | 6    | 10.6 | 12.73 | 7.24 | 37.9 | 0.02 | 100   | 117 | 367 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 6  | 7.5  | 10.8 | 12.22 | 7.22 | 37.7 | 0.02 | 100.9 | 55  | 366 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 5  | 9    | 10.7 | 12    | 7.19 | 37   | 0.02 | 99.3  | 132 | 366 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 4  | 10.5 | 10.5 | 11.79 | 7.13 | 36.7 | 0.02 | 97.4  | 46  | 363 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 3  | 12   | 10.7 | 10.77 | 7.1  | 35.9 | 0.02 | 96.8  | 130 | 363 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 2  | 13.5 | 10.3 | 9.94  | 7    | 37.9 | 0.02 | 91.1  | 158 | 362 |
| Conkling Point | 2398 | DTJLAS | 6/8/98  | 13:00:22 | 2.5 | 1  | 15.1 | 9.16 | 8.71  | 6.85 | 42.2 | 0.03 | 78.7  | 242 | 361 |
|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 16 | 0.3  | 10.1 | 19.54 | 7.54 | 42.2 | 0.03 | 109.9 | 28  | 355 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 15 | 1.1  | 10.1 | 19.28 | 7.54 | 42.1 | 0.03 | 109.2 | 100 | 356 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 14 | 2.1  | 10.3 | 18.39 | 7.57 | 41.9 | 0.03 | 109.3 | 57  | 356 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 13 | 3.1  | 10.5 | 17.56 | 7.61 | 41.7 | 0.03 | 109.5 | 135 | 356 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 12 | 4.1  | 10.6 | 16.63 | 7.6  | 42   | 0.03 | 108.7 | 115 | 357 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 11 | 5.1  | 10.8 | 16.09 | 7.61 | 51   | 0.03 | 109.8 | 214 | 358 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 10 | 6.1  | 10.9 | 15.25 | 7.46 | 49.8 | 0.03 | 108.8 | 159 | 362 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 9  | 7.1  | 10.9 | 14.79 | 7.45 | 43.6 | 0.03 | 108.1 | 53  | 360 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 8  | 8.1  | 11   | 14.06 | 7.36 | 42.9 | 0.03 | 106.6 | 108 | 362 |

|                |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|----------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 7  | 9.1  | 11   | 13.81 | 7.31 | 41.7 | 0.03 | 106.2 | 206 | 362 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 6  | 10.1 | 11   | 13.41 | 7.26 | 41.5 | 0.03 | 105.8 | 131 | 362 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 5  | 11.1 | 10.8 | 12.99 | 7.12 | 40   | 0.03 | 102.7 | 114 | 363 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 4  | 12.1 | 10.2 | 12.57 | 6.98 | 37.5 | 0.02 | 95.6  | 153 | 364 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 3  | 13.1 | 10.2 | 12.33 | 6.95 | 38.1 | 0.02 | 95.7  | 110 | 362 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 2  | 14.1 | 9.61 | 11.64 | 6.87 | 37.4 | 0.02 | 88.5  | 227 | 362 |
| Conkling Point | 2598 | DTASBH | 6/22/98 | 12:17:57 | 4.3 | 1  | 15.1 | 8.96 | 11.16 | 6.81 | 37.8 | 0.02 | 81.5  | 207 | 376 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 16 | 0.3  | 10.1 | 24.35 | 8.72 | 51.2 | 0.03 | 120.3 | 56  | 307 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 15 | 1.3  | 9.63 | 20.88 | 7.56 | 50.7 | 0.03 | 107.3 | 108 | 334 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 14 | 2.3  | 9.85 | 19.32 | 7.46 | 50.9 | 0.03 | 106.4 | 115 | 335 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 13 | 3.3  | 10.5 | 18.17 | 7.7  | 49.9 | 0.03 | 111   | 130 | 332 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 12 | 4.3  | 10.6 | 17.57 | 7.81 | 50   | 0.03 | 110.9 | 118 | 328 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 11 | 5.3  | 10.7 | 16.39 | 7.71 | 48.5 | 0.03 | 108.5 | 314 | 331 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 10 | 6.3  | 10.6 | 16.12 | 7.61 | 48.2 | 0.03 | 107.5 | 145 | 332 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 9  | 7.2  | 10.5 | 15.86 | 7.5  | 47.5 | 0.03 | 105.7 | 229 | 332 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 8  | 8.3  | 10.3 | 15.44 | 7.34 | 46.8 | 0.03 | 102.9 | 116 | 333 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 7  | 9.3  | 10.2 | 15.04 | 7.24 | 46.4 | 0.03 | 100.5 | 123 | 333 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 6  | 10.3 | 10.1 | 14.94 | 7.2  | 46.3 | 0.03 | 99.7  | 50  | 333 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 5  | 11.3 | 10.1 | 14.72 | 7.14 | 46.5 | 0.03 | 98.6  | 123 | 332 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 4  | 12.3 | 9.79 | 14.21 | 7.04 | 45.8 | 0.03 | 95.1  | 120 | 332 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 3  | 13.3 | 9.63 | 13.86 | 6.96 | 44.9 | 0.03 | 92.8  | 136 | 330 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 2  | 14.3 | 8.31 | 13.13 | 6.78 | 45   | 0.03 | 78.8  | 116 | 328 |
| Conkling Point | 2798 | DTASBH | 7/7/98  | 12:07:14 | 3.2 | 1  | 15.3 | 7.28 | 11.98 | 6.71 | 45.5 | 0.03 | 67.2  | 211 | 316 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 15 | 0.9  | 8.96 | 23.31 | 7.81 | 51.3 | 0.03 | 104.7 | 143 | 354 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 14 | 1.9  | 9.12 | 22.92 | 7.92 | 51.4 | 0.03 | 105.8 | 117 | 354 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 13 | 2.9  | 9.14 | 22.59 | 7.86 | 51.2 | 0.03 | 105.3 | 112 | 356 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 12 | 3.9  | 9.91 | 21.23 | 7.94 | 49.4 | 0.03 | 111.3 | 126 | 355 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 11 | 4.9  | 10.2 | 20.55 | 7.99 | 48.2 | 0.03 | 112.5 | 103 | 355 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 10 | 5.9  | 10.3 | 19.8  | 7.95 | 47.5 | 0.03 | 112.8 | 118 | 357 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 9  | 6.9  | 10.5 | 18.58 | 7.77 | 46.7 | 0.03 | 111.4 | 128 | 360 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 8  | 7.9  | 10.4 | 16.93 | 7.48 | 45.5 | 0.03 | 107.4 | 125 | 366 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 7  | 8.9  | 10.1 | 16.14 | 7.27 | 44.8 | 0.03 | 102.6 | 132 | 369 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 6  | 9.9  | 9.34 | 14.94 | 7    | 44.3 | 0.03 | 92.2  | 132 | 372 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 5  | 10.9 | 8.95 | 13.78 | 6.86 | 43.2 | 0.03 | 86.1  | 116 | 373 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 4  | 11.9 | 7.93 | 12.54 | 6.72 | 43   | 0.03 | 74.2  | 55  | 373 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 3  | 12.9 | 7.84 | 11.88 | 6.7  | 42.5 | 0.03 | 72.3  | 112 | 372 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 2  | 13.9 | 6.95 | 10.97 | 6.63 | 43.8 | 0.03 | 62.7  | 132 | 372 |
| Conkling Point | 2998 | ASBH   | 7/20/98 | 11:44:42 | 6.3 | 1  | 14.9 | 6.66 | 10.25 | 6.64 | 44.5 | 0.03 | 59.1  | 304 | 369 |
| Conkling Point | 3298 | DBAS   | 8/10/98 | 12:10:22 | 9.2 | 16 | 0.3  | 8.29 | 24.51 | 7.53 | 59   | 0.04 | 99    | 30  | 343 |
| Conkling Point | 3298 | DBAS   | 8/10/98 | 12:10:22 | 9.2 | 15 | 1    | 8.28 | 24.09 | 7.56 | 59.1 | 0.04 | 98.1  | 35  | 343 |
| Conkling Point | 3298 | DBAS   | 8/10/98 | 12:10:22 | 9.2 | 14 | 2.1  | 8.27 | 23.8  | 7.55 | 59.1 | 0.04 | 97.4  | 29  | 345 |

|                |      |      |         |          |     |    |      |      |       |      |      |      |       |     |     |
|----------------|------|------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 13 | 3.2  | 8.26 | 23.29 | 7.55 | 58.6 | 0.04 | 96.4  | 35  | 346 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 12 | 4.3  | 8.44 | 22.84 | 7.49 | 57.1 | 0.04 | 97.7  | 48  | 347 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 11 | 5.4  | 9.08 | 22.2  | 7.52 | 55.7 | 0.04 | 103.8 | 35  | 347 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 10 | 6.5  | 9.19 | 21.74 | 7.49 | 54.3 | 0.03 | 104.2 | 109 | 348 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 9  | 7.4  | 9.46 | 20.91 | 7.44 | 52.3 | 0.03 | 105.5 | 122 | 349 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 8  | 8.5  | 9.58 | 20.34 | 7.35 | 51.3 | 0.03 | 105.6 | 44  | 352 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 7  | 9.5  | 9.61 | 19.13 | 7.27 | 49.7 | 0.03 | 103.5 | 131 | 353 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 6  | 10.5 | 8.5  | 18.27 | 7.04 | 50.3 | 0.03 | 89.9  | 101 | 356 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 5  | 11.5 | 9.45 | 17.4  | 7.1  | 49.1 | 0.03 | 98.2  | 100 | 356 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 4  | 12.4 | 8.77 | 15.66 | 6.93 | 48.1 | 0.03 | 87.8  | 128 | 358 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 3  | 13.2 | 8.29 | 14.08 | 6.81 | 47.2 | 0.03 | 80.3  | 118 | 359 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 2  | 14.4 | 5.74 | 12.61 | 6.63 | 49.2 | 0.03 | 53.8  | 120 | 359 |
| Conkling Point | 3298 | DBAS | 8/10/98 | 12:10:22 | 9.2 | 1  | 15.5 | 4.81 | 10.31 | 6.61 | 52   | 0.03 | 42.8  | 253 | 359 |
|                |      |      |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 16 | 0.5  | 9.01 | 21.81 | 7.87 | 54.1 | 0.03 | 102.8 | 30  | 362 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 15 | 1.5  | 9.01 | 21.72 | 7.84 | 54   | 0.03 | 102.6 | 107 | 365 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 14 | 2.5  | 8.99 | 21.76 | 7.86 | 54.1 | 0.03 | 102.4 | 125 | 364 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 13 | 3.5  | 8.99 | 21.63 | 7.82 | 53.8 | 0.03 | 102.1 | 110 | 367 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 12 | 4.5  | 8.98 | 21.53 | 7.81 | 53.8 | 0.03 | 101.9 | 234 | 368 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 11 | 5.5  | 8.83 | 21.46 | 7.71 | 53.6 | 0.03 | 100   | 137 | 372 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 10 | 6.5  | 8.76 | 21.44 | 7.59 | 54.5 | 0.03 | 99.2  | 147 | 376 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 9  | 7.5  | 8.5  | 21.29 | 7.45 | 54.8 | 0.04 | 95.9  | 138 | 381 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 8  | 8.5  | 7.61 | 20.76 | 7.09 | 53.4 | 0.03 | 85.1  | 208 | 391 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 7  | 9.5  | 8.61 | 19.68 | 7.03 | 47.9 | 0.03 | 94.1  | 112 | 393 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 6  | 10.5 | 8.18 | 18.42 | 6.88 | 45.7 | 0.03 | 87.3  | 221 | 397 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 5  | 11.5 | 8.34 | 16.63 | 6.84 | 44.1 | 0.03 | 85.7  | 316 | 398 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 4  | 12.5 | 6.9  | 14.91 | 6.63 | 44.2 | 0.03 | 68.3  | 203 | 401 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 3  | 13.5 | 6.86 | 13.55 | 6.54 | 43.8 | 0.03 | 66    | 154 | 401 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 2  | 14.5 | 4.84 | 12.85 | 6.44 | 45.8 | 0.03 | 45.8  | 235 | 401 |
| Conkling Point | 3498 | ASGL | 8/26/98 | 13:49:11 | 5.3 | 1  | 15.4 | 3.48 | 11.23 | 6.36 | 49.2 | 0.03 | 31.8  | 434 | 404 |
|                |      |      |         |          |     |    |      |      |       |      |      |      |       |     |     |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 16 | 0.4  | 9.04 | 22.58 | 7.83 | 59.2 | 0.04 | 104.8 | 33  | 290 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 15 | 1.6  | 9.11 | 21.56 | 7.86 | 59.7 | 0.04 | 103.5 | 52  | 289 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 14 | 2.6  | 9.07 | 21.42 | 7.78 | 58.9 | 0.04 | 102.8 | 41  | 292 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 13 | 3.6  | 9.15 | 21.28 | 7.8  | 58.7 | 0.04 | 103.5 | 42  | 292 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 12 | 4.6  | 9.18 | 21.14 | 7.75 | 58.8 | 0.04 | 103.5 | 40  | 294 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 11 | 5.6  | 9.09 | 20.97 | 7.65 | 58.6 | 0.04 | 102.2 | 37  | 298 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 10 | 6.6  | 9.08 | 20.97 | 7.55 | 58.5 | 0.04 | 102   | 54  | 303 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 9  | 7.6  | 8.87 | 20.48 | 7.26 | 56.9 | 0.04 | 98.7  | 101 | 312 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 8  | 8.6  | 8.6  | 19.66 | 6.96 | 55.2 | 0.04 | 94.2  | 44  | 321 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 7  | 9.6  | 7.21 | 17.76 | 6.74 | 49.4 | 0.03 | 76    | 36  | 326 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 6  | 10.6 | 7.06 | 16.14 | 6.7  | 47.4 | 0.03 | 72    | 100 | 326 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 5  | 11.6 | 6.4  | 14.56 | 6.62 | 47.8 | 0.03 | 63    | 100 | 328 |
| Conkling Point | 3598 | DTAS | 9/3/98  | 13:43:10 | 5.7 | 4  | 12.6 | 5.45 | 13.25 | 6.54 | 49.1 | 0.03 | 52.1  | 113 | 327 |

|                |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
|----------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| Conkling Point | 3598 | DTAS   | 9/3/98   | 13:43:10 | 5.7 | 3  | 13.6 | 4.68 | 12.77 | 6.51 | 49.4 | 0.03 | 44.3 | 35  | 323 |
| Conkling Point | 3598 | DTAS   | 9/3/98   | 13:43:10 | 5.7 | 2  | 14.6 | 5.79 | 11.47 | 6.52 | 48.2 | 0.03 | 53.2 | 123 | 320 |
| Conkling Point | 3598 | DTAS   | 9/3/98   | 13:43:10 | 5.7 | 1  | 15.6 | 2.78 | 10.69 | 6.49 | 53.3 | 0.03 | 25.1 | 210 | 313 |
|                |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 16 | 0.3  | 9.07 | 15.24 | 7.42 | 58.7 | 0.04 | 90.2 | 50  | 325 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 15 | 0.7  | 9.07 | 15.23 | 7.4  | 58.6 | 0.04 | 90.2 | 43  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 14 | 1.7  | 9.05 | 15.24 | 7.38 | 58.5 | 0.04 | 90   | 48  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 13 | 2.7  | 9.02 | 15.22 | 7.38 | 58.6 | 0.04 | 89.7 | 58  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 12 | 3.7  | 8.96 | 15.18 | 7.35 | 58.8 | 0.04 | 89   | 58  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 11 | 4.7  | 8.91 | 15.19 | 7.35 | 58.5 | 0.04 | 88.6 | 54  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 10 | 5.8  | 8.93 | 15.17 | 7.34 | 58.5 | 0.04 | 88.7 | 56  | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 9  | 6.6  | 8.96 | 15.16 | 7.34 | 59   | 0.04 | 89   | 134 | 326 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 8  | 7.7  | 8.98 | 15.16 | 7.35 | 58.3 | 0.04 | 89.2 | 114 | 324 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 7  | 8.7  | 9    | 15.14 | 7.35 | 58.4 | 0.04 | 89.3 | 142 | 323 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 6  | 9.7  | 9.08 | 15.14 | 7.36 | 59.2 | 0.04 | 90.1 | 123 | 321 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 5  | 10.7 | 9.15 | 15.04 | 7.36 | 58.6 | 0.04 | 90.7 | 135 | 321 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 4  | 11.7 | 9.1  | 14.91 | 7.33 | 58.6 | 0.04 | 89.8 | 109 | 320 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 3  | 12.7 | 9.12 | 14.88 | 7.31 | 58.3 | 0.04 | 90   | 125 | 319 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 2  | 13.7 | 9.2  | 14.84 | 7.29 | 58.6 | 0.04 | 90.7 | 156 | 318 |
| Conkling Point | 4098 | KBDB   | 10/9/98  | 11:50:00 | 4.5 | 1  | 14.7 | 8.64 | 14.59 | 7.13 | 59.7 | 0.04 | 84.7 | 344 | 318 |
|                |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 15 | 0.5  | 9.13 | 12.92 | 7.39 | 58.6 | 0.04 | 85.5 | 115 | 343 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 14 | 1.5  | 9.14 | 12.81 | 7.41 | 58.6 | 0.04 | 85.4 | 47  | 342 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 13 | 2.5  | 9.15 | 12.71 | 7.38 | 58.7 | 0.04 | 85.3 | 53  | 343 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 12 | 3.5  | 9.13 | 12.63 | 7.34 | 58.7 | 0.04 | 84.9 | 114 | 346 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 11 | 4.5  | 9.07 | 12.58 | 7.33 | 58.8 | 0.04 | 84.3 | 125 | 346 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 10 | 5.5  | 9.01 | 12.56 | 7.32 | 58.3 | 0.04 | 83.7 | 41  | 345 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 9  | 6.5  | 9.01 | 12.55 | 7.31 | 58.4 | 0.04 | 83.7 | 52  | 345 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 8  | 7.5  | 8.98 | 12.51 | 7.32 | 58.7 | 0.04 | 83.3 | 126 | 343 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 7  | 8.5  | 9.05 | 12.46 | 7.33 | 58.3 | 0.04 | 83.8 | 39  | 343 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 6  | 9.5  | 9.06 | 12.4  | 7.33 | 58.7 | 0.04 | 83.8 | 156 | 343 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 5  | 10.5 | 9.11 | 12.27 | 7.34 | 58.8 | 0.04 | 84.1 | 110 | 341 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 4  | 11.5 | 9.17 | 12.17 | 7.36 | 58.7 | 0.04 | 84.4 | 58  | 339 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 3  | 12.5 | 9.49 | 11.91 | 7.39 | 58.4 | 0.04 | 86.8 | 57  | 338 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 2  | 13.5 | 9.51 | 11.46 | 7.32 | 59   | 0.04 | 86.2 | 43  | 339 |
| Conkling Point | 4298 | DTASJB | 10/19/98 | 11:49:36 | 4.8 | 1  | 14.5 | 9.26 | 11.35 | 7.28 | 59   | 0.04 | 83.6 | 645 | 340 |
|                |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 15 | 0.3  | 8.23 | 8.32  | 7.16 | 54.7 | 0.04 | 70.7 | 44  | 327 |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 14 | 1    | 8.21 | 8.33  | 7.15 | 54.9 | 0.04 | 70.6 | 46  | 328 |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 13 | 2    | 8.21 | 8.33  | 7.17 | 54.7 | 0.04 | 70.6 | 48  | 326 |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 12 | 3    | 8.21 | 8.32  | 7.11 | 54.5 | 0.03 | 70.5 | 57  | 329 |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 11 | 4    | 8.19 | 8.32  | 7.12 | 54.6 | 0.03 | 70.4 | 101 | 4   |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 10 | 5    | 8.15 | 8.3   | 7.09 | 54.7 | 0.04 | 70   | 41  | 329 |
| Conkling Point | 4698 | KBAS   | 11/17/98 | 12:18:00 | 4.1 | 9  | 6    | 8.11 | 8.29  | 7.09 | 54.4 | 0.03 | 71.9 | 8   | 328 |

|                |      |      |          |          |     |   |      |      |      |      |      |      |      |     |     |
|----------------|------|------|----------|----------|-----|---|------|------|------|------|------|------|------|-----|-----|
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 8 | 7    | 8.07 | 8.22 | 7.06 | 54.2 | 0.03 | 69.2 | 51  | 328 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 7 | 8    | 7.86 | 8.11 | 7.01 | 54.1 | 0.03 | 67.2 | 120 | 327 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 6 | 9    | 7.48 | 8.04 | 6.96 | 54.4 | 0.03 | 63.8 | 116 | 326 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 5 | 10   | 7.61 | 7.99 | 7    | 54   | 0.03 | 64.9 | 112 | 324 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 4 | 10.9 | 7.97 | 7.88 | 7.05 | 54.5 | 0.03 | 67.8 | 123 | 321 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 3 | 12   | 8.29 | 7.76 | 0    | 54.1 | 0.03 | 70.3 | 149 | 319 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 2 | 13   | 8.9  | 6.73 | 7.15 | 54.8 | 0.04 | 73.6 | 44  | 316 |
| Conkling Point | 4698 | KBAS | 11/17/98 | 12:18:00 | 4.1 | 1 | 14   | 8.9  | 6.73 | 7.16 | 54.5 | 0.03 | 73.6 | 555 | 315 |

| Location    | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 10       | 0.3       | 12.2                    | 7.79            | 7.06 | 37.8                 | 0.02 | 103.2                      | 43                 | 389   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 9        | 0.3       | 12.2                    | 7.79            | 7.04 | 37.8                 | 0.02 | 103.2                      | 105                | 390   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 8        | 1         | 12.4                    | 6.69            | 7.08 | 37.7                 | 0.02 | 101.6                      | 124                | 388   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 7        | 2         | 12.5                    | 6.26            | 7.08 | 37.9                 | 0.02 | 101.2                      | 46                 | 387   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 6        | 3         | 12.4                    | 6.18            | 7.04 | 37.6                 | 0.02 | 100.9                      | 114                | 389   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 5        | 4         | 12.4                    | 6.15            | 7    | 37.9                 | 0.02 | 100.7                      | 143                | 390   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 4        | 5         | 12.5                    | 6.1             | 6.94 | 37.7                 | 0.02 | 100.7                      | 144                | 391   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 3        | 6         | 12.5                    | 6.03            | 6.96 | 38                   | 0.02 | 100.5                      | 134                | 389   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 2        | 7         | 12.5                    | 5.98            | 6.91 | 37.9                 | 0.02 | 100.4                      | 137                | 389   |
| Hidden Lake | 1398  | DBRP     | 4/3/98  | 10:57:54 | 2          | 1        | 7.9       | 12.2                    | 6.22            | 6.92 | 37.6                 | 0.02 | 99                         | 400                | 387   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 9        | 0.3       | 12                      | 10.09           | 7.15 | 41.9                 | 0.03 | 106.2                      | 121                | 398   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 8        | 1.3       | 11.9                    | 9.69            | 7.25 | 41.7                 | 0.03 | 104.5                      | 207                | 394   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 7        | 2.3       | 12.1                    | 8.89            | 7.18 | 41.5                 | 0.03 | 103.9                      | 46                 | 394   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 6        | 3.3       | 12                      | 8.68            | 7.15 | 41.5                 | 0.03 | 102.7                      | 111                | 394   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 5        | 4.3       | 11.9                    | 8.4             | 7.11 | 41.4                 | 0.03 | 101.3                      | 132                | 393   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 4        | 5.4       | 11.9                    | 8               | 7.07 | 40.9                 | 0.03 | 100                        | 121                | 391   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 3        | 6.4       | 11.7                    | 7.76            | 6.99 | 41                   | 0.03 | 97.5                       | 141                | 389   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 2        | 7.4       | 10.3                    | 7.36            | 6.79 | 41.3                 | 0.03 | 85.1                       | 55                 | 388   |
| Hidden Lake | 1698  | DBASRP   | 4/21/98 | 13:37:37 | 2.2        | 1        | 7.3       | 10.6                    | 7.48            | 6.84 | 41.1                 | 0.03 | 87.8                       | 216                | 391   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 9        | 0.3       | 11.1                    | 9.51            | 6.95 | 32.5                 | 0.02 | 98.4                       | 40                 | 339   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 8        | 1.3       | 11.1                    | 9.42            | 6.92 | 32.8                 | 0.02 | 98.2                       | 36                 | 337   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 7        | 2.3       | 11.1                    | 9.33            | 6.96 | 32.9                 | 0.02 | 98.1                       | 43                 | 331   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 6        | 3.3       | 11.2                    | 9.3             | 6.94 | 33.1                 | 0.02 | 98.1                       | 35                 | 328   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 5        | 4.3       | 11.1                    | 9.28            | 6.86 | 33.1                 | 0.02 | 97.8                       | 39                 | 328   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 4        | 5.3       | 11.1                    | 9.25            | 6.96 | 33                   | 0.02 | 97.4                       | 39                 | 316   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 3        | 6.3       | 11.1                    | 9.27            | 6.81 | 32.9                 | 0.02 | 97.5                       | 37                 | 318   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 2        | 7.3       | 11.1                    | 9.25            | 6.95 | 32.8                 | 0.02 | 97.3                       | 37                 | 309   |
| Hidden Lake | 2198  | ASJLDT   | 5/26/98 | 14:05:25 | 2.1        | 1        | 8.3       | 11.1                    | 9.22            | 6.95 | 32.8                 | 0.02 | 97.1                       | 107                | 318   |
| Hidden Lake | 2398  | DTJLAS   | 6/8/98  | 13:26:09 | 2.4        | 11       | 0.4       | 11                      | 15.62           | 7.51 | 38.7                 | 0.02 | 110.5                      | 29                 | 352   |
| Hidden Lake | 2398  | DTJLAS   | 6/8/98  | 13:26:09 | 2.4        | 10       | 1.2       | 11.1                    | 14.27           | 7.47 | 38.6                 | 0.02 | 108.4                      | 44                 | 355   |

|             |      |        |         |          |     |    |     |      |       |      |      |      |       |     |     |
|-------------|------|--------|---------|----------|-----|----|-----|------|-------|------|------|------|-------|-----|-----|
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 9  | 2   | 11.1 | 13.89 | 7.45 | 38.2 | 0.02 | 107.4 | 103 | 355 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 8  | 2.7 | 11   | 13.51 | 7.34 | 38.2 | 0.02 | 105.1 | 108 | 357 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 7  | 3.5 | 11   | 12.57 | 7.22 | 36.7 | 0.02 | 103.6 | 103 | 358 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 6  | 4.2 | 10.6 | 12.11 | 7.11 | 36.3 | 0.02 | 98.2  | 59  | 359 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 5  | 5   | 10.4 | 11.85 | 7.02 | 36.3 | 0.02 | 95.9  | 39  | 359 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 4  | 5.7 | 9.68 | 11.46 | 6.94 | 36.4 | 0.02 | 88.8  | 129 | 359 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 3  | 6.5 | 9.13 | 10.85 | 6.87 | 36.5 | 0.02 | 82.5  | 130 | 357 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 2  | 7.2 | 8.76 | 10.62 | 6.83 | 36.3 | 0.02 | 78.7  | 122 | 353 |
| Hidden Lake | 2398 | DTJLAS | 6/8/98  | 13:26:09 | 2.4 | 1  | 8   | 7.97 | 10.21 | 6.77 | 37   | 0.02 | 71    | 233 | 347 |
|             |      |        |         |          |     |    |     |      |       |      |      |      |       |     |     |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 10 | 0.3 | 10.1 | 19.32 | 7.68 | 40.6 | 0.03 | 110   | 36  | 336 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 9  | 1.2 | 10.2 | 18.75 | 7.64 | 40.7 | 0.03 | 109.2 | 112 | 338 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 8  | 2   | 10.2 | 18.29 | 7.6  | 40.8 | 0.03 | 108.4 | 105 | 339 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 7  | 2.7 | 10.2 | 17.41 | 7.53 | 40   | 0.03 | 106.8 | 117 | 340 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 6  | 3.5 | 10.2 | 16.31 | 7.34 | 39.5 | 0.03 | 104.3 | 156 | 344 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 5  | 4.2 | 10.6 |       | 7.15 | 37.2 | 0.02 | 105.9 | 135 | 348 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 4  | 5   | 8.81 | 12.82 | 6.81 | 35.9 | 0.02 | 83.3  | 143 | 354 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 3  | 5.7 | 7.14 | 12.03 | 6.6  | 36.3 | 0.02 | 66.3  | 146 | 356 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 2  | 6.5 | 4.53 | 11.28 | 6.47 | 37.7 | 0.02 | 41.3  | 236 | 354 |
| Hidden Lake | 2598 | DTASBH | 6/22/98 | 12:17:57 | 3.2 | 1  | 7.2 | 4.15 | 11.18 | 6.45 | 38   | 0.02 | 37.8  | 506 | 344 |
|             |      |        |         |          |     |    |     |      |       |      |      |      |       |     |     |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 11 | 0.4 | 9.9  | 24.13 | 8.46 | 49.2 | 0.03 | 117.3 | 151 | 320 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 10 | 1   | 9.91 | 23.67 | 8.48 | 49.2 | 0.03 | 116.4 | 139 | 322 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 9  | 1.7 | 10.2 | 22.84 | 8.52 | 48.8 | 0.03 | 117.6 | 117 | 323 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 8  | 2.5 | 10.9 | 21.28 | 8.48 | 47.9 | 0.03 | 122.8 | 110 | 326 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 7  | 3.2 | 11.1 | 19.38 | 8.19 | 45.9 | 0.03 | 119.5 | 141 | 334 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 6  | 4   | 10.7 | 17.54 | 7.58 | 44.7 | 0.03 | 111.6 | 111 | 349 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 5  | 4.7 | 10.8 | 15.86 | 7.19 | 40.8 | 0.03 | 108.3 | 39  | 353 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 4  | 5.5 | 8.58 | 14.59 | 6.79 | 39.9 | 0.03 | 84    | 112 | 359 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 3  | 6.2 | 6.08 | 13.13 | 6.52 | 41.1 | 0.03 | 57.6  | 109 | 361 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 2  | 7   | 4.01 | 12.46 | 6.46 | 42.2 | 0.03 | 37.5  | 121 | 359 |
| Hidden Lake | 2798 | DTASBH | 7/7/98  | 11:36:34 | 4.6 | 1  | 7.6 | 2.9  | 12.33 | 6.45 | 45.9 | 0.03 | 27    | 320 | 358 |
|             |      |        |         |          |     |    |     |      |       |      |      |      |       |     |     |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 9  | 0.4 | 9.31 | 24.51 | 8.39 | 49.4 | 0.03 | 111.2 | 123 | 258 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 8  | 1.4 | 9.16 | 24.27 | 8.31 | 49.3 | 0.03 | 109   | 122 | 255 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 7  | 2.4 | 9.23 | 23.89 | 8.28 | 49.4 | 0.03 | 109   | 138 | 250 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 6  | 3.4 | 9.19 | 23.84 | 8.18 | 49.2 | 0.03 | 108.4 | 158 | 245 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 5  | 4.4 | 10.9 | 19.8  | 7.47 | 42.1 | 0.03 | 118.8 | 111 | 259 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 4  | 5.4 | 11.4 | 17.3  | 7.21 | 40.4 | 0.03 | 117.9 | 133 | 259 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 3  | 6.4 | 7.73 | 14.94 | 6.66 | 42.1 | 0.03 | 76.3  | 136 | 256 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 2  | 7.4 | 3.6  | 13.65 | 6.46 | 44.2 | 0.03 | 34.6  | 133 | 222 |
| Hidden Lake | 2998 | ASBH   | 7/21/98 | 9:51:57  | 6.8 | 1  | 8.4 | 0.25 | 13.27 | 6.47 | 57.6 | 0.04 | 2.3   | 202 | 168 |
|             |      |        |         |          |     |    |     |      |       |      |      |      |       |     |     |
| Hidden Lake | 3298 | DBAS   | 8/11/98 | 13:00:48 | 5.6 | 13 | 0.3 | 9.19 | 25.06 | 8.36 | 53.4 | 0.03 | 110.8 | 50  | 253 |
| Hidden Lake | 3298 | DBAS   | 8/11/98 | 13:00:48 | 5.6 | 12 | 0.8 | 9.27 | 24.79 | 8.4  | 53.3 | 0.03 | 111.2 | 131 | 250 |

|             |      |      |         |          |     |    |     |      |       |      |       |      |       |     |     |
|-------------|------|------|---------|----------|-----|----|-----|------|-------|------|-------|------|-------|-----|-----|
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 11 | 1.4 | 9.34 | 24.66 | 8.42 | 53.2  | 0.03 | 111.8 | 50  | 248 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 10 | 2   | 9.4  | 24.49 | 8.46 | 53.3  | 0.03 | 112.2 | 44  | 244 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 9  | 2.6 | 9.49 | 24.42 | 8.45 | 53.4  | 0.03 | 113.1 | 132 | 243 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 8  | 3.2 | 9.37 | 24.13 | 8.39 | 53.1  | 0.03 | 111   | 133 | 237 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 7  | 3.9 | 9.61 | 24.02 | 8.5  | 53.8  | 0.03 | 113.7 | 111 | 228 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 6  | 4.6 | 10.5 | 23.35 | 8.45 | 51.9  | 0.03 | 122.9 | 157 | 222 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 5  | 5.3 | 13.3 | 21.42 | 8.52 | 44.9  | 0.03 | 149.5 | 132 | 202 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 4  | 6   | 13.2 | 19.3  | 8.42 | 48.7  | 0.03 | 142.7 | 104 | 160 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 3  | 6.6 | 0.21 | 16.81 | 6.51 | 56    | 0.04 | 2.1   | 111 | 163 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 2  | 7.1 | 0.19 | 15.35 | 6.55 | 56.7  | 0.04 | 1.9   | 130 | 134 |
| Hidden Lake | 3298 | DBAS | 8/11/98 | 13:00:48 | 5.6 | 1  | 7.7 | 0.23 | 14.04 | 6.65 | 76.9  | 0.05 | 2.2   | 323 | 99  |
|             |      |      |         |          |     |    |     |      |       |      |       |      |       |     |     |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 12 | 0.3 | 9.41 | 22.01 | 8.34 | 53.3  | 0.03 | 107.7 | 141 | 231 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 11 | 1   | 9.43 | 21.97 | 8.32 | 53.3  | 0.03 | 107.8 | 129 | 230 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 10 | 1.7 | 9.43 | 21.85 | 8.28 | 53.3  | 0.03 | 107.6 | 117 | 229 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 9  | 2.5 | 9.42 | 21.62 | 8.29 | 53.3  | 0.03 | 107   | 112 | 223 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 8  | 3.2 | 9.42 | 21.58 | 8.28 | 53    | 0.03 | 106.9 | 203 | 221 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 7  | 4   | 9.41 | 21.55 | 8.27 | 53.1  | 0.03 | 106.8 | 129 | 214 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 6  | 4.7 | 9.37 | 21.55 | 8.25 | 53.1  | 0.03 | 106.3 | 149 | 202 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 5  | 5.5 | 9.69 | 21.32 | 8.37 | 53.9  | 0.03 | 109.4 | 132 | 176 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 4  | 6.2 | 2.34 | 20.16 | 6.86 | 58.8  | 0.04 | 25.9  | 203 | 185 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 3  | 7   | 0.13 | 17.95 | 6.7  | 67.5  | 0.04 | 1.4   | 201 | 123 |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 2  | 7.7 | 0.12 | 15.42 | 6.89 | 83.7  | 0.05 | 1.2   | 132 | 55  |
| Hidden Lake | 3498 | ASGL | 8/26/98 | 11:45:08 | 3.8 | 1  | 8.5 | 0.15 | 13.55 | 7.04 | 129.1 | 0.08 | 1.4   | 318 | 32  |
|             |      |      |         |          |     |    |     |      |       |      |       |      |       |     |     |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 9  | 0.3 | 9.46 | 22.13 | 8.26 | 57.9  | 0.04 | 108.7 | 40  | 238 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 8  | 1   | 9.48 | 21.97 | 8.27 | 57.9  | 0.04 | 108.6 | 41  | 232 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 7  | 2   | 9.48 | 21.74 | 8.21 | 57.7  | 0.04 | 108.2 | 42  | 231 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 6  | 3   | 9.47 | 21.6  | 8.17 | 57.7  | 0.04 | 107.7 | 151 | 224 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 5  | 4   | 9.42 | 21.53 | 8.02 | 58.1  | 0.04 | 107   | 38  | 209 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 4  | 5   | 9.34 | 21.46 | 7.9  | 57.8  | 0.04 | 106   | 129 | 206 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 3  | 6   | 4.92 | 20.83 | 7    | 59.6  | 0.04 | 55.2  | 125 | 217 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 2  | 7   | 0.12 | 17.86 | 6.64 | 77.7  | 0.05 | 1.3   | 58  | 184 |
| Hidden Lake | 3598 | DTAS | 9/3/98  | 13:09:26 | 6.2 | 1  | 7.8 | 0.17 | 15.2  | 6.93 | 116.9 | 0.07 | 1.7   | 309 | 90  |
|             |      |      |         |          |     |    |     |      |       |      |       |      |       |     |     |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 12 | 0.3 | 9.83 | 14.5  | 7.84 | 57.7  | 0.04 | 96.2  | 53  | 313 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 11 | 0.9 | 9.88 | 14.45 | 7.83 | 57.7  | 0.04 | 96.6  | 59  | 314 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 10 | 1.7 | 9.84 | 14.27 | 7.75 | 57.6  | 0.04 | 95.9  | 43  | 316 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 9  | 2.4 | 9.64 | 14.23 | 7.69 | 57.7  | 0.04 | 93.8  | 54  | 317 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 8  | 3.1 | 9.48 | 14.21 | 7.61 | 57.6  | 0.04 | 92.2  | 103 | 320 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 7  | 3.8 | 9.48 | 14.22 | 7.65 | 57.8  | 0.04 | 92.2  | 111 | 317 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 6  | 4.5 | 9.46 | 14.2  | 7.63 | 57.7  | 0.04 | 92    | 122 | 318 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 5  | 5.2 | 9.47 | 14.18 | 7.65 | 57.9  | 0.04 | 92.1  | 55  | 317 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 4  | 5.9 | 9.69 | 14.13 | 7.68 | 57.2  | 0.04 | 94.1  | 153 | 317 |
| Hidden Lake | 4098 | KBDB | 10/9/98 | 12:22:00 | 2.7 | 3  | 6.6 | 9.76 | 14.08 | 7.66 | 57.4  | 0.04 | 94.7  | 113 | 317 |

|             |      |        |          |          |     |    |     |      |       |      |      |      |      |     |     |
|-------------|------|--------|----------|----------|-----|----|-----|------|-------|------|------|------|------|-----|-----|
| Hidden Lake | 4098 | KBDB   | 10/9/98  | 12:22:00 | 2.7 | 2  | 7.3 | 9.76 | 14.03 | 7.63 | 57.7 | 0.04 | 94.6 | 143 | 317 |
| Hidden Lake | 4098 | KBDB   | 10/9/98  | 12:22:00 | 2.7 | 1  | 7.9 | 9.94 | 13.86 | 7.58 | 57.6 | 0.04 | 95.9 | 401 | 323 |
|             |      |        |          |          |     |    |     |      |       |      |      |      |      |     |     |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 13 | 0.4 | 10.4 | 11.18 | 7.89 | 57.8 | 0.04 | 93.4 | 39  | 335 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 12 | 1   | 10.4 | 11.13 | 7.9  | 57.9 | 0.04 | 93.2 | 52  | 335 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 11 | 1.5 | 10.4 | 11    | 7.9  | 57.5 | 0.04 | 92.9 | 42  | 336 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 10 | 2   | 10.4 | 10.95 | 7.86 | 57.8 | 0.04 | 92.9 | 116 | 338 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 9  | 2.5 | 10.4 | 10.79 | 7.86 | 57.5 | 0.04 | 92.5 | 38  | 337 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 8  | 3   | 10.4 | 10.77 | 7.82 | 57.6 | 0.04 | 92.3 | 56  | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 7  | 3.5 | 10.3 | 10.76 | 7.8  | 57.5 | 0.04 | 92   | 119 | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 6  | 4   | 10.3 | 10.71 | 7.75 | 57.6 | 0.04 | 91.7 | 47  | 340 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 5  | 4.5 | 10.3 | 10.68 | 7.76 | 57.3 | 0.04 | 91.4 | 38  | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 4  | 5   | 10.3 | 10.66 | 7.76 | 57.5 | 0.04 | 91.6 | 37  | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 3  | 5.5 | 10.4 | 10.64 | 7.77 | 57.4 | 0.04 | 92.3 | 46  | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 2  | 6   | 10.4 | 10.58 | 7.76 | 57.6 | 0.04 | 92.5 | 119 | 339 |
| Hidden Lake | 4298 | DTASJB | 10/19/98 | 11:09:03 | 2.7 | 1  | 6.5 | 10.4 | 10.58 | 7.69 | 57.6 | 0.04 | 92.5 | 228 | 346 |
|             |      |        |          |          |     |    |     |      |       |      |      |      |      |     |     |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 8  | 0.3 | 9.83 | 6.86  | 7.7  | 53.4 | 0.03 | 81.4 | 100 | 277 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 7  | 1.2 | 9.81 | 6.85  | 7.73 | 53.7 | 0.03 | 81.3 | 125 | 273 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 6  | 2.2 | 9.75 | 6.82  | 7.68 | 53.4 | 0.03 | 80.8 | 130 | 272 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 5  | 3.3 | 9.72 | 6.79  | 7.64 | 53.4 | 0.03 | 80.4 | 121 | 270 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 4  | 4.2 | 9.72 | 6.79  | 7.61 | 53.6 | 0.03 | 80.4 | 55  | 267 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 3  | 5.2 | 9.7  | 6.78  | 7.61 | 53.4 | 0.03 | 80.2 | 134 | 264 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 2  | 6.2 | 9.63 | 6.79  | 7.56 | 53.5 | 0.03 | 79.7 | 115 | 261 |
| Hidden Lake | 4698 | KBAS   | 11/17/98 | 11:23:00 | 3.3 | 1  | 7.2 | 9.6  | 6.79  | 7.54 | 53.3 | 0.03 | 79.4 | 713 | 256 |

| Location   | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Round Lake | 1398  | DBRP     | 4/3/98  | 13:04:20 | 1.3        | 4        | 0.2       | 12.5                    | 6.13            | 7.04 | 41.8                 | 0.03 | 101.5                      | 46                 | 396   |
| Round Lake | 1398  | DBRP     | 4/3/98  | 13:04:20 | 1.3        | 3        | 0.5       | 12.6                    | 6.05            | 7.02 | 41.7                 | 0.03 | 101.4                      | 57                 | 399   |
| Round Lake | 1398  | DBRP     | 4/3/98  | 13:04:20 | 1.3        | 2        | 0.9       | 12.5                    | 6.1             | 7.02 | 41.6                 | 0.03 | 101.4                      | 154                | 401   |
| Round Lake | 1398  | DBRP     | 4/3/98  | 13:04:20 | 1.3        | 1        | 1.3       | 12.7                    | 6.36            | 6.98 | 41.8                 | 0.03 | 103.5                      | 153                | 405   |
|            |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Round Lake | 1698  | DBASRP   | 4/21/98 | 12:28:46 | 1.0        | 4        | 0.4       | 11.8                    | 9.13            | 7.05 | 44.1                 | 0.03 | 102.4                      | 57                 | 376   |
| Round Lake | 1698  | DBASRP   | 4/21/98 | 12:28:46 | 1.0        | 3        | 0.6       | 11.8                    | 8.97            | 7.1  | 44                   | 0.03 | 101.9                      | 54                 | 374   |
| Round Lake | 1698  | DBASRP   | 4/21/98 | 12:28:46 | 1.0        | 2        | 0.8       | 11.8                    | 8.99            | 7.07 | 44                   | 0.03 | 101.6                      | 213                | 376   |
| Round Lake | 1698  | DBASRP   | 4/21/98 | 12:28:46 | 1.0        | 1        | 1         | 11.7                    | 9.2             | 7.08 | 43.7                 | 0.03 | 101.1                      | 250                | 395   |
|            |       |          |         |          |            |          |           |                         |                 |      |                      |      |                            |                    |       |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 6        | 0.3       | 11.9                    | 9.68            | 7.04 | 35.3                 | 0.02 | 104.4                      | 54                 | 410   |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 5        | 0.3       | 11.9                    | 9.76            | 7.07 | 35.2                 | 0.02 | 105.1                      | 50                 | 410   |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 4        | 1         | 11.9                    | 9.66            | 7.06 | 35.3                 | 0.02 | 104.9                      | 58                 | 410   |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 3        | 1         | 11.9                    | 9.64            | 7.02 | 35.2                 | 0.02 | 105                        | 102                | 414   |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 2        | 1.7       | 11.9                    | 9.66            | 7.03 | 35.3                 | 0.02 | 105.1                      | 57                 | 414   |
| Round Lake | 2098  | ASJLDT   | 5/21/98 | 14:04:17 | 1.7        | 1        | 1.7       | 11.9                    | 9.68            | 7.05 | 35.3                 | 0.02 | 105                        | 340                | 413   |



|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
|------------|------|--------|---------|----------|-----|---|-----|------|-------|------|------|------|-------|-----|-----|
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 8 | 0.3 | 10.7 | 12.62 | 7.3  | 39.2 | 0.03 | 100.9 | 35  | 368 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 7 | 0.3 | 10.7 | 12.62 | 7.3  | 39.2 | 0.03 | 100.3 | 44  | 369 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 6 | 0.7 | 10.7 | 12.58 | 7.27 | 39.2 | 0.03 | 100.8 | 47  | 371 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 5 | 0.7 | 10.7 | 12.64 | 7.29 | 39.3 | 0.03 | 100.4 | 122 | 370 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 4 | 1.2 | 10.8 | 12.46 | 7.28 | 39.2 | 0.03 | 100.9 | 43  | 371 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 3 | 1.2 | 10.8 | 12.48 | 7.28 | 39.2 | 0.03 | 101   | 116 | 372 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 2 | 1.8 | 10.9 | 12.49 | 7.28 | 39.3 | 0.03 | 101.9 | 44  | 372 |
| Round Lake | 2398 | DTJLAS | 6/8/98  | 14:10:07 | 1.8 | 1 | 1.8 | 10.9 | 12.54 | 7.27 | 39.2 | 0.03 | 102.1 | 215 | 372 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 6 | 0.3 | 10.3 | 16.71 | 7.36 | 43.9 | 0.03 | 106.2 | 123 | 343 |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 5 | 0.3 | 10.3 | 16.36 | 7.33 | 44   | 0.03 | 105.1 | 49  | 343 |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 4 | 0.9 | 9.99 | 15.66 | 7.28 | 43.6 | 0.03 | 100.6 | 213 | 345 |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 3 | 0.9 | 10.1 | 15.83 | 7.28 | 43.6 | 0.03 | 101.8 | 143 | 343 |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 2 | 1.6 | 10   | 15.39 | 7.25 | 43.6 | 0.03 | 100.5 | 112 | 343 |
| Round Lake | 2598 | DTASBH | 6/22/98 | 10:57:51 | 1.6 | 1 | 1.6 | 10.1 | 15.4  | 7.25 | 43.8 | 0.03 | 100.6 | 323 | 343 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 2798 | DTASBH | 7/7/98  | 9:38:04  | 2.0 | 4 | 0.4 | 9.08 | 20.76 | 7.32 | 53.3 | 0.03 | 100.9 | 135 | 340 |
| Round Lake | 2798 | DTASBH | 7/7/98  | 9:38:04  | 2.0 | 3 | 1   | 9.13 | 20.08 | 7.27 | 53.1 | 0.03 | 100.2 | 236 | 343 |
| Round Lake | 2798 | DTASBH | 7/7/98  | 9:38:04  | 2.0 | 2 | 1.5 | 9.18 | 19.9  | 7.26 | 53   | 0.03 | 100.4 | 203 | 342 |
| Round Lake | 2798 | DTASBH | 7/7/98  | 9:38:04  | 2.0 | 1 | 2   | 9.23 | 19.9  | 7.3  | 52.8 | 0.03 | 100.9 | 403 | 350 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 2998 | ASBH   | 7/21/98 | 10:20:54 | 1.5 | 4 | 0.4 | 8.75 | 23.47 | 7.41 | 54.5 | 0.03 | 102.6 | 56  | 306 |
| Round Lake | 2998 | ASBH   | 7/21/98 | 10:20:54 | 1.5 | 3 | 0.4 | 8.86 | 23.29 | 7.38 | 54.6 | 0.03 | 103.5 | 126 | 307 |
| Round Lake | 2998 | ASBH   | 7/21/98 | 10:20:54 | 1.5 | 2 | 1.6 | 9.43 | 22.04 | 7.47 | 54.1 | 0.03 | 107.6 | 104 | 303 |
| Round Lake | 2998 | ASBH   | 7/21/98 | 10:20:54 | 1.5 | 1 | 1.5 | 9.38 | 22.02 | 7.42 | 53.9 | 0.03 | 107   | 311 | 304 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 3298 | DBAS   | 8/11/98 | 10:29:32 | 1.2 | 3 | 0.4 | 8.24 | 23.86 | 7.4  | 60.7 | 0.04 | 97.2  | 53  | 306 |
| Round Lake | 3298 | DBAS   | 8/11/98 | 10:29:32 | 1.2 | 2 | 0.8 | 8.43 | 23.64 | 7.4  | 60.6 | 0.04 | 99    | 132 | 306 |
| Round Lake | 3298 | DBAS   | 8/11/98 | 10:29:32 | 1.2 | 1 | 1.2 | 8.46 | 23.49 | 7.42 | 60.7 | 0.04 | 99.1  | 25  | 306 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 3498 | ASGL   | 8/26/98 | 12:18:56 | .8  | 4 | 0.3 | 9.37 | 22.9  | 7.69 | 60.2 | 0.04 | 109.1 | 33  | 258 |
| Round Lake | 3498 | ASGL   | 8/26/98 | 12:18:56 | .8  | 3 | 0.3 | 9.41 | 22.9  | 7.7  | 60.2 | 0.04 | 109.6 | 53  | 256 |
| Round Lake | 3498 | ASGL   | 8/26/98 | 12:18:56 | .8  | 2 | 0.8 | 9.53 | 22.83 | 7.7  | 60.2 | 0.04 | 110.8 | 41  | 256 |
| Round Lake | 3498 | ASGL   | 8/26/98 | 12:18:56 | .8  | 1 | 0.8 | 9.55 | 22.83 | 7.7  | 60.2 | 0.04 | 111.1 | 216 | 255 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 3598 | DTAS   | 9/3/98  | 11:19:20 | 1.1 | 4 | 0.2 | 8.49 | 21.95 | 7.38 | 65.4 | 0.04 | 97.3  | 120 | 258 |
| Round Lake | 3598 | DTAS   | 9/3/98  | 11:19:20 | 1.1 | 3 | 0.2 | 8.42 | 21.78 | 7.34 | 65.4 | 0.04 | 96.1  | 109 | 258 |
| Round Lake | 3598 | DTAS   | 9/3/98  | 11:19:20 | 1.1 | 2 | 1.1 | 8.35 | 21.35 | 7.35 | 65.3 | 0.04 | 94.5  | 227 | 254 |
| Round Lake | 3598 | DTAS   | 9/3/98  | 11:19:20 | 1.1 | 1 | 1.1 | 8.45 | 21.37 | 7.38 | 65.2 | 0.04 | 95.7  | 142 | 251 |
|            |      |        |         |          |     |   |     |      |       |      |      |      |       |     |     |
| Round Lake | 4098 | KBDB   | 10/9/98 | 13:55:00 | 1.0 | 3 | 0.3 | 11.2 | 14.38 | 8.2  | 61.5 | 0.04 | 109.2 | 104 | 287 |
| Round Lake | 4098 | KBDB   | 10/9/98 | 13:55:00 | 1.0 | 2 | 0.6 | 11.6 | 14.21 | 8.39 | 61.3 | 0.04 | 112.9 | 131 | 284 |
| Round Lake | 4098 | KBDB   | 10/9/98 | 13:55:00 | 1.0 | 1 | 1.1 | 12.6 | 13.78 | 8.63 | 61   | 0.04 | 121.5 | 101 | 274 |

|            |      |        |          |          |    |   |     |      |      |      |      |      |       |     |     |
|------------|------|--------|----------|----------|----|---|-----|------|------|------|------|------|-------|-----|-----|
| Round Lake | 4298 | DTASJB | 10/19/98 | 10:22:53 | .7 | 3 | 0.7 | 11.9 | 9.63 | 8.28 | 61.1 | 0.04 | 103.1 | 40  | 334 |
| Round Lake | 4298 | DTASJB | 10/19/98 | 10:22:53 | .7 | 2 | 0.7 | 11.9 | 9.63 | 8.29 | 61   | 0.04 | 103.4 | 43  | 335 |
| Round Lake | 4298 | DTASJB | 10/19/98 | 10:22:53 | .7 | 1 | 0.7 | 12.1 | 9.58 | 8.3  | 61.1 | 0.04 | 104.7 | 114 | 336 |

| Location          | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 11       | 0.3       | 13.5                    | 5.65            | 7.07 | 45.1                 | 0.03 | 106.2                      | 37                 | 390   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 10       | 1.1       | 13.4                    | 5.46            | 7.1  | 45                   | 0.03 | 105.4                      | 47                 | 388   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 9        | 2.1       | 13.3                    | 5.04            | 7.07 | 44.6                 | 0.03 | 103.8                      | 119                | 389   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 8        | 3.1       | 13.3                    | 4.93            | 6.99 | 44.5                 | 0.03 | 102.8                      | 55                 | 391   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 7        | 4.1       | 13.3                    | 4.59            | 7    | 44.3                 | 0.03 | 102.1                      | 113                | 389   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 6        | 5.1       | 13.3                    | 4.34            | 7.07 | 44.8                 | 0.03 | 101.2                      | 103                | 385   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 5        | 6.1       | 13.3                    | 4.34            | 6.97 | 45                   | 0.03 | 101.2                      | 55                 | 389   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 4        | 7.1       | 13.3                    | 4.31            | 7    | 45.1                 | 0.03 | 101.3                      | 116                | 386   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 3        | 8.1       | 13.3                    | 4.31            | 6.95 | 45.5                 | 0.03 | 101.3                      | 139                | 387   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 2        | 9.1       | 13.3                    | 4.24            | 6.84 | 45.3                 | 0.03 | 101.1                      | 125                | 391   |
| Chatcolet Lake DP | 1098  | DBRP     | 3/13/98 | 13:19:09 | 1.0        | 1        | 10.1      | 13                      | 4.34            | 6.79 | 44.8                 | 0.03 | 99.3                       | 237                | 392   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 11       | 0.3       | 12.3                    | 7.43            | 7.1  | 39.2                 | 0.03 | 102.3                      | 54                 | 398   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 10       | 1.1       | 12.3                    | 7.43            | 7.04 | 39.4                 | 0.03 | 102.2                      | 38                 | 402   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 9        | 2.1       | 12.3                    | 7.4             | 7.07 | 39.1                 | 0.03 | 102.2                      | 56                 | 401   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 8        | 3.1       | 12.3                    | 7.33            | 7.13 | 39.1                 | 0.03 | 101.9                      | 53                 | 398   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 7        | 4.1       | 12.3                    | 7.24            | 7.06 | 39.1                 | 0.03 | 101.6                      | 59                 | 401   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 6        | 5.2       | 12.3                    | 7.19            | 7.09 | 38.9                 | 0.02 | 101.6                      | 40                 | 398   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 5        | 6.2       | 12.3                    | 7.04            | 7.12 | 38.6                 | 0.02 | 101.3                      | 100                | 396   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 4        | 7.2       | 12.2                    | 7.02            | 7.01 | 38.9                 | 0.02 | 100.9                      | 50                 | 401   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 3        | 8.1       | 12.2                    | 6.89            | 7.03 | 38.7                 | 0.02 | 100.6                      | 37                 | 398   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 2        | 9.2       | 12.2                    | 6.39            | 6.95 | 38.3                 | 0.02 | 99                         | 29                 | 401   |
| Chatcolet Lake DP | 1498  | DBAS     | 4/8/98  | 14:53:37 | 2.0        | 1        | 10.1      | 12.2                    | 6.36            | 6.98 | 38                   | 0.02 | 99                         | 204                | 403   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 11       | 0.4       | 11.9                    | 10.53           | 7.14 | 42.5                 | 0.03 | 106.4                      | 50                 | 392   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 10       | 1.4       | 11.9                    | 10.33           | 7.21 | 42.4                 | 0.03 | 106                        | 111                | 391   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 9        | 2.4       | 12                      | 9.79            | 7.16 | 42.3                 | 0.03 | 105                        | 100                | 394   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 8        | 3.5       | 11.9                    | 9.14            | 7.16 | 42.1                 | 0.03 | 103                        | 102                | 394   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 7        | 4.4       | 11.8                    | 8.61            | 7.08 | 42.8                 | 0.03 | 101.2                      | 117                | 397   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 6        | 5.4       | 11.8                    | 8.46            | 7.08 | 43.1                 | 0.03 | 100                        | 111                | 395   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 5        | 6.4       | 11.6                    | 7.95            | 6.98 | 43                   | 0.03 | 97.3                       | 105                | 400   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 4        | 7.4       | 11.6                    | 7.43            | 6.97 | 42.4                 | 0.03 | 96                         | 114                | 399   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 3        | 8.4       | 11.3                    | 6.9             | 6.91 | 41.1                 | 0.03 | 92.8                       | 134                | 401   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 2        | 9.4       | 11.2                    | 6.89            | 6.92 | 41.1                 | 0.03 | 91.6                       | 146                | 401   |
| Chatcolet Lake DP | 1698  | DBASRP   | 4/21/98 | 13:04:35 | 2.6        | 1        | 10.4      | 11                      | 7.27            | 6.95 | 40.9                 | 0.03 | 90.9                       | 218                | 401   |

|                   |      |        |         |          |     |    |      |      |       |      |      |      |       |      |     |
|-------------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|------|-----|
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 12 | 0.3  | 11.7 | 11.76 | 7.17 | 34.3 | 0.02 | 108.1 | 152  | 417 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 11 | 1.3  | 11.7 | 11.76 | 7.23 | 34.4 | 0.02 | 108.1 | 107  | 415 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 10 | 2.3  | 11.7 | 11.64 | 7.13 | 34.3 | 0.02 | 108.1 | 101  | 421 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 9  | 3.3  | 11.7 | 11.56 | 7.12 | 34.3 | 0.02 | 108.1 | 115  | 421 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 8  | 4.3  | 11.7 | 11.4  | 7.13 | 34.4 | 0.02 | 107.4 | 107  | 418 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 7  | 5.3  | 11.6 | 10.95 | 6.9  | 34.6 | 0.02 | 105.7 | 47   | 428 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 6  | 6.3  | 11.4 | 9.73  | 6.88 | 34.6 | 0.02 | 100.5 | 156  | 427 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 5  | 7.3  | 11.3 | 9.37  | 6.85 | 33.8 | 0.02 | 99.2  | 151  | 427 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 4  | 8.4  | 11.1 | 8.97  | 6.82 | 33.7 | 0.02 | 95.9  | 247  | 426 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 3  | 9.2  | 11.1 | 8.97  | 6.76 | 33.6 | 0.02 | 96.2  | 206  | 427 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 2  | 10.3 | 11.1 | 8.91  | 6.7  | 33.7 | 0.02 | 96.1  | 155  | 428 |
| Chatcolet Lake DP | 2098 | ASJLDT | 5/21/98 | 12:25:30 | 2.6 | 1  | 11.3 | 10.9 | 8.65  | 6.62 | 33.6 | 0.02 | 93.5  | 437  | 429 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 12 | 0.4  | 10.7 | 16.06 | 7.46 | 39.6 | 0.03 | 108.5 | 25   | 355 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 11 | 1.1  | 10.7 | 16.06 | 7.45 | 39.6 | 0.03 | 108.6 | 28   | 356 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 10 | 2.1  | 10.5 | 16.03 | 7.45 | 39.7 | 0.03 | 106.8 | 34   | 355 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 9  | 3.1  | 10.7 | 15.34 | 7.43 | 39.4 | 0.03 | 107.4 | 56   | 355 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 8  | 4.1  | 10.9 | 12.86 | 7.3  | 38.4 | 0.02 | 103   | 128  | 358 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 7  | 5.1  | 10.6 | 12.3  | 7.17 | 37.8 | 0.02 | 98.6  | 55   | 359 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 6  | 6.1  | 10.2 | 11.75 | 7.09 | 37.4 | 0.02 | 93.8  | 51   | 359 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 5  | 7.1  | 10.1 | 11.2  | 7.04 | 37   | 0.02 | 91.8  | 48   | 358 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 4  | 8.1  | 9.76 | 10.56 | 6.98 | 36.7 | 0.02 | 87.6  | 48   | 356 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 3  | 9.1  | 9.73 | 10.21 | 6.96 | 36.4 | 0.02 | 86.6  | 37   | 354 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 2  | 10.1 | 9.67 | 10.1  | 6.94 | 36.3 | 0.02 | 85.8  | 40   | 352 |
| Chatcolet Lake DP | 2398 | DTJLAS | 6/8/98  | 15:35:20 | 2.9 | 1  | 11.1 | 9.17 | 9.38  | 6.9  | 36.6 | 0.02 | 80.1  | 225  | 371 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 11 | 0.4  | 10.5 | 19.09 | 7.66 | 41.9 | 0.03 | 113.2 | 1021 | 339 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 10 | 1.4  | 10.7 | 18.1  | 7.62 | 41.7 | 0.03 | 113.6 | 131  | 343 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 9  | 2.4  | 11.1 | 16.44 | 7.59 | 41.6 | 0.03 | 113.2 | 52   | 346 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 8  | 3.4  | 10.3 | 15.52 | 7.3  | 40.9 | 0.03 | 103.3 | 227  | 352 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 7  | 4.4  | 9.42 | 14.44 | 7.07 | 39.9 | 0.03 | 92.4  | 117  | 356 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 6  | 5.4  | 9.42 | 13.86 | 7.03 | 39.9 | 0.03 | 91.2  | 105  | 356 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 5  | 6.4  | 8.9  | 13.18 | 6.88 | 38.6 | 0.02 | 84.8  | 55   | 358 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 4  | 7.4  | 8.13 | 12.15 | 6.78 | 37.3 | 0.02 | 75.7  | 120  | 359 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 3  | 8.4  | 7.63 | 11.33 | 6.72 | 36.8 | 0.02 | 69.7  | 200  | 358 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 2  | 9.4  | 7.14 | 11.11 | 6.7  | 37.3 | 0.02 | 64.9  | 141  | 355 |
| Chatcolet Lake DP | 2598 | DTASBH | 6/22/98 | 11:40:57 | 3.1 | 1  | 10.4 | 6.46 | 10.57 | 6.7  | 37.3 | 0.02 | 58    | 228  | 351 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 11 | 0.4  | 10.1 | 23.28 | 8.48 | 50.5 | 0.03 | 117.5 | 39   | 329 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 10 | 1.7  | 10.4 | 21.67 | 8.51 | 49.8 | 0.03 | 117.7 | 152  | 333 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 9  | 2.7  | 10.6 | 20.77 | 8.37 | 49.3 | 0.03 | 117.3 | 134  | 337 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 8  | 3.7  | 10.2 | 19.44 | 7.61 | 49.5 | 0.03 | 110.3 | 151  | 354 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 7  | 4.7  | 10.2 | 17.79 | 7.55 | 47.2 | 0.03 | 106.8 | 231  | 356 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 6  | 5.7  | 9.25 | 15.62 | 7.14 | 45.1 | 0.03 | 92.6  | 204  | 364 |

|                   |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|-------------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 5  | 6.7  | 7.9  | 14.66 | 6.9  | 44.1 | 0.03 | 77.4  | 304 | 367 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 4  | 7.7  | 7.16 | 13.71 | 6.74 | 43.3 | 0.03 | 68.8  | 102 | 369 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 3  | 8.7  | 5.81 | 12.85 | 6.64 | 42.6 | 0.03 | 54.7  | 247 | 369 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 2  | 9.7  | 4.78 | 12.16 | 6.58 | 42.9 | 0.03 | 44.4  | 126 | 367 |
| Chatcolet Lake DP | 2798 | DTASBH | 7/7/98  | 10:48:54 | 3.6 | 1  | 10.7 | 4.21 | 12.16 | 6.59 | 43.4 | 0.03 | 39    | 309 | 364 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 11 | 0.6  | 9.15 | 25.45 | 8.49 | 51.1 | 0.03 | 111.3 | 144 | 283 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 10 | 1.6  | 9.13 | 24    | 8.28 | 50.2 | 0.03 | 108.1 | 101 | 287 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 9  | 2.6  | 9.15 | 23.53 | 8.26 | 50.1 | 0.03 | 107.4 | 120 | 289 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 8  | 3.6  | 9.15 | 23.22 | 8.13 | 49.8 | 0.03 | 106.7 | 128 | 293 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 7  | 4.6  | 9.16 | 22.63 | 7.86 | 49   | 0.03 | 105.7 | 121 | 302 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 6  | 5.6  | 9.15 | 19.82 | 7.21 | 45.9 | 0.03 | 100   | 102 | 315 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 5  | 6.6  | 7.22 | 16.49 | 6.78 | 43.3 | 0.03 | 73.7  | 120 | 322 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 4  | 7.6  | 5.34 | 13.88 | 6.57 | 42.4 | 0.03 | 51.5  | 120 | 324 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 3  | 8.6  | 3.24 | 13.02 | 6.48 | 43.9 | 0.03 | 30.7  | 122 | 323 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 2  | 9.6  | 1.63 | 12.25 | 6.45 | 46   | 0.03 | 15.1  | 130 | 320 |
| Chatcolet Lake DP | 2998 | ASBH   | 7/21/98 | 12:25:45 | 6.8 | 1  | 10.6 | 0.97 | 12.44 | 6.49 | 47.9 | 0.03 | 9     | 528 | 317 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 11 | 0.3  | 9.4  | 23.84 | 8.81 | 54.5 | 0.03 | 110.8 | 35  | 245 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 10 | 1.3  | 9.49 | 23.76 | 8.84 | 54.6 | 0.03 | 111.8 | 213 | 243 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 9  | 2.3  | 9.62 | 23.6  | 8.9  | 55   | 0.04 | 112.9 | 53  | 234 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 8  | 3.3  | 9.6  | 23.55 | 8.86 | 55   | 0.04 | 112.5 | 58  | 231 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 7  | 4.4  | 9.42 | 23.38 | 8.78 | 54.2 | 0.03 | 110.1 | 120 | 228 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 6  | 5.3  | 8.93 | 23.2  | 8.66 | 53.8 | 0.03 | 104   | 325 | 221 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 5  | 6.4  | 5.71 | 18.37 | 6.71 | 46.6 | 0.03 | 60.6  | 212 | 238 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 4  | 7.4  | 2.52 | 15.69 | 6.54 | 46.5 | 0.03 | 25.3  | 235 | 212 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 3  | 8.4  | 0.79 | 14.32 | 6.51 | 48.6 | 0.03 | 7.7   | 130 | 167 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 2  | 9.4  | 0.16 | 13.28 | 6.55 | 55.3 | 0.04 | 1.5   | 116 | 131 |
| Chatcolet Lake DP | 3298 | DBAS   | 8/11/98 | 9:27:45  | 4.8 | 1  | 10.5 | 0.16 | 12.94 | 6.59 | 61.1 | 0.04 | 1.5   | 818 | 118 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 12 | 0.3  | 8.99 | 21.11 | 8.36 | 51.6 | 0.03 | 100.7 | 26  | 192 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 11 | 0.9  | 8.92 | 21.12 | 8.3  | 51.7 | 0.03 | 100   | 55  | 192 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 10 | 1.9  | 8.92 | 21.11 | 8.29 | 51.6 | 0.03 | 99.9  | 59  | 187 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 9  | 2.9  | 8.92 | 21.01 | 8.22 | 51.5 | 0.03 | 99.7  | 51  | 182 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 8  | 3.9  | 8.79 | 20.95 | 8.13 | 51.4 | 0.03 | 98.2  | 56  | 177 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 7  | 4.9  | 8.7  | 20.67 | 7.89 | 51.5 | 0.03 | 96.7  | 115 | 174 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 6  | 5.9  | 7.11 | 20.08 | 7.27 | 50.9 | 0.03 | 78.1  | 135 | 179 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 5  | 6.9  | 5.12 | 19.13 | 6.9  | 50.5 | 0.03 | 55.2  | 151 | 157 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 4  | 7.9  | 0.13 | 15.84 | 6.61 | 53.7 | 0.03 | 1.3   | 118 | 101 |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 3  | 8.9  | 0.16 | 13.73 | 6.66 | 61.6 | 0.04 | 1.5   | 107 | 73  |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 2  | 9.9  | 0.2  | 13.38 | 6.68 | 66.9 | 0.04 | 1.9   | 46  | 73  |
| Chatcolet Lake DP | 3498 | DTAS   | 8/26/98 | 13:25:04 | 4.0 | 1  | 10.9 | 0.23 | 13    | 6.68 | 73.2 | 0.05 | 2.2   | 212 | 77  |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98  | 13:51:50 | 6.1 | 12 | 0.3  | 9.23 | 22.17 | 8.35 | 57.8 | 0.04 | 106.1 | 35  | 205 |

|                   |      |        |          |          |     |    |      |      |       |      |      |      |       |      |     |
|-------------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|-------|------|-----|
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 11 | 1    | 9.3  | 21.74 | 8.38 | 58   | 0.04 | 106.1 | 140  | 202 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 10 | 1.9  | 9.24 | 21.49 | 8.34 | 57.8 | 0.04 | 104.9 | 142  | 198 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 9  | 2.9  | 9.07 | 21.04 | 8.25 | 57.3 | 0.04 | 102.1 | 153  | 193 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 8  | 3.9  | 8.92 | 20.9  | 8.09 | 57.1 | 0.04 | 100.2 | 140  | 188 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 7  | 4.9  | 7.84 | 20.63 | 7.65 | 56.5 | 0.04 | 87.6  | 235  | 190 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 6  | 5.9  | 7.39 | 20.32 | 7.27 | 56.6 | 0.04 | 82    | 145  | 176 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 5  | 6.9  | 2.71 | 19.35 | 6.84 | 55.3 | 0.04 | 29.5  | 129  | 135 |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 4  | 7.9  | 0.1  | 16.96 | 6.71 | 61.7 | 0.04 | 1.1   | 120  | 74  |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 3  | 8.9  | 0.11 | 14.47 | 6.74 | 70.2 | 0.04 | 1.1   | 220  | 42  |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 2  | 9.9  | 0.14 | 12.87 | 6.82 | 86   | 0.06 | 1.3   | 238  | 36  |
| Chatcolet Lake DP | 3598 | DTAS   | 9/2/98   | 13:51:50 | 6.1 | 1  | 10.9 | 0.17 | 12.62 | 6.81 | 92.6 | 0.06 | 1.6   | 345  | 45  |
|                   |      |        |          |          |     |    |      |      |       |      |      |      |       |      |     |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 12 | 0.3  | 9.63 | 14.91 | 7.82 | 58.4 | 0.04 | 95.2  | 34   | 302 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 11 | 0.6  | 9.63 | 14.91 | 7.79 | 58.3 | 0.04 | 95.2  | 42   | 303 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 10 | 1.6  | 9.59 | 14.66 | 7.72 | 58.1 | 0.04 | 94.2  | 57   | 304 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 9  | 2.6  | 9.17 | 14.53 | 7.62 | 58.1 | 0.04 | 89.9  | 134  | 307 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 8  | 3.5  | 8.87 | 14.49 | 7.53 | 58.2 | 0.04 | 86.8  | 44   | 308 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 7  | 4.5  | 8.72 | 14.43 | 7.49 | 58.3 | 0.04 | 85.2  | 50   | 309 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 6  | 5.6  | 8.53 | 14.38 | 7.48 | 58.7 | 0.04 | 83.2  | 15   | 308 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 5  | 6.6  | 8.4  | 14.36 | 7.53 | 58.6 | 0.04 | 82    | 52   | 307 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 4  | 7.4  | 8.44 | 14.35 | 7.61 | 58.4 | 0.04 | 82.4  | 54   | 304 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 3  | 8.5  | 9    | 14.31 | 7.79 | 58.7 | 0.04 | 87.8  | 108  | 300 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 2  | 9.5  | 9.11 | 14    | 7.97 | 58.8 | 0.04 | 88.2  | 123  | 297 |
| Chatcolet Lake DP | 4098 | KBDB   | 10/9/98  | 14:29:00 | 2.7 | 1  | 10.5 | 8.8  | 13.68 | 8.14 | 58.5 | 0.04 | 84.6  | 358  | 292 |
|                   |      |        |          |          |     |    |      |      |       |      |      |      |       |      |     |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 11 | 0.5  | 9.92 | 11.4  | 7.69 | 57.8 | 0.04 | 89.7  | 48   | 359 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 10 | 1.3  | 9.91 | 11.4  | 7.66 | 57.8 | 0.04 | 89.6  | 37   | 361 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 9  | 2.3  | 10.2 | 11.38 | 7.65 | 57.8 | 0.04 | 91.8  | 11   | 361 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 8  | 3.3  | 9.82 | 11.38 | 7.63 | 58.1 | 0.04 | 88.8  | 48   | 362 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 7  | 4.3  | 9.82 | 11.38 | 7.61 | 58.1 | 0.04 | 88.8  | 37   | 362 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 6  | 5.3  | 9.83 | 11.38 | 7.63 | 58.1 | 0.04 | 88.9  | 44   | 361 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 5  | 6.3  | 9.82 | 11.36 | 7.62 | 57.6 | 0.04 | 88.7  | 41   | 361 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 4  | 7.3  | 9.83 | 11.31 | 7.62 | 57.7 | 0.04 | 88.7  | 44   | 361 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 3  | 8.3  | 9.89 | 0     | 7.64 | 57.8 | 0.04 | 89.1  | 0    | 359 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 2  | 9.3  | 9.92 | 11.2  | 7.64 | 58.1 | 0.04 | 89.3  | 110  | 360 |
| Chatcolet Lake DP | 4298 | DTASJB | 10/19/98 | 10:05    | 2.3 | 1  | 10.3 | 9.82 | 0     | 7.58 | 57.7 | 0.04 | 88.4  | 1837 | 360 |
|                   |      |        |          |          |     |    |      |      |       |      |      |      |       |      |     |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 11 | 0.5  | 9.87 | 7.25  | 7.96 | 53.5 | 0.03 | 82.6  | 102  | 383 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 10 | 1.5  | 9.86 | 7.27  | 7.94 | 53.5 | 0.03 | 82.5  | 100  | 384 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 9  | 2.5  | 9.86 | 7.24  | 7.88 | 53.5 | 0.03 | 82.5  | 104  | 388 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 8  | 3.5  | 9.84 | 7.22  | 7.86 | 53.3 | 0.03 | 82.3  | 108  | 389 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 7  | 4.5  | 9.82 | 7.22  | 7.82 | 53.8 | 0.03 | 82.1  | 59   | 390 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 6  | 5.5  | 9.79 | 7.2   | 7.8  | 53.6 | 0.03 | 81.8  | 54   | 391 |
| Chatcolet Lake DP | 4698 | KBAS   | 11/17/98 | 10:40:00 | 2.7 | 5  | 6.5  | 9.8  | 7.17  | 7.79 | 53.6 | 0.03 | 104.1 | 4    | 391 |

|                   |      |      |          |          |     |   |     |      |      |      |      |      |      |     |     |
|-------------------|------|------|----------|----------|-----|---|-----|------|------|------|------|------|------|-----|-----|
| Chatcolet Lake DP | 4698 | KBAS | 11/17/98 | 10:40:00 | 2.7 | 4 | 6.5 | 9.79 | 0    | 7.78 | 53.8 | 0.03 | 81.8 | 40  | 392 |
| Chatcolet Lake DP | 4698 | KBAS | 11/17/98 | 10:40:00 | 2.7 | 3 | 7.6 | 9.79 | 7.14 | 7.76 | 53.3 | 0.03 | 81.7 | 150 | 392 |
| Chatcolet Lake DP | 4698 | KBAS | 11/17/98 | 10:40:00 | 2.7 | 2 | 8.5 | 9.71 | 7.17 | 7.73 | 53.3 | 0.03 | 81.1 | 120 | 393 |
| Chatcolet Lake DP | 4698 | KBAS | 11/17/98 | 10:40:00 | 2.7 | 1 | 9.5 | 9.75 | 7.1  | 7.7  | 52.9 | 0.03 | 81.3 | 913 | 394 |

| Location          | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|-------------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Chatcolet Lake SH | 2098  | ASJLDT   | 5/21/98 | 12:15:55 | 0.7        | 4        | 0.3       | 11.8                    | 9.35            | 6.95 | 35                   | 0.02 | 102.8                      | 47                 | 407   |
| Chatcolet Lake SH | 2098  | ASJLDT   | 5/21/98 | 12:15:55 | 0.7        | 3        | 0.3       | 11.8                    | 9.35            | 6.96 | 35                   | 0.02 | 102.7                      | 109                | 408   |
| Chatcolet Lake SH | 2098  | ASJLDT   | 5/21/98 | 12:15:55 | 0.7        | 2        | 0.7       | 11.8                    | 9.37            | 6.98 | 35.7                 | 0.02 | 103.1                      | 122                | 409   |
| Chatcolet Lake SH | 2098  | ASJLDT   | 5/21/98 | 12:15:55 | 0.7        | 1        | 0.7       | 11.9                    | 9.37            | 6.9  | 35.2                 | 0.02 | 103.9                      | 203                | 415   |
| Chatcolet Lake SH | 2398  | DTJLAS   | 6/8/98  | 15:18:55 | 0.7        | 4        | 0.3       | 10.8                    | 12.77           | 7.28 | 39.3                 | 0.03 | 102.2                      | 25                 | 365   |
| Chatcolet Lake SH | 2398  | DTJLAS   | 6/8/98  | 15:18:55 | 0.7        | 3        | 0.3       | 10.8                    | 12.79           | 7.27 | 39.3                 | 0.03 | 102                        | 32                 | 365   |
| Chatcolet Lake SH | 2398  | DTJLAS   | 6/8/98  | 15:18:55 | 0.7        | 2        | 0.7       | 10.8                    | 12.79           | 7.26 | 39.3                 | 0.03 | 102.2                      | 54                 | 373   |
| Chatcolet Lake SH | 2398  | DTJLAS   | 6/8/98  | 15:18:55 | 0.7        | 1        | 0.7       | 11                      | 12.77           | 7.24 | 39.3                 | 0.03 | 103.4                      | 155                | 375   |
| Chatcolet Lake SH | 2598  | DTASBH   | 6/22/98 | 12:03    | 0.8        | 4        | 0.2       | 10.1                    | 16.16           | 7.3  | 43.8                 | 0.03 | 102.8                      | 57                 | 320   |
| Chatcolet Lake SH | 2598  | DTASBH   | 6/22/98 | 12:03    | 0.8        | 3        | 0.2       | 10.1                    |                 | 7.31 | 43.7                 | 0.03 | 102.8                      | 246                | 316   |
| Chatcolet Lake SH | 2598  | DTASBH   | 6/22/98 | 12:03    | 0.8        | 2        | 0.8       | 10.2                    | 16.17           | 7.3  | 43.8                 | 0.03 | 103.4                      | 709                | 339   |
| Chatcolet Lake SH | 2598  | DTASBH   | 6/22/98 | 12:03    | 0.8        | 1        | 0.8       | 10.4                    | 16.29           | 7.25 | 43.7                 | 0.03 | 105.6                      | 142                | 333   |
| Chatcolet Lake SH | 2798  | DTASBH   | 7/7/98  | 10:37:02 | 0.8        | 4        | 0.3       | 11.3                    | 22.86           | 9.01 | 52.1                 | 0.03 | 130.6                      | 101                | 302   |
| Chatcolet Lake SH | 2798  | DTASBH   | 7/7/98  | 10:37:02 | 0.8        | 3        | 0.3       | 11.4                    | 22.65           | 9.02 | 52.2                 | 0.03 | 131.3                      | 137                | 303   |
| Chatcolet Lake SH | 2798  | DTASBH   | 7/7/98  | 10:37:02 | 0.8        | 2        | 0.8       | 11.8                    | 22.15           | 8.95 | 52.4                 | 0.03 | 134.3                      | 102                | 308   |
| Chatcolet Lake SH | 2798  | DTASBH   | 7/7/98  | 10:37:02 | 0.8        | 1        | 0.8       | 11.6                    | 22.06           | 8.88 | 52.2                 | 0.03 | 132.5                      | 510                | 314   |
| Chatcolet Lake SH | 2998  | ASBH     | 7/21/98 | 12:09:39 | 0.9        | 4        | 0.3       | 12.1                    | 26.62           | 9.3  | 56.2                 | 0.04 | 149.7                      | 56                 | 259   |
| Chatcolet Lake SH | 2998  | ASBH     | 7/21/98 | 12:09:39 | 0.9        | 3        | 0.3       | 11.9                    | 26.43           | 9.3  | 56.4                 | 0.04 | 146.9                      | 57                 | 257   |
| Chatcolet Lake SH | 2998  | ASBH     | 7/21/98 | 12:09:39 | 0.9        | 2        | 0.8       | 12.1                    | 22.66           | 9.02 | 54.3                 | 0.03 | 139.8                      | 159                | 266   |
| Chatcolet Lake SH | 2998  | ASBH     | 7/21/98 | 12:09:39 | 0.9        | 1        | 0.9       | 12.6                    | 23.06           | 9.34 | 54.2                 | 0.03 | 146                        | 249                | 261   |
| Chatcolet Lake SH | 3298  | DBAS     | 8/11/98 | 9:53:29  | 0.9        | 3        | 0.3       | 8.63                    | 23.31           | 8.12 | 62.3                 | 0.04 | 100.7                      | 57                 | 280   |
| Chatcolet Lake SH | 3298  | DBAS     | 8/11/98 | 9:53:29  | 0.9        | 2        | 0.6       | 8.32                    | 23.11           | 7.83 | 62.3                 | 0.04 | 96.8                       | 110                | 284   |
| Chatcolet Lake SH | 3298  | DBAS     | 8/11/98 | 9:53:29  | 0.9        | 1        | 0.9       | 7.92                    | 22.86           | 7.62 | 63.1                 | 0.04 | 91.8                       | 232                | 287   |
| Chatcolet Lake SH | 3498  | DTAS     | 8/27/98 | 8:52:58  | 1.0        | 4        | 0.3       | 8.05                    | 20.02           | 7.28 | 59.2                 | 0.04 | 88.3                       | 28                 | 237   |
| Chatcolet Lake SH | 3498  | DTAS     | 8/27/98 | 8:52:58  | 1.0        | 3        | 0.3       | 8.05                    | 20.06           | 7.25 | 59.3                 | 0.04 | 88.3                       | 104                | 237   |
| Chatcolet Lake SH | 3498  | DTAS     | 8/27/98 | 8:52:58  | 1.0        | 2        | 1         | 8.04                    | 20.04           | 7.3  | 59                   | 0.04 | 88.2                       | 48                 | 232   |
| Chatcolet Lake SH | 3498  | DTAS     | 8/27/98 | 8:52:58  | 1.0        | 1        | 1         | 8.11                    | 19.94           | 7.27 | 59.1                 | 0.04 | 88.8                       | 103                | 235   |
| Chatcolet Lake SH | 3598  | DTAS     | 9/3/98  | 11:06:09 | 0.8        | 4        | 0.2       | 10.3                    | 22.34           | 8.62 | 64.7                 | 0.04 | 119                        | 51                 | 225   |
| Chatcolet Lake SH | 3598  | DTAS     | 9/3/98  | 11:06:09 | 0.8        | 3        | 0.2       | 10.2                    | 22.33           | 8.59 | 64.5                 | 0.04 | 117.8                      | 57                 | 223   |

|                   |      |        |          |          |     |   |     |      |       |      |      |      |       |     |     |
|-------------------|------|--------|----------|----------|-----|---|-----|------|-------|------|------|------|-------|-----|-----|
| Chatcolet Lake SH | 3598 | DTAS   | 9/3/98   | 11:06:09 | 0.8 | 2 | 0.8 | 10.4 | 21.92 | 8.45 | 64.6 | 0.04 | 119.1 | 114 | 218 |
| Chatcolet Lake SH | 3598 | DTAS   | 9/3/98   | 11:06:09 | 0.8 | 1 | 0.8 | 10.3 | 21.95 | 8.37 | 64.5 | 0.04 | 118.2 | 325 | 215 |
| Chatcolet Lake SH | 4098 | KBDB   | 10/9/98  | 14:10:00 | 1.1 | 3 | 0.3 | 9.54 | 14.06 | 7.47 | 61.9 | 0.04 | 92.5  | 103 | 307 |
| Chatcolet Lake SH | 4098 | KBDB   | 10/9/98  | 14:10:00 | 1.1 | 2 | 0.7 | 9.9  | 13.93 | 7.49 | 61.7 | 0.04 | 95.7  | 120 | 308 |
| Chatcolet Lake SH | 4098 | KBDB   | 10/9/98  | 14:10:00 | 1.1 | 1 | 1.1 | 10.2 | 13.72 | 7.5  | 61.4 | 0.04 | 98.1  | 29  | 309 |
| Chatcolet Lake SH | 4298 | DTASJB | 10/19/98 | 10:13:52 | 0.9 | 3 | 0.9 | 13.7 | 8.37  | 9.36 | 59.8 | 0.04 | 115.2 | 107 | 313 |
| Chatcolet Lake SH | 4298 | DTASJB | 10/19/98 | 10:13:52 | 0.9 | 2 | 0.9 | 13.7 |       | 9.33 | 59.7 | 0.04 | 115.3 | 56  | 314 |
| Chatcolet Lake SH | 4298 | DTASJB | 10/19/98 | 10:13:52 | 0.9 | 1 | 0.9 | 13.7 | 8.29  | 9.33 | 59.9 | 0.04 | 115.3 | 158 | 314 |

| Location     | Phase | Samplers | Date    | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (µs/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|--------------|-------|----------|---------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 10       | 0.2       | 11.88                   | 10.32           | 6.95 | 34.2                 | 0.02 | 106.5                      | 43                 | 395   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 9        | 0.5       | 11.77                   | 10.3            | 7.09 | 34.1                 | 0.02 | 105.6                      | 114                | 388   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 8        | 1         | 11.75                   | 10.25           | 7.12 | 34.1                 | 0.02 | 105.2                      | 52                 | 386   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 7        | 1.5       | 11.72                   | 10.25           | 6.96 | 33.9                 | 0.02 | 105                        | 120                | 394   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 6        | 2         | 11.44                   | 9.82            | 6.98 | 34.1                 | 0.02 | 101.4                      | 125                | 391   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 5        | 2.5       | 11.48                   | 8.78            | 6.93 | 33.9                 | 0.02 | 99.2                       | 54                 | 392   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 4        | 3         | 11.3                    | 8.47            | 6.91 | 33.9                 | 0.02 | 97                         | 58                 | 391   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 3        | 3.5       | 11.26                   | 8.35            | 6.81 | 34                   | 0.02 | 96.4                       | 49                 | 395   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 2        | 4         | 11.13                   | 8.17            | 6.78 | 33.9                 | 0.02 | 94.8                       | 107                | 396   |
| Benewah Lake | 1398  | DBRP     | 4/3/98  | 12:37:53 | 0.95       | 1        | 4.4       | 10.68                   | 7.76            | 6.72 | 34.1                 | 0.02 | 90.1                       | 214                | 403   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 8        | 0.3       | 10.84                   | 14.46           | 6.96 | 36.8                 | 0.02 | 106                        | 222                | 384   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 7        | 1         | 10.84                   | 12.38           | 6.82 | 36.4                 | 0.02 | 101.2                      | 237                | 398   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 6        | 1.5       | 10.81                   | 12.05           | 6.9  | 36.4                 | 0.02 | 100.2                      | 202                | 391   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 5        | 2         | 10.64                   | 11.76           | 6.82 | 36.5                 | 0.02 | 97.9                       | 212                | 392   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 4        | 2.5       | 10.53                   | 11.29           | 6.75 | 36.3                 | 0.02 | 95.9                       | 104                | 389   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 3        | 3         | 10.09                   | 10.18           | 6.6  | 36.7                 | 0.02 | 89.5                       | 140                | 392   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 2        | 3.5       | 9.31                    | 9.26            | 6.56 | 36.8                 | 0.02 | 80.8                       | 204                | 390   |
| Benewah Lake | 1698  | DBASRP   | 4/21/98 | 11:47:22 | 1.0        | 1        | 4         | 7.42                    | 8.82            | 6.53 | 37.1                 | 0.02 | 63.7                       | 302                | 400   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 9        | 0.4       | 10.3                    | 15.33           | 7.01 | 37.1                 | 0.02 | 103.1                      | 35                 | 408   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 8        | 1         | 10.28                   | 15.21           | 7.01 | 37.1                 | 0.02 | 102.7                      | 39                 | 408   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 7        | 1.5       | 10.27                   | 15.14           | 7.01 | 37.1                 | 0.02 | 102.5                      | 58                 | 407   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 6        | 2         | 10.27                   | 15.09           | 6.98 | 37.2                 | 0.02 | 102.3                      | 48                 | 407   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 5        | 2.5       | 9.62                    | 13.77           | 6.71 | 38.1                 | 0.02 | 93.1                       | 57                 | 419   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 4        | 3         | 9.71                    | 12.2            | 6.73 | 37.4                 | 0.02 | 90.7                       | 125                | 414   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 3        | 3.5       | 8.55                    | 11.41           | 6.52 | 37.8                 | 0.02 | 78.5                       | 57                 | 420   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 2        | 4         | 7.96                    | 11.07           | 6.52 | 38.7                 | 0.02 | 72.4                       | 427                | 418   |
| Benewah Lake | 2098  | ASJLDT   | 5/21/98 | 13:35:46 | 3.2        | 1        | 4.5       | 5                       | 10.51           | 6.21 | 40.1                 | 0.03 | 44.9                       | 331                | 429   |

|              |      |        |         |          |     |    |     |       |       |      |      |      |       |     |     |
|--------------|------|--------|---------|----------|-----|----|-----|-------|-------|------|------|------|-------|-----|-----|
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 10 | 0.4 | 9.88  | 19.44 | 7.32 | 40.3 | 0.03 | 107.5 | 28  | 365 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 9  | 1   | 9.82  | 17.07 | 7.25 | 39.9 | 0.03 | 101.8 | 39  | 369 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 8  | 1.5 | 9.77  | 15.99 | 7.09 | 39.9 | 0.03 | 99.1  | 44  | 374 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 7  | 2   | 9.55  | 13.82 | 6.93 | 39.1 | 0.03 | 92.4  | 28  | 377 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 6  | 2.5 | 8.69  | 13.09 | 6.82 | 38.9 | 0.02 | 82.7  | 35  | 379 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 5  | 3   | 7.03  | 11.87 | 6.68 | 38.7 | 0.02 | 65    | 110 | 381 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 4  | 3.5 | 6.5   | 10.66 | 6.61 | 38.9 | 0.02 | 58.5  | 21  | 380 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 3  | 4   | 6.4   | 10.45 | 6.6  | 38.6 | 0.02 | 57.3  | 42  | 378 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 2  | 4.5 | 4.95  | 9.89  | 6.55 | 40   | 0.03 | 43.8  | 135 | 376 |
| Benewah Lake | 2398 | DTJLAS | 6/8/98  | 15:00:18 | 1.4 | 1  | 4.8 | 4.13  | 9.84  | 6.56 | 41   | 0.03 | 36.4  | 205 | 388 |
|              |      |        |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 8  | 0.4 | 10.17 | 20.46 | 7.71 | 40.9 | 0.03 | 112.9 | 38  | 330 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 7  | 1   | 10.41 | 19.56 | 7.69 | 40.8 | 0.03 | 113.5 | 126 | 331 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 6  | 1.5 | 10.87 | 17.71 | 7.55 | 40.8 | 0.03 | 114.3 | 203 | 334 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 5  | 2   | 10.53 | 16.95 | 7.36 | 41   | 0.03 | 109   | 208 | 336 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 4  | 2.5 | 10.42 | 15.99 | 7.22 | 41.1 | 0.03 | 105.6 | 201 | 336 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 3  | 3   | 10.39 | 15.47 | 7.18 | 41.2 | 0.03 | 104.2 | 232 | 335 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 2  | 3.5 | 7.77  | 14.18 | 6.79 | 40.3 | 0.03 | 75.7  | 320 | 336 |
| Benewah Lake | 2598 | DTASBH | 6/22/98 | 10:35    | 2.8 | 1  | 4.5 | 0.38  | 11.25 | 6.29 | 50.4 | 0.03 | 3.5   | 519 | 332 |
|              |      |        |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 10 | 0.3 | 10.28 | 23.49 | 8.55 | 49.7 | 0.03 | 120.5 | 40  | 356 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 9  | 0.7 | 10.19 | 23.15 | 8.49 | 49.3 | 0.03 | 118.6 | 111 | 358 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 8  | 1.2 | 10.12 | 22.93 | 8.44 | 49.4 | 0.03 | 117.3 | 48  | 359 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 7  | 1.7 | 10.2  | 22.52 | 8.25 | 50.1 | 0.03 | 117.3 | 125 | 365 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 6  | 2.2 | 10.85 | 21.85 | 8.21 | 49.9 | 0.03 | 123.2 | 107 | 366 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 5  | 2.7 | 11    | 20.35 | 7.76 | 50.1 | 0.03 | 121.3 | 140 | 372 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 4  | 3.2 | 10.47 | 18.18 | 7.41 | 48.5 | 0.03 | 110.6 | 153 | 375 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 3  | 3.7 | 10.6  | 15.96 | 7.37 | 47.3 | 0.03 | 106.9 | 309 | 369 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 2  | 4.2 | 3.79  | 13.73 | 6.56 | 50.1 | 0.03 | 36.4  | 204 | 368 |
| Benewah Lake | 2798 | DTASBH | 7/7/98  | 6:47:21  | 3.2 | 1  | 4.7 | 0.31  | 12.44 | 6.35 | 60.1 | 0.04 | 2.9   | 232 | 355 |
|              |      |        |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 10 | 0.3 | 9.2   | 25.92 | 8.41 | 50.1 | 0.03 | 112.8 | 103 | 254 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 9  | 0.8 | 9.22  | 25.86 | 8.41 | 50.1 | 0.03 | 112.9 | 131 | 251 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 8  | 1.3 | 9.28  | 25.36 | 8.47 | 50.3 | 0.03 | 112.6 | 59  | 245 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 7  | 1.8 | 9.59  | 24.9  | 8.53 | 50.1 | 0.03 | 115.4 | 109 | 238 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 6  | 2.3 | 9.24  | 24.64 | 8.45 | 50.1 | 0.03 | 110.7 | 20  | 232 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 5  | 2.8 | 9.22  | 24.35 | 8.51 | 50.2 | 0.03 | 109.9 | 110 | 218 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 4  | 3.5 | 10.19 | 20.83 | 7.63 | 48.2 | 0.03 | 113.5 | 113 | 229 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 3  | 3.8 | 10.48 | 18.85 | 7.86 | 46.8 | 0.03 | 112.4 | 115 | 209 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 2  | 4.3 | 9.77  | 16.69 | 7.75 | 49.5 | 0.03 | 100.2 | 248 | 199 |
| Benewah Lake | 2998 | ASBH   | 7/21/98 | 10:47:28 | 4.1 | 1  | 4.8 | 0.35  | 13.6  | 6.48 | 69.4 | 0.04 | 3.3   | 314 | 186 |
|              |      |        |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 3298 | DBAS   | 8/11/98 | 11:35:09 | 2.7 | 10 | 0.4 | 10.35 | 25.42 | 9.11 | 54.8 | 0.04 | 125.5 | 101 | 248 |
| Benewah Lake | 3298 | DBAS   | 8/11/98 | 11:35:09 | 2.7 | 9  | 0.9 | 10.77 | 23.91 | 9.17 | 54.5 | 0.03 | 127.1 | 102 | 245 |



|              |      |      |         |          |     |    |     |       |       |      |      |      |       |     |     |
|--------------|------|------|---------|----------|-----|----|-----|-------|-------|------|------|------|-------|-----|-----|
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 8  | 1.4 | 10.7  | 23.6  | 9.14 | 54.5 | 0.03 | 125.6 | 40  | 240 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 7  | 1.9 | 10.51 | 23.53 | 9.09 | 54.4 | 0.03 | 123.2 | 54  | 238 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 6  | 2.4 | 10.6  | 23.47 | 9.1  | 54.4 | 0.03 | 124.2 | 52  | 231 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 5  | 2.9 | 10.42 | 23.37 | 9.01 | 54.1 | 0.03 | 121.8 | 58  | 225 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 4  | 3.4 | 9.89  | 23.15 | 8.74 | 53.8 | 0.03 | 115.1 | 108 | 219 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 3  | 3.9 | 9     | 22.75 | 7.65 | 53.7 | 0.03 | 104   | 128 | 230 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 2  | 4.4 | 4.94  | 19.14 | 6.9  | 56.8 | 0.04 | 53.2  | 321 | 248 |
| Benewah Lake | 3298 | DBAS | 8/11/98 | 11:35:09 | 2.7 | 1  | 4.9 | 0.17  | 17.42 | 6.57 | 69   | 0.04 | 1.8   | 359 | 196 |
|              |      |      |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 10 | 0.4 | 9.3   | 20.88 | 8.53 | 52   | 0.03 | 103.7 | 31  | 224 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 9  | 1   | 9.29  | 20.67 | 8.49 | 51.9 | 0.03 | 103.2 | 108 | 226 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 8  | 1.5 | 9.24  | 20.62 | 8.47 | 51.8 | 0.03 | 102.6 | 49  | 222 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 7  | 2   | 9.32  | 20.56 | 8.47 | 51.8 | 0.03 | 103.4 | 49  | 218 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 6  | 2.5 | 9.38  | 20.48 | 8.44 | 52.2 | 0.03 | 103.7 | 112 | 214 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 5  | 3   | 8.95  | 20.28 | 8.04 | 52.2 | 0.03 | 98.7  | 111 | 216 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 4  | 3.5 | 8.71  | 20.23 | 7.76 | 52.6 | 0.03 | 95.9  | 128 | 220 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 3  | 4   | 7.72  | 20.04 | 7.28 | 52.9 | 0.03 | 84.7  | 124 | 225 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 2  | 4.5 | 2.98  | 19.71 | 6.79 | 54.8 | 0.04 | 32.4  | 223 | 218 |
| Benewah Lake | 3498 | DTAS | 8/27/98 | 9:04:12  | 3.3 | 1  | 4.9 | 0.14  | 18.46 | 6.6  | 75.7 | 0.05 | 1.4   | 322 | 163 |
|              |      |      |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 9  | 0.4 | 9.9   | 22.22 | 8.5  | 57.9 | 0.04 | 114   | 101 | 228 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 8  | 1   | 9.87  | 21.76 | 8.41 | 57.9 | 0.04 | 112.6 | 41  | 224 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 7  | 1.5 | 9.85  | 21.46 | 8.35 | 57.7 | 0.04 | 111.8 | 58  | 221 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 6  | 2   | 9.56  | 21.37 | 8.13 | 57.6 | 0.04 | 108.3 | 103 | 221 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 5  | 2.5 | 9.44  | 21.23 | 7.93 | 57.7 | 0.04 | 106.6 | 102 | 220 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 4  | 3   | 9.22  | 21.09 | 7.66 | 57.5 | 0.04 | 103.8 | 41  | 221 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 3  | 3.5 | 9.07  | 20.93 | 7.57 | 57.7 | 0.04 | 101.9 | 124 | 218 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 2  | 4   | 8.84  | 20.53 | 7.52 | 58.9 | 0.04 | 98.6  | 241 | 204 |
| Benewah Lake | 3598 | DTAS | 9/3/98  | 11:34:53 | 4.1 | 1  | 4.7 | 2.07  | 19.28 | 6.67 | 69.3 | 0.04 | 22.5  | 329 | 207 |
|              |      |      |         |          |     |    |     |       |       |      |      |      |       |     |     |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 7  | 0.3 | 9.7   | 13.39 | 7.37 | 55.4 | 0.04 | 92.7  | 120 | 299 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 6  | 1   | 9.47  | 12.96 | 7.34 | 55.3 | 0.04 | 89.5  | 117 | 305 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 5  | 1.6 | 9.2   | 12.89 | 7.27 | 54.9 | 0.04 | 86.9  | 53  | 305 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 4  | 2.2 | 9.23  | 12.81 | 7.27 | 55.2 | 0.04 | 87    | 105 | 304 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 3  | 2.6 | 9.22  | 12.78 | 7.26 | 55.1 | 0.04 | 86.8  | 110 | 303 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 2  | 3.2 | 9.62  | 12.73 | 7.28 | 55.3 | 0.04 | 90.5  | 244 | 302 |
| Benewah Lake | 4098 | KBDB | 10/9/98 | 13:33:00 | 2.0 | 1  | 3.7 | 8.88  | 12.69 | 7.09 | 55.1 | 0.04 | 83.5  | 320 | 301 |

| Location      | Phase | Samplers | Date   | Time     | Secchi (m) | Sequence | Depth (m) | Dissolved Oxygen (mg/L) | Temperature (C) | pH   | Conductivity (us/cm) | TDS  | Saturated Dissolved Oxygen | Stabilization Time | Redox |
|---------------|-------|----------|--------|----------|------------|----------|-----------|-------------------------|-----------------|------|----------------------|------|----------------------------|--------------------|-------|
| St. Joe River | 1398  | DBRP     | 4/3/98 | 11:51:33 | 3.5        | 12       | 0.4       | 12.6                    | 5.98            | 7.03 | 41.8                 | 0.03 | 101.7                      | 46                 | 392   |
| St. Joe River | 1398  | DBRP     | 4/3/98 | 11:51:33 | 3.5        | 11       | 2.1       | 12.4                    | 5.98            | 7.04 | 41.9                 | 0.03 | 100.1                      | 629                | 392   |
| St. Joe River | 1398  | DBRP     | 4/3/98 | 11:51:33 | 3.5        | 10       | 3.2       | 12.4                    | 6               | 7.02 | 41.9                 | 0.03 | 100.1                      | 458                | 396   |

|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|---------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 9  | 4.2  | 12.4 | 5.98  | 7.02 | 41.9 | 0.03 | 100.3 | 821 | 398 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 8  | 5.1  | 12.7 | 5.97  | 7.03 | 42   | 0.03 | 102.3 | 26  | 402 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 7  | 6.1  | 12.6 | 5.97  | 7.04 | 41.9 | 0.03 | 101.8 | 44  | 403 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 6  | 7.1  | 12.6 | 5.97  | 7.02 | 41.9 | 0.03 | 101.5 | 50  | 404 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 5  | 8.3  | 12.6 | 5.98  | 7.01 | 42   | 0.03 | 101.5 | 104 | 405 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 4  | 9.2  | 12.7 | 5.98  | 7.02 | 41.9 | 0.03 | 102.2 | 128 | 406 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 3  | 10.3 | 12.7 | 5.97  | 7.02 | 42.1 | 0.03 | 102.1 | 34  | 408 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 2  | 11.2 | 12.8 | 5.97  | 7.03 | 42   | 0.03 | 102.8 | 25  | 408 |
| St. Joe River | 1398 | DBRP   | 4/3/98  | 11:51:33 | 3.5 | 1  | 12.3 | 12.8 | 5.97  | 7.05 | 42   | 0.03 | 103.4 | 131 | 408 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 11 | 0.4  | 11.8 | 7.94  | 6.97 | 43.7 | 0.03 | 98.7  | 142 | 412 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 10 | 1.3  | 11.8 | 7.92  | 6.93 | 43.7 | 0.03 | 99.1  | 53  | 419 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 9  | 2.2  | 11.8 | 7.94  | 6.92 | 43.8 | 0.03 | 99.2  | 126 | 421 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 8  | 3.3  | 11.8 | 7.94  | 6.93 | 43.7 | 0.03 | 99.2  | 104 | 422 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 7  | 4.3  | 11.8 | 7.91  | 6.93 | 44   | 0.03 | 99.1  | 103 | 424 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 6  | 5.3  | 11.8 | 7.91  | 6.9  | 43.7 | 0.03 | 99.1  | 52  | 427 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 5  | 6.2  | 11.8 | 7.91  | 6.88 | 43.7 | 0.03 | 99.1  | 125 | 429 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 4  | 6.9  | 11.8 | 7.91  | 6.87 | 43.7 | 0.03 | 99    | 124 | 431 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 3  | 8.1  | 11.8 | 7.92  | 6.85 | 43.9 | 0.03 | 99.4  | 159 | 433 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 2  | 9.6  | 11.9 | 7.94  | 6.79 | 43.6 | 0.03 | 99.7  | 134 | 437 |
| St. Joe River | 1698 | DBASRP | 4/21/98 | 11:16:26 | 4.5 | 1  | 10.3 | 11.9 | 7.94  | 6.72 | 43.7 | 0.03 | 100.1 | 212 | 441 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 12 | 0.4  | 11.8 | 9.76  | 7.05 | 35.4 | 0.02 | 104   | 134 | 400 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 11 | 1.3  | 11.7 | 9.76  | 7.06 | 35.4 | 0.02 | 103.3 | 109 | 402 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 10 | 2.3  | 11.7 | 9.78  | 7.11 | 35.4 | 0.02 | 103.4 | 55  | 402 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 9  | 3.3  | 11.7 | 9.76  | 7.08 | 35.4 | 0.02 | 103.5 | 115 | 404 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 8  | 4.3  | 11.7 | 9.74  | 7.08 | 35.4 | 0.02 | 103.6 | 102 | 405 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 7  | 5.3  | 11.8 | 9.74  | 7.06 | 35.5 | 0.02 | 103.7 | 113 | 406 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 6  | 6.3  | 11.7 | 9.74  | 7.04 | 35.5 | 0.02 | 103.1 | 129 | 407 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 5  | 7.3  | 11.7 | 9.73  | 7.07 | 35.4 | 0.02 | 103.5 | 116 | 406 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 4  | 8.2  | 11.7 | 9.76  | 7.06 | 35.5 | 0.02 | 103.4 | 142 | 406 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 3  | 9.2  | 11.8 | 9.73  | 7.07 | 35.4 | 0.02 | 103.7 | 139 | 406 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 2  | 10.2 | 11.7 | 9.76  | 7.01 | 35.4 | 0.02 | 103.5 | 255 | 409 |
| St. Joe River | 2098 | ASJLDT | 5/21/98 | 12:56:22 | 4.0 | 1  | 11.2 | 11.8 | 9.73  | 7    | 35.5 | 0.02 | 104.1 | 337 | 411 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 14 | 0.3  | 11   | 12.27 | 7.34 | 39.4 | 0.03 | 103.1 | 28  | 370 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 13 | 1.1  | 11   | 12.23 | 7.35 | 39.4 | 0.03 | 102.9 | 42  | 371 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 12 | 2.1  | 11   | 12.21 | 7.34 | 39.4 | 0.03 | 102.6 | 57  | 372 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 11 | 3.1  | 11   | 12.26 | 7.34 | 39.5 | 0.03 | 102.6 | 38  | 373 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 10 | 4.1  | 11   | 12.2  | 7.34 | 39.5 | 0.03 | 102.5 | 53  | 373 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 9  | 5.1  | 11   | 12.18 | 7.34 | 39.4 | 0.03 | 102.4 | 133 | 373 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 8  | 6.1  | 11   | 12.13 | 7.32 | 39.5 | 0.03 | 102.3 | 44  | 374 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 7  | 7.1  | 11   | 12.12 | 7.33 | 39.6 | 0.03 | 102.4 | 48  | 374 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 6  | 8.1  | 11   | 12.16 | 7.33 | 39.6 | 0.03 | 102.1 | 59  | 374 |

|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
|---------------|------|--------|---------|----------|-----|----|------|------|-------|------|------|------|-------|-----|-----|
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 5  | 9.1  | 11   | 12.11 | 7.32 | 39.4 | 0.03 | 102.4 | 47  | 375 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 4  | 10.1 | 11   | 12.15 | 7.32 | 39.6 | 0.03 | 102.1 | 122 | 375 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 3  | 11.1 | 11   | 12.13 | 7.31 | 39.7 | 0.03 | 102.6 | 35  | 376 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 2  | 12.1 | 11   | 12.15 | 7.31 | 39.6 | 0.03 | 102.4 | 43  | 377 |
| St. Joe River | 2398 | DTJLAS | 6/8/98  | 14:34:30 | 4.4 | 1  | 13.1 | 10.9 | 12.16 | 7.3  | 39.6 | 0.03 | 101.6 | 223 | 377 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 12 | 0.3  | 9.85 | 15.1  | 7.24 | 43.9 | 0.03 | 98    | 36  | 360 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 11 | 1.8  | 9.84 | 15.02 | 7.24 | 44   | 0.03 | 97.7  | 57  | 362 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 10 | 2.8  | 9.84 | 14.97 | 7.23 | 44   | 0.03 | 97.6  | 130 | 362 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 9  | 3.8  | 9.85 |       | 7    | 44   | 0.03 | 97.5  | 128 | 362 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 8  | 4.8  | 9.84 | 14.87 | 7.23 | 43.9 | 0.03 | 97.3  | 205 | 362 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 7  | 5.8  | 9.84 | 14.86 | 7.22 | 43.8 | 0.03 | 97.4  | 220 | 363 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 6  | 6.8  | 9.85 | 14.86 | 7.22 | 44   | 0.03 | 97.5  | 122 | 363 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 5  | 7.8  | 9.87 | 14.86 | 7.22 | 43.9 | 0.03 | 97.7  | 143 | 363 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 4  | 8.8  | 9.88 | 14.84 | 7.22 | 43.8 | 0.03 | 97.7  | 224 | 364 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 3  | 9.8  | 9.89 | 14.87 | 7.22 | 43.9 | 0.03 | 97.9  | 205 | 364 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 2  | 10.8 | 9.91 | 14.87 | 7.21 | 43.9 | 0.03 | 98.1  | 122 | 364 |
| St. Joe River | 2598 | DTASBH | 6/22/98 | 10:22:51 | 4.1 | 1  | 11.8 | 9.96 | 14.84 | 7.2  | 43.9 | 0.03 | 98.5  | 227 | 365 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 13 | 0.3  | 9.13 | 20.97 | 7.39 | 53.5 | 0.03 | 101.9 | 134 | 362 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 12 | 1.2  | 9.24 | 20.51 | 7.34 | 53.3 | 0.03 | 102.2 | 124 | 367 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 11 | 2.2  | 9.21 | 20.15 | 7.31 | 53.2 | 0.03 | 101.2 | 134 | 370 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 10 | 3.2  | 9.16 | 19.94 | 7.31 | 53.2 | 0.03 | 100.2 | 140 | 372 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 9  | 4.2  | 9.16 | 19.9  | 7.29 | 53.2 | 0.03 | 100.1 | 235 | 373 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 8  | 5.2  | 9.17 | 19.85 | 7.29 | 53   | 0.03 | 100.1 | 130 | 374 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 7  | 6.2  | 9.16 | 19.8  | 7.29 | 53.3 | 0.03 | 99.9  | 146 | 375 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 6  | 7.2  | 9.07 | 19.38 | 7.26 | 53.1 | 0.03 | 98.1  | 130 | 377 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 5  | 8.2  | 9    | 19.19 | 7.24 | 53.2 | 0.03 | 97.1  | 119 | 378 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 4  | 9.2  | 9.01 | 19.19 | 7.23 | 53.6 | 0.03 | 97.2  | 103 | 380 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 3  | 10.2 | 9.03 | 19.21 | 7.23 | 53   | 0.03 | 97.3  | 112 | 381 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 2  | 11.2 | 9.01 | 19.19 | 7.23 | 53   | 0.03 | 97.2  | 201 | 382 |
| St. Joe River | 2798 | DTASBH | 7/7/98  | 9:26:52  | 4.0 | 1  | 12.2 | 9.06 | 19.19 | 7.23 | 53   | 0.03 | 97.7  | 220 | 383 |
|               |      |        |         |          |     |    |      |      |       |      |      |      |       |     |     |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 13 | 0.5  | 8.52 | 22.77 | 7.28 | 54.9 | 0.04 | 98.6  | 140 | 325 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 12 | 1.5  | 8.51 | 21.99 | 7.27 | 54.7 | 0.04 | 96.9  | 117 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 11 | 2.5  | 8.46 | 21.79 | 7.26 | 54.7 | 0.04 | 96.1  | 130 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 10 | 3.5  | 8.43 | 21.69 | 7.25 | 54.5 | 0.03 | 95.5  | 120 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 9  | 4.5  | 8.36 | 21.51 | 7.24 | 54.8 | 0.04 | 94.4  | 117 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 8  | 5.5  | 8.25 | 21.37 | 7.2  | 54.6 | 0.03 | 92.9  | 122 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 7  | 6.5  | 8.17 | 21.26 | 7.18 | 54.6 | 0.03 | 91.9  | 141 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 6  | 7.5  | 8.12 | 21.18 | 7.15 | 54.6 | 0.04 | 91.1  | 114 | 327 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 5  | 8.5  | 7.95 | 20.95 | 7.13 | 54.4 | 0.03 | 88.8  | 129 | 326 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 4  | 9.5  | 7.74 | 20.63 | 7.08 | 54.2 | 0.03 | 86    | 58  | 327 |
| St. Joe River | 2998 | ASBH   | 7/21/98 | 10:31:40 | 3.9 | 3  | 10.5 | 7.54 | 20.48 | 7.04 | 54.6 | 0.03 | 83.5  | 57  | 327 |

|               |      |      |         |          |     |    |      |      |       |      |      |      |      |     |     |
|---------------|------|------|---------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| St. Joe River | 2998 | ASBH | 7/21/98 | 10:31:40 | 3.9 | 2  | 11.5 | 7.53 | 20.28 | 7.03 | 54.2 | 0.03 | 83   | 111 | 327 |
| St. Joe River | 2998 | ASBH | 7/21/98 | 10:31:40 | 3.9 | 1  | 12.5 | 7.26 | 20.21 | 6.99 | 54.6 | 0.03 | 79.9 | 414 | 327 |
|               |      |      |         |          |     |    |      |      |       |      |      |      |      |     |     |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 13 | 0.4  | 7.89 | 22.59 | 7.31 | 60.8 | 0.04 | 91.3 | 56  | 290 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 12 | 1.2  | 7.54 | 22.15 | 7.22 | 60.9 | 0.04 | 86.5 | 122 | 294 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 11 | 2.2  | 7.38 | 21.83 | 7.2  | 60.8 | 0.04 | 84.2 | 139 | 294 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 10 | 3.2  | 7.28 | 21.67 | 7.17 | 60.6 | 0.04 | 82.8 | 143 | 295 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 9  | 4.2  | 7.23 | 21.62 | 7.19 | 60.6 | 0.04 | 82.2 | 218 | 293 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 8  | 5.2  | 7.22 | 21.58 | 7.18 | 60.8 | 0.04 | 82   | 219 | 291 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 7  | 6.2  | 7.2  | 21.53 | 7.16 | 60.9 | 0.04 | 81.6 | 58  | 290 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 6  | 7.2  | 7.17 | 21.51 | 7.16 | 60.8 | 0.04 | 81.3 | 145 | 289 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 5  | 8.2  | 7.16 | 21.49 | 7.16 | 60.8 | 0.04 | 81.2 | 141 | 288 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 4  | 9.2  | 7.08 | 21.48 | 7.15 | 61   | 0.04 | 80.2 | 123 | 286 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 3  | 10.2 | 7.03 | 21.46 | 7.14 | 60.9 | 0.04 | 79.6 | 126 | 285 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 2  | 11.2 | 6.96 | 21.44 | 7.13 | 61   | 0.04 | 78.8 | 133 | 283 |
| St. Joe River | 3498 | ASGL | 8/26/98 | 12:44:53 | 4.0 | 1  | 12.2 | 6.92 | 21.42 | 7.13 | 60.6 | 0.04 | 78.3 | 529 | 281 |
|               |      |      |         |          |     |    |      |      |       |      |      |      |      |     |     |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 13 | 0.3  | 7.89 | 21.6  | 7.25 | 65.2 | 0.04 | 89.7 | 19  | 281 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 12 | 1.7  | 7.84 | 20.93 | 7.22 | 65   | 0.04 | 88   | 21  | 281 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 11 | 2.7  | 7.7  | 20.72 | 7.21 | 64.9 | 0.04 | 86.2 | 20  | 281 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 10 | 3.7  | 7.68 | 20.7  | 7.16 | 65.1 | 0.04 | 85.8 | 17  | 283 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 9  | 4.7  | 7.65 | 20.7  | 7.15 | 65.2 | 0.04 | 85.5 | 28  | 283 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 8  | 5.7  | 7.55 | 20.65 | 7.13 | 64.8 | 0.04 | 84.3 | 15  | 282 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 7  | 6.7  | 7.5  | 20.65 | 7.14 | 65.2 | 0.04 | 83.8 | 19  | 281 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 6  | 7.7  | 7.35 | 20.65 | 7.12 | 65   | 0.04 | 82.1 | 29  | 281 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 5  | 8.7  | 7.33 | 20.63 | 7.11 | 65.3 | 0.04 | 81.9 | 44  | 280 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 4  | 9.7  | 7.18 | 20.6  | 7.08 | 65.3 | 0.04 | 80.1 | 42  | 280 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 3  | 10.7 | 6.83 | 20.53 | 7.03 | 65.4 | 0.04 | 76.1 | 54  | 278 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 2  | 11.7 | 6.35 | 20.39 | 6.99 | 65.9 | 0.04 | 70.6 | 54  | 276 |
| St. Joe River | 3598 | DTAS | 9/3/98  | 12:00:53 | 4.1 | 1  | 12.7 | 5.75 | 20.28 | 6.93 | 67.1 | 0.04 | 63.8 | 157 | 274 |
|               |      |      |         |          |     |    |      |      |       |      |      |      |      |     |     |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 14 | 0.3  | 9.26 | 14.01 | 7.33 | 61.7 | 0.04 | 89.7 | 48  | 329 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 13 | 0.6  | 9.21 | 14    | 7.31 | 61.7 | 0.04 | 89.1 | 130 | 330 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 12 | 1.6  | 9.11 | 13.93 | 7.28 | 61.5 | 0.04 | 88   | 49  | 331 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 11 | 2.6  | 8.97 | 13.77 | 7.23 | 61.5 | 0.04 | 86.4 | 110 | 333 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 10 | 3.5  | 8.96 | 13.76 | 7.23 | 61.6 | 0.04 | 86.3 | 54  | 333 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 9  | 4.5  | 8.96 | 13.75 | 7.22 | 61.6 | 0.04 | 86.3 | 105 | 333 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 8  | 5.4  | 8.96 | 13.75 | 7.22 | 61.5 | 0.04 | 86.3 | 126 | 333 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 7  | 6.5  | 8.96 | 13.75 | 7.22 | 61.8 | 0.04 | 86.3 | 121 | 332 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 6  | 7.6  | 8.96 | 13.75 | 7.22 | 61.6 | 0.04 | 86.3 | 137 | 331 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 5  | 8.6  | 8.96 | 13.75 | 7.22 | 61.4 | 0.04 | 86.3 | 113 | 331 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 4  | 9.5  | 8.94 | 13.75 | 7.23 | 61.8 | 0.04 | 86.1 | 120 | 330 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 3  | 10.1 | 8.94 | 13.73 | 7.22 | 61.5 | 0.04 | 86   | 217 | 331 |
| St. Joe River | 4098 | KBDB | 10/9/98 | 12:58:00 | 2.8 | 2  | 11.6 | 8.95 | 13.73 | 7.21 | 61.4 | 0.04 | 86.2 | 155 | 331 |

|               |      |        |          |          |     |    |      |      |       |      |      |      |      |     |     |
|---------------|------|--------|----------|----------|-----|----|------|------|-------|------|------|------|------|-----|-----|
| St. Joe River | 4098 | KBDB   | 10/9/98  | 12:58:00 | 2.8 | 1  | 12.6 | 8.96 | 13.73 | 7.19 | 61.8 | 0.04 | 86.3 | 338 | 332 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 12 | 0.5  | 9.79 | 10.66 | 7.29 | 60   | 0.04 | 87   | 129 | 355 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 11 | 1.6  | 9.78 | 0     | 7.27 | 60.1 | 0.04 | 86.5 | 31  | 357 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 10 | 2.6  | 9.75 | 10.4  | 7.27 | 59.9 | 0.04 | 86.2 | 42  | 357 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 9  | 3.6  | 9.74 | 10.4  | 7.26 | 59.9 | 0.04 | 86.1 | 34  | 358 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 8  | 4.6  | 9.73 | 10.36 | 7.23 | 60.2 | 0.04 | 85.9 | 32  | 359 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 7  | 5.6  | 9.73 | 10.35 | 7.25 | 59.9 | 0.04 | 85.8 | 32  | 358 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 6  | 6.6  | 9.72 | 10.35 | 7.24 | 60   | 0.04 | 85.7 | 35  | 358 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 5  | 7.6  | 9.73 | 10.33 | 7.25 | 60.2 | 0.04 | 85.9 | 34  | 357 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 4  | 8.6  | 9.73 | 10.35 | 7.25 | 60.1 | 0.04 | 85.8 | 31  | 357 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 3  | 9.6  | 9.74 | 10.33 | 7.25 | 60.2 | 0.04 | 86   | 42  | 356 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 2  | 10.6 | 9.63 | 10.2  | 7.25 | 59.9 | 0.04 | 84.7 | 517 | 356 |
| St. Joe River | 4298 | DTASJB | 10/19/98 | 10:41:51 | 2.6 | 1  | 10.6 | 9.82 | 10.22 | 7.33 | 59.9 | 0.04 | 86.4 | 122 | 353 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 10 | 0.5  | 9.68 | 5.24  | 7.19 | 53.9 | 0.03 | 77   | 110 | 401 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 9  | 1.7  | 9.68 | 5.26  | 7.19 | 54   | 0.03 | 77   | 49  | 402 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 8  | 2.7  | 9.66 | 5.29  | 7.16 | 53.9 | 0.03 | 77   | 48  | 405 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 7  | 3.7  | 9.66 | 5.29  | 7.15 | 53.8 | 0.03 | 77   | 103 | 405 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 6  | 4.7  | 9.67 | 5.27  | 7.16 | 53.8 | 0.03 | 77   | 52  | 405 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 5  | 5.7  | 9.68 | 5.29  | 7.16 | 53.8 | 0.03 | 77.1 | 49  | 405 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 4  | 6.7  | 9.7  | 5.24  | 7.16 | 54.2 | 0.03 | 77.1 | 100 | 406 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 3  | 7.7  | 9.7  | 5.24  | 7.16 | 54.1 | 0.03 | 77.1 | 107 | 407 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 2  | 8.7  | 9.72 | 5.22  | 7.15 | 54.1 | 0.03 | 77.3 | 117 | 409 |
| St. Joe River | 4698 | KB AS  | 11/17/98 | 9:48:00  | 3.6 | 1  | 9.7  | 9.65 | 5.29  | 7.14 | 54   | 0.03 | 76.9 | 857 | 410 |

## **Appendix B**

Water quality results from Coeur d'Alene Lake during 1997 and 1998.

Appendix B.1 Nitrate ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | 113      | 53      |
| Windy Bay Shallow      | <5       | 59      |
| Windy Bay Deep         | <5       | 53      |
| Coeur d'Alene River    | <5       | <5      |
| Mid Lake Coeur d'Alene | <5       | 49      |
| Carey Bay              | <5       | <5      |
| Conkling Point         | <5       | <5      |
| Hidden Lake            | <5       | <5      |
| Round Lake             | <5       | 67      |
| Chatcolet Lake Deep    | <5       | <5      |
| Chatcolet Lake Shallow | <5       | 43      |
| Benewah Lake           | <5       | -       |
| St. Joe River          | <5       | 66      |

- No samples taken to Spokane Tribal Laboratory.

Appendix B.2 Nitrate ( $\mu\text{g/L}$ ) results from the hypolimnion taken at ten stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | <5       | 50      |
| Windy Bay Shallow      | <5       | 63      |
| Windy Bay Deep         | 131      | 75      |
| Coeur d'Alene River    | <5       | 101     |
| Mid Lake Coeur d'Alene | 54       | 128     |
| Carey Bay              | <5       | 92      |
| Conkling Point         | <5       | 125     |
| Hidden Lake            | <5       | <5      |
| Chatcolet Lake Deep    | <5       | <5      |
| St. Joe River          | <5       | <5      |

Appendix B.3 Nitrate ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 33      | <5      | <5     | <5              | <5     | <5       | <5       |
| Windy Bay Shallow      | 53      | <5      | <5     | <5              | <5     | <5       | <5       |
| Windy Bay Deep         | 37      | <5      | <5     | <5              | <5     | <5       | <5       |
| Coeur d'Alene River    | 8       | <5      | <5     | <5              | <5     | 39       | 6        |
| Mid Lake Coeur d'Alene | <5      | <5      | <5     | <5              | <5     | <5       | 7        |
| Carey Bay              | 16      | <5      | <5     | <5              | <5     | <5       | 35       |
| Conkling Point         | 57      | <5      | <5     | <5              | <5     | <5       | 18       |
| Hidden Lake            | 210     | <5      | <5     | <5              | <5     | <5       | <5       |
| Round Lake             | 8       | <5      | <5     | <5              | <5     | <5       | -        |
| Chatcolet Lake Deep    | <5      | <5      | <5     | <5              | <5     | <5       | <5       |
| Chatcolet Lake Shallow | -       | <5      | <5     | <5              | <5     | <5       | -        |
| Benewah Lake           | <5      | <5      | <5     | <5              | <5     | -        | -        |
| St. Joe River          | <5      | <5      | <5     | <5              | <5     | <5       | 6        |

-No samples taken to Spokane Tribal Lab.

Appendix B.4 Nitrate ( $\mu\text{g/L}$ ) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 50      | <5      | <5     | <5              | <5     | <5       | <5       |
| Windy Bay Shallow      | 52      | <5      | <5     | <5              | <5     | <5       | <5       |
| Windy Bay Deep         | 49      | <5      | <5     | <5              | 29     | 39       | 51       |
| Coeur d'Alene River    | 19      | <5      | <5     | <5              | <5     | <5       | 28       |
| Mid Lake Coeur d'Alene | 62      | <5      | <5     | <5              | <5     | 78       | 65       |
| Carey Bay              | 8       | <5      | <5     | <5              | <5     | <5       | 11       |
| Conkling Point         | 36      | <5      | <5     | <5              | <5     | <5       | 16       |
| Hidden Lake            | <5      | <5      | <5     | <5              | <5     | -        | <5       |
| Chatcolet Lake Deep    | 21      | <5      | <5     | <5              | <5     | <5       | <5       |
| Benewah Lake           | <5      | <5      | <5     | <5              | <5     | -        | -        |
| St. Joe River          | <5      | <5      | <5     | <5              | <5     | <5       | 6        |

-No samples taken to Spokane Tribal Lab.



Appendix B.5 Nitrite ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | <29      | <29     |
| Windy Bay Shallow      | <29      | <29     |
| Windy Bay Deep         | <29      | <29     |
| Coeur d'Alene River    | <29      | <29     |
| Mid Lake Coeur d'Alene | <29      | <29     |
| Carey Bay              | <29      | <29     |
| Conkling Point         | <29      | <29     |
| Hidden Lake            | <29      | <29     |
| Round Lake             | <29      | <29     |
| Chatcolet Lake Deep    | <29      | <29     |
| Chatcolet Lake Shallow | <29      | <29     |
| Benewah Lake           | <29      | -       |
| St. Joe River          | <29      | <29     |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.6 Nitrite ( $\mu\text{g/L}$ ) results from the hypolimnion taken at ten stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | <29      | <29     |
| Windy Bay Shallow      | <29      | <29     |
| Windy Bay Deep         | <29      | <29     |
| Coeur d'Alene River    | <29      | <29     |
| Mid Lake Coeur d'Alene | <29      | <29     |
| Carey Bay              | <29      | <29     |
| Conkling Point         | <29      | <29     |
| Hidden Lake            | <29      | <29     |
| Chatcolet Lake Deep    | <29      | <29     |
| St. Joe River          | <29      | <29     |

Appendix B.7 Nitrite ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Windy Bay Shallow      | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Windy Bay Deep         | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Coeur d'Alene River    | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Mid Lake Coeur d'Alene | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Carey Bay              | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Conkling Point         | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Hidden Lake            | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Round Lake             | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Chatcolet Lake Deep    | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Chatcolet Lake Shallow | -       | <29     | <10    | <10             | <10    | <10      | <10      |
| Benewah Lake           | <29     | <29     | <10    | <10             | <10    | -        | -        |
| St. Joe River          | <29     | <29     | <10    | <10             | <10    | <10      | <10      |

-No samples taken to Spokane Tribal Lab.

Appendix B.8 Nitrite ( $\mu\text{g/L}$ ) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Windy Bay Shallow      | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Windy Bay Deep         | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Coeur d'Alene River    | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Mid Lake Coeur d'Alene | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Carey Bay              | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Conkling Point         | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Hidden Lake            | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Chatcolet Lake Deep    | <29     | <29     | <10    | <10             | <10    | <10      | <10      |
| Benewah Lake           | <29     | <29     | <10    | <10             | <10    | -        | -        |
| St. Joe River          | <29     | <29     | <10    | <10             | <10    | <10      | <10      |

-No samples taken to Spokane Tribal Lab.

Appendix B.9 Total kjeldahl nitrogen ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | Date   |         |          |
|------------------------|--------|---------|----------|
|                        | 7/8/98 | 8/10/98 | 10/20/98 |
| Rockford Bay           | 200    | 132     | 160      |
| Windy Bay Shallow      | <120   | 124     | 227      |
| Windy Bay Deep         | 140    | <120    | 212      |
| Coeur d'Alene River    | <120   | 232     | 157      |
| Mid Lake Coeur d'Alene | <120   | <120    | <120     |
| Carey Bay              | 240    | <120    | <120     |
| Conkling Point         | 136    | <120    | <120     |
| Hidden Lake            | 256    | <120    | <120     |
| Round Lake             | <120   | 120     | <120     |
| Chatcolet Lake Deep    | 172    | <120    | <120     |
| Chatcolet Lake Shallow | 228    | <120    | <120     |
| Benewah Lake           | <120   | 160     | -        |
| St. Joe River          | <120   | <120    | <120     |

-No samples taken to Spokane Tribal Lab.

Appendix B.10 Total kjeldahl nitrogen ( $\mu\text{g/L}$ ) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | Date   |         |          |
|------------------------|--------|---------|----------|
|                        | 7/8/98 | 8/10/98 | 10/20/98 |
| Rockford Bay           | 316    | <120    | 156      |
| Windy Bay Shallow      | <120   | 124     | <120     |
| Windy Bay Deep         | <120   | 188     | 157      |
| Coeur d'Alene River    | <120   | 132     | 124      |
| Mid Lake Coeur d'Alene | <120   | <120    | 489      |
| Carey Bay              | 248    | <120    | <120     |
| Conkling Point         | 148    | <120    | <120     |
| Hidden Lake            | 312    | <120    | -        |
| Chatcolet Lake Deep    | 232    | 120     | <120     |
| Benewah Lake           | <120   | 136     | -        |
| St. Joe River          | <120   | <120    | <120     |

-No samples taken to Spokane Tribal Lab.

Appendix B.11 Ortho-Phosphate ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | <26      | <26     |
| Windy Bay Shallow      | <26      | <26     |
| Windy Bay Deep         | <26      | <26     |
| Coeur d'Alene River    | <26      | <26     |
| Mid Lake Coeur d'Alene | <26      | <26     |
| Carey Bay              | <26      | <26     |
| Conkling Point         | <26      | <26     |
| Hidden Lake            | <26      | <26     |
| Round Lake             | <26      | <26     |
| Chatcolet Lake Deep    | <26      | <26     |
| Chatcolet Lake Shallow | <26      | <26     |
| Benewah Lake           | <26      | -       |
| St. Joe River          | <26      | <26     |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.12 Ortho-Phosphate ( $\mu\text{g/L}$ ) results from the hypolimnion taken at ten stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date     |         |
|------------------------|----------|---------|
|                        | 10/20/97 | 11/4/97 |
| Rockford Bay           | <26      | <26     |
| Windy Bay Shallow      | <26      | <26     |
| Windy Bay Deep         | <26      | <26     |
| Coeur d'Alene River    | <26      | <26     |
| Mid Lake Coeur d'Alene | <26      | <26     |
| Carey Bay              | <26      | <26     |
| Conkling Point         | <26      | <26     |
| Hidden Lake            | <26      | <26     |
| Chatcolet Lake Deep    | <26      | <26     |
| St. Joe River          | <26      | <26     |

Appendix B.13 Ortho-phosphate ( $\mu\text{g/L}$ ) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/25/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Windy Bay Shallow      | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Windy Bay Deep         | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Coeur d'Alene River    | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Mid Lake Coeur d'Alene | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Carey Bay              | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Conkling Point         | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Hidden Lake            | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Round Lake             | <26     | <26     | <20    | <20             | <20    | <20      | -        |
| Chatcolet Lake Deep    | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Chatcolet Lake Shallow | -       | <26     | <20    | <20             | <20    | <20      | -        |
| Benewah Lake           | <26     | <26     | <20    | <20             | <20    | -        | -        |
| St. Joe River          | <26     | <26     | <20    | <20             | <20    | <20      | <20      |

-No sample taken to Spokane Tribal Lab.

Appendix B.14 Ortho-phosphate ( $\mu\text{g/L}$ ) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Windy Bay Shallow      | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Windy Bay Deep         | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Coeur d'Alene River    | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Mid Lake Coeur d'Alene | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Carey Bay              | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Conkling Point         | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Hidden Lake            | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Chatcolet Lake Deep    | <26     | <26     | <20    | <20             | <20    | <20      | <20      |
| Benewah Lake           | <26     | <26     | <20    | <20             | <20    | -        | -        |
| St. Joe River          | <26     | <26     | <20    | <20             | <20    | <20      | <20      |

-No sample taken to Spokane Tribal Lab.



Appendix B.17 Chlorophyll<sub>a</sub> (µg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date    |              |          |                      |
|------------------------|---------|--------------|----------|----------------------|
|                        | 8/13/97 | 9/16/97      | 10/20/97 | 11/4/97 <sup>a</sup> |
| Rockford Bay           | <0.01   | 0.70         | <0.01    | 3.30                 |
| Windy Bay Shallow      | <0.01   | <0.01        | 0.74     | 2.14                 |
| Windy Bay Deep         | <0.01   | 1.34         | 2.23     | 1.38                 |
| Coeur d'Alene River    | 0.65    | <0.01        | -        | 2.96                 |
| Mid Lake Coeur d'Alene | 1.42    | 0.64         | 1.90     | 2.91                 |
| Carey Bay              | 0.64    | <0.01        | -        | 1.28                 |
| Conkling Point         | <0.01   | 2.16         | 1.38     | 3.39                 |
| Hidden Lake            | <0.01   | Contaminated | 3.62     | 6.89                 |
| Round Lake             | <0.01   | 1.54         | 0.67     | 0.64                 |
| Chatcolet Lake Deep    | 3.14    | 2.81         | 4.26     | 7.66                 |
| Chatcolet Lake Shallow | 5.32    | 4.89         | 2.78     | 0.69                 |
| Benewah Lake           | 0.64    | 25.79        | 3.54     | -                    |
| St. Joe River          | -       | Contaminated | 1.40     | <0.01                |

- No sample taken to Spokane Tribal Laboratory.

<sup>a</sup> Coeur d'Alene Lake was isothermal.

Appendix B.18 Chlorophyll<sub>a</sub> (µg/L) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date    |              |          |                      |
|------------------------|---------|--------------|----------|----------------------|
|                        | 8/13/97 | 9/16/97      | 10/20/97 | 11/4/97 <sup>a</sup> |
| Rockford Bay           | 1.40    | 2.09         | -        | -                    |
| Windy Bay Shallow      | 2.14    | 0.68         | -        | -                    |
| Windy Bay Deep         | 0.64    | 0.64         | <0.01    | -                    |
| Coeur d'Alene River    | 0.69    | 1.32         | <0.01    | -                    |
| Mid Lake Coeur d'Alene | 0.69    | 2.07         | 1.38     | -                    |
| Carey Bay              | <0.01   | 1.33         | 1.4      | -                    |
| Conkling Point         | 0.005   | 1.37         | <0.01    | -                    |
| Hidden Lake            | 11.92   | 20.25        | 1.98     | -                    |
| Chatcolet Lake Deep    | 6.03    | Contaminated | 14.04    | -                    |
| Benewah Lake           | 34.14   | 28.78        | -        | -                    |
| St. Joe River          | -       | 0.70         | <0.01    | -                    |

- No sample taken to Spokane Tribal Laboratory.

<sup>a</sup> Coeur d'Alene Lake was isothermal.

Appendix B.19 Chlorophyll<sub>a</sub> (µg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 1.240   | 1.49    | <0.01  | <0.01           | 1.31   | 1.89     | 0.71     |
| Windy Bay Shallow      | 2.080   | 2.11    | <0.01  | 0.71            | 0.066  | <0.01    | 1.85     |
| Windy Bay Deep         | 1.290   | 0.7     | 2.85   | 2.01            | 2.25   | <0.01    | 1.24     |
| Coeur d'Alene River    | 1.310   | 1.22    | 0.73   | <0.01           | <0.01  | <0.01    | 2.46     |
| Mid Lake Coeur d'Alene | 1.890   | 2.22    | 0.69   | <0.01           | 1.43   | 0.7      | 0.66     |
| Carey Bay              | 1.920   | 3.14    | 1.79   | 2               | 2.09   | 0.69     | 1.26     |
| Conkling Point         | <0.01   | 1.4     | 2.58   | <0.01           | 2.78   | 0.63     | 2.63     |
| Hidden Lake            | <0.01   | 1.88    | 3.57   | 30.84           | 1.38   | 6.04     | 2.87     |
| Round Lake             | <0.01   | <0.01   | 1.29   | 0.67            | 3.49   | 3.25     | -        |
| Chatcolet Lake Deep    | <0.01   | 2.55    | 4.69   | 4.09            | 3.42   | 0.69     | 7.21     |
| Chatcolet Lake Shallow | -       | 0.75    | 8.96   | 2.61            | 2.77   | 8.38     | -        |
| Benewah Lake           | 1.470   | 4.05    | 23.6   | 61.41           | 0.65   | -        | -        |
| St. Joe River          | 0.640   | 1.47    | 0.67   | <0.01           | 2.68   | 3.91     | 1.38     |

-No samples taken to Spokane Tribal Laboratory.

Appendix B.20 Chlorophyll<sub>a</sub> (µg/L) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 1.930   | 0.7     | 2.1    | 4.09            | 1.41   | 2.17     | 0.65     |
| Windy Bay Shallow      | 3.440   | <0.01   | 1.44   | 2.07            | 0.71   | <0.01    | <0.01    |
| Windy Bay Deep         | 1.250   | 2.96    | 0.67   | <0.01           | 2.86   | <0.01    | 0.66     |
| Coeur d'Alene River    | <0.01   | <0.01   | 1.99   | 0.65            | 2.67   | 1.42     | 2.84     |
| Mid Lake Coeur d'Alene | 0.590   | 0.7     | <0.01  | 0.62            | 2.02   | <0.01    | <0.01    |
| Carey Bay              | 1.260   | <0.01   | 0.72   | 2.08            | 2.65   | 1.97     | 1.26     |
| Conkling Point         | <0.01   | <0.01   | 0.7    | 0.68            | 2.24   | 3.43     | 2.75     |
| Hidden Lake            | 5.490   | 6.16    | 16.84  | 80.92           | 18.72  | -        | 5.22     |
| Chatcolet Lake Deep    | 2.030   | <0.01   | 3.31   | 4.42            | 3.06   | 8.47     | 6.76     |
| Benewah Lake           | 2.05    | 9.18    | 15.26  | 54.46           | 15.95  | -        | -        |
| St. Joe River          | <0.01   | 1.27    | 0.74   | 4.69            | <0.01  | 5.83     | 0.63     |

-No samples taken to Spokane Tribal Laboratory.



Appendix B.21 Total suspended solids (mg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho 1997.

| Location               | Date    |         |          |         |
|------------------------|---------|---------|----------|---------|
|                        | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| Rockford Bay           | <2      | 2       | <0.5     | <0.5    |
| Windy Bay Shallow      | <2      | 2       | <0.5     | <0.5    |
| Windy Bay Deep         | <2      | 2       | <0.5     | <0.5    |
| Coeur d'Alene River    | 2       | 2       | <0.5     | <0.5    |
| Mid Lake Coeur d'Alene | <2      | 2       | <0.5     | 1.5     |
| Carey Bay              | <2      | 2       | <0.5     | 4.0     |
| Conkling Point         | <2      | 2       | <0.5     | 6.5     |
| Hidden Lake            | 10      | 97      | 2.5      | <0.5    |
| Round Lake             | <2      | 2       | 1.5      | 16.0    |
| Chatcolet Lake Deep    | <2      | 2       | 2.0      | <0.5    |
| Chatcolet Lake Shallow | 10      | 16      | 6.5      | 4.0     |
| Benewah Lake           | 10      | 3       | 8.0      | -       |
| St. Joe River          | -       | 2       | 5.0      | 3.0     |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.22 Total suspended solids (mg/L) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date    |         |              |         |
|------------------------|---------|---------|--------------|---------|
|                        | 8/13/97 | 9/16/97 | 10/20/97     | 11/4/97 |
| Rockford Bay           | <2      | 2.0     | <0.5         | <0.5    |
| Windy Bay Shallow      | <2      | 2.0     | <0.5         | <0.5    |
| Windy Bay Deep         | 2       | 2.0     | <0.5         | 29.0    |
| Coeur d'Alene River    | 3       | 2.0     | <0.5         | 1.5     |
| Mid Lake Coeur d'Alene | <2      | 2.0     | 1.0          | 3.0     |
| Carey Bay              | 210     | 2.0     | 2.5          | 4.0     |
| Conkling Point         | <2      | 2.0     | <0.5         | 7.5     |
| Hidden Lake            | 10      | 2.000   | 3.000<br>3.0 | <0.5    |
| Chatcolet Lake Deep    | 20      | 5.0     | 12.5         | <0.5    |
| Benewah Lake           | 20      | 7       | -            | -       |
| St. Joe River          | -       | 2.0     | 4.5          | <0.5    |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.23 Total suspended solids (mg/L) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | <2      | 2       | 3      | <2              | <2     | <2       | <2       |
| Windy Bay Shallow      | <2      | 3       | 2      | 2               | <2     | <2       | <2       |
| Windy Bay Deep         | 5.5     | 2       | 3      | 3               | <2     | <2       | <2       |
| Coeur d'Alene River    | 3       | 3       | 2      | <2              | <2     | <2       | <2       |
| Mid Lake Coeur d'Alene | 3       | <2      | 2      | <2              | <2     | <2       | 2.5      |
| Carey Bay              | 4       | <2      | 2      | 10              | <2     | 2        | <2       |
| Conkling Point         | 3       | <2      | 3      | 2               | <2     | <2       | <2       |
| Hidden Lake            | 3       | <2      | 4      | 7.5             | <2     | 6.5      | 3.5      |
| Round Lake             | 4.5     | 4       | <2     | <2              | 2.5    | 9        | -        |
| Chatcolet Lake Deep    | <2      | 2       | 2      | <2              | <2     | 5.5      | 6.75     |
| Chatcolet Lake Shallow | -       | 3       | <2     | <2              | <2     | 10.500   | -        |
| Benewah Lake           | 4       | 4       | 8      | 4.5             | <2     | -        | -        |
| St. Joe River          | 3       | 2       | 2      | 2.5             | <2     | 2.5      | 3.75     |

-No samples taken to Spokane Tribal Lab.

Appendix B.24 Total suspended solids (mg/L) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 10      | 4       | 3      | 4.5             | 2      | <2       | <2       |
| Windy Bay Shallow      | <2      | 3       | <2     | 12              | 3      | 2.5      | <2       |
| Windy Bay Deep         | 27      | <2      | 3      | 2.5             | <2     | <2       | <2       |
| Coeur d'Alene River    | 8       | 3       | 4      | 2               | 2.5    | 7.75     | 3.75     |
| Mid Lake Coeur d'Alene | <2      | <2      | 14     | <2              | <2     | 4.25     | 3.25     |
| Carey Bay              | <2      | <2      | 4      | 10              | 11     | 5.5      | <2       |
| Conkling Point         | <2      | <2      | 3      | <2              | 2      | 4.5      | 4.5      |
| Hidden Lake            | 47      | 4       | 9      | 11              | 14.5   | -        | 3.25     |
| Chatcolet Lake Deep    | 28.5    | 141     | 4      | 3               | 3      | 3        | 7        |
| Benewah Lake           | 4.5     | 16      | 7      | <2              | 3.5    | -        | -        |
| St. Joe River          | 3.5     | 2       | 9      | 3               | <2     | 2.5      | 3        |

-No samples taken to Spokane Tribal Lab.

Appendix B.25 Turbidity (NTU) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date    |         |          |         |
|------------------------|---------|---------|----------|---------|
|                        | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| Rockford Bay           | 0.35    | 0.30    | 0.27     | 0.24    |
| Windy Bay Shallow      | 0.63    | 0.33    | 0.27     | 0.25    |
| Windy Bay Deep         | 0.28    | 0.4     | 0.23     | 0.27    |
| Coeur d'Alene River    | 0.87    | 0.51    | 0.51     | 0.43    |
| Mid Lake Coeur d'Alene | 0.33    | 1.99    | 0.84     | 0.38    |
| Carey Bay              | 0.56    | 0.84    | 0.88     | 0.49    |
| Conkling Point         | 0.54    | 0.96    | 0.76     | 0.86    |
| Hidden Lake            | 4.76    | 1.56    | 0.96     | 1.42    |
| Round Lake             | 3.820   | 0.45    | 1.88     | 0.93    |
| Chatcolet Lake Deep    | 1.01    | 1.39    | 1.08     | 1.61    |
| Chatcolet Lake Shallow | 3.69    | 0.77    | 2.01     | 11.20   |
| Benewah Lake           | 5.81    | 1.55    | 1.60     | -       |
| St. Joe River          | -       | 0.62    | 1.87     | 7.86    |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.26 Turbidity (NTU) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1997.

| Location               | Date    |         |          |         |
|------------------------|---------|---------|----------|---------|
|                        | 8/13/97 | 9/16/97 | 10/20/97 | 11/4/97 |
| Rockford Bay           | 0.6     | 0.37    | 0.24     | 0.27    |
| Windy Bay Shallow      | 0.92    | 0.42    | 0.28     | 0.85    |
| Windy Bay Deep         | 0.86    | 0.87    | 0.67     | 0.74    |
| Coeur d'Alene River    | 1.02    | 0.51    | 0.85     | 0.88    |
| Mid Lake Coeur d'Alene | 1.3     | 0.82    | 1.13     | 0.76    |
| Carey Bay              | 1.32    | 1.34    | 2.23     | 0.70    |
| Conkling Point         | 1.87    | 2.00    | 0.92     | 1.01    |
| Hidden Lake            | 1.31    | 3.60    | 1.06     | 1.39    |
| Chatcolet Lake Deep    | 3.52    | 1.90    | 2.63     | 3.68    |
| Benewah Lake           | 0.81    | 2.01    | -        | -       |
| St. Joe River          | -       | 0.76    | 1.83     | 7.68    |

- No sample taken to Spokane Tribal Laboratory.

Appendix B.27 Turbidity (NTU) results from the epilimnion taken at thirteen stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 0.790   | 1.65    | 0.507  | 1.06            | 0.668  | 0.811    | 0.404    |
| Windy Bay Shallow      | 1.700   | 2.61    | 0.492  | 0.505           | 0.956  | 0.832    | 0.536    |
| Windy Bay Deep         | 1.560   | 0.89    | 0.667  | 0.316           | 0.592  | 0.68     | 0.403    |
| Coeur d'Alene River    | 1.920   | 1.48    | 1.26   | 1.06            | 0.563  | 0.752    | 0.775    |
| Mid Lake Coeur d'Alene | 1.690   | 0.782   | 0.935  | 0.441           | 2.3    | 0.89     | 0.902    |
| Carey Bay              | 1.800   | 1.06    | 0.577  | 0.44            | 1.21   | 1.03     | 1.33     |
| Conkling Point         | 1.720   | 1.09    | 1.19   | 0.464           | 0.862  | 0.946    | 1.12     |
| Hidden Lake            | 2.630   | 1.1     | 1.96   | 5.53            | 0.85   | 2.76     | 2.02     |
| Round Lake             | 1.700   | 2.31    | 0.915  | 1.24            | 1.53   | 2        | -        |
| Chatcolet Lake Deep    | 2.430   | 1.98    | 1.74   | 1.73            | 1.33   | 2.88     | 0.963    |
| Chatcolet Lake Shallow | -       | 1.7     | 2.08   | 2.3             | 2.1    | 4.28     | -        |
| Benewah Lake           | 6.84    | 2.41    | 4.39   | 3.37            | 1.72   | -        | -        |
| St. Joe River          | 1.34    | 1.82    | 0.865  | 1.18            | 1.57   | 2.57     | 2.21     |

-No samples taken to Spokane Tribal Lab.

Appendix B.28 Turbidity (NTU) results from the hypolimnion taken at eleven stations on Coeur d'Alene Lake, Idaho, 1998.

| Location               | 4/28/98 | 6/26/98 | 7/8/98 | Date<br>8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
|------------------------|---------|---------|--------|-----------------|--------|----------|----------|
| Rockford Bay           | 1.890   | 3.61    | 2.24   | 1.73            | 1.19   | 0.806    | 0.465    |
| Windy Bay Shallow      | 1.130   | 3.8     | 4.72   | 6.72            | 1.24   | 1.06     | 0.597    |
| Windy Bay Deep         | 12.000  | 1.53    | 1.26   | 0.54            | 1.07   | 0.752    | 0.335    |
| Coeur d'Alene River    | 2.680   | 2.85    | 2.16   | 1.31            | 1.65   | 4.21     | 1.74     |
| Mid Lake Coeur d'Alene | 1.340   | 2.88    | 0.947  | 0.89            | 1.98   | 3.2      | 1.42     |
| Carey Bay              | 2.090   | 1.6     | 2.06   | 0.722           | 5.06   | 2.27     | 1.21     |
| Conkling Point         | 1.700   | 1.93    | 1.38   | 1               | 1.76   | 3.48     | 2.23     |
| Hidden Lake            | 6.200   | 4.45    | 4.56   | 9.7             | 18.5   | -        | 1.69     |
| Chatcolet Lake Deep    | 9.700   | 43.1    | 1.78   | 1.75            | 4.06   | 0.886    | 2.95     |
| Benewah Lake           | 8.23    | 12.4    | 9.35   | 4.02            | 3.19   | -        | -        |
| St. Joe River          | 1.920   | 1.96    | 1.1    | 1.9             | 1.6    | 2.19     | 2.01     |

-No samples taken to Spokane Tribal Lab.

Appendix B.29 Results from trace metal analysis (mg/L). Water samples taken from the epilimnion at three sites on Coeur d'Alene Lake, Idaho, 1997.

| Metals | Location            | Date     |         |
|--------|---------------------|----------|---------|
|        |                     | 10/20/97 | 11/4/97 |
| As     | Site 1 <sup>a</sup> | <0.050   | <0.050  |
|        | Site 2 <sup>b</sup> | <0.050   | <0.050  |
|        | Site 3 <sup>c</sup> | <0.050   | <0.050  |
| Ca     | Site 1 <sup>a</sup> | 5.92     | 5.81    |
|        | Site 2 <sup>b</sup> | 6.70     | 6.94    |
|        | Site 3 <sup>c</sup> | 6.73     | 6.94    |
| Cd     | Site 1 <sup>a</sup> | <0.010   | <0.010  |
|        | Site 2 <sup>b</sup> | <0.010   | <0.010  |
|        | Site 3 <sup>c</sup> | <0.010   | <0.010  |
| Fe     | Site 1 <sup>a</sup> | 0.054    | <0.020  |
|        | Site 2 <sup>b</sup> | 0.043    | 0.108   |
|        | Site 3 <sup>c</sup> | 0.065    | 0.102   |
| K      | Site 1 <sup>a</sup> | 0.888    | <0.850  |
|        | Site 2 <sup>b</sup> | <0.850   | <0.850  |
|        | Site 3 <sup>c</sup> | 1.21     | 1.16    |
| Mg     | Site 1 <sup>a</sup> | 1.690    | 1.67    |
|        | Site 2 <sup>b</sup> | 1.950    | 1.91    |
|        | Site 3 <sup>c</sup> | 1.85     | 1.87    |
| Mn     | Site 1 <sup>a</sup> | <0.005   | <0.005  |
|        | Site 2 <sup>b</sup> | 0.018    | 0.022   |
|        | Site 3 <sup>c</sup> | 0.007    | 0.022   |
| Na     | Site 1 <sup>a</sup> | 1.51     | 1.48    |
|        | Site 2 <sup>b</sup> | 1.61     | 1.59    |
|        | Site 3 <sup>c</sup> | 1.58     | 1.68    |
| Pb     | Site 1 <sup>a</sup> | <0.001   | <0.001  |
|        | Site 2 <sup>b</sup> | 0.004    | 0.005   |
|        | Site 3 <sup>c</sup> | 0.002    | 0.003   |
| Si     | Site 1 <sup>a</sup> | 5.120    | 5.24    |
|        | Site 2 <sup>b</sup> | 5.070    | 4.46    |
|        | Site 3 <sup>c</sup> | 4.52     | 4.69    |
| Zn     | Site 1 <sup>a</sup> | 0.094    | 0.101   |
|        | Site 2 <sup>b</sup> | 0.132    | 0.105   |
|        | Site 3 <sup>c</sup> | 0.097    | 0.090   |

<sup>a</sup> Station Windy Bay Deep.

<sup>b</sup> Station Coeur d'Alene River.

<sup>c</sup> Station Mid Lake Coeur d'Alene.

Appendix B.30 Results from trace metal analysis (mg/L). Water samples taken from the hypolimnion at three sites on Coeur d'Alene Lake, Idaho, 1997.

| Metals | Location            | Date     |         |
|--------|---------------------|----------|---------|
|        |                     | 10/20/97 | 11/4/97 |
| As     | Site 1 <sup>a</sup> | <0.050   | <0.050  |
|        | Site 2 <sup>b</sup> | <0.050   | <0.050  |
|        | Site 3 <sup>c</sup> | <0.050   | <0.050  |
| Ca     | Site 1 <sup>a</sup> | 5.48     | 6.07    |
|        | Site 2 <sup>b</sup> | 7.38     | 6.92    |
|        | Site 3 <sup>c</sup> | 6.95     | 5.59    |
| Cd     | Site 1 <sup>a</sup> | <0.010   | <0.010  |
|        | Site 2 <sup>b</sup> | <0.010   | <0.010  |
|        | Site 3 <sup>c</sup> | <0.010   | <0.010  |
| Fe     | Site 1 <sup>a</sup> | 0.066    | 0.681   |
|        | Site 2 <sup>b</sup> | 0.134    | 0.145   |
|        | Site 3 <sup>c</sup> | 0.108    | 0.140   |
| K      | Site 1 <sup>a</sup> | <0.850   | <0.850  |
|        | Site 2 <sup>b</sup> | <0.850   | <0.850  |
|        | Site 3 <sup>c</sup> | <0.850   | <0.850  |
| Mg     | Site 1 <sup>a</sup> | 1.660    | 1.81    |
|        | Site 2 <sup>b</sup> | 2.320    | 2.34    |
|        | Site 3 <sup>c</sup> | 1.86     | 1.67    |
| Mn     | Site 1 <sup>a</sup> | 0.013    | 0.183   |
|        | Site 2 <sup>b</sup> | 0.050    | 0.058   |
|        | Site 3 <sup>c</sup> | 0.019    | 0.037   |
| Na     | Site 1 <sup>a</sup> | 1.72     | 1.64    |
|        | Site 2 <sup>b</sup> | 1.78     | 1.53    |
|        | Site 3 <sup>c</sup> | 1.65     | 1.67    |
| Pb     | Site 1 <sup>a</sup> | 0.003    | 0.032   |
|        | Site 2 <sup>b</sup> | 0.013    | 0.016   |
|        | Site 3 <sup>c</sup> | <0.001   | 0.004   |
| Si     | Site 1 <sup>a</sup> | 6.740    | 5.97    |
|        | Site 2 <sup>b</sup> | 4.710    | 4.68    |
|        | Site 3 <sup>c</sup> | 4.78     | 6.42    |
| Zn     | Site 1 <sup>a</sup> | 0.133    | 0.159   |
|        | Site 2 <sup>b</sup> | 0.185    | 0.226   |
|        | Site 3 <sup>c</sup> | 0.078    | 0.124   |

<sup>a</sup> Station Windy Bay Deep.

<sup>b</sup> Station Coeur d'Alene River.

<sup>c</sup> Station Mid Lake Coeur d'Alene.

**Appendix B.31 Results from trace metal analysis. Water samples taken from the epilimnion at three sites on Co**

| Metals | Location            | Date    |         |        |         |        |          |          |
|--------|---------------------|---------|---------|--------|---------|--------|----------|----------|
|        |                     | 4/28/98 | 6/26/98 | 7/8/98 | 8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
| As     | Site 1 <sup>a</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
|        | Site 2 <sup>b</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
|        | Site 3 <sup>c</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
| Ca     | Site 1 <sup>a</sup> | 6.007   | 5.955   | 5.768  | 6.306   | 6.176  | 6.609    | 6.231    |
|        | Site 2 <sup>b</sup> | 4.923   | 6.16    | 6.48   | 7.58    | 6.317  | 7.722    | 7        |
|        | Site 3 <sup>c</sup> | 5.623   | 5.638   | 5.842  | 6.844   | 6.532  | 7.473    | 6.591    |
| Cd     | Site 1 <sup>a</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | <0.005   |
|        | Site 2 <sup>b</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | 0.0059   |
|        | Site 3 <sup>c</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | <0.005   |
| Fe     | Site 1 <sup>a</sup> | 0.130   | 0.013   | 0.1291 | 0.0177  | 0.0376 | 0.0252   | 0.0214   |
|        | Site 2 <sup>b</sup> | 0.175   | <0.010  | 0.1048 | 0.0984  | 0.0379 | 0.0912   | 0.074    |
|        | Site 3 <sup>c</sup> | 0.162   | 0.0104  | 0.0664 | 0.0399  | 0.0707 | 0.0558   | 0.0783   |
| K      | Site 1 <sup>a</sup> | <0.900  | <0.900  | <0.900 | <0.900  | <0.900 | <0.900   | <0.900   |
|        | Site 2 <sup>b</sup> | <0.900  | <0.900  | <0.900 | <0.900  | <0.900 | <0.900   | 1.023    |
|        | Site 3 <sup>c</sup> | <0.900  | <0.900  | <0.900 | <0.900  | <0.900 | 1.535    | <0.900   |
| Mg     | Site 1 <sup>a</sup> | 1.869   | 1.54    | 1.643  | 1.768   | 1.757  | 1.956    | 1.765    |
|        | Site 2 <sup>b</sup> | 1.688   | 1.448   | 1.803  | 2.178   | 1.776  | 2.306    | 1.92     |
|        | Site 3 <sup>c</sup> | 1.431   | 1.459   | 1.532  | 1.804   | 1.768  | 2.092    | 1.734    |
| Mn     | Site 1 <sup>a</sup> | 0.017   | <0.003  | 0.0371 | 0.0044  | 0.0071 | 0.0072   | 0.0083   |
|        | Site 2 <sup>b</sup> | 0.053   | <0.003  | 0.0274 | 0.0355  | 0.0089 | 0.0273   | 0.0298   |
|        | Site 3 <sup>c</sup> | 0.017   | <0.003  | 0.0125 | 0.0073  | 0.0134 | 0.0054   | 0.035    |
| Na     | Site 1 <sup>a</sup> | 1.848   | 1.355   | 1.469  | 1.58    | 1.559  | 1.597    | 1.586    |
|        | Site 2 <sup>b</sup> | 1.256   | 1.315   | 1.495  | 1.717   | 1.588  | 1.783    | 1.819    |
|        | Site 3 <sup>c</sup> | 1.458   | 1.348   | 1.413  | 1.641   | 1.605  | 1.886    | 1.505    |
| Pb     | Site 1 <sup>a</sup> | 0.005   | 0.002   | 0.007  | <0.001  | <0.001 | 0.002    | <0.001   |
|        | Site 2 <sup>b</sup> | 0.021   | 0.002   | 0.007  | 0.011   | 0.002  | 0.007    | 0.002    |
|        | Site 3 <sup>c</sup> | 0.002   | <0.001  | 0.002  | <0.001  | 0.004  | 0.003    | 0.002    |
| Si     | Site 1 <sup>a</sup> | 5.323   | 4.246   | 4.17   | 3.956   | 3.816  | 3.659    | 3.43     |
|        | Site 2 <sup>b</sup> | 4.646   | 4.458   | 4.154  | 3.927   | 3.492  | 3.392    | 3.27     |
|        | Site 3 <sup>c</sup> | 5.423   | 4.193   | 4.066  | 4.1     | 3.681  | 3.311    | 3.029    |
| Zn     | Site 1 <sup>a</sup> | 0.114   | 0.0706  | 0.0847 | 0.0668  | 0.0537 | 0.0782   | 0.0702   |
|        | Site 2 <sup>b</sup> | 0.140   | 0.0393  | 0.0728 | 0.0914  | 0.0498 | 0.124    | 0.0693   |
|        | Site 3 <sup>c</sup> | 0.037   | 0.055   | 0.0532 | 0.0583  | 0.0546 | 0.0664   | 0.0611   |

<sup>a</sup> Station Windy Bay Deep.

<sup>b</sup> Station Coeur d'Alene River.

<sup>c</sup> Station Mid Lake Coeur d'Alene.

**Appendix B.32 Results from trace metal analysis. Water samples taken from the hypolimnion**

at three sites on Coeur d'Alene Lake, Idaho, 1998.

| Metals | Location            | Date    |         |        |         |        |          |          |
|--------|---------------------|---------|---------|--------|---------|--------|----------|----------|
|        |                     | 4/28/98 | 6/26/98 | 7/8/98 | 8/10/98 | 9/2/98 | 10/20/98 | 11/12/98 |
| As     | Site 1 <sup>a</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
|        | Site 2 <sup>b</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
|        | Site 3 <sup>c</sup> | <0.050  | <0.050  | <0.050 | <0.050  | <0.050 | <0.050   | <0.050   |
| Ca     | Site 1 <sup>a</sup> | 5.909   | 5.744   | 5.66   | 5.909   | 5.576  | 6.105    | 6.21     |
|        | Site 2 <sup>b</sup> | 5.062   | 5.984   | 5.594  | 5.727   | 6.888  | 8.355    | 7.218    |
|        | Site 3 <sup>c</sup> | 6.085   | 5.691   | 5.313  | 6.339   | 5.539  | 6.068    | 6.485    |
| Cd     | Site 1 <sup>a</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | <0.005   |
|        | Site 2 <sup>b</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | <0.005   |
|        | Site 3 <sup>c</sup> | <0.005  | <0.005  | <0.005 | <0.005  | <0.005 | <0.005   | 0.0078   |
| Fe     | Site 1 <sup>a</sup> | 0.126   | 0.0152  | 0.0737 | 0.0839  | 0.1562 | 0.0273   | 0.0205   |
|        | Site 2 <sup>b</sup> | 0.369   | 0.0152  | 0.1564 | 0.0882  | 0.1475 | 0.453    | 0.0645   |
|        | Site 3 <sup>c</sup> | 0.115   | 0.0133  | 0.2067 | 0.0502  | 0.0536 | 0.2738   | 0.0825   |
| K      | Site 1 <sup>a</sup> | <0.900  | 0.908   | <0.900 | <0.900  | <0.900 | 1.025    | <0.900   |
|        | Site 2 <sup>b</sup> | <0.900  | <0.900  | <0.900 | <0.900  | <0.900 | 1.407    | <0.900   |
|        | Site 3 <sup>c</sup> | <0.900  | <0.900  | <0.900 | <0.900  | <0.900 | 1.03     | <0.900   |
| Mg     | Site 1 <sup>a</sup> | 1.866   | 1.58    | 1.713  | 1.715   | 1.667  | 1.845    | 1.758    |
|        | Site 2 <sup>b</sup> | 1.744   | 1.441   | 1.607  | 1.62    | 1.906  | 2.651    | 1.978    |
|        | Site 3 <sup>c</sup> | 1.895   | 1.527   | 1.51   | 1.72    | 1.551  | 1.816    | 1.865    |
| Mn     | Site 1 <sup>a</sup> | 0.033   | <0.003  | 0.0081 | 0.0274  | 0.0526 | 0.0126   | 0.0136   |
|        | Site 2 <sup>b</sup> | 0.080   | <0.003  | 0.0395 | 0.0371  | 0.0294 | 0.0992   | 0.0309   |
|        | Site 3 <sup>c</sup> | 0.012   | <0.003  | 0.0419 | 0.0564  | 0.0406 | 0.2789   | 0.0768   |
| Na     | Site 1 <sup>a</sup> | 1.728   | 1.536   | 1.671  | 1.66    | 1.645  | 1.623    | 1.634    |
|        | Site 2 <sup>b</sup> | 1.290   | 1.389   | 1.421  | 1.502   | 1.638  | 2.093    | 1.666    |
|        | Site 3 <sup>c</sup> | 1.730   | 1.537   | 1.492  | 1.626   | 1.521  | 1.683    | 1.724    |
| Pb     | Site 1 <sup>a</sup> | 0.005   | 0.002   | 0.001  | 0.005   | 0.003  | 0.002    | <0.001   |
|        | Site 2 <sup>b</sup> | 0.058   | 0.007   | 0.015  | 0.013   | 0.013  | 0.063    | 0.002    |
|        | Site 3 <sup>c</sup> | 0.005   | <0.001  | 0.001  | 0.002   | 0.004  | 0.006    | 0.003    |
| Si     | Site 1 <sup>a</sup> | 5.225   | 4.687   | 4.759  | 4.476   | 4.605  | 4.522    | 4.044    |
|        | Site 2 <sup>b</sup> | 4.663   | 4.616   | 4.092  | 3.773   | 3.765  | 3.467    | 3.347    |
|        | Site 3 <sup>c</sup> | 5.223   | 4.703   | 4.659  | 4.456   | 4.367  | 5.317    | 4.217    |
| Zn     | Site 1 <sup>a</sup> | 0.131   | 0.1011  | 0.1077 | 0.1213  | 0.1261 | 0.1077   | 0.0965   |
|        | Site 2 <sup>b</sup> | 0.159   | 0.0585  | 0.0873 | 0.0881  | 0.0681 | 0.2367   | 0.0734   |
|        | Site 3 <sup>c</sup> | 0.122   | 0.0748  | 0.0788 | 0.0949  | 0.0962 | 0.1147   | 0.1096   |

<sup>a</sup> Station Windy Bay Deep.

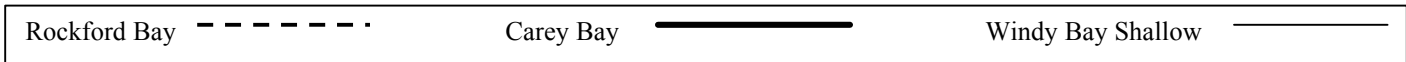
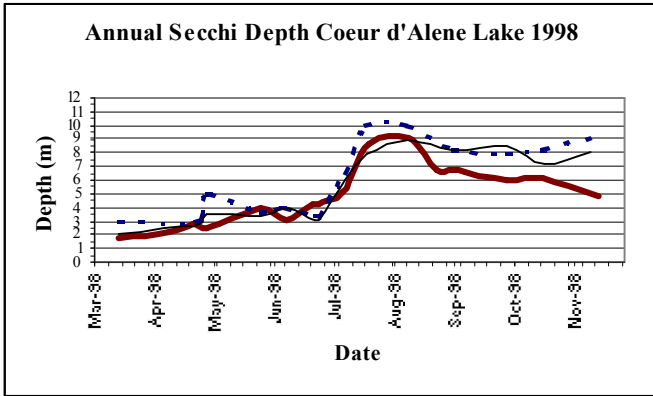
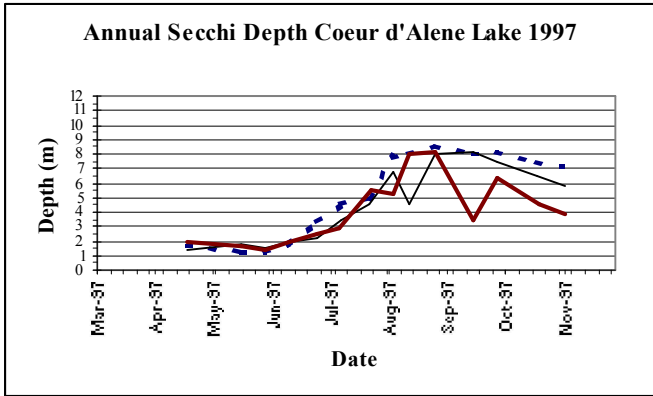
<sup>b</sup> Station Coeur d'Alene River.

<sup>c</sup> Station Mid Lake Coeur d'Alene.

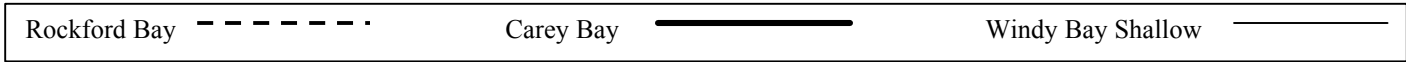
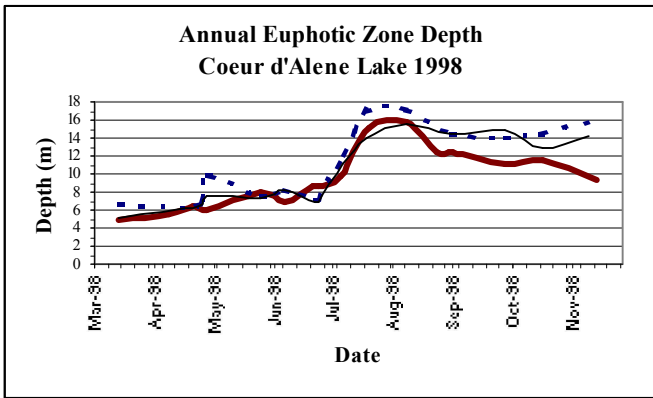
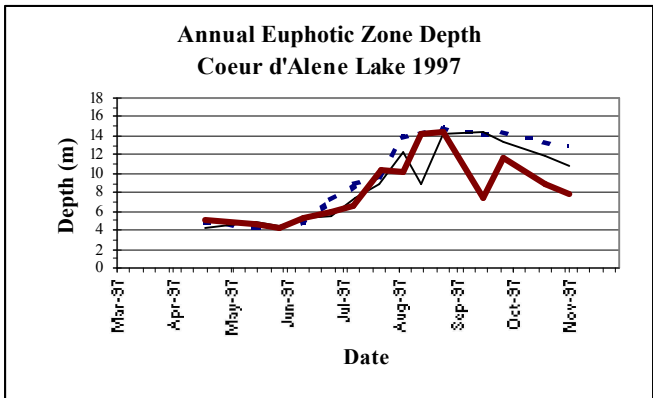


## **Appendix C**

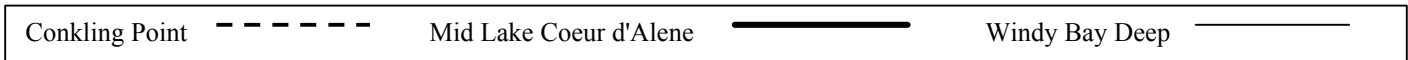
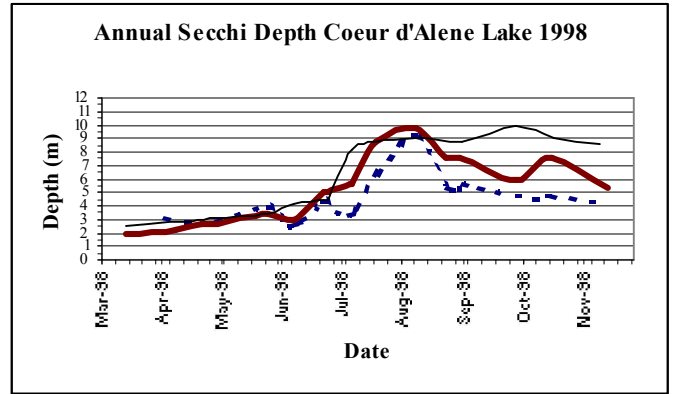
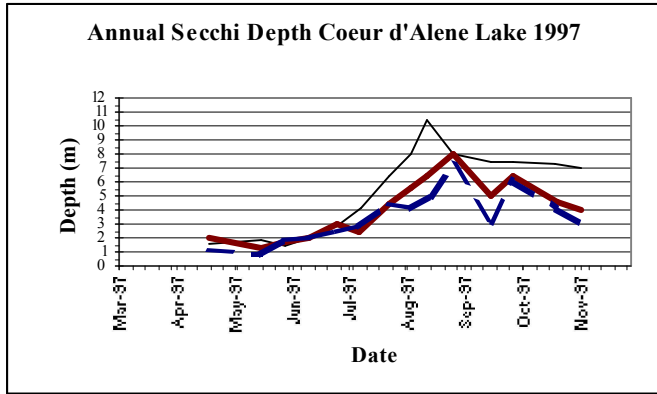
Secchi readings and empirically derived estimates of euphotic zone depth for thirteen stations in Coeur d'Alene Lake, 1997 and 1998.



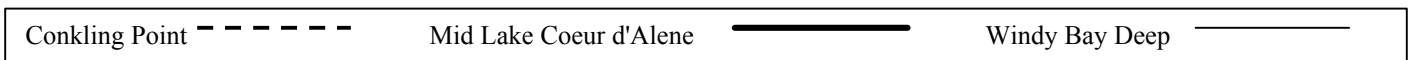
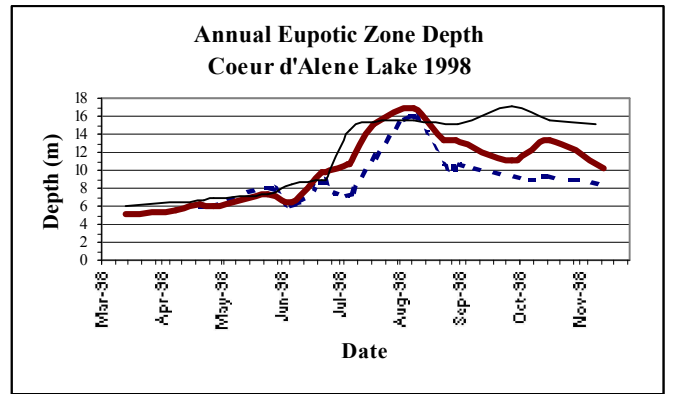
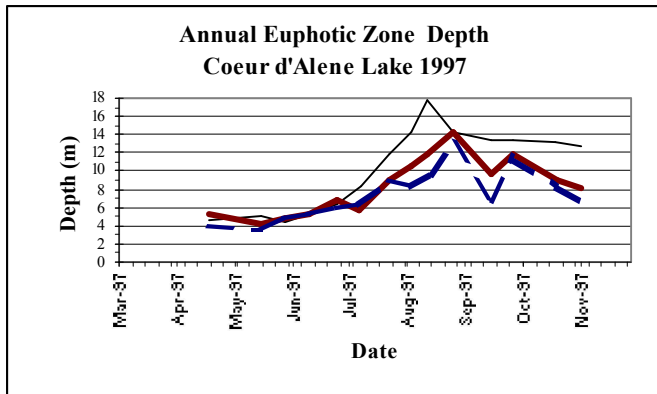
Appendix C.1 Annual secchi readings (SD) for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.



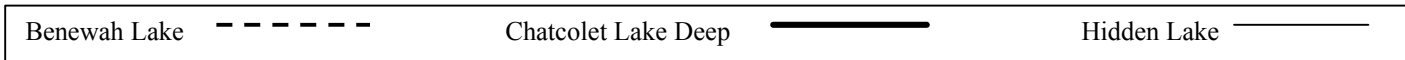
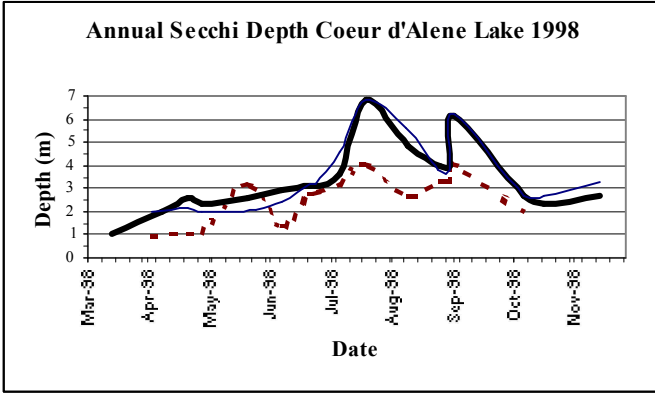
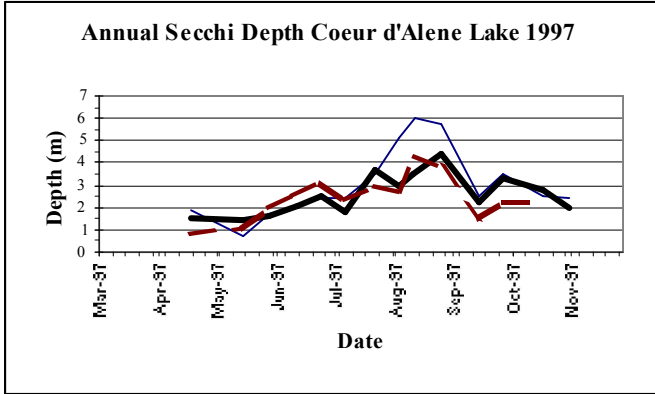
Appendix C.2 Empirically derived estimate of euphotic zone depth (EZD) during 1997 and 1998. Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ ).



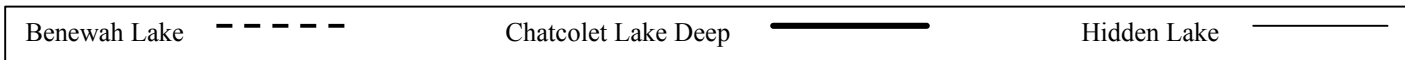
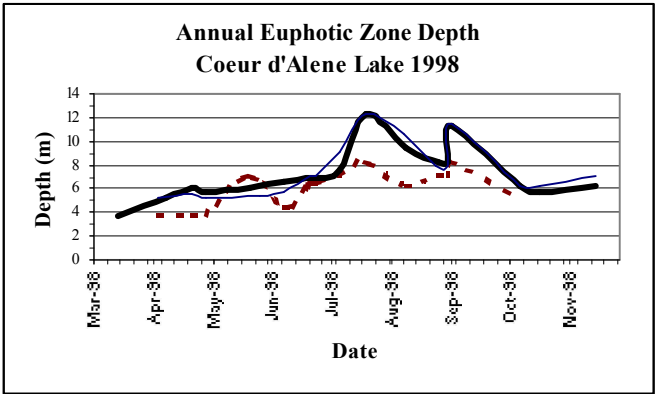
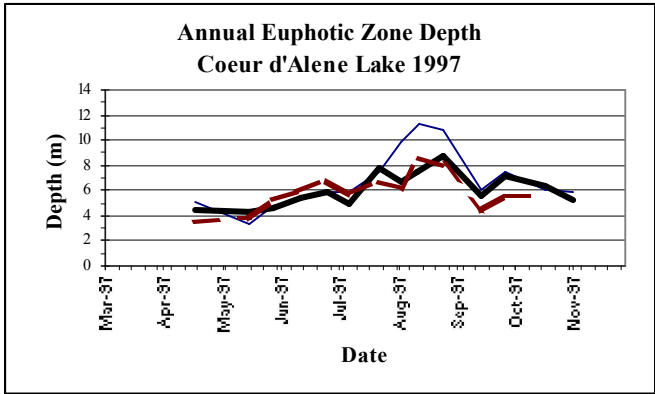
Appendix C.3 Annual secchi readings (SD) for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.



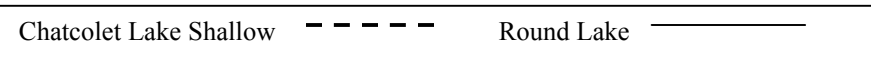
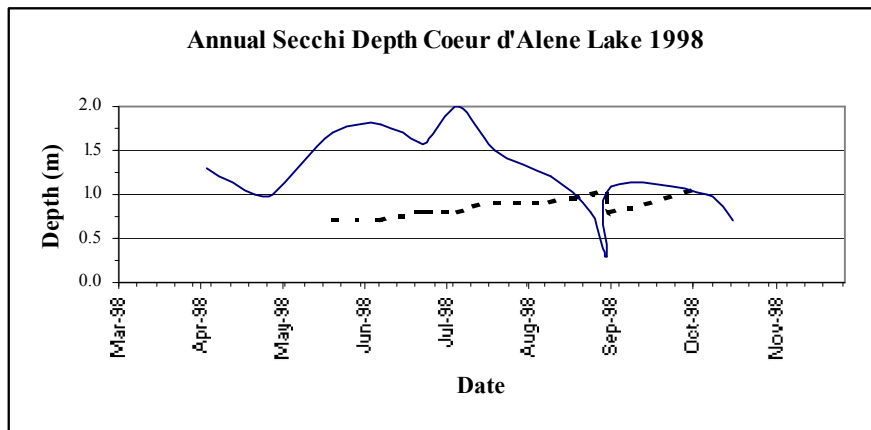
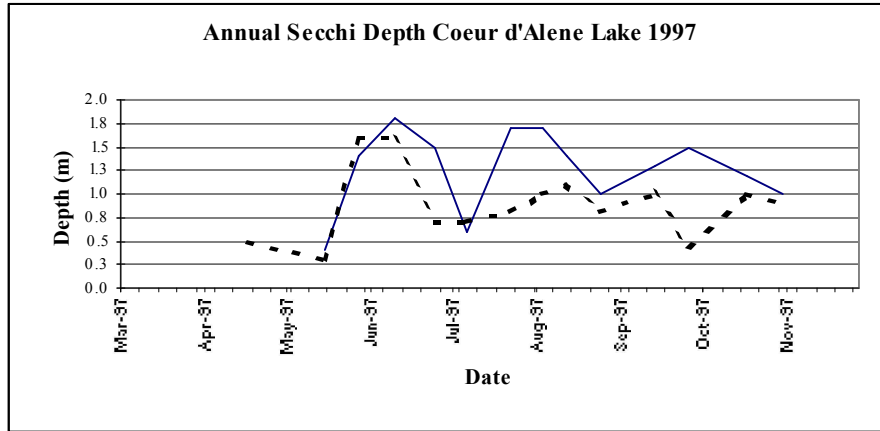
Appendix C.4 Empirically derived estimate of euphotic zone depth (EZD) during 1997 and 1998. Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ ).



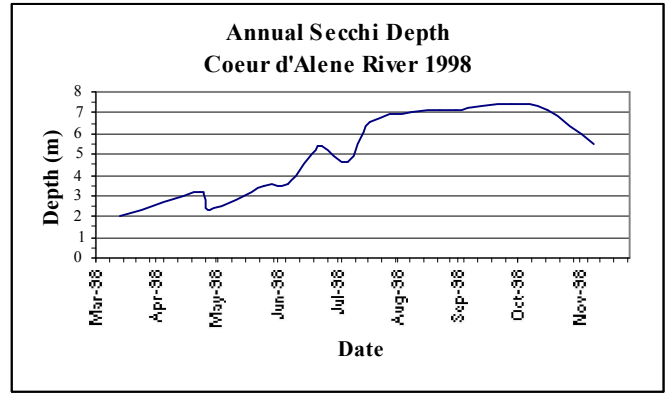
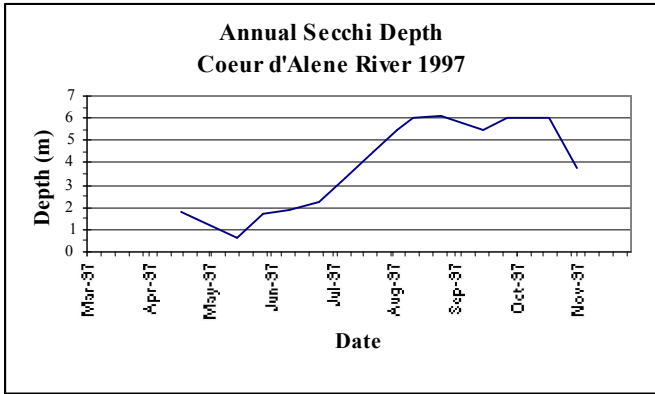
Appendix C.5 Annual secchi readings (SD) for three geomorphology similar sampling locations on Coeur d'Alene Lake during 1997 and 1998.



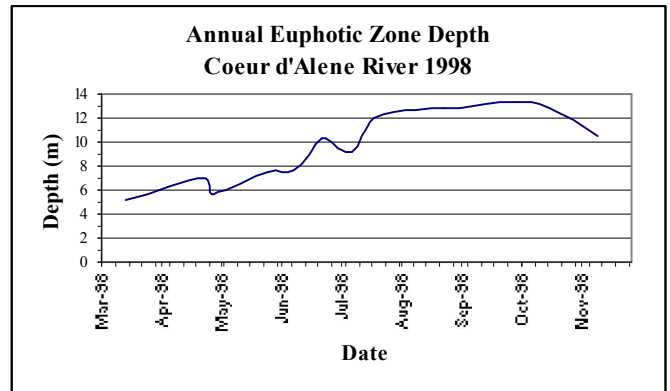
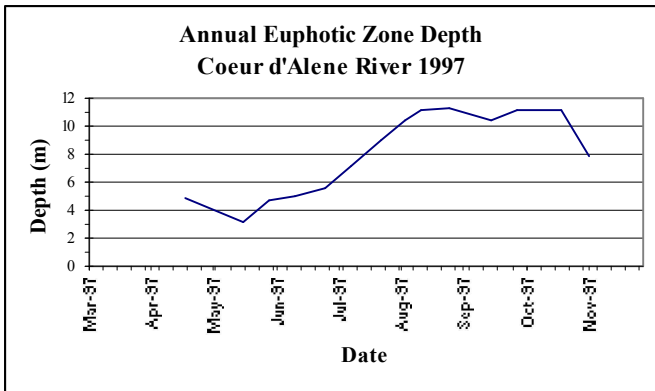
Appendix C.6 Empirically derived estimate of euphotic zone depth (EZD) during 1997 and 1998. Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ ).



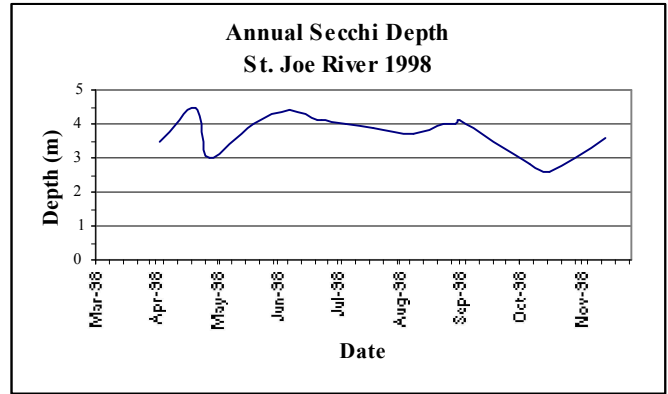
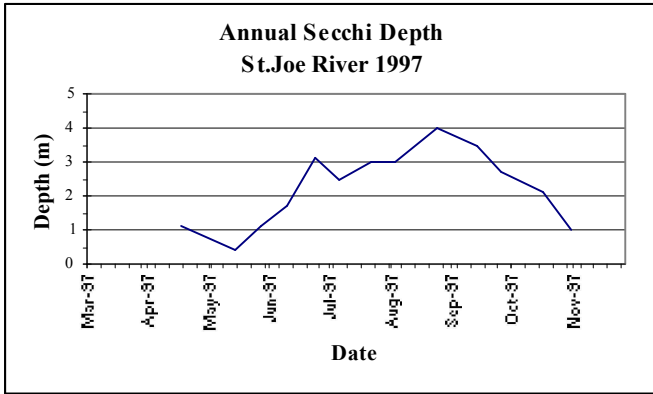
Appendix C.7 Annual secchi readings (SD) for two geomorphically similar sampling locations on Coeur d'Alene Lake during 1997 and 1998. Due to the shallowness of these lakes, the euphotic zone depth encompasses the entire water column.



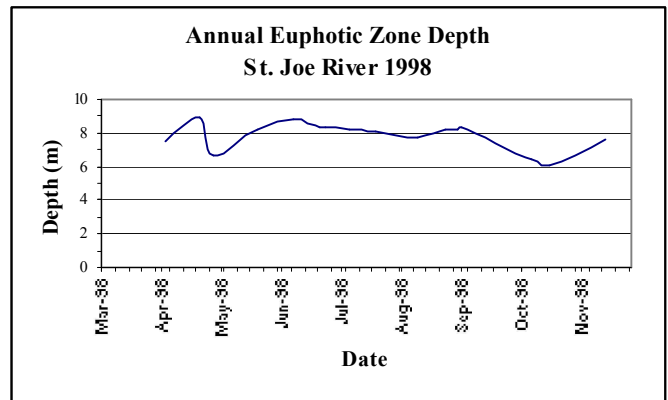
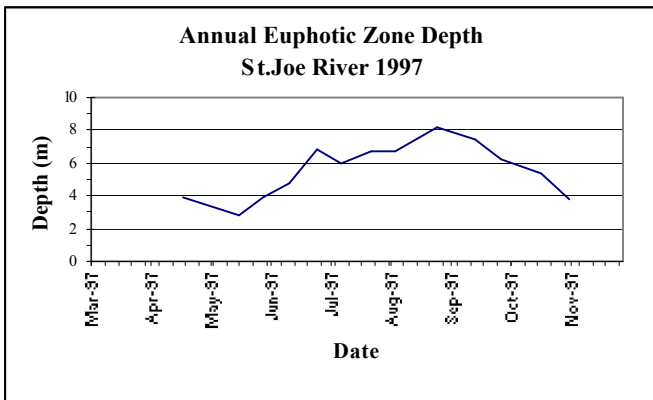
Appendix C.8 Annual secchi readings (SD) for the mouth of the Coeur d'Alene River during 1997 and 1998.



Appendix C.9 Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD=2.2303 + 1.4914SD$  ( $r^2 = .78$ ).



Appendix C.10 Annual secchi readings (SD) for the mouth of the St. Joe River during 1997 and 1998.

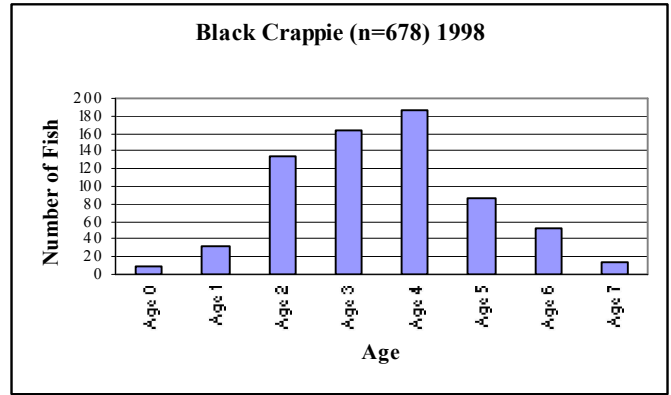
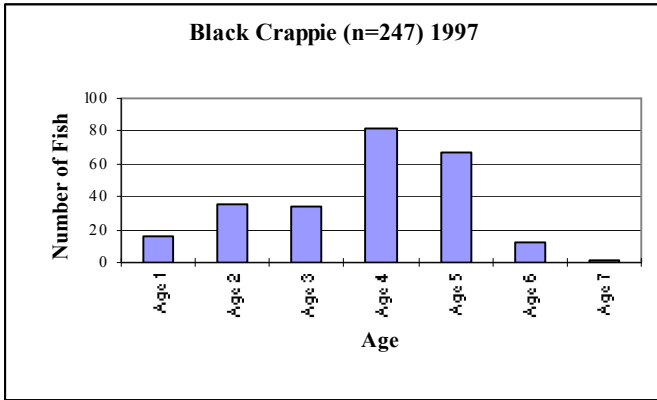


Appendix C.11 Empirically derived estimate of euphotic zone depth (EZD). Values are calculated using the following regression equation published by Alaska Department of Fish and Game (1987):  $EZD = 2.2303 + 1.4914SD$  ( $r^2 = .78$ ).

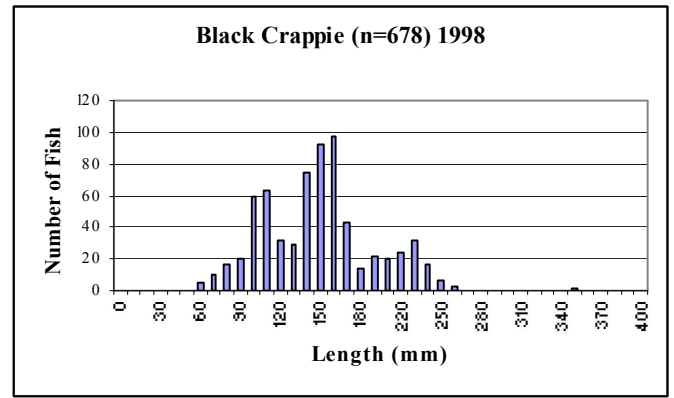
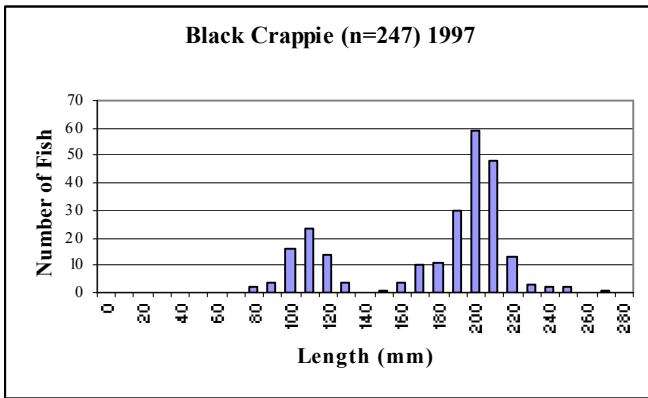
## **Appendix D**

Age distribution, length and weight frequencies and length and weight regression graphs for aged fish species sampled in Coeur d'Alene Lake, 1997 and 1998.

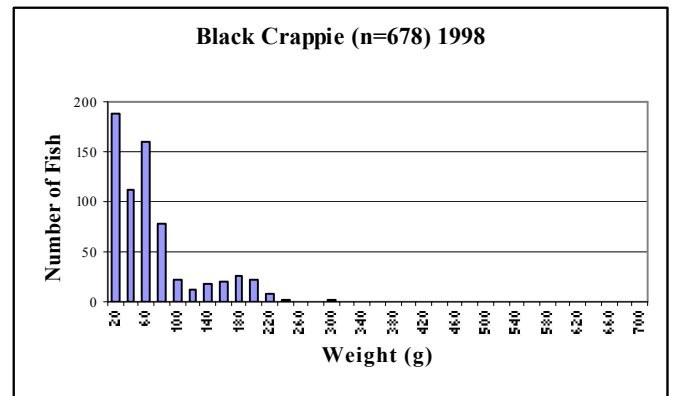
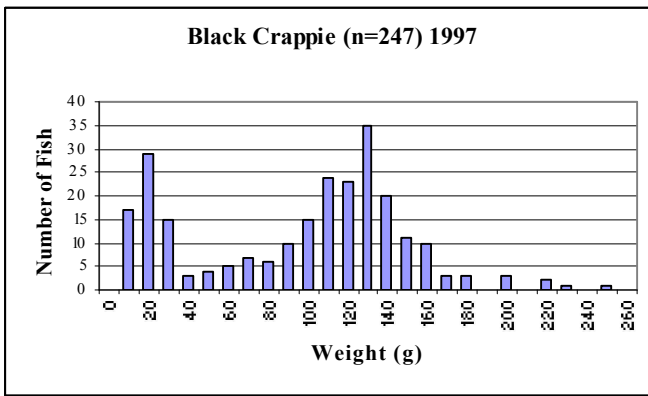




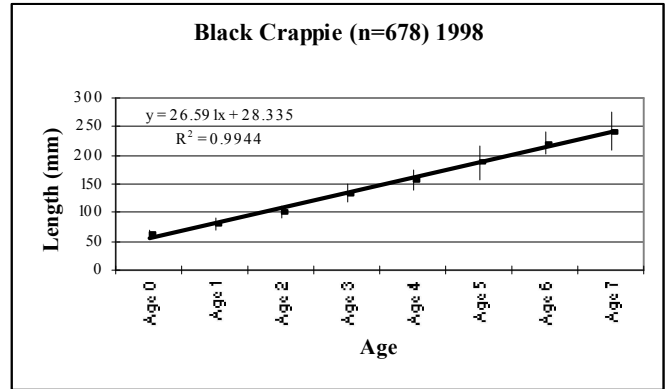
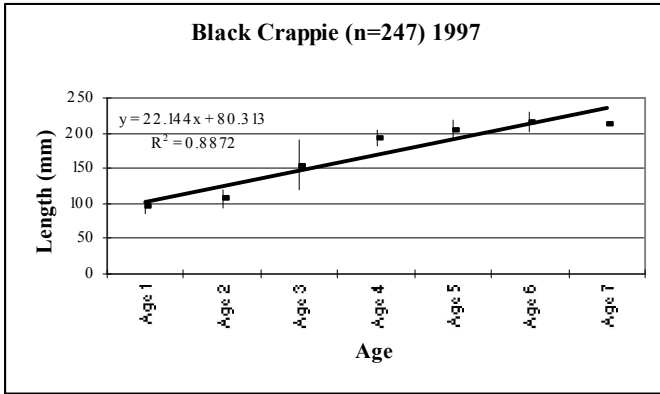
Appendix D.1 Age distribution of the number of Black Crappie versus age, in 1997 and 1998.



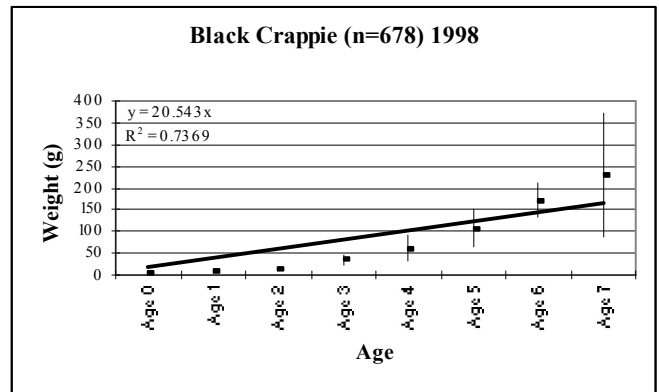
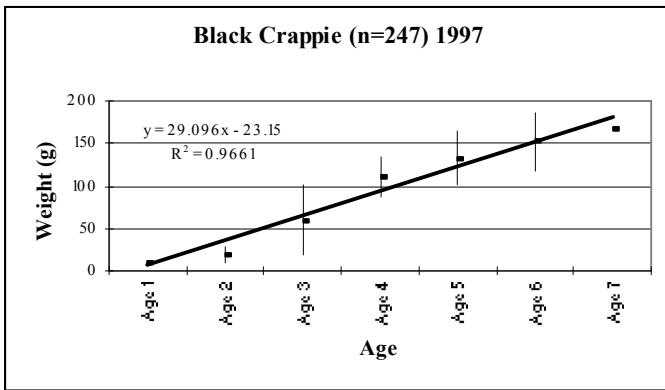
Appendix D.2 Frequency distribution of the number of Black Crappie sampled versus body length in 1997 and 1998.



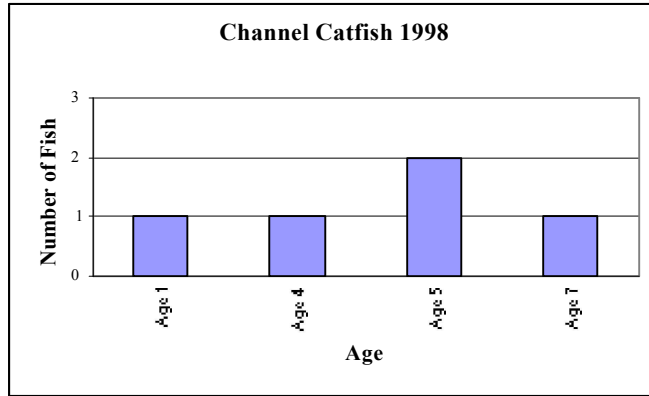
Appendix D.3 Frequency distribution of the number of Black Crappie sampled versus weight in 1997 and 1998.



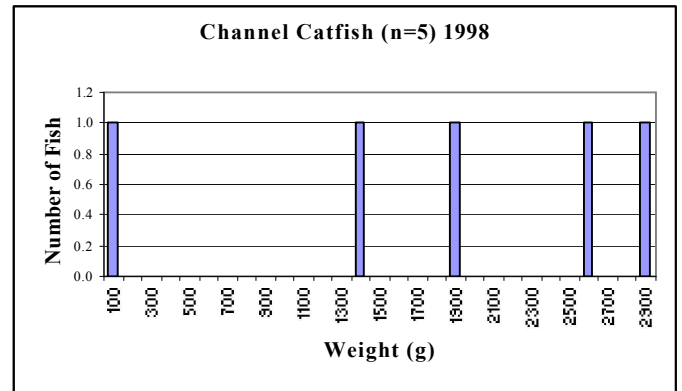
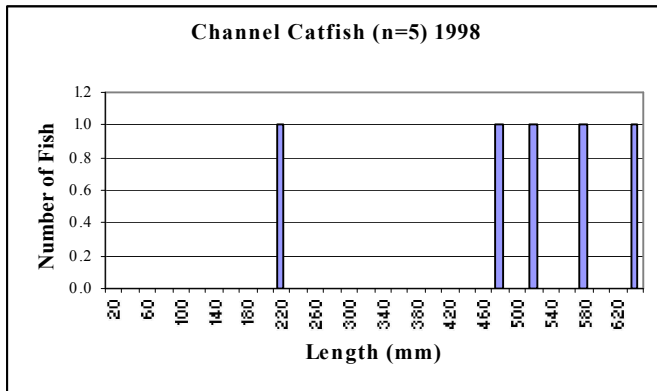
Appendix D.4 Regression equations of body length versus age for Black Crappie, in 1997 and 1998.



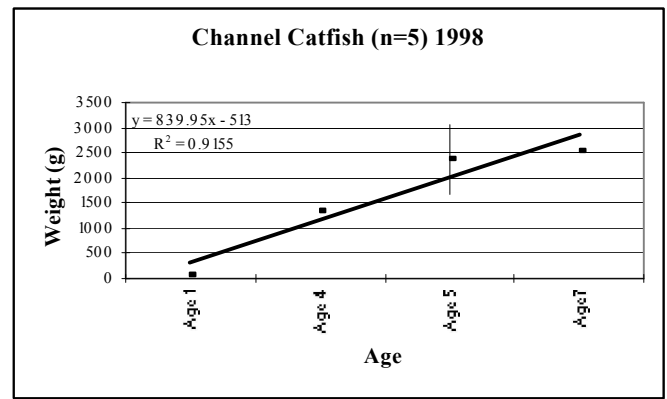
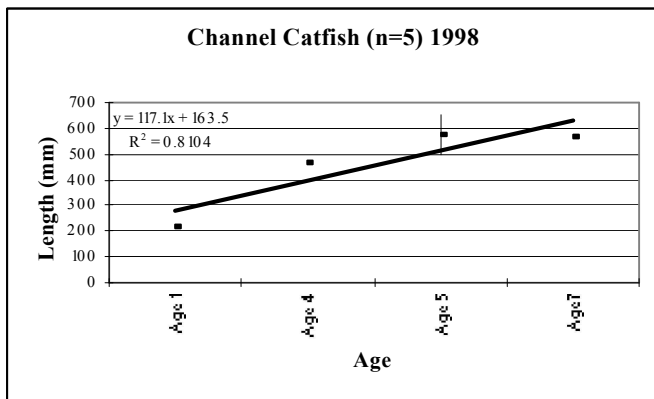
Appendix D.5 Regression equations of weight versus age for Black Crappie, in 1997 and 1998.



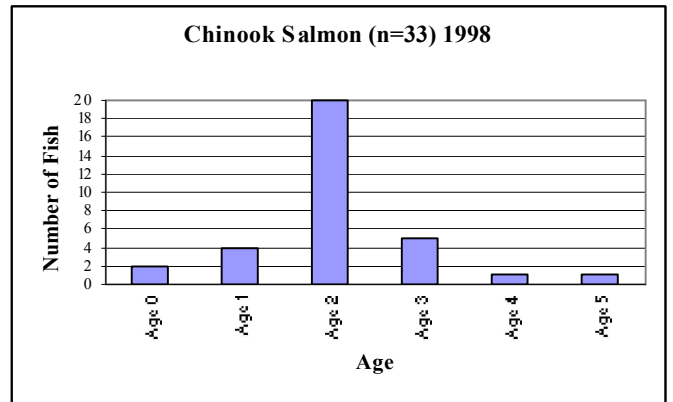
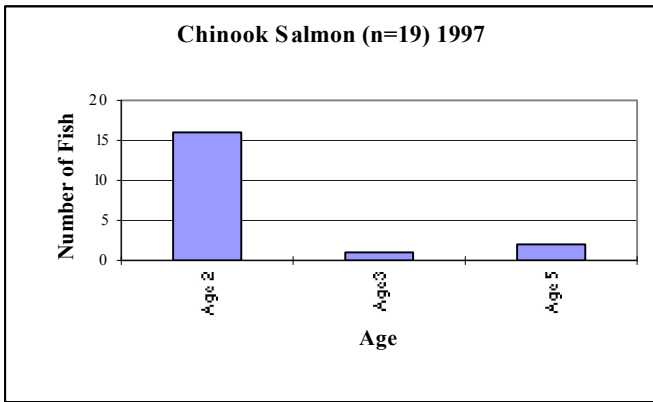
Appendix D.6 Age distribution of the number of Channel Catfish sampled versus age in 1998.



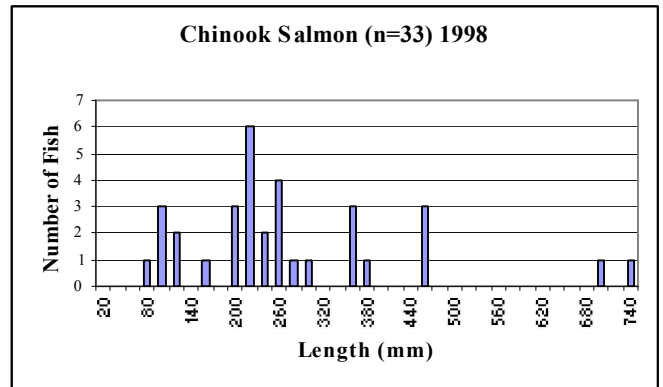
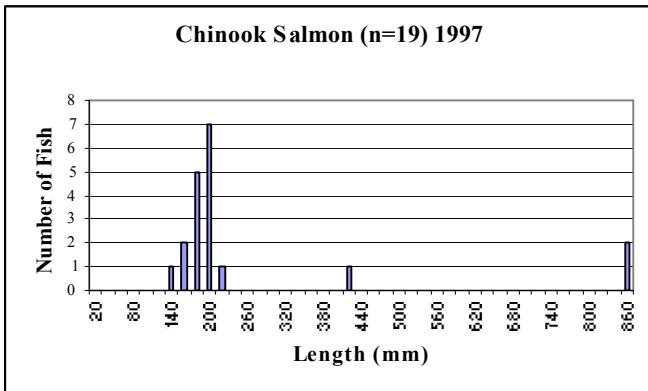
Appendix D.7 Frequency distribution of the number of Channel Catfish sampled versus body length and weight in 1998.



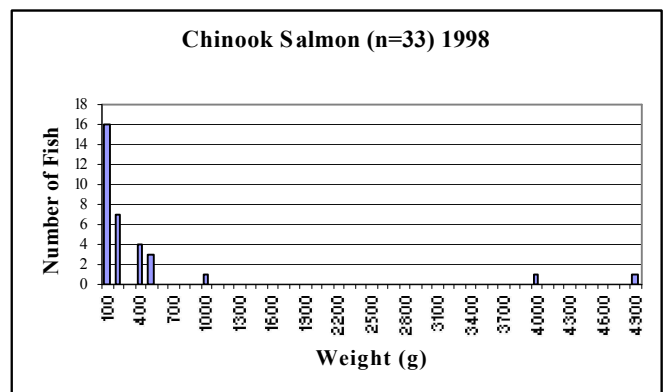
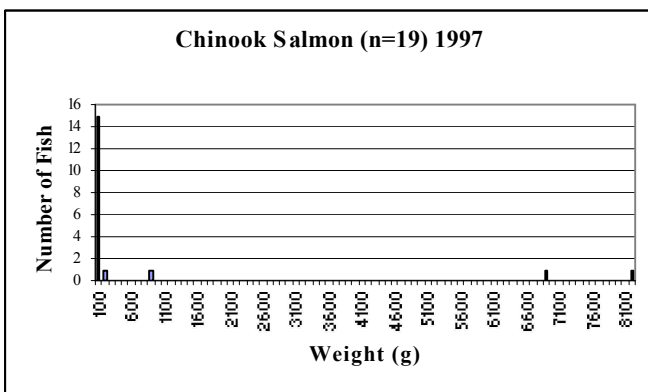
Appendix D.8 Regression equations of body length and weight versus age for Channel Catfish, in 1998.



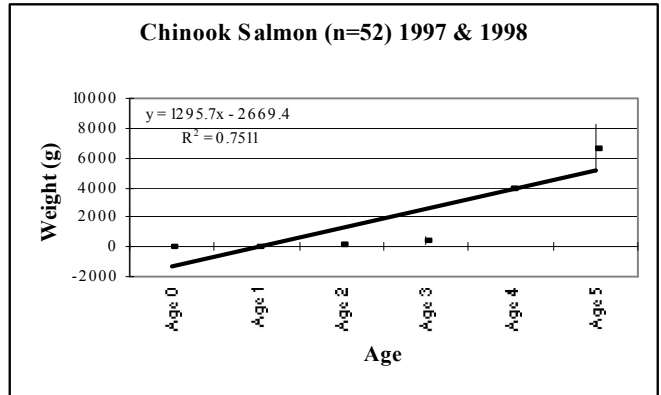
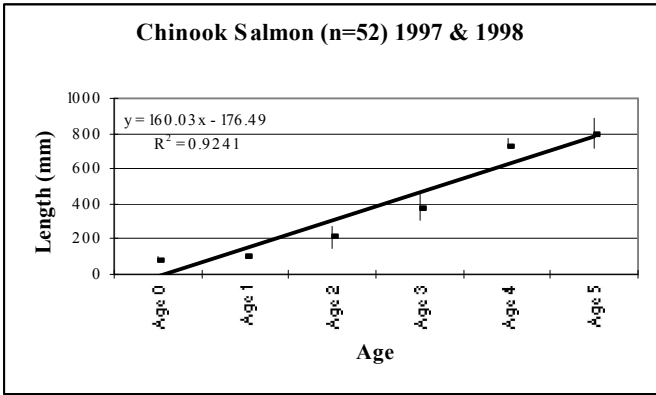
Appendix D.9 Age distribution of the number of Chinook Salmon sampled versus age in 1997 and 1998.



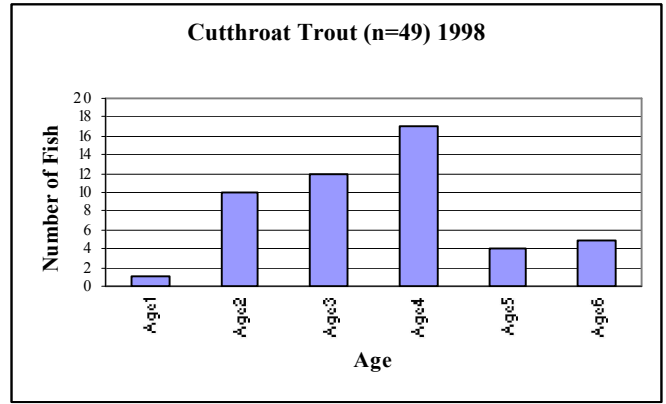
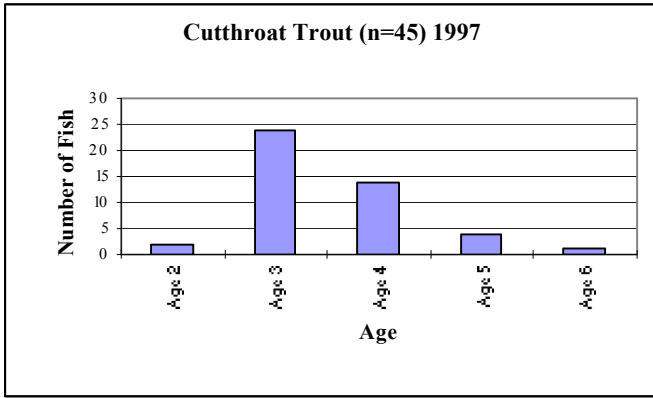
Appendix D.10 Frequency distribution of the number of Chinook Salmon sampled versus body length in 1997 and 1998.



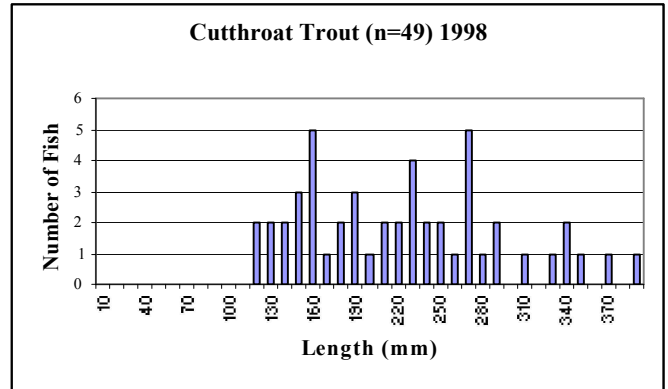
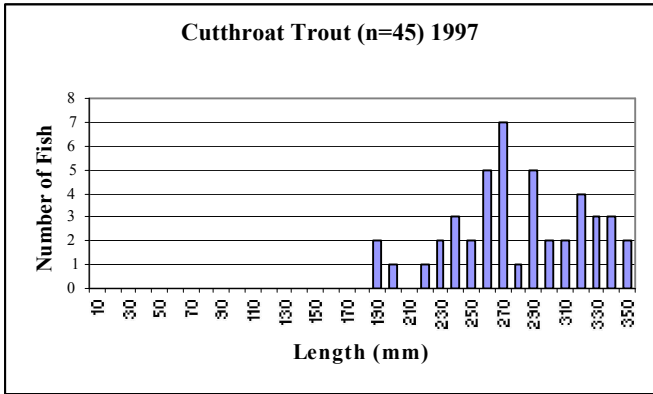
Appendix D.11 Frequency distribution of the number of Chinook Salmon sampled versus weight in 1997 and 1998.



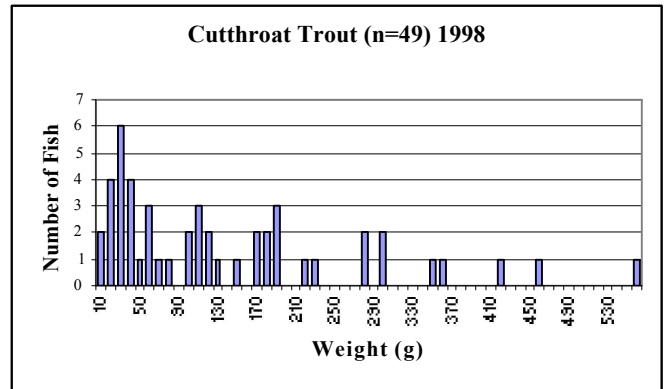
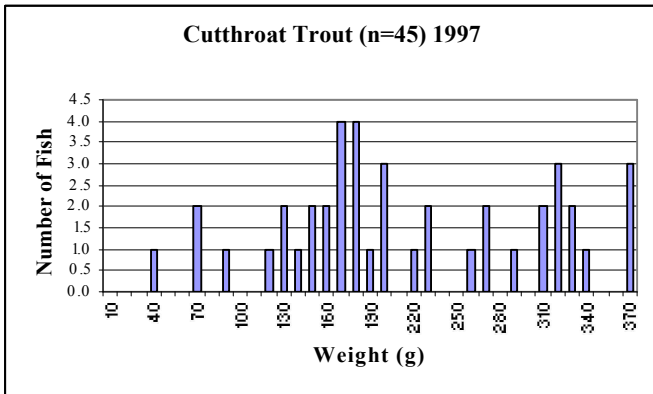
Appendix D.12 Regression equations of body length and weight versus age for Chinook Salmon, combined 1997 and 1998 data.



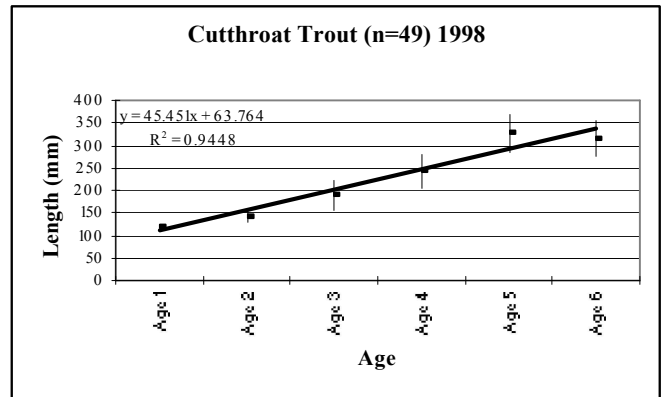
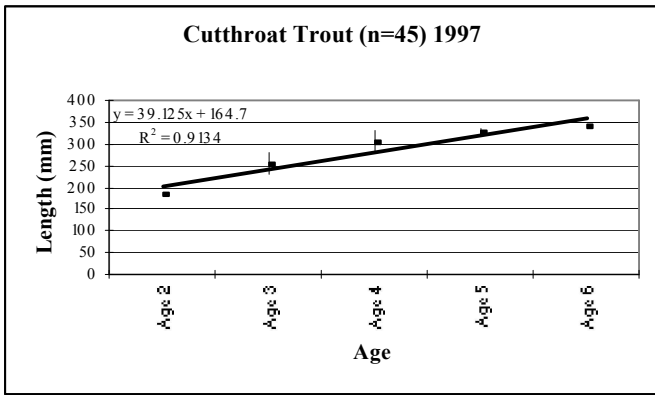
Appendix D.13 Age distribution of the number of Cutthroat Trout sampled versus age in 1997 and 1998.



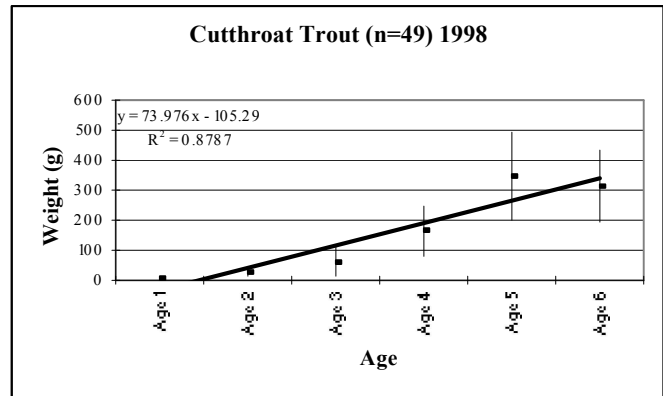
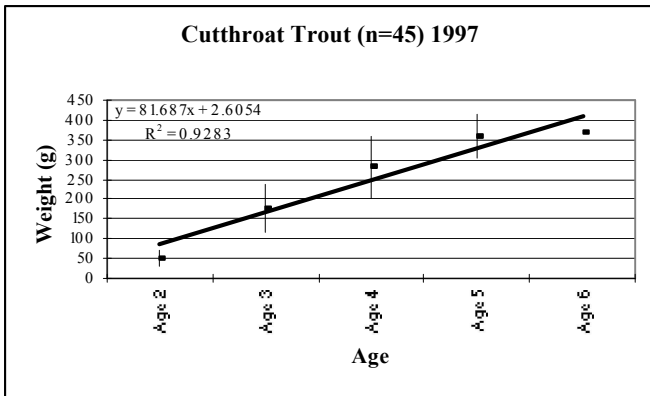
Appendix D.14 Frequency distribution of the number of Cutthroat Trout sampled versus body length in 1997 and 1998.



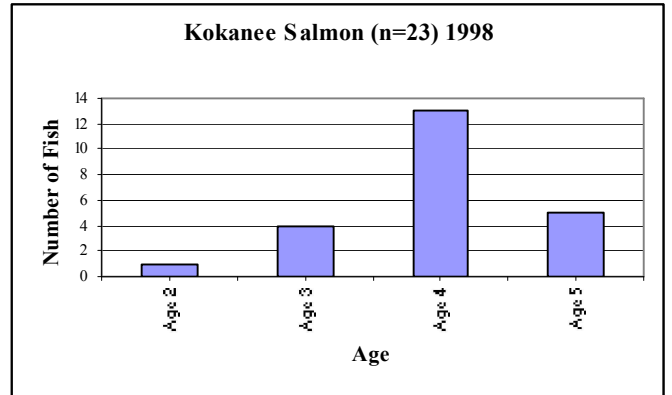
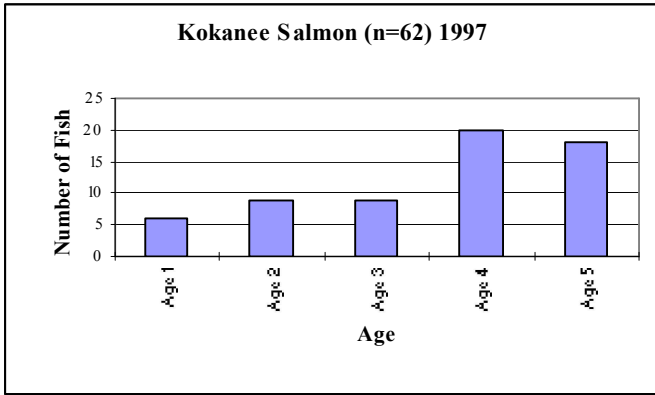
Appendix D.15 Frequency distribution of the number of Cutthroat Trout sampled versus weight in 1997 and 1998.



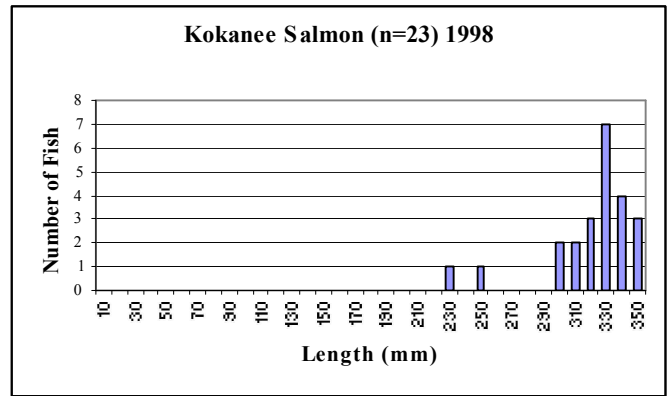
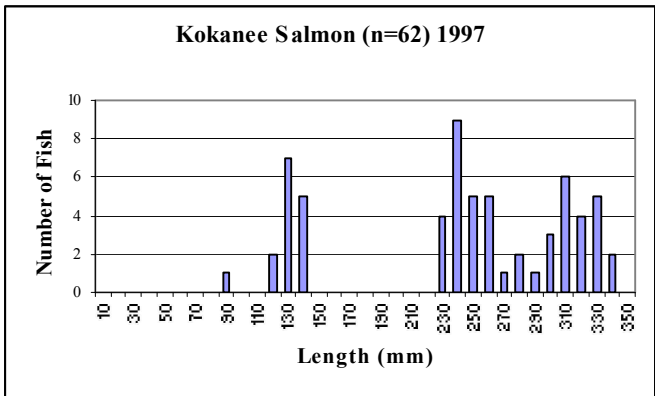
Appendix D.16 Regression equations of body length versus age for Cutthroat Trout, in 1997 and 1998.



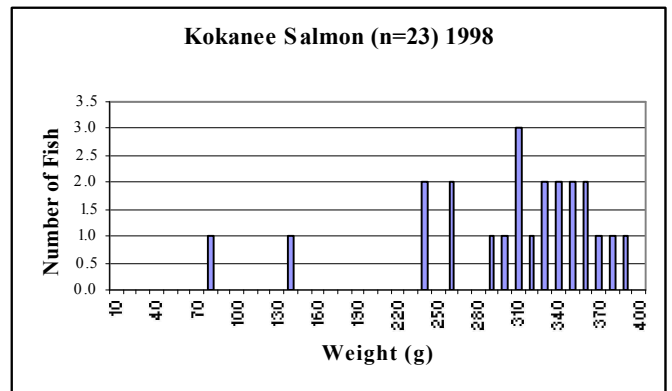
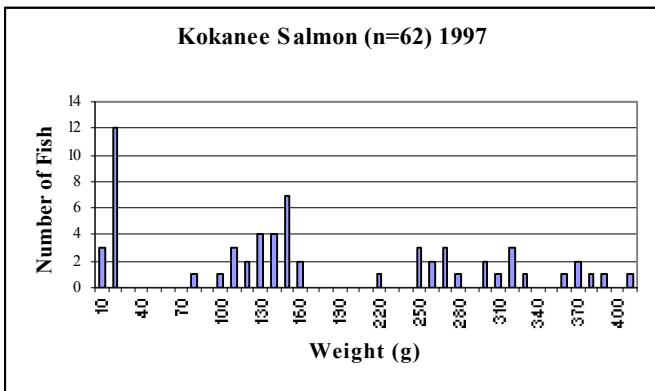
Appendix D.17 Regression equations of and weight versus age for Cutthroat Trout, in 1997 and 1998.



Appendix D.18 Age distribution of the number of Kokanee Salmon sampled versus age in 1997 and 1998.

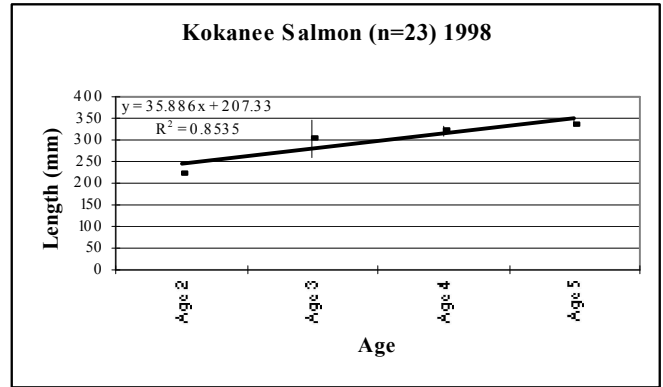
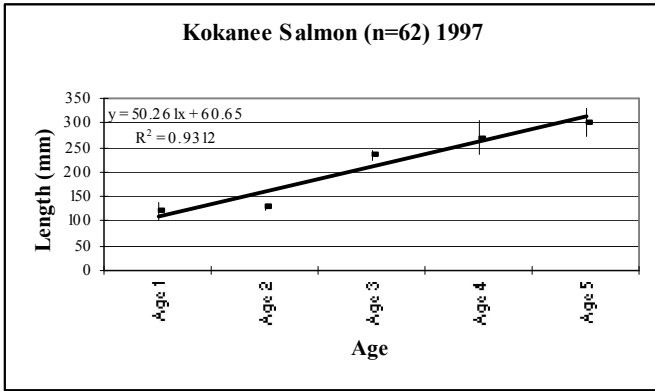


Appendix D.19 Frequency distribution of the number of Kokanee Salmon sampled versus body length in 1997 and 1998.

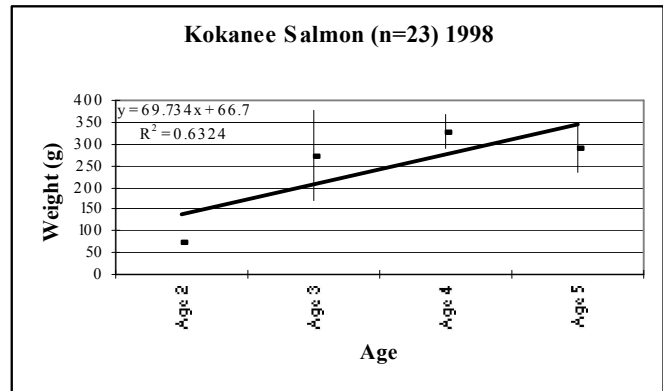
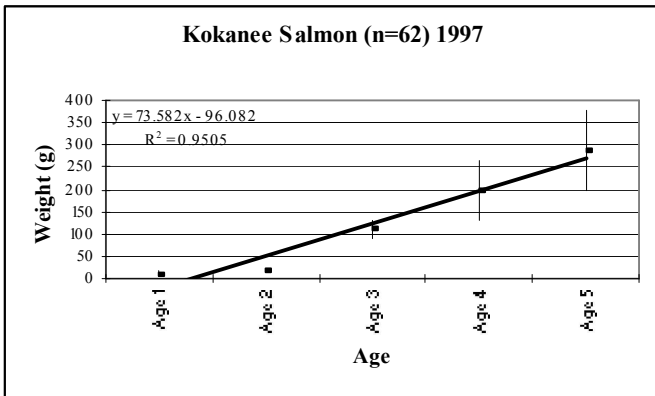


Appendix D.20 Frequency distribution of the number of Kokanee Salmon sampled versus weight in 1997 and 1998.

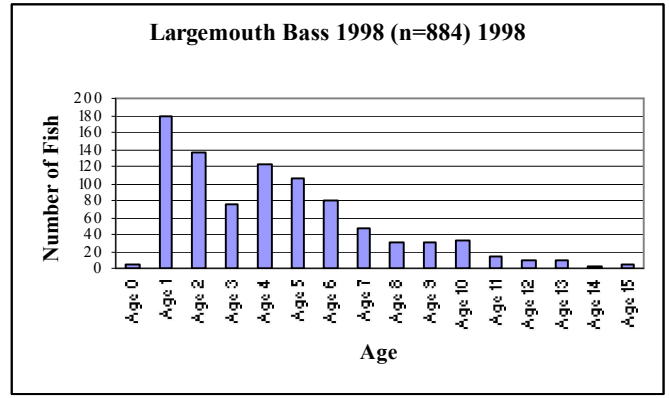
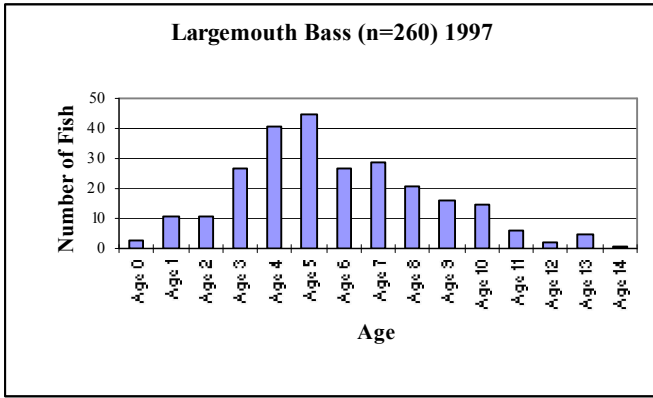




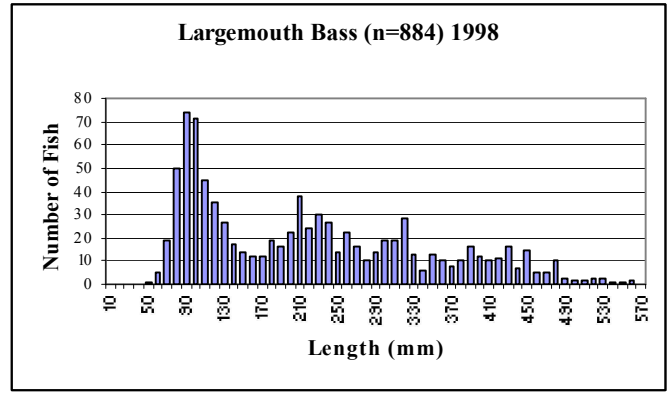
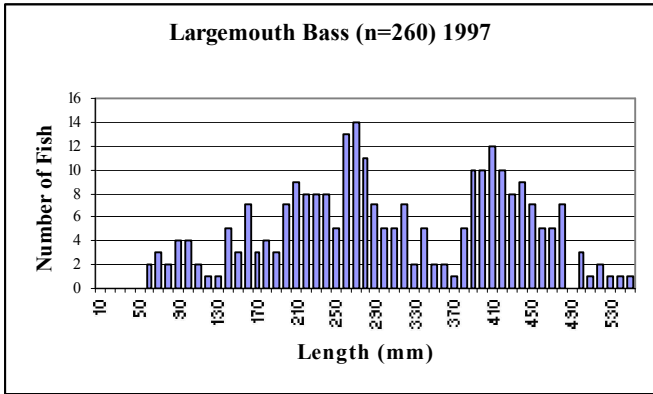
Appendix D.21 Regression equations of body length versus age for Kokanee Salmon, in 1997 and 1998.



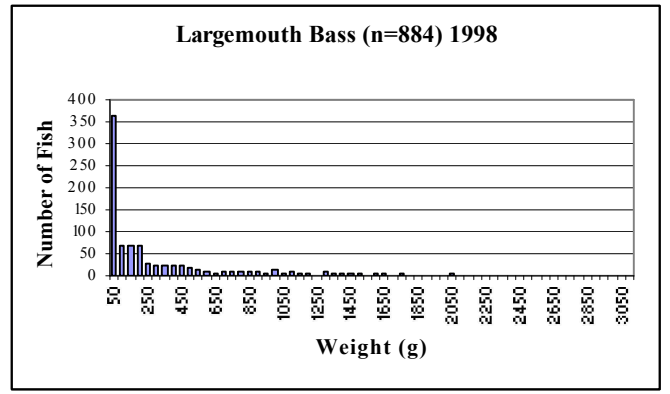
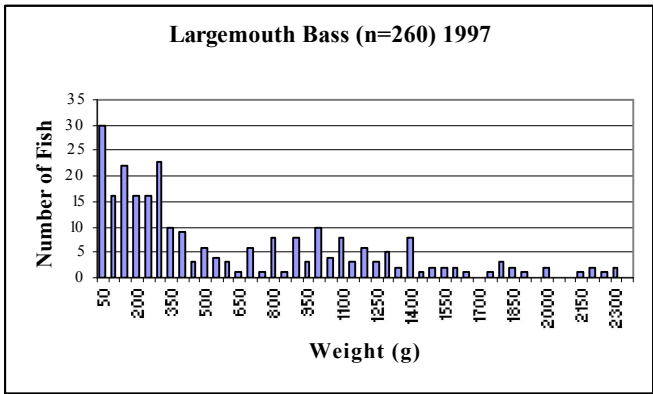
Appendix D.22 Regression equations of weight versus age for Kokanee Salmon, in 1997 and 1998.



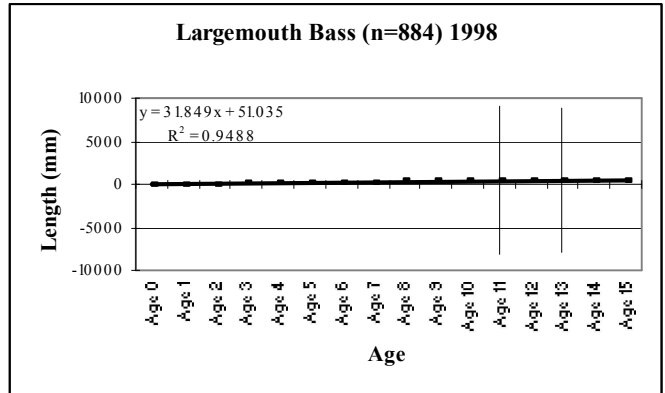
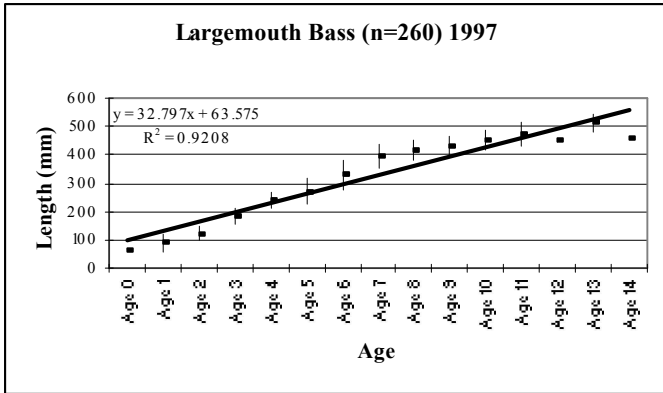
Appendix D.23 Age distribution of the number of Largemouth Bass sampled versus age in 1997 and 1998.



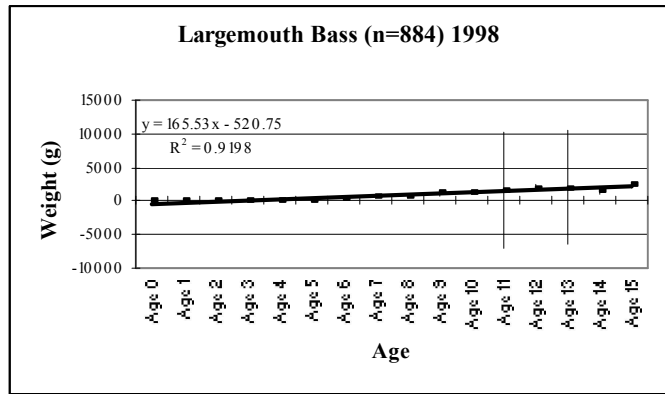
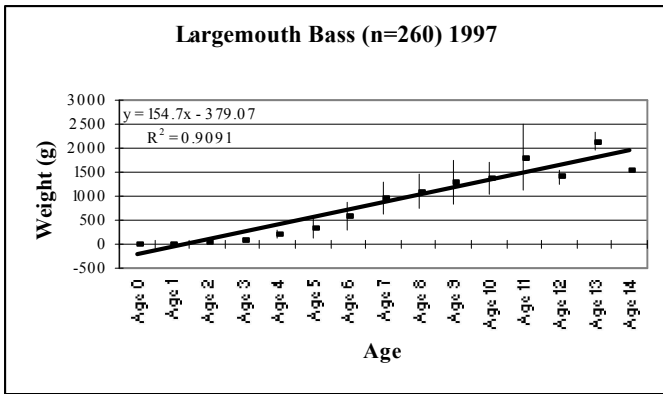
Appendix D.24 Frequency distribution of the number of Largemouth Bass sampled versus body length in 1997 and 1998.



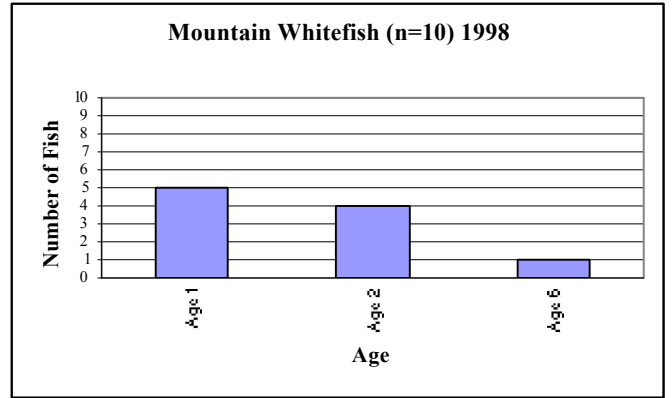
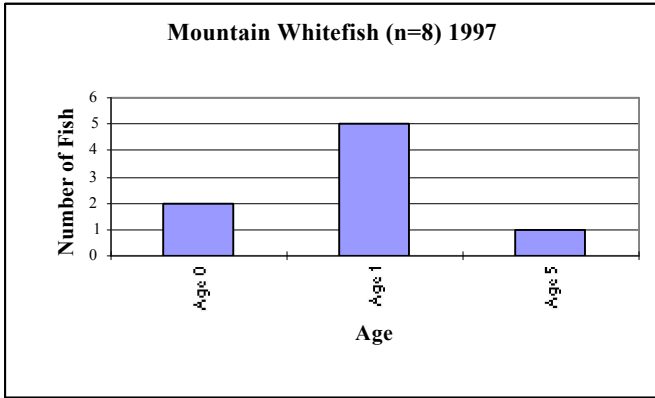
Appendix D.25 Frequency distribution of the number of Largemouth Bass sampled versus weight in 1997 and 1998.



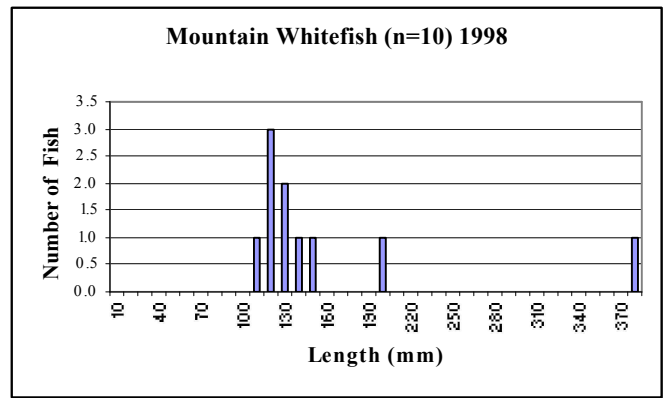
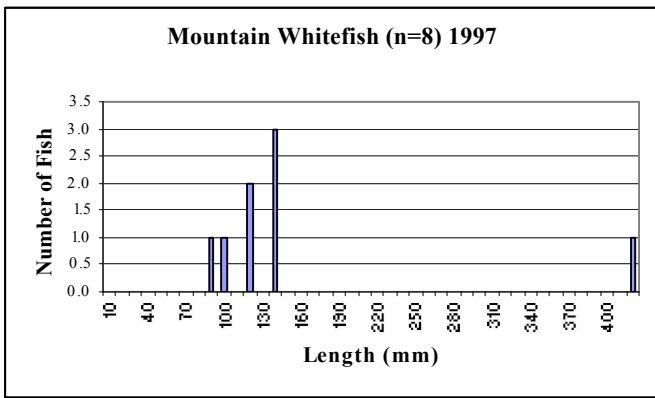
Appendix D.26 Regression equations of body length versus age for Largemouth Bass, in 1997 and 1998.



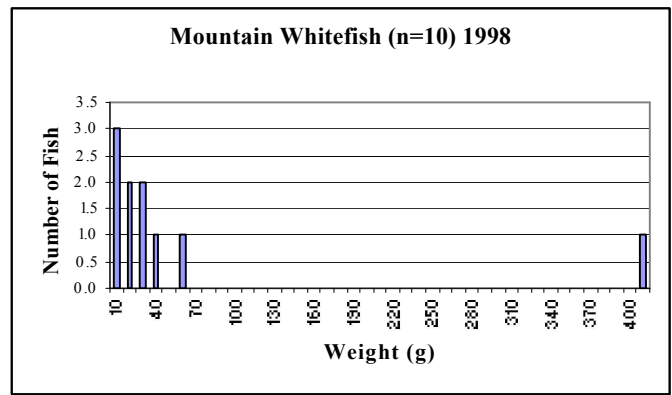
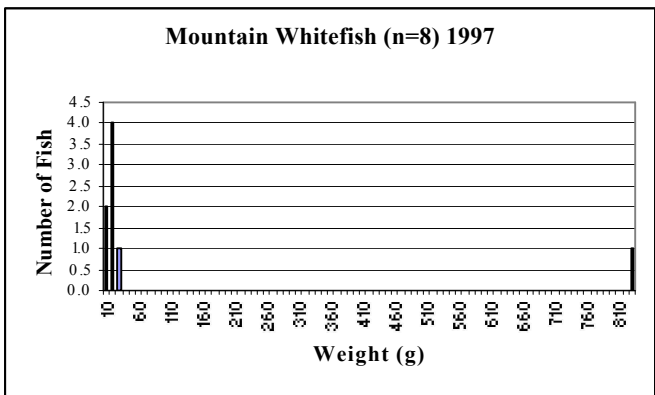
Appendix D.27 Regression equations of weight versus age for Largemouth Bass, in 1997 and 1998.



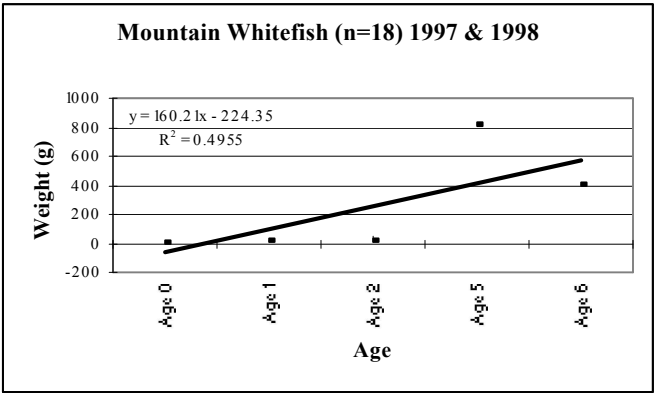
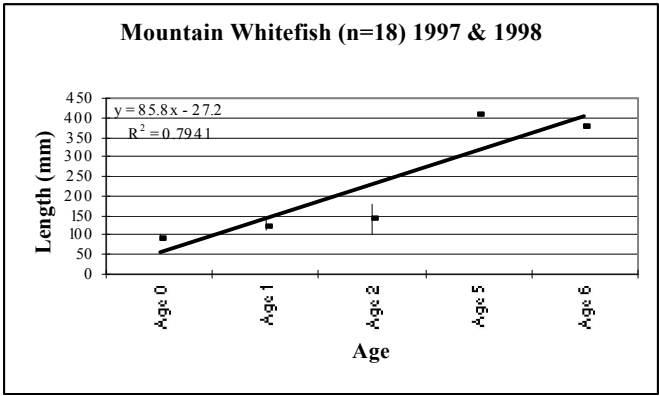
Appendix D.28 Age distribution of the number of Mountain Whitefish sampled versus age in 1997 and 1998.



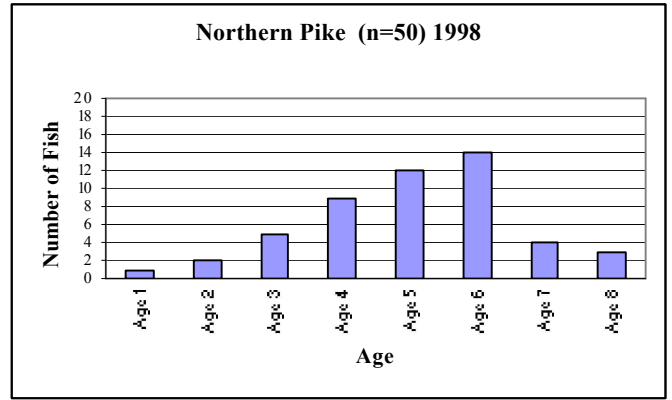
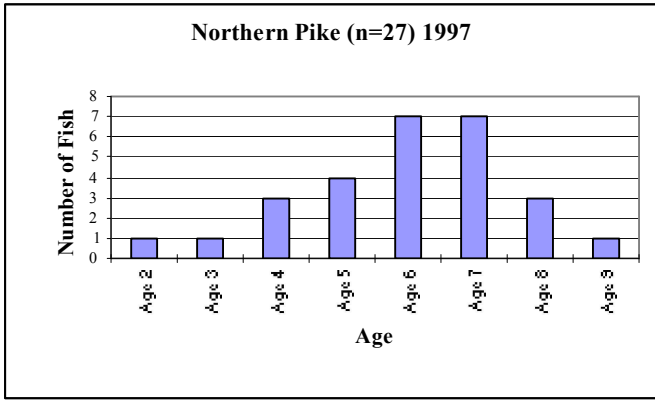
Appendix D.29 Frequency distribution of the number of Mountain Whitefish sampled versus body length in 1997 and 1998.



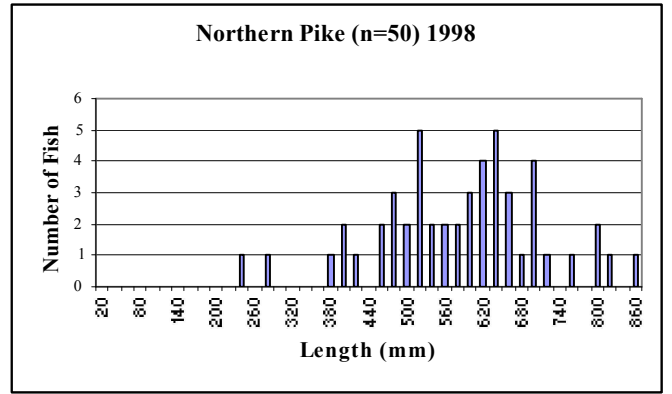
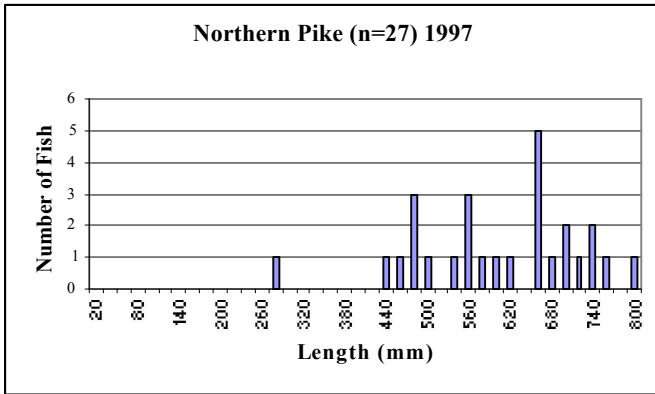
Appendix D.30 Frequency distribution of the number of Mountain Whitefish sampled versus weight in 1997 and 1998.



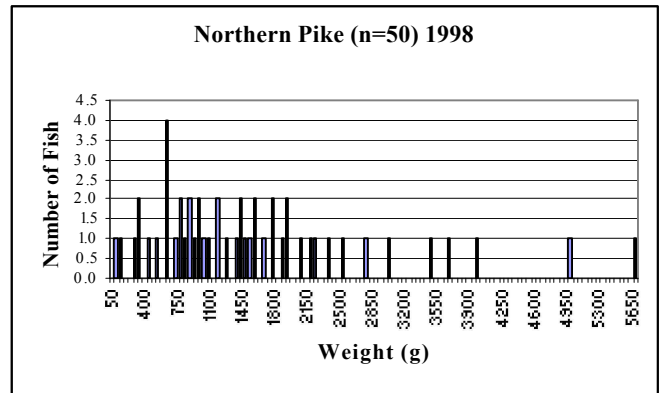
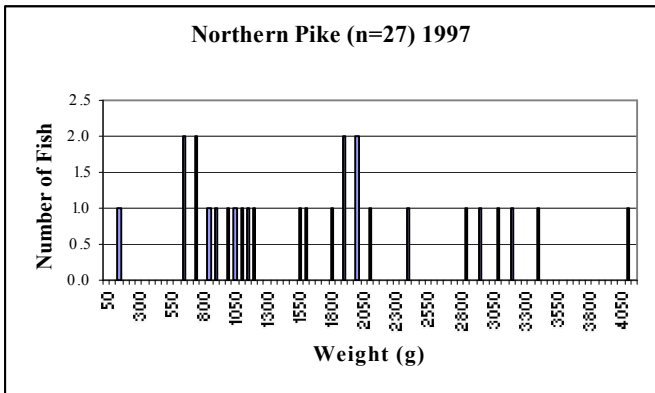
Appendix D.31 Regression equations of body length and weight versus age for Mountain Whitefish, combined 1997 and 1998 data.



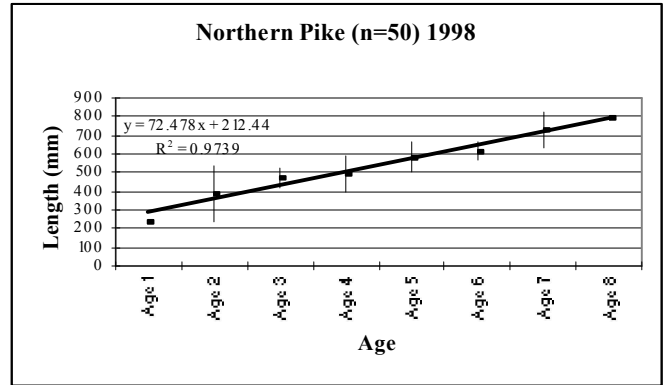
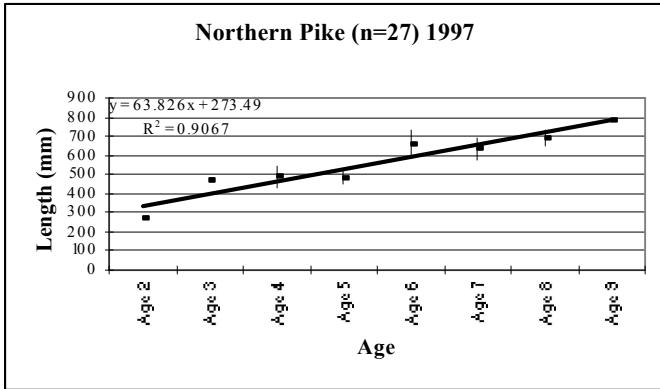
Appendix D.32 Age distribution of the number of Northern Pike sampled versus age in 1997 and 1998.



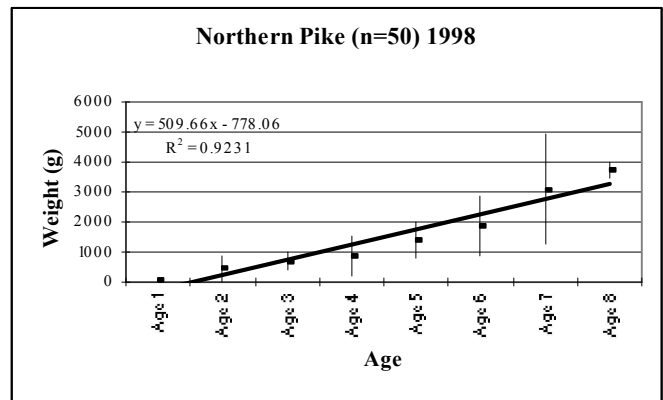
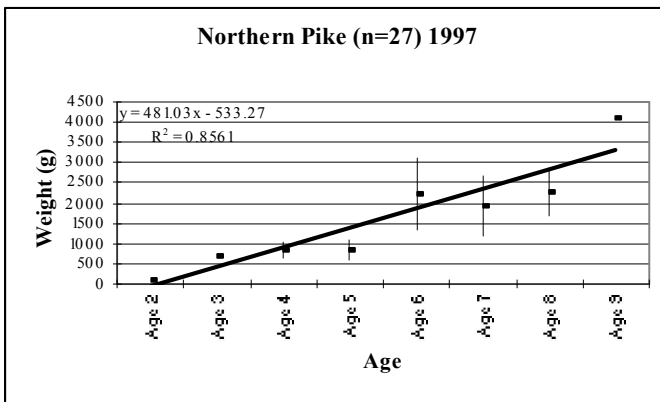
Appendix D.33 Frequency distribution of the number of Northern Pike sampled versus body length in 1997 and 1998.



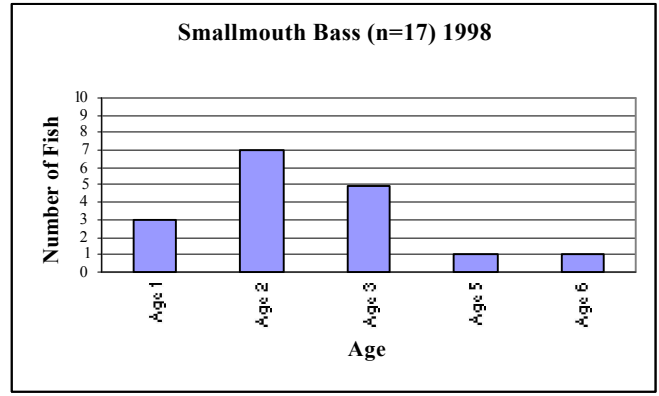
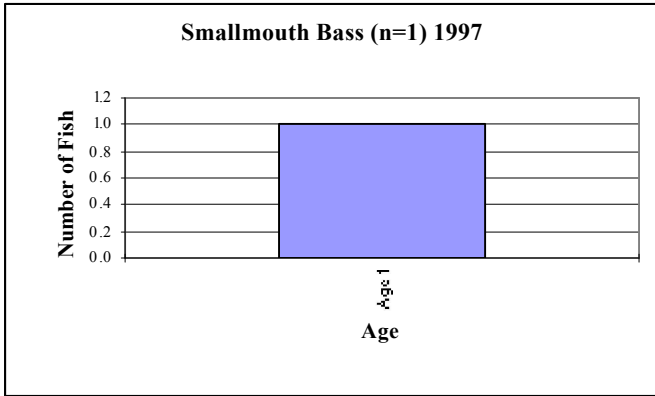
Appendix D.34 Frequency distribution of the number of Northern Pike sampled versus weight in 1997 and 1998.



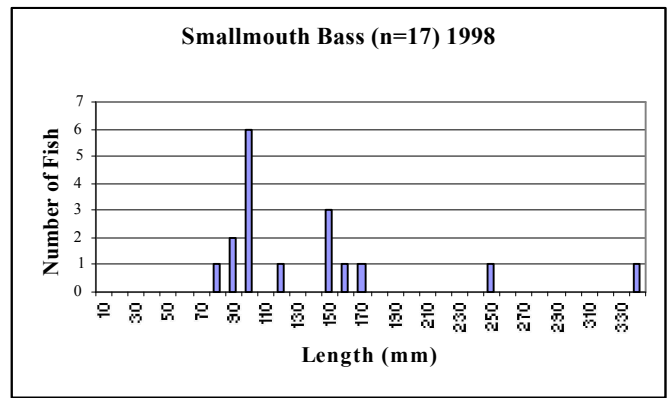
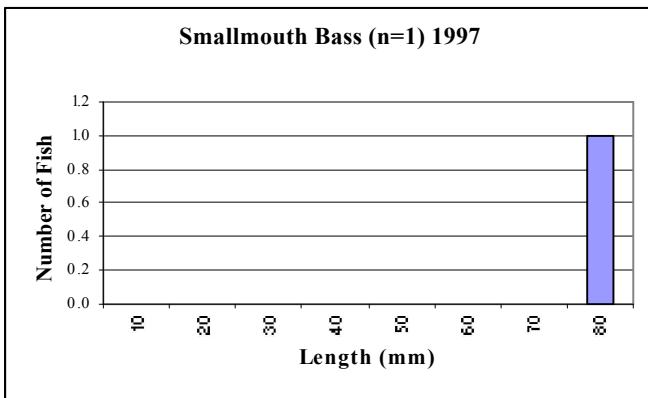
Appendix D.35 Regression equations of body length versus age for Northern Pike, in 1997 and 1998.



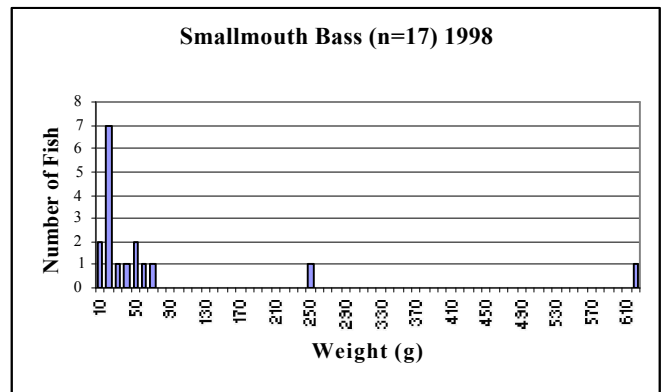
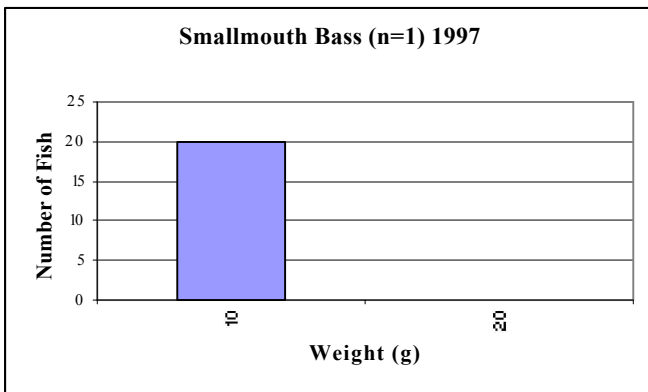
Appendix D.36 Regression equations of weight versus age for Northern Pike, in 1997 and 1998.



Appendix D.37 Age distribution of the number of Smallmouth Bass sampled versus age during 1997 and 1998.

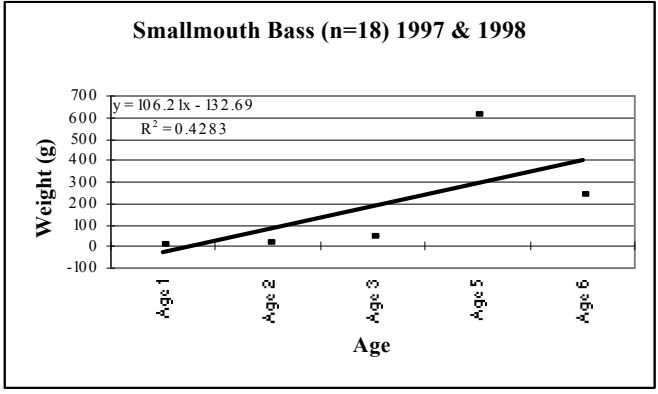
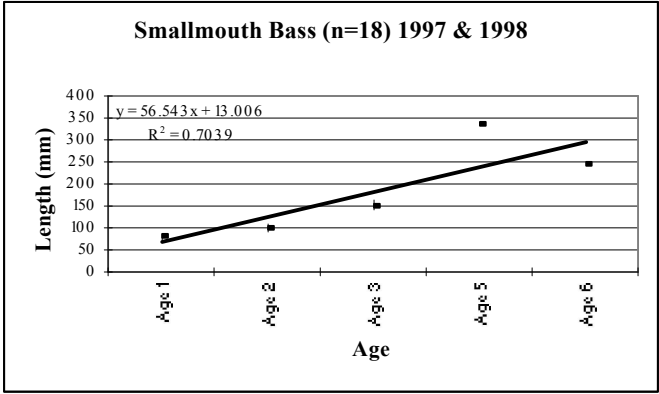


Appendix D.38 Frequency distribution of the number of Smallmouth Bass sampled versus body length during 1997 and 1998.

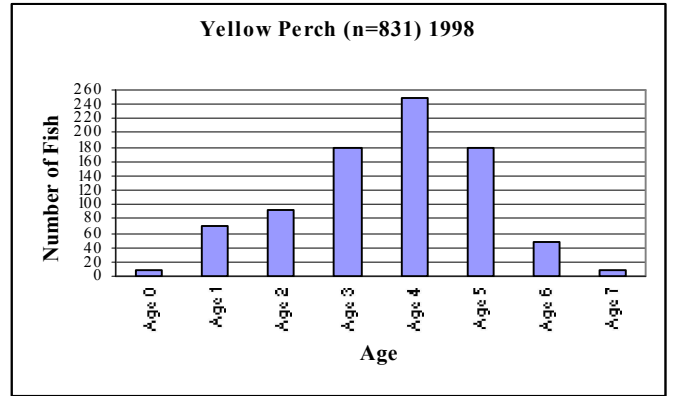
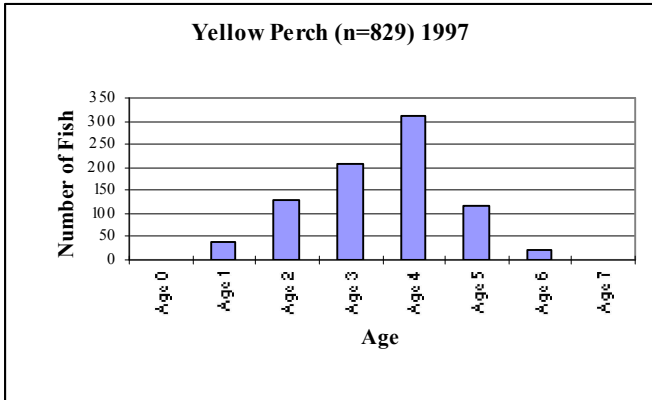


Appendix D.39 Frequency distribution of the number of Smallmouth Bass sampled versus weight during 1997 and 1998.

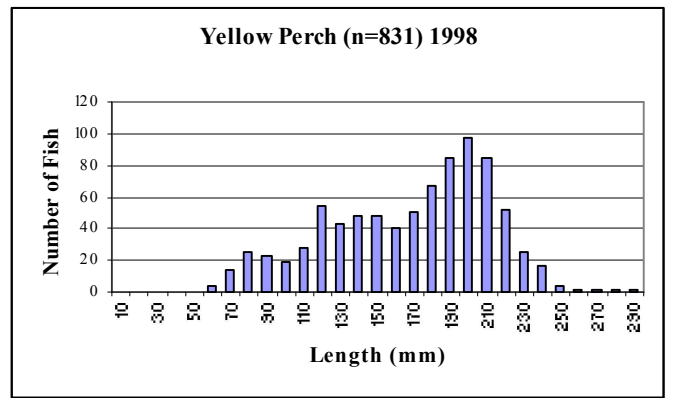
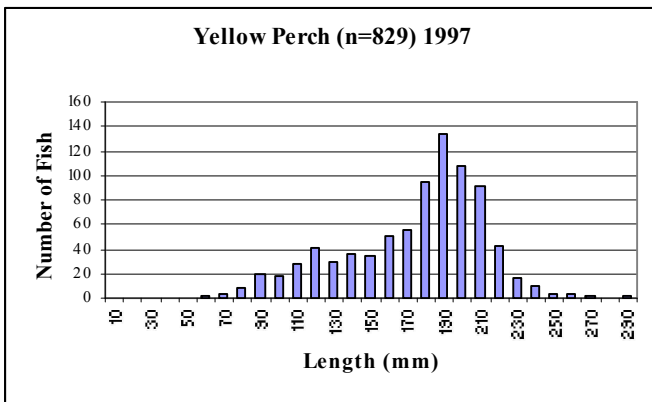




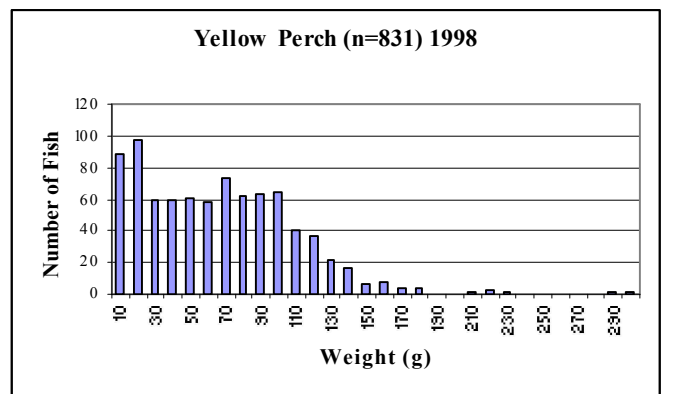
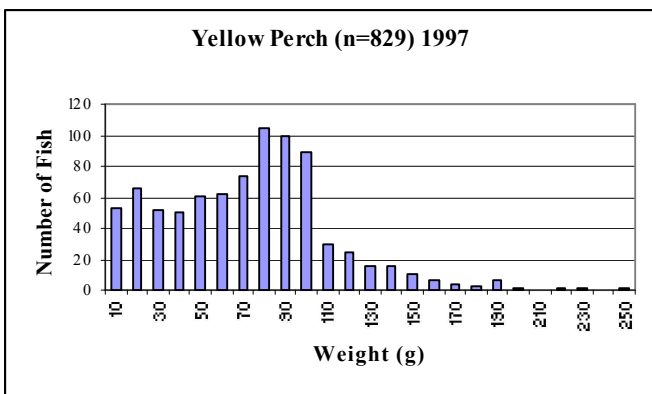
Appendix D.40 Regression equations of body length and weight versus age for Smallmouth Bass, combined 1997 and 1998 data.



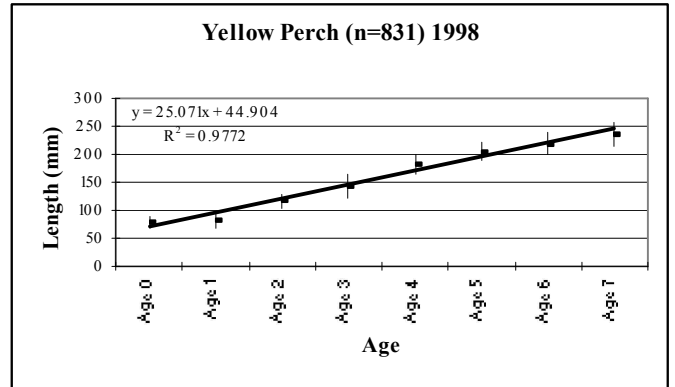
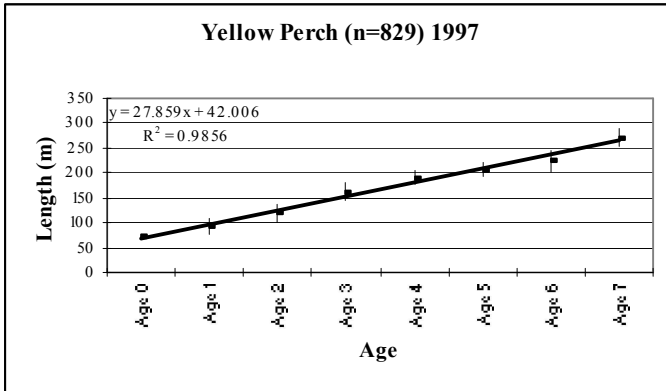
Appendix D.41 Age distribution of the number of Yellow Perch sampled versus age in 1997 and 1998.



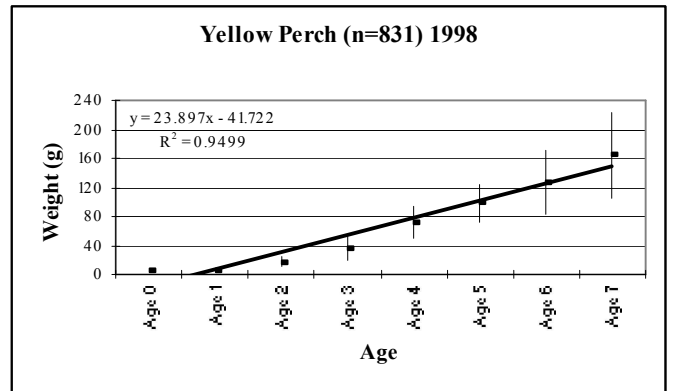
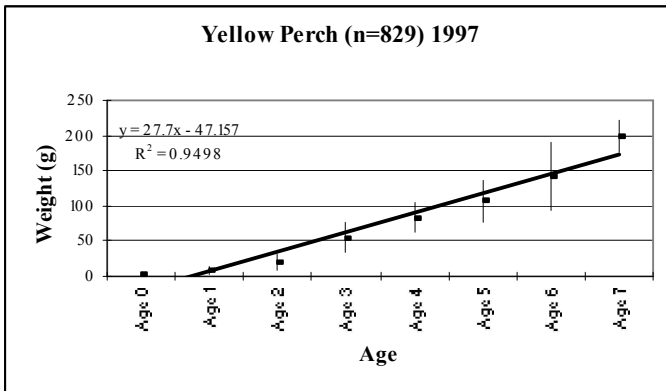
Appendix D.42 Frequency distribution of the number of Yellow Perch sampled versus body length in 1997 and 1998.



Appendix D.43 Frequency distribution of the number of Yellow Perch sampled versus weight in 1997 and 1998.



Appendix D.44 Regression equations of body length versus age for Yellow Perch, in 1997 and 1998.



Appendix D.45 Regression equations of weight versus age for Yellow Perch, in 1997 and 1998.

## **Appendix E**

Statistics of mean length and weight, standard deviation and range for all aged species sampled in Coeur d'Alene Lake, 1997 and 1998.

Appendix E.1 Mean length and weight, standard deviation and range for aged Black Crappie in 1997 and 1998.

| Age   | Statistics         | 1997 (n=247) |            | 1998 (n=678) |            |
|-------|--------------------|--------------|------------|--------------|------------|
|       |                    | Length (mm)  | Weight (g) | Length (mm)  | Weight (g) |
| Age 0 | Mean               | -            | -          | 62.60        | 3.00       |
|       | Standard Deviation | -            | -          | 6.85         | 1.33       |
|       | Range              | -            | -          | 53-75        | 1-5        |
| Age 1 | Mean               | 95.25        | 9.63       | 79.47        | 7.25       |
|       | Standard Deviation | 9.64         | 3.03       | 10.46        | 2.65       |
|       | Range              | 77-107       | 6-17       | 56-102       | 4-13       |
| Age 2 | Mean               | 106.60       | 19.46      | 102.35       | 13.75      |
|       | Standard Deviation | 13.89        | 9.68       | 10.42        | 6.70       |
|       | Range              | 81-118       | 8-62       | 75-150       | 3-54       |
| Age 3 | Mean               | 154.56       | 59.79      | 133.99       | 35.73      |
|       | Standard Deviation | 34.85        | 41.39      | 14.63        | 12.42      |
|       | Range              | 34.85        | 41.39      | 97-183       | 10-70      |
| Age 4 | Mean               | 192.61       | 110.16     | 156.36       | 61.41      |
|       | Standard Deviation | 11.61        | 23.15      | 18.05        | 30.08      |
|       | Range              | 160-214      | 50-191     | 115-229      | 26-314     |
| Age 5 | Mean               | 204.87       | 132.94     | 187.03       | 106.15     |
|       | Standard Deviation | 14.25        | 31.62      | 28.33        | 43.43      |
|       | Range              | 14.25        | 31.62      | 148-240      | 50-207     |
| Age 6 | Mean               | 216.33       | 152.67     | 221.15       | 171.38     |
|       | Standard Deviation | 13.85        | 33.93      | 18.59        | 38.91      |
|       | Range              | 199-247      | 100-226    | 170-260      | 76-297     |
| Age 7 | Mean               | 212.00       | 168.00     | 241.00       | 230.69     |
|       | Standard Deviation | -            | -          | 33.02        | 141.99     |
|       | Range              | -            | -          | 200-342      | 118-690    |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.2 Mean length and weight, standard deviation and range for aged Channel Catfish during 1998.

| Age   | Statistics         | 1998 (n=5)  |            |
|-------|--------------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) |
| Age 1 | Mean               | 216         | 75         |
|       | Standard Deviation | -           | -          |
|       | Range              | -           | -          |
| Age 4 | Mean               | 465         | 1365       |
|       | Standard Deviation | -           | -          |
|       | Range              | -           | -          |
| Age 5 | Mean               | 574.00      | 2366.50    |
|       | Standard Deviation | 79.20       | 702.16     |
|       | Range              | 518-630     | 1870-2863  |
| Age 7 | Mean               | 570         | 2541       |
|       | Standard Deviation | -           | -          |
|       | Range              | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.3 Mean length and weight, standard deviation and range for aged Chinook Salmon in 1997 and 1998.

| Age   | Statistics         | 1997 (n=19) |            | 1998 (n=33) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 0 | Mean               | -           | -          | 85          | 4.5        |
|       | Standard Deviation | -           | -          | 14.14       | 0.71       |
|       | Range              | -           | -          | 75-95       | 4-5        |
| Age 1 | Mean               | -           | -          | 100.00      | 5.25       |
|       | Standard Deviation | -           | -          | 12.25       | 2.63       |
|       | Range              | -           | -          | 85-110      | 3-8        |
| Age 2 | Mean               | 177.69      | 64.88      | 245.35      | 188.20     |
|       | Standard Deviation | 19.30       | 36.25      | 68.48       | 216.32     |
|       | Range              | 138-210     | 27-185     | 143-443     | 30-972     |
| Age 3 | Mean               | 413         | 830        | 374         | 353.2      |
|       | Standard Deviation | -           | -          | 83.20       | 141.91     |
|       | Range              | -           | -          | 250-449     | 116-500    |
| Age 4 | Mean               | -           | -          | 722         | 4000       |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 5 | Mean               | 848.5       | 7483.5     | 700         | 4886       |
|       | Standard Deviation | 9.19        | 962.37     | -           | -          |
|       | Range              | 842-855     | 6803-8164  | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.4 Mean length and weight, standard deviation and range for aged Cutthroat Trout in 1997 and 1998.

| Age   | Statistics         | 1997 (n=45) |            | 1998 (n=49) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 1 | Mean               | -           | -          | 120         | 10         |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 2 | Mean               | 182.50      | 51.50      | 141.40      | 24.70      |
|       | Standard Deviation | 0.71        | 19.09      | 13.53       | 8.67       |
|       | Range              | 182-183     | 38-65      | 120-160     | 10-36      |
| Age 3 | Mean               | 254.75      | 176.88     | 189.08      | 62.00      |
|       | Standard Deviation | 27.12       | 60.70      | 33.46       | 49.00      |
|       | Range              | 194-315     | 63-312     | 152-267     | 16-186     |
| Age 4 | Mean               | 304.14      | 281.21     | 243.82      | 163.76     |
|       | Standard Deviation | 25.80       | 76.46      | 37.28       | 83.32      |
|       | Range              | 260-344     | 154-374    | 185-290     | 59-300     |
| Age 5 | Mean               | 327.00      | 360.75     | 326.75      | 347.50     |
|       | Standard Deviation | 10.61       | 55.16      | 43.15       | 148.68     |
|       | Range              | 312-335     | 308-420    | 280-382     | 215-555    |
| Age 6 | Mean               | 342         | 368        | 316.00      | 313.80     |
|       | Standard Deviation | -           | -          | 41.70       | 118.56     |
|       | Range              | -           | -          | 265-367     | 184-455    |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.5 Mean length and weight, standard deviation and range for aged Kokanee Salmon in 1997 and 1998.

| Age   | Statistics         | 1997 (n=62) |            | 1998 (n=23) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 1 | Mean               | 120.83      | 11.00      | -           | -          |
|       | Standard Deviation | 17.88       | 5.10       | -           | -          |
|       | Range              | 87-140      | 4-15       | -           | -          |
| Age 2 | Mean               | 129.22      | 15.78      | 225         | 75         |
|       | Standard Deviation | 6.12        | 2.68       | -           | -          |
|       | Range              | 120-135     | 14-19      | -           | -          |
| Age 3 | Mean               | 235.11      | 110.67     | 303.25      | 272        |
|       | Standard Deviation | 9.91        | 19.09      | 42.14       | 103.94     |
|       | Range              | 222-250     | 74-135     | 245-338     | 132-352    |
| Age 4 | Mean               | 235.11      | 110.67     | 321.31      | 328.54     |
|       | Standard Deviation | 9.91        | 19.09      | 11.44       | 37.63      |
|       | Range              | 222-250     | 74-135     | 300-343     | 257-390    |
| Age 5 | Mean               | 301.50      | 287.72     | 338.60      | 288.60     |
|       | Standard Deviation | 28.11       | 89.05      | 5.90        | 52.19      |
|       | Range              | 251-340     | 140-402    | 333-345     | 232-338    |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.6 Mean length and weight, standard deviation and range for aged Largemouth Bass in 1997 and 1998.

| Age    | Statistics         | 1997 (n=260) |            | 1998 (n=884) |            |
|--------|--------------------|--------------|------------|--------------|------------|
|        |                    | Length (mm)  | Weight (g) | Length (mm)  | Weight (g) |
| Age 0  | Mean               | 64.33        | 4.67       | 66.00        | 3.50       |
|        | Standard Deviation | 4.04         | 2.52       | 14.94        | 1.29       |
|        | Range              | 60-68        | 2-7        | 49-85        | 2-5        |
| Age 1  | Mean               | 90.55        | 14.91      | 84.92        | 8.03       |
|        | Standard Deviation | 31.55        | 25.82      | 14.46        | 3.35       |
|        | Range              | 58-175       | 1-92       | 54-196       | 1-18       |
| Age 2  | Mean               | 122.18       | 29.55      | 107.98       | 16.74      |
|        | Standard Deviation | 23.65        | 15.51      | 16.15        | 8.88       |
|        | Range              | 90-156       | 11-59      | 80-175       | 5-64       |
| Age 3  | Mean               | 180.15       | 82.81      | 158.41       | 53.45      |
|        | Standard Deviation | 28.16        | 43.66      | 29.19        | 31.06      |
|        | Range              | 134-270      | 34-250     | 100-240      | 12-156     |
| Age 4  | Mean               | 240.80       | 209.95     | 212.17       | 137.23     |
|        | Standard Deviation | 29.91        | 81.56      | 39.58        | 75.90      |
|        | Range              | 159-290      | 56-415     | 147-421      | 38-457     |
| Age 5  | Mean               | 271.40       | 325.47     | 252.13       | 231.49     |
|        | Standard Deviation | 47.97        | 202.41     | 33.34        | 104.71     |
|        | Range              | 162-404      | 120-1000   | 186-346      | 79-630     |
| Age 6  | Mean               | 330.59       | 572.93     | 302.90       | 444.00     |
|        | Standard Deviation | 53.87        | 281.38     | 42.32        | 220.24     |
|        | Range              | 249-410      | 222-1040   | 210-380      | 116-1255   |
| Age 7  | Mean               | 392.69       | 966.72     | 338.02       | 598.81     |
|        | Standard Deviation | 42.99        | 342.13     | 42.73        | 285.62     |
|        | Range              | 310-470      | 400-1820   | 250-460      | 241-1600   |
| Age 8  | Mean               | 413.48       | 1091.71    | 372.10       | 796.87     |
|        | Standard Deviation | 35.58        | 353.08     | 37.09        | 232.48     |
|        | Range              | 338-473      | 544-2000   | 290-445      | 342-1465   |
| Age 9  | Mean               | 428.69       | 1291.00    | 422.26       | 1161.32    |
|        | Standard Deviation | 35.33        | 456.60     | 31.18        | 346.08     |
|        | Range              | 361-480      | 700-2200   | 365-492      | 411-2016   |
| Age 10 | Mean               | 449.40       | 1385.67    | 428.50       | 1281.63    |
|        | Standard Deviation | 35.58        | 335.54     | 26.11        | 292.96     |
|        | Range              | 406-515      | 884-1850   | 380-494      | 705-1960   |
| Age 11 | Mean               | 476.00       | 1812.50    | 452.93       | 1507.79    |
|        | Standard Deviation | 42.09        | 676.71     | 8676.44      | 8676.44    |
|        | Range              | 428-536      | 1075-2600  | 396-560      | 452-2832   |
| Age 12 | Mean               | 455.00       | 1400.00    | 489.80       | 1953.70    |
|        | Standard Deviation | 7.07         | 141.42     | 27.35        | 477.91     |
|        | Range              | 450-460      | 1300-1500  | 460-530      | 1396-2800  |
| Age 13 | Mean               | 512.00       | 2140.00    | 495.44       | 2031.44    |
|        | Standard Deviation | 28.84        | 181.66     | 8445.62      | 8445.62    |
|        | Range              | 468-546      | 1900-2300  | 410-560      | 1183-3073  |
| Age 14 | Mean               | 462          | 1550       | 460.00       | 1528.50    |
|        | Standard Deviation | -            | -          | 28.28        | 535.28     |
|        | Range              | -            | -          | 440-480      | 1150-1907  |
| Age 15 | Mean               | -            | -          | 504.50       | 2426.00    |
|        | Standard Deviation | -            | -          | 21.93        | 380.59     |
|        | Range              | -            | -          | 484-532      | 1893-2744  |

- Only one fish was caught in this age class or zero fish were sampled.



Appendix E.7 Mean length and weight, standard deviation and range for aged Mountain Whitefish in 1997 and 1998.

| Age   | Statistics         | 1997 (n=8)  |            | 1998 (n=10) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 0 | Mean               | 94.00       | 9.50       | -           | -          |
|       | Standard Deviation | 8.49        | 2.12       | -           | -          |
|       | Range              | 88-100      | 8-11       | -           | -          |
| Age 1 | Mean               | 126.20      | 15.80      | 123.80      | 16.00      |
|       | Standard Deviation | 11.90       | 5.40       | 14.41       | 11.07      |
|       | Range              | 112-138     | 7-21       | 104-141     | 4-32       |
| Age 2 | Mean               | -           | -          | 141.00      | 25.00      |
|       | Standard Deviation | -           | -          | 36.99       | 20.35      |
|       | Range              | -           | -          | 117-195     | 9-54       |
| Age 5 | Mean               | 411         | 825        | -           | -          |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 6 | Mean               | -           | -          | 380         | 406        |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.8 Mean length and weight, standard deviation and range for aged Northern Pike in 1997 and 1998.

| Age   | Statistics         | 1997 (n=27) |            | 1998 (n=50) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 1 | Mean               | -           | -          | 240         | 74         |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 2 | Mean               | 267.00      | 120.00     | 385.50      | 445.00     |
|       | Standard Deviation | -           | -          | 149.20      | 431.34     |
|       | Range              | -           | -          | 280-491     | 140-750    |
| Age 3 | Mean               | 470.00      | 710.00     | 472.60      | 679.00     |
|       | Standard Deviation | -           | -          | 53.42       | 308.90     |
|       | Range              | -           | -          | 400-550     | 350-1191   |
| Age 4 | Mean               | 491.67      | 833.33     | 492.44      | 866.89     |
|       | Standard Deviation | 57.50       | 201.08     | 91.76       | 667.08     |
|       | Range              | 434-549     | 610-1000   | 380-690     | 271-2520   |
| Age 5 | Mean               | 485.75      | 845.00     | 583.75      | 1387.58    |
|       | Standard Deviation | 38.80       | 250.53     | 83.89       | 610.26     |
|       | Range              | 448-540     | 620-1200   | 415-705     | 431-2377   |
| Age 6 | Mean               | 655.57      | 2226.43    | 615.36      | 1860.07    |
|       | Standard Deviation | 74.47       | 881.83     | 49.02       | 1008.42    |
|       | Range              | 550-750     | 1120-3390  | 505-691     | 983-4960   |
| Age 7 | Mean               | 636.29      | 1932.86    | 728.75      | 3092.50    |
|       | Standard Deviation | 57.74       | 750.48     | 98.17       | 1845.63    |
|       | Range              | 545-716     | 1030-3200  | 620-851     | 1442-5690  |
| Age 8 | Mean               | 689.33      | 2283.33    | 790.33      | 3718.33    |
|       | Standard Deviation | 39.00       | 579.51     | 9.24        | 272.96     |
|       | Range              | 650-728     | 1900-2950  | 785-801     | 3455-4000  |
| Age 9 | Mean               | 790         | 4100       | -           | -          |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.9 Mean length and weight, standard deviation and range for aged Rainbow Trout in 1997 and 1998.

| Age   | Statistics         | 1997 (n=1)  |            | 1998 (n=1)  |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 3 | Mean               | -           | -          | 430         | 630        |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 6 | Mean               | 370         | 508        | -           | -          |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.10 Mean length and weight, standard deviation and range for aged Smallmouth Bass in 1997 and 1998.

| Age   | Statistics         | 1997 (n=1)  |            | 1998 (n=17) |            |
|-------|--------------------|-------------|------------|-------------|------------|
|       |                    | Length (mm) | Weight (g) | Length (mm) | Weight (g) |
| Age 1 | Mean               | 75          | 6          | 83.00       | 10.67      |
|       | Standard Deviation | -           | -          | 5.57        | 6.35       |
|       | Range              | -           | -          | 77-88       | 7-18       |
| Age 2 | Mean               | -           | -          | 99.57       | 16.86      |
|       | Standard Deviation | -           | -          | 9.29        | 5.61       |
|       | Range              | -           | -          | 92-120      | 12-28      |
| Age 3 | Mean               | -           | -          | 151.60      | 48.40      |
|       | Standard Deviation | -           | -          | 10.71       | 10.36      |
|       | Range              | -           | -          | 143-168     | 35-62      |
| Age 5 | Mean               | -           | -          | 335         | 612        |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |
| Age 6 | Mean               | -           | -          | 246         | 243        |
|       | Standard Deviation | -           | -          | -           | -          |
|       | Range              | -           | -          | -           | -          |

- Only one fish was caught in this age class or zero fish were sampled.

Appendix E.11 Mean length and weight, standard deviation and range for aged Yellow Perch in 1997 and 1998.

| Age   | Statistics         | 1997 (n=829) |            | 1998 (n=831) |            |
|-------|--------------------|--------------|------------|--------------|------------|
|       |                    | Length (mm)  | Weight (g) | Length (mm)  | Weight (g) |
| Age 0 | Mean               | 73           | 3          | 78.00        | 4.57       |
|       | Standard Deviation | -            | -          | 10.86        | 1.99       |
|       | Range              | -            | -          | 61-90        | 2-7        |
| Age 1 | Mean               | 91.74        | 9.29       | 80.49        | 6.11       |
|       | Standard Deviation | 16.34        | 5.58       | 13.12        | 3.21       |
|       | Range              | 59-124       | 0-20       | 57-115       | 1-18       |
| Age 2 | Mean               | 119.23       | 20.81      | 116.15       | 17.03      |
|       | Standard Deviation | 18.66        | 11.41      | 13.37        | 6.78       |
|       | Range              | 80-159       | 4-55       | 85-154       | 4-38       |
| Age 3 | Mean               | 162.62       | 54.86      | 144.19       | 36.25      |
|       | Standard Deviation | 19.38        | 21.49      | 21.06        | 17.35      |
|       | Range              | 116-208      | 15-165     | 98-215       | 10-116     |
| Age 4 | Mean               | 190.72       | 83.72      | 183.46       | 72.34      |
|       | Standard Deviation | 14.05        | 20.94      | 17.82        | 22.78      |
|       | Range              | 125-230      | 24-180     | 126-234      | 18-175     |
| Age 5 | Mean               | 205.95       | 106.73     | 203.72       | 98.07      |
|       | Standard Deviation | 14.78        | 28.64      | 16.18        | 25.85      |
|       | Range              | 170-251      | 40-190     | 154-244      | 38-180     |
| Age 6 | Mean               | 224.22       | 141.52     | 219.04       | 127.15     |
|       | Standard Deviation | 22.00        | 47.65      | 18.95        | 44.57      |
|       | Range              | 186-267      | 72-242     | 190-280      | 63-284     |
| Age 7 | Mean               | 271.50       | 200.00     | 236.75       | 165.00     |
|       | Standard Deviation | 16.26        | 22.63      | 21.02        | 58.91      |
|       | Range              | 260-283      | 184-216    | 215-283      | 123-293    |

- Only one fish was caught in this age class or zero fish were sampled.

## **Appendix F**

Analysis of spawning gravel beds in select tributaries of the Coeur d'Alene Reservation, 1998.

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 5.0  
 2 6.0  
 3 4.0  
 4 6.0  
 5 5.0  
 6 8.0  
 7 6.0  
 8 4.0  
 9 7.0  
 10 9.0  
 Avg. Width (ft.) = 6

**McNeil Core Sample**

| Date Processed:         | 12-Nov-98     | Total Dry Weight (g): | 1412       |        |     |
|-------------------------|---------------|-----------------------|------------|--------|-----|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 3                     | 0          | 0      | 0   |
| Fine Sand               | 0.250         | 3.5                   | 0          | 0      | 0   |
| Medium Sand             | 0.500         | 4.5                   | 0          | 0      | 1   |
| Coarse Sand             | 0.850         | 2                     | 0          | 0      | 1   |
| Vry Coarse Sand         | 1.0           | 15.5                  | 1          | 1      | 2   |
| Vry Fine Gravel         | 2.0           | 99.5                  | 7          | 7      | 9   |
| Fine Gravel             | 4.75          | 79.5                  | 6          | 13     | 15  |
| Fine Gravel             | 6.3           | 85.5                  | 6          | 19     | 21  |
| Medium Gravel           | 8.0           | 409.5                 | 29         | 48     | 50  |
| Coarse Gravel           | 16.0          | 655.5                 | 47         | 95     | 96  |
| Vry Coarse Gravel       | 31.5          | 49.5                  | 4          | 99     | 100 |
| Small Cobble            | 63.5          | 0                     | 0          | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>1407.5</b>         | <b>100</b> |        |     |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count                       | %Total | %Cumm. |
|-----------------|----------------------------------|--------|--------|
| 0-2             | 10                               | 10     | 10     |
| 2-4             | 15                               | 15     | 25     |
| 4-8             | 20                               | 20     | 45     |
| 8-16            | 30                               | 30     | 75     |
| 16-32           | 20                               | 20     | 95     |
| 32-64           | 5                                | 5      | 100    |
| 64-128          | 0                                | 0      | 100    |
| 128-256         | 0                                | 0      | 100    |
| 256-512         | 0                                | 0      | 100    |
| 512-1024        | 0                                | 0      | 100    |
| 1024-2048       | 0                                | 0      | 100    |
| 2048-4096       | 0                                | 0      | 100    |
| <b>Total</b>    | <input type="text" value="100"/> |        |        |

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="41"/> | 11 | <input type="text" value="69"/> | 21 | <input type="text" value="34"/> |
| 2  | <input type="text" value="29"/> | 12 | <input type="text" value="80"/> | 22 | <input type="text" value="20"/> |
| 3  | <input type="text" value="55"/> | 13 | <input type="text" value="71"/> | 23 | <input type="text" value="28"/> |
| 4  | <input type="text" value="31"/> | 14 | <input type="text" value="47"/> | 24 | <input type="text" value="48"/> |
| 5  | <input type="text" value="28"/> | 15 | <input type="text" value="29"/> | 25 | <input type="text" value="50"/> |
| 6  | <input type="text" value="62"/> | 16 | <input type="text" value="35"/> | 26 | <input type="text" value="51"/> |
| 7  | <input type="text" value="68"/> | 17 | <input type="text" value="46"/> | 27 | <input type="text" value="67"/> |
| 8  | <input type="text" value="53"/> | 18 | <input type="text" value="72"/> | 28 | <input type="text" value="30"/> |
| 9  | <input type="text" value="43"/> | 19 | <input type="text" value="68"/> | 29 | <input type="text" value="35"/> |
| 10 | <input type="text" value="70"/> | 20 | <input type="text" value="60"/> | 30 | <input type="text" value="40"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.00019L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 \cdot P_1 + \dots + D_i \cdot P_i$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The Fredex index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 14.0  
 2 10.0  
 3 10.0  
 4 8.0  
 5 6.0  
 6 11.0  
 7 13.0  
 8 14.0  
 9 10.0  
 10 8.0  
 Avg. Width (ft.) = 10.4

**McNeil Core Sample**

| Date Processed:         | 12-Nov-98     | Total Dry Weight (g): | 3468       | %Total | %Cumm. |
|-------------------------|---------------|-----------------------|------------|--------|--------|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |        |
| Vry Fine Sand           | 0.125         | 11                    | 0          | 0      |        |
| Fine Sand               | 0.250         | 19.5                  | 1          | 1      |        |
| Medium Sand             | 0.500         | 32                    | 1          | 2      |        |
| Coarse Sand             | 0.850         | 12                    | 0          | 2      |        |
| Vry Coarse Sand         | 1.0           | 94                    | 3          | 5      |        |
| Vry Fine Gravel         | 2.0           | 246.5                 | 7          | 12     |        |
| Fine Gravel             | 4.75          | 147                   | 4          | 16     |        |
| Fine Gravel             | 6.3           | 147.5                 | 4          | 20     |        |
| Medium Gravel           | 8.0           | 645.5                 | 19         | 39     |        |
| Coarse Gravel           | 16.0          | 1019.5                | 29         | 69     |        |
| Vry Coarse Gravel       | 31.5          | 1086.5                | 31         | 100    |        |
| Small Cobble            | 63.5          | 0                     | 0          | 100    |        |
| <b>Total Weight (g)</b> |               | <b>3461</b>           | <b>100</b> |        |        |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 1          | 1      | 1      |
| 2-4             | 10         | 10     | 11     |
| 4-8             | 15         | 15     | 26     |
| 8-16            | 25         | 25     | 50     |
| 16-32           | 20         | 20     | 70     |
| 32-64           | 20         | 20     | 90     |
| 64-128          | 10         | 10     | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                  |    |                                 |
|----|----------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="50"/>  | 11 | <input type="text" value="97"/>  | 21 | <input type="text" value="41"/> |
| 2  | <input type="text" value="74"/>  | 12 | <input type="text" value="120"/> | 22 | <input type="text" value="32"/> |
| 3  | <input type="text" value="91"/>  | 13 | <input type="text" value="40"/>  | 23 | <input type="text" value="44"/> |
| 4  | <input type="text" value="38"/>  | 14 | <input type="text" value="72"/>  | 24 | <input type="text" value="77"/> |
| 5  | <input type="text" value="69"/>  | 15 | <input type="text" value="40"/>  | 25 | <input type="text" value="62"/> |
| 6  | <input type="text" value="78"/>  | 16 | <input type="text" value="62"/>  | 26 | <input type="text" value="90"/> |
| 7  | <input type="text" value="44"/>  | 17 | <input type="text" value="59"/>  | 27 | <input type="text" value="71"/> |
| 8  | <input type="text" value="50"/>  | 18 | <input type="text" value="81"/>  | 28 | <input type="text" value="63"/> |
| 9  | <input type="text" value="50"/>  | 19 | <input type="text" value="70"/>  | 29 | <input type="text" value="47"/> |
| 10 | <input type="text" value="110"/> | 20 | <input type="text" value="40"/>  | 30 | <input type="text" value="55"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Throuw and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>0.00729L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$$D_g = \sum_{i=1}^n P_i \cdot D_i^3 / \sum_{i=1}^n P_i \cdot D_i^2$$

where D<sub>g</sub> = the geometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;

P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.

\*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>, where S<sub>i</sub> = a sorting coefficient



**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>3</sup>):   
 Bankfull Width  
 1 14.0  
 2 12.0  
 3 10.0  
 4 10.0  
 5 12.0  
 6 14.0  
 7 13.0  
 8 9.0  
 9 10.0  
 10 7.0  
 Avg. Width (ft.) = 11.1

**McNeil Core Sample**

| Date Processed:         | 13-Nov-98     | Total Dry Weight (g): | 1573       |        |  |
|-------------------------|---------------|-----------------------|------------|--------|--|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |  |
| Vry Fine Sand           | 0.125         | 2.5                   | 0          | 0      |  |
| Fine Sand               | 0.250         | 3                     | 0          | 0      |  |
| Medium Sand             | 0.500         | 5                     | 0          | 1      |  |
| Coarse Sand             | 0.850         | 3                     | 0          | 1      |  |
| Vry Coarse Sand         | 1.0           | 35                    | 2          | 3      |  |
| Vry Fine Gravel         | 2.0           | 153.5                 | 10         | 13     |  |
| Fine Gravel             | 4.75          | 89.5                  | 6          | 19     |  |
| Fine Gravel             | 6.3           | 79                    | 5          | 24     |  |
| Medium Gravel           | 8.0           | 294.5                 | 19         | 43     |  |
| Coarse Gravel           | 16.0          | 368.5                 | 24         | 66     |  |
| Vry Coarse Gravel       | 31.5          | 527.5                 | 34         | 100    |  |
| Small Cobble            | 63.5          | 0                     | 0          | 100    |  |
| <b>Total Weight (g)</b> |               | <b>1561</b>           | <b>100</b> |        |  |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 18         | 18     | 18     |
| 2-4             | 20         | 20     | 38     |
| 4-8             | 20         | 20     | 58     |
| 8-16            | 20         | 20     | 78     |
| 16-32           | 15         | 15     | 93     |
| 32-64           | 7          | 7      | 100    |
| 64-128          | 0          | 0      | 100    |
| 128-256         | 0          | 0      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                 |
|----|----------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="46"/>  | 11 | <input type="text" value="65"/> | 21 | <input type="text" value="74"/> |
| 2  | <input type="text" value="72"/>  | 12 | <input type="text" value="60"/> | 22 | <input type="text" value="62"/> |
| 3  | <input type="text" value="101"/> | 13 | <input type="text" value="51"/> | 23 | <input type="text" value="90"/> |
| 4  | <input type="text" value="80"/>  | 14 | <input type="text" value="50"/> | 24 | <input type="text" value="81"/> |
| 5  | <input type="text" value="67"/>  | 15 | <input type="text" value="70"/> | 25 | <input type="text" value="64"/> |
| 6  | <input type="text" value="60"/>  | 16 | <input type="text" value="94"/> | 26 | <input type="text" value="48"/> |
| 7  | <input type="text" value="49"/>  | 17 | <input type="text" value="38"/> | 27 | <input type="text" value="41"/> |
| 8  | <input type="text" value="58"/>  | 18 | <input type="text" value="40"/> | 28 | <input type="text" value="52"/> |
| 9  | <input type="text" value="34"/>  | 19 | <input type="text" value="29"/> | 29 | <input type="text" value="57"/> |
| 10 | <input type="text" value="38"/>  | 20 | <input type="text" value="70"/> | 30 | <input type="text" value="30"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Throw and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.0005L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$$D_g = \frac{D_1 P_1 + 2 D_2 P_2 + \dots + D_n P_n}{P_1 + 2 P_2 + \dots + P_n}$$

where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient



**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 5.0  
 2 6.0  
 3 8.0  
 4 7.0  
 5 4.0  
 6 14.0  
 7 10.0  
 8 12.0  
 9 4.0  
 10 5.0  
 Avg. Width (ft.) = 8

**McNeil Core Sample**

Date Processed: 13-Nov-98 Total Dry Weight (g):   

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 1              | 0          | 0      |
| Fine Sand               | 0.250         | 1              | 0          | 0      |
| Medium Sand             | 0.500         | 1              | 0          | 0      |
| Coarse Sand             | 0.850         | 1              | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 4              | 0          | 0      |
| Vry Fine Gravel         | 2.0           | 29             | 1          | 2      |
| Fine Gravel             | 4.75          | 28.5           | 1          | 3      |
| Fine Gravel             | 6.3           | 42             | 2          | 5      |
| Medium Gravel           | 8.0           | 259            | 12         | 17     |
| Coarse Gravel           | 16.0          | 602.5          | 27         | 44     |
| Vry Coarse Gravel       | 31.5          | 909            | 41         | 86     |
| Small Cobble            | 63.5          | 318            | 14         | 100    |
| <b>Total Weight (g)</b> |               | <b>2196</b>    | <b>100</b> |        |

$D_g$  =  % < 6.3mm   
 $F_i$  =  % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 10         | 10     | 10     |
| 2-4             | 10         | 10     | 20     |
| 4-8             | 15         | 15     | 35     |
| 8-16            | 15         | 15     | 50     |
| 16-32           | 18         | 18     | 68     |
| 32-64           | 15         | 15     | 83     |
| 64-128          | 15         | 15     | 98     |
| 128-256         | 2          | 2      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |     |    |     |    |     |
|----|-----|----|-----|----|-----|
| 1  | 61  | 11 | 50  | 21 | 20  |
| 2  | 90  | 12 | 39  | 22 | 65  |
| 3  | 71  | 13 | 41  | 23 | 81  |
| 4  | 55  | 14 | 59  | 24 | 110 |
| 5  | 62  | 15 | 72  | 25 | 52  |
| 6  | 80  | 16 | 93  | 26 | 66  |
| 7  | 49  | 17 | 102 | 27 | 49  |
| 8  | 71  | 18 | 40  | 28 | 81  |
| 9  | 100 | 19 | 35  | 29 | 54  |
| 10 | 120 | 20 | 29  | 30 | 30  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E=82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$$D_g = \frac{D_1 \cdot P_1 + \dots + D_i \cdot P_i}{P_1 + \dots + P_i}; \text{ where } D_i = \text{the meometric mean (mm); } D_i = \text{the mean diameter (mm) of material retained on sieve } i;$$

$$P_i = \text{the proportion by weight of the entire sample made up of material retained on sieve } i.$$

\*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 27.0 |
| 2                  | 24.0 |
| 3                  | 14.0 |
| 4                  | 12.0 |
| 5                  | 18.0 |
| 6                  | 20.0 |
| 7                  | 12.0 |
| 8                  | 11.0 |
| 9                  | 10.0 |
| 10                 | 12.0 |
| Avg. Width (ft.) = | 16   |

**McNeil Core Sample**

| Date Processed:         | 02-Sep-98     | Total Dry Weight (g): | 2816.5     |        |  |
|-------------------------|---------------|-----------------------|------------|--------|--|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |  |
| Vry Fine Sand           | 0.125         | 8.5                   | 0          | 0      |  |
| Fine Sand               | 0.250         | 21                    | 1          | 1      |  |
| Medium Sand             | 0.500         | 42                    | 2          | 3      |  |
| Coarse Sand             | 0.850         | 15.5                  | 1          | 3      |  |
| Vry Coarse Sand         | 1.0           | 132.5                 | 5          | 8      |  |
| Vry Fine Gravel         | 2.0           | 168                   | 6          | 14     |  |
| Fine Gravel             | 4.75          | 66                    | 2          | 16     |  |
| Fine Gravel             | 6.3           | 84                    | 3          | 19     |  |
| Medium Gravel           | 8.0           | 707.5                 | 26         | 45     |  |
| Coarse Gravel           | 16.0          | 762                   | 28         | 73     |  |
| Vry Coarse Gravel       | 31.5          | 753                   | 27         | 100    |  |
| Small Cobble            | 63.5          | 0                     | 0          | 100    |  |
| <b>Total Weight (g)</b> |               | <b>2760</b>           | <b>100</b> |        |  |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |  |
|-----------------|------------|--------|--------|--|
| 0-2             | 8          | 8      | 8      |  |
| 2-4             | 13         | 13     | 21     |  |
| 4-8             | 15         | 15     | 36     |  |
| 8-16            | 17         | 17     | 53     |  |
| 16-32           | 25         | 25     | 78     |  |
| 32-64           | 15         | 15     | 93     |  |
| 64-128          | 7          | 7      | 100    |  |
| 128-256         |            | 0      | 100    |  |
| 256-512         |            | 0      | 100    |  |
| 512-1024        |            | 0      | 100    |  |
| 1024-2048       |            | 0      | 100    |  |
| 2048-4096       |            | 0      | 100    |  |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                  |
|----|----------------------------------|----|---------------------------------|----|----------------------------------|
| 1  | <input type="text" value="32"/>  | 11 | <input type="text" value="52"/> | 21 | <input type="text" value="104"/> |
| 2  | <input type="text" value="64"/>  | 12 | <input type="text" value="12"/> | 22 | <input type="text" value="112"/> |
| 3  | <input type="text" value="200"/> | 13 | <input type="text" value="18"/> | 23 | <input type="text" value="15"/>  |
| 4  | <input type="text" value="164"/> | 14 | <input type="text" value="25"/> | 24 | <input type="text" value="94"/>  |
| 5  | <input type="text" value="12"/>  | 15 | <input type="text" value="16"/> | 25 | <input type="text" value="99"/>  |
| 6  | <input type="text" value="29"/>  | 16 | <input type="text" value="13"/> | 26 | <input type="text" value="86"/>  |
| 7  | <input type="text" value="30"/>  | 17 | <input type="text" value="14"/> | 27 | <input type="text" value="74"/>  |
| 8  | <input type="text" value="42"/>  | 18 | <input type="text" value="18"/> | 28 | <input type="text" value="75"/>  |
| 9  | <input type="text" value="63"/>  | 19 | <input type="text" value="29"/> | 29 | <input type="text" value="26"/>  |
| 10 | <input type="text" value="75"/>  | 20 | <input type="text" value="78"/> | 30 | <input type="text" value="29"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E = 82.63e<sup>-0.0001L</sup>), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where D<sub>n</sub> = the meometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;  
 P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.  
 \*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location :   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 14.0 |
| 2                  | 10.0 |
| 3                  | 12.0 |
| 4                  | 18.0 |
| 5                  | 9.0  |
| 6                  | 8.0  |
| 7                  | 6.0  |
| 8                  | 12.0 |
| 9                  | 13.0 |
| 10                 | 6.0  |
| Avg. Width (ft.) = | 10.8 |

**McNeil Core Sample**

|                         |                      |                       |               |               |  |
|-------------------------|----------------------|-----------------------|---------------|---------------|--|
| Date Processed:         | 20-Aug-98            | Total Dry Weight (g): | 4108          |               |  |
| <b>Particle</b>         | <b>Diameter (mm)</b> | <b>Weight (grams)</b> | <b>%Total</b> | <b>%Cumm.</b> |  |
| Vry Fine Sand           | 0.125                | 37.5                  | 1             | 1             |  |
| Fine Sand               | 0.250                | 127.5                 | 3             | 4             |  |
| Medium Sand             | 0.500                | 248                   | 6             | 10            |  |
| Coarse Sand             | 0.850                | 66.5                  | 2             | 12            |  |
| Vry Coarse Sand         | 1.0                  | 258.5                 | 6             | 18            |  |
| Vry Fine Gravel         | 2.0                  | 297.5                 | 7             | 25            |  |
| Fine Gravel             | 4.75                 | 108.5                 | 3             | 28            |  |
| Fine Gravel             | 6.3                  | 96                    | 2             | 30            |  |
| Medium Gravel           | 8.0                  | 517.5                 | 13            | 43            |  |
| Coarse Gravel           | 16.0                 | 999                   | 24            | 67            |  |
| Vry Coarse Gravel       | 31.5                 | 933                   | 23            | 90            |  |
| Small Cobble            | 63.5                 | 418.5                 | 10            | 100           |  |
| <b>Total Weight (g)</b> |                      | <b>4108</b>           | <b>100</b>    |               |  |

$D_g =$        % < 6.3mm   
 $F_i =$        % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count                       | %Total | %Cumm. |
|-----------------|----------------------------------|--------|--------|
| 0-2             | 1                                | 1      | 1      |
| 2-4             | 7                                | 7      | 8      |
| 4-8             | 16                               | 16     | 24     |
| 8-16            | 20                               | 20     | 44     |
| 16-32           | 25                               | 25     | 69     |
| 32-64           | 20                               | 20     | 89     |
| 64-128          | 10                               | 10     | 99     |
| 128-256         | 1                                | 1      | 100    |
| 256-512         |                                  | 0      | 100    |
| 512-1024        |                                  | 0      | 100    |
| 1024-2048       |                                  | 0      | 100    |
| 2048-4096       |                                  | 0      | 100    |
| <b>Total</b>    | <input type="text" value="100"/> |        |        |

**30 Count Results**

|    |                                  |    |                                  |    |                                  |
|----|----------------------------------|----|----------------------------------|----|----------------------------------|
| 1  | <input type="text" value="90"/>  | 11 | <input type="text" value="70"/>  | 21 | <input type="text" value="64"/>  |
| 2  | <input type="text" value="95"/>  | 12 | <input type="text" value="80"/>  | 22 | <input type="text" value="47"/>  |
| 3  | <input type="text" value="100"/> | 13 | <input type="text" value="97"/>  | 23 | <input type="text" value="60"/>  |
| 4  | <input type="text" value="103"/> | 14 | <input type="text" value="83"/>  | 24 | <input type="text" value="56"/>  |
| 5  | <input type="text" value="86"/>  | 15 | <input type="text" value="79"/>  | 25 | <input type="text" value="115"/> |
| 6  | <input type="text" value="80"/>  | 16 | <input type="text" value="58"/>  | 26 | <input type="text" value="80"/>  |
| 7  | <input type="text" value="78"/>  | 17 | <input type="text" value="100"/> | 27 | <input type="text" value="80"/>  |
| 8  | <input type="text" value="74"/>  | 18 | <input type="text" value="87"/>  | 28 | <input type="text" value="85"/>  |
| 9  | <input type="text" value="41"/>  | 19 | <input type="text" value="78"/>  | 29 | <input type="text" value="64"/>  |
| 10 | <input type="text" value="83"/>  | 20 | <input type="text" value="50"/>  | 30 | <input type="text" value="87"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = D_1 P_1 + D_2 P_2 + \dots + D_n P_n$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i / S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 8.0  |
| 2  | 7.0  |
| 3  | 8.5  |
| 4  | 9.0  |
| 5  | 8.0  |
| 6  | 9.0  |
| 7  | 16.0 |
| 8  | 12.0 |
| 9  | 10.0 |
| 10 | 10.0 |

Avg. Width (ft.) =

**McNeil Core Sample**

| Date Processed:         | 20-Aug-98     | Total Dry Weight (g): | 4313       |        |  |
|-------------------------|---------------|-----------------------|------------|--------|--|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |  |
| Vry Fine Sand           | 0.125         | 65.5                  | 1          | 1      |  |
| Fine Sand               | 0.250         | 112                   | 2          | 4      |  |
| Medium Sand             | 0.500         | 125.5                 | 3          | 7      |  |
| Coarse Sand             | 0.850         | 40                    | 1          | 8      |  |
| Vry Coarse Sand         | 1.0           | 522                   | 12         | 19     |  |
| Vry Fine Gravel         | 2.0           | 520                   | 11         | 31     |  |
| Fine Gravel             | 4.75          | 250.5                 | 6          | 36     |  |
| Fine Gravel             | 6.3           | 226                   | 5          | 41     |  |
| Medium Gravel           | 8.0           | 885                   | 20         | 61     |  |
| Coarse Gravel           | 16.0          | 764                   | 17         | 77     |  |
| Vry Coarse Gravel       | 31.5          | 539.5                 | 12         | 89     |  |
| Small Cobble            | 63.5          | 483.5                 | 11         | 100    |  |
| <b>Total Weight (g)</b> |               | <b>4533</b>           | <b>100</b> |        |  |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 8          | 8      | 8      |
| 2-4             | 7          | 7      | 15     |
| 4-8             | 8          | 8      | 24     |
| 8-16            | 9          | 9      | 33     |
| 16-32           | 8          | 8      | 41     |
| 32-64           | 9          | 9      | 51     |
| 64-128          | 16         | 16     | 67     |
| 128-256         | 12         | 12     | 79     |
| 256-512         | 10         | 10     | 90     |
| 512-1024        | 10         | 10     | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="50"/> | 11 | <input type="text" value="93"/> | 21 | <input type="text" value="30"/> |
| 2  | <input type="text" value="58"/> | 12 | <input type="text" value="60"/> | 22 | <input type="text" value="36"/> |
| 3  | <input type="text" value="49"/> | 13 | <input type="text" value="34"/> | 23 | <input type="text" value="30"/> |
| 4  | <input type="text" value="53"/> | 14 | <input type="text" value="45"/> | 24 | <input type="text" value="33"/> |
| 5  | <input type="text" value="65"/> | 15 | <input type="text" value="28"/> | 25 | <input type="text" value="27"/> |
| 6  | <input type="text" value="40"/> | 16 | <input type="text" value="40"/> | 26 | <input type="text" value="38"/> |
| 7  | <input type="text" value="90"/> | 17 | <input type="text" value="40"/> | 27 | <input type="text" value="40"/> |
| 8  | <input type="text" value="50"/> | 18 | <input type="text" value="35"/> | 28 | <input type="text" value="24"/> |
| 9  | <input type="text" value="41"/> | 19 | <input type="text" value="41"/> | 29 | <input type="text" value="60"/> |
| 10 | <input type="text" value="56"/> | 20 | <input type="text" value="34"/> | 30 | <input type="text" value="40"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{1.0033L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where  $D_i$  = the meometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 5.0  |
| 2  | 7.0  |
| 3  | 10.0 |
| 4  | 13.0 |
| 5  | 9.0  |
| 6  | 5.0  |
| 7  | 5.0  |
| 8  | 8.0  |
| 9  | 8.0  |
| 10 | 12.0 |

Avg. Width (ft.) = 8.2

**McNeil Core Sample**

Date Processed: 13-Nov-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 7.5            | 0          | 0      |
| Fine Sand               | 0.250         | 18.5           | 1          | 1      |
| Medium Sand             | 0.500         | 43             | 1          | 2      |
| Coarse Sand             | 0.850         | 20             | 1          | 3      |
| Vry Coarse Sand         | 1.0           | 136            | 4          | 7      |
| Vry Fine Gravel         | 2.0           | 455            | 13         | 20     |
| Fine Gravel             | 4.75          | 300            | 9          | 29     |
| Fine Gravel             | 6.3           | 274.5          | 8          | 37     |
| Medium Gravel           | 8.0           | 1470.5         | 43         | 80     |
| Coarse Gravel           | 16.0          | 701.5          | 20         | 100    |
| Vry Coarse Gravel       | 31.5          | 0              | 0          | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>3426.5</b>  | <b>100</b> |        |

$D_g$  =  % < 6.3mm

$F_i$  =  % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 0          | 0      | 0      |
| 2-4             | 11         | 11     | 11     |
| 4-8             | 14         | 14     | 25     |
| 8-16            | 21         | 21     | 46     |
| 16-32           | 26         | 26     | 72     |
| 32-64           | 20         | 20     | 92     |
| 64-128          | 4          | 4      | 96     |
| 128-256         | 4          | 4      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                  |    |                                 |
|----|---------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="80"/> | 11 | <input type="text" value="101"/> | 21 | <input type="text" value="42"/> |
| 2  | <input type="text" value="65"/> | 12 | <input type="text" value="90"/>  | 22 | <input type="text" value="62"/> |
| 3  | <input type="text" value="46"/> | 13 | <input type="text" value="60"/>  | 23 | <input type="text" value="75"/> |
| 4  | <input type="text" value="74"/> | 14 | <input type="text" value="37"/>  | 24 | <input type="text" value="31"/> |
| 5  | <input type="text" value="30"/> | 15 | <input type="text" value="41"/>  | 25 | <input type="text" value="30"/> |
| 6  | <input type="text" value="48"/> | 16 | <input type="text" value="29"/>  | 26 | <input type="text" value="51"/> |
| 7  | <input type="text" value="51"/> | 17 | <input type="text" value="25"/>  | 27 | <input type="text" value="58"/> |
| 8  | <input type="text" value="52"/> | 18 | <input type="text" value="60"/>  | 28 | <input type="text" value="64"/> |
| 9  | <input type="text" value="76"/> | 19 | <input type="text" value="57"/>  | 29 | <input type="text" value="88"/> |
| 10 | <input type="text" value="35"/> | 20 | <input type="text" value="38"/>  | 30 | <input type="text" value="41"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>0.0005L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$$D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n; \text{ where } D_g = \text{the meometric mean (mm); } D_i = \text{the mean diameter (mm) of material retained on sieve } i;$$

$$P_i = \text{the proportion by weight of the entire sample made up of material retained on sieve } i.$$

\*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>g</sub>/S<sub>i</sub>, where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Analysis and Results**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 6.0  |
| 2                  | 6.0  |
| 3                  | 10.0 |
| 4                  | 12.0 |
| 5                  | 4.0  |
| 6                  | 5.0  |
| 7                  | 4.0  |
| 8                  | 10.0 |
| 9                  | 9.0  |
| 10                 | 10.0 |
| Avg. Width (ft.) = | 7.2  |

**McNeil Core Sample**

Date Processed: 25-Aug-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 2              | 0          | 0      |
| Fine Sand               | 0.250         | 6.5            | 0          | 0      |
| Medium Sand             | 0.500         | 15             | 0          | 1      |
| Coarse Sand             | 0.850         | 7              | 0          | 1      |
| Vry Coarse Sand         | 1.0           | 60.5           | 2          | 3      |
| Vry Fine Gravel         | 2.0           | 245.5          | 7          | 10     |
| Fine Gravel             | 4.75          | 178            | 5          | 15     |
| Fine Gravel             | 6.3           | 208.5          | 6          | 22     |
| Medium Gravel           | 8.0           | 890            | 27         | 48     |
| Coarse Gravel           | 16.0          | 770.5          | 23         | 71     |
| Vry Coarse Gravel       | 31.5          | 969.5          | 29         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>3353</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 18         | 18     | 18     |
| 2-4             | 27         | 27     | 45     |
| 4-8             | 26         | 26     | 71     |
| 8-16            | 22         | 22     | 93     |
| 16-32           | 7          | 7      | 100    |
| 32-64           | 0          | 0      | 100    |
| 64-128          | 0          | 0      | 100    |
| 128-256         | 0          | 0      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="29"/> | 11 | <input type="text" value="40"/> | 21 | <input type="text" value="19"/> |
| 2  | <input type="text" value="34"/> | 12 | <input type="text" value="59"/> | 22 | <input type="text" value="27"/> |
| 3  | <input type="text" value="42"/> | 13 | <input type="text" value="65"/> | 23 | <input type="text" value="38"/> |
| 4  | <input type="text" value="58"/> | 14 | <input type="text" value="60"/> | 24 | <input type="text" value="42"/> |
| 5  | <input type="text" value="50"/> | 15 | <input type="text" value="71"/> | 25 | <input type="text" value="50"/> |
| 6  | <input type="text" value="49"/> | 16 | <input type="text" value="76"/> | 26 | <input type="text" value="61"/> |
| 7  | <input type="text" value="31"/> | 17 | <input type="text" value="30"/> | 27 | <input type="text" value="68"/> |
| 8  | <input type="text" value="28"/> | 18 | <input type="text" value="34"/> | 28 | <input type="text" value="70"/> |
| 9  | <input type="text" value="25"/> | 19 | <input type="text" value="28"/> | 29 | <input type="text" value="77"/> |
| 10 | <input type="text" value="37"/> | 20 | <input type="text" value="20"/> | 30 | <input type="text" value="32"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 13.0 |
| 2  | 10.0 |
| 3  | 12.0 |
| 4  | 9.0  |
| 5  | 8.0  |
| 6  | 10.0 |
| 7  | 6.0  |
| 8  | 8.0  |
| 9  | 4.0  |
| 10 | 10.0 |

Avg. Width (ft.) =

**McNeil Core Sample**

Date Processed: 12-Nov-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 0.5            | 0          | 0      |
| Fine Sand               | 0.250         | 1              | 0          | 0      |
| Medium Sand             | 0.500         | 1              | 0          | 0      |
| Coarse Sand             | 0.850         | 0.5            | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 2              | 0          | 0      |
| Vry Fine Gravel         | 2.0           | 48             | 2          | 2      |
| Fine Gravel             | 4.75          | 1.5            | 0          | 2      |
| Fine Gravel             | 6.3           | 6              | 0          | 2      |
| Medium Gravel           | 8.0           | 91.5           | 3          | 6      |
| Coarse Gravel           | 16.0          | 755            | 28         | 34     |
| Vry Coarse Gravel       | 31.5          | 1774.5         | 66         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>2681.5</b>  | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 4          | 4      | 4      |
| 2-4             | 6          | 6      | 10     |
| 4-8             | 20         | 20     | 30     |
| 8-16            | 32         | 32     | 62     |
| 16-32           | 18         | 18     | 80     |
| 32-64           | 20         | 20     | 100    |
| 64-128          |            | 0      | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                  |    |                                 |
|----|---------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="27"/> | 11 | <input type="text" value="16"/>  | 21 | <input type="text" value="28"/> |
| 2  | <input type="text" value="26"/> | 12 | <input type="text" value="66"/>  | 22 | <input type="text" value="36"/> |
| 3  | <input type="text" value="65"/> | 13 | <input type="text" value="124"/> | 23 | <input type="text" value="16"/> |
| 4  | <input type="text" value="72"/> | 14 | <input type="text" value="106"/> | 24 | <input type="text" value="12"/> |
| 5  | <input type="text" value="14"/> | 15 | <input type="text" value="60"/>  | 25 | <input type="text" value="10"/> |
| 6  | <input type="text" value="19"/> | 16 | <input type="text" value="58"/>  | 26 | <input type="text" value="17"/> |
| 7  | <input type="text" value="26"/> | 17 | <input type="text" value="52"/>  | 27 | <input type="text" value="21"/> |
| 8  | <input type="text" value="24"/> | 18 | <input type="text" value="40"/>  | 28 | <input type="text" value="19"/> |
| 9  | <input type="text" value="8"/>  | 19 | <input type="text" value="32"/>  | 29 | <input type="text" value="16"/> |
| 10 | <input type="text" value="4"/>  | 20 | <input type="text" value="24"/>  | 30 | <input type="text" value="24"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = \frac{\sum D_i \cdot P_i \cdot X_i}{\sum P_i \cdot X_i}$ ; where  $D_i$  = the geometric mean (mm),  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 12.0 |
| 2  | 7.0  |
| 3  | 9.0  |
| 4  | 10.0 |
| 5  | 9.0  |
| 6  | 10.0 |
| 7  | 7.0  |
| 8  | 6.0  |
| 9  | 5.0  |
| 10 | 6.0  |

Avg. Width (ft.) =

**McNeil Core Sample**

Date Processed: 10-Nov-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 10             | 0          | 0      |
| Fine Sand               | 0.250         | 25             | 1          | 1      |
| Medium Sand             | 0.500         | 62.5           | 1          | 2      |
| Coarse Sand             | 0.850         | 33             | 1          | 3      |
| Vry Coarse Sand         | 1.0           | 224.5          | 5          | 7      |
| Vry Fine Gravel         | 2.0           | 435            | 9          | 16     |
| Fine Gravel             | 4.75          | 185            | 4          | 20     |
| Fine Gravel             | 6.3           | 165.5          | 3          | 23     |
| Medium Gravel           | 8.0           | 618.5          | 13         | 36     |
| Coarse Gravel           | 16.0          | 1319.5         | 27         | 62     |
| Vry Coarse Gravel       | 31.5          | 1670           | 34         | 96     |
| Small Cobble            | 63.5          | 193.5          | 4          | 100    |
| <b>Total Weight (g)</b> |               | <b>4942</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 7          | 7      | 7      |
| 2-4             | 5          | 5      | 12     |
| 4-8             | 13         | 13     | 25     |
| 8-16            | 16         | 16     | 41     |
| 16-32           | 22         | 22     | 63     |
| 32-64           | 27         | 27     | 90     |
| 64-128          | 10         | 10     | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                  |
|----|----------------------------------|----|---------------------------------|----|----------------------------------|
| 1  | <input type="text" value="150"/> | 11 | <input type="text" value="14"/> | 21 | <input type="text" value="72"/>  |
| 2  | <input type="text" value="121"/> | 12 | <input type="text" value="26"/> | 22 | <input type="text" value="80"/>  |
| 3  | <input type="text" value="46"/>  | 13 | <input type="text" value="13"/> | 23 | <input type="text" value="66"/>  |
| 4  | <input type="text" value="43"/>  | 14 | <input type="text" value="35"/> | 24 | <input type="text" value="145"/> |
| 5  | <input type="text" value="35"/>  | 15 | <input type="text" value="7"/>  | 25 | <input type="text" value="203"/> |
| 6  | <input type="text" value="10"/>  | 16 | <input type="text" value="6"/>  | 26 | <input type="text" value="26"/>  |
| 7  | <input type="text" value="2"/>   | 17 | <input type="text" value="7"/>  | 27 | <input type="text" value="30"/>  |
| 8  | <input type="text" value="105"/> | 18 | <input type="text" value="14"/> | 28 | <input type="text" value="32"/>  |
| 9  | <input type="text" value="84"/>  | 19 | <input type="text" value="21"/> | 29 | <input type="text" value="4"/>   |
| 10 | <input type="text" value="300"/> | 20 | <input type="text" value="24"/> | 30 | <input type="text" value="16"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>-0.00011L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

D<sub>g</sub> = D<sub>1</sub><sup>P<sub>1</sub></sup> x ... x D<sub>i</sub><sup>P<sub>i</sub></sup>; where D<sub>g</sub> = the geometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;

P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.

\*\*\*\*\* The Fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient



**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 11.0 |
| 2                  | 12.0 |
| 3                  | 14.0 |
| 4                  | 10.0 |
| 5                  | 10.0 |
| 6                  | 9.0  |
| 7                  | 8.0  |
| 8                  | 12.0 |
| 9                  | 11.0 |
| 10                 | 10.0 |
| Avg. Width (ft.) = | 10.7 |

**McNeil Core Sample**

|                         |                      |                       |                                   |               |
|-------------------------|----------------------|-----------------------|-----------------------------------|---------------|
| Date Processed:         | 10-Nov-98            | Total Dry Weight (g): | <input type="text" value="4127"/> |               |
| <b>Particle</b>         | <b>Diameter (mm)</b> | <b>Weight (grams)</b> | <b>%Total</b>                     | <b>%Cumm.</b> |
| Vry Fine Sand           | 0.125                | 5.5                   | 0                                 | 0             |
| Fine Sand               | 0.250                | 7.5                   | 0                                 | 0             |
| Medium Sand             | 0.500                | 10                    | 0                                 | 1             |
| Coarse Sand             | 0.850                | 4                     | 0                                 | 1             |
| Vry Coarse Sand         | 1.0                  | 43                    | 1                                 | 2             |
| Vry Fine Gravel         | 2.0                  | 98.5                  | 2                                 | 4             |
| Fine Gravel             | 4.75                 | 53                    | 1                                 | 5             |
| Fine Gravel             | 6.3                  | 56                    | 1                                 | 6             |
| Medium Gravel           | 8.0                  | 375.5                 | 9                                 | 15            |
| Coarse Gravel           | 16.0                 | 1253.5                | 29                                | 44            |
| Vry Coarse Gravel       | 31.5                 | 2430                  | 56                                | 100           |
| Small Cobble            | 63.5                 | 0                     | 0                                 | 100           |
| <b>Total Weight (g)</b> |                      | <b>4336.5</b>         | <b>100</b>                        |               |

$D_g =$        % < 6.3mm   
 $F_i =$        % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 5          | 5      | 5      |
| 2-4             | 11         | 11     | 16     |
| 4-8             | 15         | 15     | 31     |
| 8-16            | 15         | 15     | 46     |
| 16-32           | 20         | 20     | 66     |
| 32-64           | 18         | 18     | 84     |
| 64-128          | 16         | 16     | 100    |
| 128-256         | 0          | 0      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                  |    |                                 |
|----|---------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="83"/> | 11 | <input type="text" value="108"/> | 21 | <input type="text" value="56"/> |
| 2  | <input type="text" value="74"/> | 12 | <input type="text" value="120"/> | 22 | <input type="text" value="91"/> |
| 3  | <input type="text" value="69"/> | 13 | <input type="text" value="80"/>  | 23 | <input type="text" value="15"/> |
| 4  | <input type="text" value="54"/> | 14 | <input type="text" value="62"/>  | 24 | <input type="text" value="74"/> |
| 5  | <input type="text" value="30"/> | 15 | <input type="text" value="75"/>  | 25 | <input type="text" value="71"/> |
| 6  | <input type="text" value="57"/> | 16 | <input type="text" value="90"/>  | 26 | <input type="text" value="64"/> |
| 7  | <input type="text" value="28"/> | 17 | <input type="text" value="30"/>  | 27 | <input type="text" value="87"/> |
| 8  | <input type="text" value="40"/> | 18 | <input type="text" value="31"/>  | 28 | <input type="text" value="21"/> |
| 9  | <input type="text" value="48"/> | 19 | <input type="text" value="70"/>  | 29 | <input type="text" value="28"/> |
| 10 | <input type="text" value="55"/> | 20 | <input type="text" value="71"/>  | 30 | <input type="text" value="23"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{-0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \sum D_i \cdot P_i \cdot x_i / \sum D_i \cdot P_i$ ; where  $D_i$  = the meometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_g / S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 9.0  |
| 2                  | 12.0 |
| 3                  | 15.0 |
| 4                  | 8.0  |
| 5                  | 7.0  |
| 6                  | 10.0 |
| 7                  | 14.0 |
| 8                  | 10.0 |
| 9                  | 8.0  |
| 10                 | 6.0  |
| Avg. Width (ft.) = | 9.9  |

**McNeil Core Sample**

Date Processed: 03-Dec-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 7.5            | 0          | 0      |
| Fine Sand               | 0.250         | 17             | 1          | 1      |
| Medium Sand             | 0.500         | 32             | 1          | 2      |
| Coarse Sand             | 0.850         | 13             | 0          | 2      |
| Vry Coarse Sand         | 1.0           | 95.5           | 3          | 6      |
| Vry Fine Gravel         | 2.0           | 240.5          | 8          | 14     |
| Fine Gravel             | 4.75          | 149            | 5          | 19     |
| Fine Gravel             | 6.3           | 207.5          | 7          | 26     |
| Medium Gravel           | 8.0           | 863.5          | 29         | 55     |
| Coarse Gravel           | 16.0          | 1197.5         | 41         | 96     |
| Vry Coarse Gravel       | 31.5          | 125            | 4          | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>2948.5</b>  | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 7          | 7      | 7      |
| 2-4             | 23         | 23     | 30     |
| 4-8             | 19         | 19     | 49     |
| 8-16            | 22         | 22     | 71     |
| 16-32           | 16         | 16     | 87     |
| 32-64           | 12         | 12     | 99     |
| 64-128          | 1          | 1      | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="15"/> | 11 | <input type="text" value="71"/> | 21 | <input type="text" value="38"/> |
| 2  | <input type="text" value="31"/> | 12 | <input type="text" value="84"/> | 22 | <input type="text" value="27"/> |
| 3  | <input type="text" value="29"/> | 13 | <input type="text" value="56"/> | 23 | <input type="text" value="20"/> |
| 4  | <input type="text" value="41"/> | 14 | <input type="text" value="50"/> | 24 | <input type="text" value="19"/> |
| 5  | <input type="text" value="50"/> | 15 | <input type="text" value="27"/> | 25 | <input type="text" value="40"/> |
| 6  | <input type="text" value="33"/> | 16 | <input type="text" value="37"/> | 26 | <input type="text" value="52"/> |
| 7  | <input type="text" value="39"/> | 17 | <input type="text" value="41"/> | 27 | <input type="text" value="62"/> |
| 8  | <input type="text" value="28"/> | 18 | <input type="text" value="50"/> | 28 | <input type="text" value="77"/> |
| 9  | <input type="text" value="60"/> | 19 | <input type="text" value="90"/> | 29 | <input type="text" value="64"/> |
| 10 | <input type="text" value="68"/> | 20 | <input type="text" value="92"/> | 30 | <input type="text" value="51"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Throw and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>-0.0001L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

D<sub>g</sub> = D<sub>1</sub><sup>P<sub>1</sub></sup> x ... x D<sub>i</sub><sup>P<sub>i</sub></sup>; where D<sub>i</sub> = the meometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;

P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.

\*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#): 22 Production Index:   
 Bankfull Area (ft<sup>2</sup>): 2620.0 Gravel Quantity (ft<sup>2</sup>): 35.2  
 Bankfull Width  
 1 13.0  
 2 15.0  
 3 18.0  
 4 12.0  
 5 10.0  
 6 16.0  
 7 15.0  
 8 10.0  
 9 12.0  
 10 10.0  
 Avg. Width (ft.) = 13.1

**McNeil Core Sample**

Date Processed: 12-Nov-98 Total Dry Weight (g):   

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 3              | 0          | 0      |
| Fine Sand               | 0.250         | 3.5            | 0          | 0      |
| Medium Sand             | 0.500         | 5.5            | 0          | 0      |
| Coarse Sand             | 0.850         | 4              | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 60             | 1          | 1      |
| Vry Fine Gravel         | 2.0           | 416            | 8          | 9      |
| Fine Gravel             | 4.75          | 294            | 5          | 15     |
| Fine Gravel             | 6.3           | 337.5          | 6          | 21     |
| Medium Gravel           | 8.0           | 1501.5         | 28         | 49     |
| Coarse Gravel           | 16.0          | 1334.5         | 25         | 73     |
| Vry Coarse Gravel       | 31.5          | 1428.5         | 27         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>5388</b>    | <b>100</b> |        |

$D_g$  =  % < 6.3mm   
 $F_i$  =  % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 7          | 7      | 7      |
| 2-4             | 11         | 11     | 18     |
| 4-8             | 12         | 12     | 30     |
| 8-16            | 18         | 18     | 48     |
| 16-32           | 17         | 17     | 64     |
| 32-64           | 10         | 10     | 74     |
| 64-128          | 15         | 15     | 89     |
| 128-256         | 11         | 11     | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                  |    |                                 |
|----|----------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="71"/>  | 11 | <input type="text" value="101"/> | 21 | <input type="text" value="55"/> |
| 2  | <input type="text" value="110"/> | 12 | <input type="text" value="40"/>  | 22 | <input type="text" value="58"/> |
| 3  | <input type="text" value="92"/>  | 13 | <input type="text" value="39"/>  | 23 | <input type="text" value="30"/> |
| 4  | <input type="text" value="64"/>  | 14 | <input type="text" value="38"/>  | 24 | <input type="text" value="44"/> |
| 5  | <input type="text" value="60"/>  | 15 | <input type="text" value="20"/>  | 25 | <input type="text" value="62"/> |
| 6  | <input type="text" value="88"/>  | 16 | <input type="text" value="36"/>  | 26 | <input type="text" value="27"/> |
| 7  | <input type="text" value="91"/>  | 17 | <input type="text" value="66"/>  | 27 | <input type="text" value="20"/> |
| 8  | <input type="text" value="47"/>  | 18 | <input type="text" value="92"/>  | 28 | <input type="text" value="18"/> |
| 9  | <input type="text" value="55"/>  | 19 | <input type="text" value="84"/>  | 29 | <input type="text" value="35"/> |
| 10 | <input type="text" value="82"/>  | 20 | <input type="text" value="60"/>  | 30 | <input type="text" value="41"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Throw and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E=82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = D_1 P_1 + \dots + D_i P_i$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 14.0 |
| 2                  | 17.0 |
| 3                  | 17.0 |
| 4                  | 12.0 |
| 5                  | 8.0  |
| 6                  | 8.0  |
| 7                  | 10.0 |
| 8                  | 14.0 |
| 9                  | 15.0 |
| 10                 | 12.0 |
| Avg. Width (ft.) = | 12.7 |

**McNeil Core Sample**

|                         |               |                       |            |        |     |
|-------------------------|---------------|-----------------------|------------|--------|-----|
| Date Processed:         | 11-Oct-98     | Total Dry Weight (g): | 4689       |        |     |
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 1                     | 0          | 0      | 0   |
| Fine Sand               | 0.250         | 2                     | 0          | 0      | 0   |
| Medium Sand             | 0.500         | 5.5                   | 0          | 0      | 0   |
| Coarse Sand             | 0.850         | 3                     | 0          | 0      | 0   |
| Vry Coarse Sand         | 1.0           | 43.5                  | 1          | 1      | 1   |
| Vry Fine Gravel         | 2.0           | 245.5                 | 5          | 6      | 6   |
| Fine Gravel             | 4.75          | 152.5                 | 3          | 9      | 9   |
| Fine Gravel             | 6.3           | 444.5                 | 9          | 18     | 18  |
| Medium Gravel           | 8.0           | 910.5                 | 18         | 37     | 37  |
| Coarse Gravel           | 16.0          | 1496                  | 30         | 67     | 67  |
| Vry Coarse Gravel       | 31.5          | 1620                  | 33         | 100    | 100 |
| Small Cobble            | 63.5          | 0                     | 0          | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>4924</b>           | <b>100</b> |        |     |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 12         | 12     | 12     |
| 2-4             | 10         | 10     | 22     |
| 4-8             | 10         | 10     | 32     |
| 8-16            | 12         | 12     | 44     |
| 16-32           | 14         | 14     | 58     |
| 32-64           | 16         | 16     | 74     |
| 64-128          | 16         | 16     | 90     |
| 128-256         | 10         | 10     | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                 |
|----|----------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="86"/>  | 11 | <input type="text" value="64"/> | 21 | <input type="text" value="61"/> |
| 2  | <input type="text" value="151"/> | 12 | <input type="text" value="79"/> | 22 | <input type="text" value="67"/> |
| 3  | <input type="text" value="73"/>  | 13 | <input type="text" value="93"/> | 23 | <input type="text" value="83"/> |
| 4  | <input type="text" value="62"/>  | 14 | <input type="text" value="57"/> | 24 | <input type="text" value="88"/> |
| 5  | <input type="text" value="77"/>  | 15 | <input type="text" value="63"/> | 25 | <input type="text" value="54"/> |
| 6  | <input type="text" value="101"/> | 16 | <input type="text" value="99"/> | 26 | <input type="text" value="61"/> |
| 7  | <input type="text" value="83"/>  | 17 | <input type="text" value="49"/> | 27 | <input type="text" value="75"/> |
| 8  | <input type="text" value="61"/>  | 18 | <input type="text" value="62"/> | 28 | <input type="text" value="80"/> |
| 9  | <input type="text" value="95"/>  | 19 | <input type="text" value="53"/> | 29 | <input type="text" value="59"/> |
| 10 | <input type="text" value="90"/>  | 20 | <input type="text" value="78"/> | 30 | <input type="text" value="70"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.00055L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 P_1 + \dots + D_n P_n$ ; where  $D_n$  = the meometric mean (mm),  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ , where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 14.0  
 2 10.0  
 3 12.0  
 4 12.0  
 5 9.0  
 6 8.0  
 7 10.0  
 8 7.0  
 9 8.0  
 10 7.0  
 Avg. Width (ft.) = 9.7

**McNeil Core Sample**

| Date Processed:         | 09-Nov-98     | Total Dry Weight (g): | 3851       |        |     |
|-------------------------|---------------|-----------------------|------------|--------|-----|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 0.5                   | 0          | 0      | 0   |
| Fine Sand               | 0.250         | 1                     | 0          | 0      | 0   |
| Medium Sand             | 0.500         | 2                     | 0          | 0      | 0   |
| Coarse Sand             | 0.850         | 0.5                   | 0          | 0      | 0   |
| Vry Coarse Sand         | 1.0           | 10.5                  | 0          | 0      | 0   |
| Vry Fine Gravel         | 2.0           | 64.5                  | 2          | 2      | 2   |
| Fine Gravel             | 4.75          | 50.5                  | 1          | 3      | 3   |
| Fine Gravel             | 6.3           | 61.5                  | 2          | 5      | 5   |
| Medium Gravel           | 8.0           | 480                   | 12         | 17     | 17  |
| Coarse Gravel           | 16.0          | 536                   | 13         | 30     | 30  |
| Vry Coarse Gravel       | 31.5          | 1591                  | 39         | 69     | 69  |
| Small Cobble            | 63.5          | 1230                  | 31         | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>4028</b>           | <b>100</b> |        |     |

$D_g =$        % < 6.3mm   
 $F_i =$        % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |     |
|-----------------|------------|--------|--------|-----|
| 0-2             | 7          | 7      | 7      | 7   |
| 2-4             | 15         | 15     | 22     | 22  |
| 4-8             | 22         | 22     | 44     | 44  |
| 8-16            | 20         | 20     | 64     | 64  |
| 16-32           | 25         | 25     | 89     | 89  |
| 32-64           | 11         | 11     | 100    | 100 |
| 64-128          |            | 0      | 100    | 100 |
| 128-256         |            | 0      | 100    | 100 |
| 256-512         |            | 0      | 100    | 100 |
| 512-1024        |            | 0      | 100    | 100 |
| 1024-2048       |            | 0      | 100    | 100 |
| 2048-4096       |            | 0      | 100    | 100 |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                  |
|----|----------------------------------|----|---------------------------------|----|----------------------------------|
| 1  | <input type="text" value="24"/>  | 11 | <input type="text" value="64"/> | 21 | <input type="text" value="7"/>   |
| 2  | <input type="text" value="32"/>  | 12 | <input type="text" value="69"/> | 22 | <input type="text" value="101"/> |
| 3  | <input type="text" value="25"/>  | 13 | <input type="text" value="52"/> | 23 | <input type="text" value="120"/> |
| 4  | <input type="text" value="21"/>  | 14 | <input type="text" value="25"/> | 24 | <input type="text" value="10"/>  |
| 5  | <input type="text" value="8"/>   | 15 | <input type="text" value="18"/> | 25 | <input type="text" value="5"/>   |
| 6  | <input type="text" value="12"/>  | 16 | <input type="text" value="74"/> | 26 | <input type="text" value="3"/>   |
| 7  | <input type="text" value="76"/>  | 17 | <input type="text" value="65"/> | 27 | <input type="text" value="24"/>  |
| 8  | <input type="text" value="84"/>  | 18 | <input type="text" value="21"/> | 28 | <input type="text" value="17"/>  |
| 9  | <input type="text" value="101"/> | 19 | <input type="text" value="18"/> | 29 | <input type="text" value="10"/>  |
| 10 | <input type="text" value="95"/>  | 20 | <input type="text" value="10"/> | 30 | <input type="text" value="12"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Throw and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.00059L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \frac{D_1 P_1 + D_2 P_2 + \dots + D_n P_n}{P_1 + P_2 + \dots + P_n}$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 10.0 |
| 2  | 10.0 |
| 3  | 12.0 |
| 4  | 9.0  |
| 5  | 8.0  |
| 6  | 10.0 |
| 7  | 7.0  |
| 8  | 7.0  |
| 9  | 7.0  |
| 10 | 6.0  |

Avg. Width (ft.) = 8.6

**McNeil Core Sample**

Date Processed:

Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 0.5            | 0          | 0      |
| Fine Sand               | 0.250         | 2              | 0          | 0      |
| Medium Sand             | 0.500         | 5              | 0          | 0      |
| Coarse Sand             | 0.850         | 3.5            | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 39             | 1          | 1      |
| Vry Fine Gravel         | 2.0           | 177            | 5          | 6      |
| Fine Gravel             | 4.75          | 89.5           | 2          | 8      |
| Fine Gravel             | 6.3           | 104            | 3          | 11     |
| Medium Gravel           | 8.0           | 469.5          | 12         | 23     |
| Coarse Gravel           | 16.0          | 812.5          | 21         | 45     |
| Vry Coarse Gravel       | 31.5          | 1444           | 38         | 83     |
| Small Cobble            | 63.5          | 644            | 17         | 100    |
| <b>Total Weight (g)</b> |               | <b>3790.5</b>  | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 8          | 8      | 8      |
| 2-4             | 10         | 10     | 18     |
| 4-8             | 18         | 18     | 36     |
| 8-16            | 20         | 20     | 56     |
| 16-32           | 26         | 26     | 82     |
| 32-64           | 14         | 14     | 96     |
| 64-128          | 4          | 4      | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                  |    |                                 |
|----|----------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="31"/>  | 11 | <input type="text" value="102"/> | 21 | <input type="text" value="62"/> |
| 2  | <input type="text" value="26"/>  | 12 | <input type="text" value="120"/> | 22 | <input type="text" value="60"/> |
| 3  | <input type="text" value="104"/> | 13 | <input type="text" value="98"/>  | 23 | <input type="text" value="96"/> |
| 4  | <input type="text" value="10"/>  | 14 | <input type="text" value="96"/>  | 24 | <input type="text" value="8"/>  |
| 5  | <input type="text" value="28"/>  | 15 | <input type="text" value="18"/>  | 25 | <input type="text" value="4"/>  |
| 6  | <input type="text" value="36"/>  | 16 | <input type="text" value="20"/>  | 26 | <input type="text" value="16"/> |
| 7  | <input type="text" value="34"/>  | 17 | <input type="text" value="26"/>  | 27 | <input type="text" value="24"/> |
| 8  | <input type="text" value="67"/>  | 18 | <input type="text" value="34"/>  | 28 | <input type="text" value="26"/> |
| 9  | <input type="text" value="70"/>  | 19 | <input type="text" value="73"/>  | 29 | <input type="text" value="68"/> |
| 10 | <input type="text" value="66"/>  | 20 | <input type="text" value="80"/>  | 30 | <input type="text" value="3"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.00032L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 P_1 + \dots + D_n P_n$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>3</sup>):   
 Bankfull Width  
 1 6.0  
 2 8.0  
 3 6.0  
 4 10.0  
 5 9.0  
 6 5.0  
 7 5.0  
 8 4.0  
 9 4.0  
 10 6.0  
 Avg. Width (ft.) = 6.3

**McNeil Core Sample**

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 1              | 0          | 0      |
| Fine Sand               | 0.250         | 2              | 0          | 0      |
| Medium Sand             | 0.500         | 4              | 0          | 0      |
| Coarse Sand             | 0.850         | 2.5            | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 20.5           | 1          | 1      |
| Vry Fine Gravel         | 2.0           | 107            | 4          | 6      |
| Fine Gravel             | 4.75          | 70             | 3          | 8      |
| Fine Gravel             | 6.3           | 87.5           | 4          | 12     |
| Medium Gravel           | 8.0           | 432            | 18         | 30     |
| Coarse Gravel           | 16.0          | 787.5          | 32         | 62     |
| Vry Coarse Gravel       | 31.5          | 936            | 38         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>2450</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 5          | 5      | 5      |
| 2-4             | 15         | 15     | 20     |
| 4-8             | 15         | 15     | 35     |
| 8-16            | 16         | 16     | 51     |
| 16-32           | 17         | 17     | 68     |
| 32-64           | 12         | 12     | 80     |
| 64-128          | 10         | 10     | 90     |
| 128-256         | 10         | 10     | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                  |
|----|----------------------------------|----|---------------------------------|----|----------------------------------|
| 1  | <input type="text" value="91"/>  | 11 | <input type="text" value="45"/> | 21 | <input type="text" value="40"/>  |
| 2  | <input type="text" value="102"/> | 12 | <input type="text" value="44"/> | 22 | <input type="text" value="51"/>  |
| 3  | <input type="text" value="80"/>  | 13 | <input type="text" value="88"/> | 23 | <input type="text" value="57"/>  |
| 4  | <input type="text" value="85"/>  | 14 | <input type="text" value="28"/> | 24 | <input type="text" value="62"/>  |
| 5  | <input type="text" value="60"/>  | 15 | <input type="text" value="76"/> | 25 | <input type="text" value="41"/>  |
| 6  | <input type="text" value="58"/>  | 16 | <input type="text" value="61"/> | 26 | <input type="text" value="140"/> |
| 7  | <input type="text" value="34"/>  | 17 | <input type="text" value="50"/> | 27 | <input type="text" value="100"/> |
| 8  | <input type="text" value="38"/>  | 18 | <input type="text" value="47"/> | 28 | <input type="text" value="38"/>  |
| 9  | <input type="text" value="20"/>  | 19 | <input type="text" value="21"/> | 29 | <input type="text" value="44"/>  |
| 10 | <input type="text" value="29"/>  | 20 | <input type="text" value="60"/> | 30 | <input type="text" value="57"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurrow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E=82.63e^{0.00022L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \frac{\sum D_i P_i}{\sum P_i}$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ , where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location :   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 11.0 |
| 2                  | 10.0 |
| 3                  | 10.0 |
| 4                  | 12.0 |
| 5                  | 11.0 |
| 6                  | 10.0 |
| 7                  | 7.0  |
| 8                  | 8.0  |
| 9                  | 8.0  |
| 10                 | 6.0  |
| Avg. Width (ft.) = | 9.3  |

**McNeil Core Sample**

Date Processed: 10-Nov-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 3.5            | 0          | 0      |
| Fine Sand               | 0.250         | 8.5            | 0          | 0      |
| Medium Sand             | 0.500         | 40             | 1          | 1      |
| Coarse Sand             | 0.850         | 27.5           | 1          | 2      |
| Vry Coarse Sand         | 1.0           | 298.5          | 8          | 11     |
| Vry Fine Gravel         | 2.0           | 839            | 23         | 34     |
| Fine Gravel             | 4.75          | 408            | 11         | 45     |
| Fine Gravel             | 6.3           | 336.5          | 9          | 55     |
| Medium Gravel           | 8.0           | 795            | 22         | 77     |
| Coarse Gravel           | 16.0          | 460            | 13         | 90     |
| Vry Coarse Gravel       | 31.5          | 367            | 10         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>3583.5</b>  | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 6          | 6      | 6      |
| 2-4             | 14         | 14     | 20     |
| 4-8             | 19         | 19     | 39     |
| 8-16            | 13         | 13     | 52     |
| 16-32           | 32         | 32     | 84     |
| 32-64           | 16         | 16     | 100    |
| 64-128          |            | 0      | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                 |
|----|----------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="19"/>  | 11 | <input type="text" value="26"/> | 21 | <input type="text" value="66"/> |
| 2  | <input type="text" value="62"/>  | 12 | <input type="text" value="38"/> | 22 | <input type="text" value="91"/> |
| 3  | <input type="text" value="11"/>  | 13 | <input type="text" value="36"/> | 23 | <input type="text" value="85"/> |
| 4  | <input type="text" value="70"/>  | 14 | <input type="text" value="34"/> | 24 | <input type="text" value="98"/> |
| 5  | <input type="text" value="90"/>  | 15 | <input type="text" value="36"/> | 25 | <input type="text" value="17"/> |
| 6  | <input type="text" value="120"/> | 16 | <input type="text" value="18"/> | 26 | <input type="text" value="10"/> |
| 7  | <input type="text" value="140"/> | 17 | <input type="text" value="96"/> | 27 | <input type="text" value="8"/>  |
| 8  | <input type="text" value="105"/> | 18 | <input type="text" value="80"/> | 28 | <input type="text" value="26"/> |
| 9  | <input type="text" value="92"/>  | 19 | <input type="text" value="70"/> | 29 | <input type="text" value="58"/> |
| 10 | <input type="text" value="17"/>  | 20 | <input type="text" value="75"/> | 30 | <input type="text" value="69"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.00033L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where  $D_i$  = the meometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient



**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|                    |      |
|--------------------|------|
| 1                  | 8.0  |
| 2                  | 10.0 |
| 3                  | 12.0 |
| 4                  | 7.0  |
| 5                  | 5.0  |
| 6                  | 7.0  |
| 7                  | 10.0 |
| 8                  | 6.0  |
| 9                  | 8.0  |
| 10                 | 5.0  |
| Avg. Width (ft.) = | 7.8  |

**McNeil Core Sample**

|                         |                      |                       |               |               |  |
|-------------------------|----------------------|-----------------------|---------------|---------------|--|
| Date Processed:         | 09-Nov-98            | Total Dry Weight (g): | 3405.5        |               |  |
| <b>Particle</b>         | <b>Diameter (mm)</b> | <b>Weight (grams)</b> | <b>%Total</b> | <b>%Cumm.</b> |  |
| Vry Fine Sand           | 0.125                | 7.5                   | 0             | 0             |  |
| Fine Sand               | 0.250                | 16                    | 0             | 1             |  |
| Medium Sand             | 0.500                | 35.5                  | 1             | 2             |  |
| Coarse Sand             | 0.850                | 15.5                  | 0             | 2             |  |
| Vry Coarse Sand         | 1.0                  | 138.5                 | 4             | 6             |  |
| Vry Fine Gravel         | 2.0                  | 266                   | 8             | 14            |  |
| Fine Gravel             | 4.75                 | 115                   | 3             | 18            |  |
| Fine Gravel             | 6.3                  | 141                   | 4             | 22            |  |
| Medium Gravel           | 8.0                  | 630                   | 19            | 40            |  |
| Coarse Gravel           | 16.0                 | 1283.5                | 38            | 78            |  |
| Vry Coarse Gravel       | 31.5                 | 733.5                 | 22            | 100           |  |
| Small Cobble            | 63.5                 | 0                     | 0             | 100           |  |
| <b>Total Weight (g)</b> |                      | <b>3382</b>           | <b>100</b>    |               |  |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 3          | 3      | 3      |
| 2-4             | 6          | 6      | 9      |
| 4-8             | 6          | 6      | 15     |
| 8-16            | 14         | 14     | 29     |
| 16-32           | 26         | 26     | 55     |
| 32-64           | 27         | 27     | 82     |
| 64-128          | 15         | 15     | 97     |
| 128-256         | 3          | 3      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |     |    |    |    |     |
|----|-----|----|----|----|-----|
| 1  | 66  | 11 | 98 | 21 | 78  |
| 2  | 15  | 12 | 91 | 22 | 85  |
| 3  | 78  | 13 | 28 | 23 | 110 |
| 4  | 21  | 14 | 7  | 24 | 65  |
| 5  | 34  | 15 | 5  | 25 | 40  |
| 6  | 36  | 16 | 10 | 26 | 16  |
| 7  | 49  | 17 | 24 | 27 | 25  |
| 8  | 76  | 18 | 29 | 28 | 16  |
| 9  | 107 | 19 | 47 | 29 | 31  |
| 10 | 124 | 20 | 56 | 30 | 28  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>-0.00011L</sup>), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 D<sub>g</sub> = D<sub>1</sub><sup>P<sub>1</sub></sup> x ... x D<sub>i</sub><sup>P<sub>i</sub></sup>; where D<sub>g</sub> = the geometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;  
 P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.  
 \*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =

Est. % Survival to Emergence:

Redd Potential (#):  Production Index:

Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):

Bankfull Width

|    |      |
|----|------|
| 1  | 14.0 |
| 2  | 10.0 |
| 3  | 8.0  |
| 4  | 6.0  |
| 5  | 25.0 |
| 6  | 30.0 |
| 7  | 15.0 |
| 8  | 9.0  |
| 9  | 10.0 |
| 10 | 8.0  |

Avg. Width (ft.) = 13.5

**McNeil Core Sample**

Date Processed: 10-Nov-98 Total Dry Weight (g):

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 1              | 0          | 0      |
| Fine Sand               | 0.250         | 1              | 0          | 0      |
| Medium Sand             | 0.500         | 1              | 0          | 0      |
| Coarse Sand             | 0.850         | 0.5            | 0          | 0      |
| Vry Coarse Sand         | 1.0           | 3.5            | 0          | 0      |
| Vry Fine Gravel         | 2.0           | 31             | 1          | 1      |
| Fine Gravel             | 4.75          | 51.5           | 1          | 2      |
| Fine Gravel             | 6.3           | 125.5          | 3          | 5      |
| Medium Gravel           | 8.0           | 977            | 25         | 30     |
| Coarse Gravel           | 16.0          | 1549           | 39         | 69     |
| Vry Coarse Gravel       | 31.5          | 1242           | 31         | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>3983</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 0          | 0      | 0      |
| 2-4             | 0          | 0      | 0      |
| 4-8             | 0          | 0      | 0      |
| 8-16            | 0          | 0      | 0      |
| 16-32           | 0          | 0      | 0      |
| 32-64           | 0          | 0      | 0      |
| 64-128          | 0          | 0      | 0      |
| 128-256         | 0          | 0      | 0      |
| 256-512         | 0          | 0      | 0      |
| 512-1024        | 0          | 0      | 0      |
| 1024-2048       | 0          | 0      | 0      |
| 2048-4096       | 0          | 0      | 0      |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                 |
|----|----------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="86"/>  | 11 | <input type="text" value="40"/> | 21 | <input type="text" value="88"/> |
| 2  | <input type="text" value="111"/> | 12 | <input type="text" value="35"/> | 22 | <input type="text" value="61"/> |
| 3  | <input type="text" value="107"/> | 13 | <input type="text" value="44"/> | 23 | <input type="text" value="39"/> |
| 4  | <input type="text" value="90"/>  | 14 | <input type="text" value="16"/> | 24 | <input type="text" value="40"/> |
| 5  | <input type="text" value="64"/>  | 15 | <input type="text" value="12"/> | 25 | <input type="text" value="49"/> |
| 6  | <input type="text" value="29"/>  | 16 | <input type="text" value="13"/> | 26 | <input type="text" value="12"/> |
| 7  | <input type="text" value="32"/>  | 17 | <input type="text" value="94"/> | 27 | <input type="text" value="7"/>  |
| 8  | <input type="text" value="36"/>  | 18 | <input type="text" value="51"/> | 28 | <input type="text" value="4"/>  |
| 9  | <input type="text" value="48"/>  | 19 | <input type="text" value="46"/> | 29 | <input type="text" value="9"/>  |
| 10 | <input type="text" value="45"/>  | 20 | <input type="text" value="49"/> | 30 | <input type="text" value="16"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurrow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E = 82.63e^{0.0001L}$ ), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;

$P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .

\*\*\*\*\* The Fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location :   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>3</sup>):   
 Bankfull Width  
 1 2.0  
 2 3.0  
 3 3.0  
 4 2.0  
 5 4.0  
 6 3.0  
 7 5.0  
 8 4.0  
 9 3.0  
 10 2.0  
 Avg. Width (ft.) = 3.1

**McNeil Core Sample**

| Particle                | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|-------------------------|---------------|----------------|------------|--------|
| Vry Fine Sand           | 0.125         | 23.5           | 1          | 1      |
| Fine Sand               | 0.250         | 112            | 5          | 6      |
| Medium Sand             | 0.500         | 350            | 15         | 21     |
| Coarse Sand             | 0.850         | 36             | 2          | 22     |
| Vry Coarse Sand         | 1.0           | 273            | 12         | 34     |
| Vry Fine Gravel         | 2.0           | 658            | 28         | 63     |
| Fine Gravel             | 4.75          | 326            | 14         | 77     |
| Fine Gravel             | 6.3           | 285            | 12         | 89     |
| Medium Gravel           | 8.0           | 259.5          | 11         | 100    |
| Coarse Gravel           | 16.0          | 0              | 0          | 100    |
| Vry Coarse Gravel       | 31.5          | 0              | 0          | 100    |
| Small Cobble            | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b> |               | <b>2323</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count                       | %Total | %Cumm. |
|-----------------|----------------------------------|--------|--------|
| 0-2             | 3                                | 3      | 3      |
| 2-4             | 12                               | 12     | 15     |
| 4-8             | 18                               | 18     | 33     |
| 8-16            | 20                               | 20     | 53     |
| 16-32           | 17                               | 17     | 70     |
| 32-64           | 30                               | 30     | 100    |
| 64-128          | 0                                | 0      | 100    |
| 128-256         | 0                                | 0      | 100    |
| 256-512         | 0                                | 0      | 100    |
| 512-1024        | 0                                | 0      | 100    |
| 1024-2048       | 0                                | 0      | 100    |
| 2048-4096       | 0                                | 0      | 100    |
| <b>Total</b>    | <input type="text" value="100"/> |        |        |

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="36"/> | 11 | <input type="text" value="16"/> | 21 | <input type="text" value="32"/> |
| 2  | <input type="text" value="32"/> | 12 | <input type="text" value="34"/> | 22 | <input type="text" value="41"/> |
| 3  | <input type="text" value="40"/> | 13 | <input type="text" value="36"/> | 23 | <input type="text" value="16"/> |
| 4  | <input type="text" value="45"/> | 14 | <input type="text" value="31"/> | 24 | <input type="text" value="34"/> |
| 5  | <input type="text" value="31"/> | 15 | <input type="text" value="41"/> | 25 | <input type="text" value="33"/> |
| 6  | <input type="text" value="10"/> | 16 | <input type="text" value="30"/> | 26 | <input type="text" value="24"/> |
| 7  | <input type="text" value="16"/> | 17 | <input type="text" value="56"/> | 27 | <input type="text" value="31"/> |
| 8  | <input type="text" value="18"/> | 18 | <input type="text" value="32"/> | 28 | <input type="text" value="24"/> |
| 9  | <input type="text" value="16"/> | 19 | <input type="text" value="41"/> | 29 | <input type="text" value="36"/> |
| 10 | <input type="text" value="32"/> | 20 | <input type="text" value="48"/> | 30 | <input type="text" value="42"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout ( $E=82.63e^{0.0002L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \frac{\sum D_i^4 P_i}{\sum D_i^3 P_i}$ ; where  $D_g$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ , where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 8.0  
 2 12.0  
 3 6.0  
 4 7.0  
 5 5.0  
 6 8.0  
 7 10.0  
 8 9.0  
 9 5.0  
 10 12.0  
 Avg. Width (ft.) = 8.2

**McNeil Core Sample**

| Date Processed:         | 10-Nov-98     | Total Dry Weight (g): | 938.5      |        |     |
|-------------------------|---------------|-----------------------|------------|--------|-----|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 14.5                  | 2          | 2      | 2   |
| Fine Sand               | 0.250         | 62                    | 7          | 8      | 8   |
| Medium Sand             | 0.500         | 91                    | 10         | 18     | 18  |
| Coarse Sand             | 0.850         | 21.5                  | 2          | 20     | 20  |
| Vry Coarse Sand         | 1.0           | 172.5                 | 18         | 39     | 39  |
| Vry Fine Gravel         | 2.0           | 472                   | 50         | 89     | 89  |
| Fine Gravel             | 4.75          | 69.5                  | 7          | 96     | 96  |
| Fine Gravel             | 6.3           | 25.5                  | 3          | 99     | 99  |
| Medium Gravel           | 8.0           | 10                    | 1          | 100    | 100 |
| Coarse Gravel           | 16.0          | 0                     | 0          | 100    | 100 |
| Vry Coarse Gravel       | 31.5          | 0                     | 0          | 100    | 100 |
| Small Cobble            | 63.5          | 0                     | 0          | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>938.5</b>          | <b>100</b> |        |     |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 8          | 8      | 8      |
| 2-4             | 21         | 21     | 29     |
| 4-8             | 31         | 31     | 60     |
| 8-16            | 30         | 30     | 90     |
| 16-32           | 10         | 10     | 100    |
| 32-64           | 0          | 0      | 100    |
| 64-128          | 0          | 0      | 100    |
| 128-256         | 0          | 0      | 100    |
| 256-512         | 0          | 0      | 100    |
| 512-1024        | 0          | 0      | 100    |
| 1024-2048       | 0          | 0      | 100    |
| 2048-4096       | 0          | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="3"/>  | 11 | <input type="text" value="14"/> | 21 | <input type="text" value="8"/>  |
| 2  | <input type="text" value="4"/>  | 12 | <input type="text" value="8"/>  | 22 | <input type="text" value="6"/>  |
| 3  | <input type="text" value="7"/>  | 13 | <input type="text" value="10"/> | 23 | <input type="text" value="13"/> |
| 4  | <input type="text" value="12"/> | 14 | <input type="text" value="24"/> | 24 | <input type="text" value="20"/> |
| 5  | <input type="text" value="4"/>  | 15 | <input type="text" value="25"/> | 25 | <input type="text" value="27"/> |
| 6  | <input type="text" value="16"/> | 16 | <input type="text" value="17"/> | 26 | <input type="text" value="31"/> |
| 7  | <input type="text" value="24"/> | 17 | <input type="text" value="11"/> | 27 | <input type="text" value="20"/> |
| 8  | <input type="text" value="32"/> | 18 | <input type="text" value="9"/>  | 28 | <input type="text" value="14"/> |
| 9  | <input type="text" value="18"/> | 19 | <input type="text" value="5"/>  | 29 | <input type="text" value="5"/>  |
| 10 | <input type="text" value="12"/> | 20 | <input type="text" value="3"/>  | 30 | <input type="text" value="4"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurrow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>0.0005L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

$$D_g = \sum D_i \cdot P_i^2 / \sum P_i^2$$

where D<sub>g</sub> = the meometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;

$$P_i = \text{the proportion by weight of the entire sample made up of material retained on sieve } i.$$

\*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>, where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

| Spawning Gravel Analysis          |                                     |                                     |                                   |        |                                 |
|-----------------------------------|-------------------------------------|-------------------------------------|-----------------------------------|--------|---------------------------------|
| % Usable Spawning Area =          |                                     | <input type="text" value="6.2"/>    |                                   |        |                                 |
| Est. % Survival to Emergence:     |                                     | <input type="text" value="26.4"/>   |                                   |        |                                 |
| Redd Potential (#):               | <input type="text" value="52"/>     | Production Index:                   | <input type="text" value="2799"/> |        |                                 |
| Bankfull Area (ft <sup>2</sup> ): | <input type="text" value="1340.0"/> | Gravel Quantity (ft <sup>2</sup> ): | <input type="text" value="83.2"/> |        |                                 |
| Bankfull Width                    |                                     |                                     |                                   |        |                                 |
|                                   | 1                                   | 12.0                                |                                   |        |                                 |
|                                   | 2                                   | 7.0                                 |                                   |        |                                 |
|                                   | 3                                   | 3.0                                 |                                   |        |                                 |
|                                   | 4                                   | 8.0                                 |                                   |        |                                 |
|                                   | 5                                   | 8.0                                 |                                   |        |                                 |
|                                   | 6                                   | 5.0                                 |                                   |        |                                 |
|                                   | 7                                   | 6.0                                 |                                   |        |                                 |
|                                   | 8                                   | 8.0                                 |                                   |        |                                 |
|                                   | 9                                   | 7.0                                 |                                   |        |                                 |
|                                   | 10                                  | 3.0                                 |                                   |        |                                 |
| Avg. Width (ft.) =                | 6.7                                 |                                     |                                   |        |                                 |
| McNeil Core Sample                |                                     |                                     |                                   |        |                                 |
| Date Processed:                   | 09-Nov-98                           | Total Dry Weight (g):               | <input type="text" value="2654"/> |        |                                 |
| Particle                          | Diameter (mm)                       | Weight (grams)                      | %Total                            | %Cumm. |                                 |
| Vry Fine Sand                     | 0.125                               | 8.5                                 | 0                                 | 0      |                                 |
| Fine Sand                         | 0.250                               | 42                                  | 2                                 | 2      |                                 |
| Medium Sand                       | 0.500                               | 118.5                               | 4                                 | 6      |                                 |
| Coarse Sand                       | 0.850                               | 36                                  | 1                                 | 8      |                                 |
| Vry Coarse Sand                   | 1.0                                 | 155.5                               | 6                                 | 13     |                                 |
| Vry Fine Gravel                   | 2.0                                 | 224.5                               | 8                                 | 22     |                                 |
| Fine Gravel                       | 4.75                                | 104                                 | 4                                 | 26     |                                 |
| Fine Gravel                       | 6.3                                 | 121.5                               | 5                                 | 30     |                                 |
| Medium Gravel                     | 8.0                                 | 494.5                               | 19                                | 49     |                                 |
| Coarse Gravel                     | 16.0                                | 1088.5                              | 41                                | 90     |                                 |
| Vry Coarse Gravel                 | 31.5                                | 278.5                               | 10                                | 100    |                                 |
| Small Cobble                      | 63.5                                | 0                                   | 0                                 | 100    |                                 |
| <b>Total Weight (g)</b>           |                                     | <b>2672</b>                         | <b>100</b>                        |        |                                 |
| $D_g =$                           | <input type="text" value="7.5"/>    | % < 6.3mm                           | <input type="text" value="30"/>   |        |                                 |
| $F_i =$                           | <input type="text" value="4.1"/>    | % < 0.88mm                          | <input type="text" value="8"/>    |        |                                 |
| Wolman Pebble Count               |                                     |                                     |                                   |        |                                 |
| Size Range (mm)                   | Item Count                          | %Total                              | %Cumm.                            |        |                                 |
| 0-2                               | 12                                  | 12                                  | 12                                |        |                                 |
| 2-4                               | 15                                  | 15                                  | 27                                |        |                                 |
| 4-8                               | 16                                  | 16                                  | 43                                |        |                                 |
| 8-16                              | 22                                  | 22                                  | 65                                |        |                                 |
| 16-32                             | 30                                  | 30                                  | 95                                |        |                                 |
| 32-64                             | 5                                   | 5                                   | 100                               |        |                                 |
| 64-128                            |                                     | 0                                   | 100                               |        |                                 |
| 128-256                           |                                     | 0                                   | 100                               |        |                                 |
| 256-512                           |                                     | 0                                   | 100                               |        |                                 |
| 512-1024                          |                                     | 0                                   | 100                               |        |                                 |
| 1024-2048                         |                                     | 0                                   | 100                               |        |                                 |
| 2048-4096                         |                                     | 0                                   | 100                               |        |                                 |
| Total                             | 100                                 |                                     |                                   |        |                                 |
| 30 Count Results                  |                                     |                                     |                                   |        |                                 |
| 1                                 | <input type="text" value="56"/>     | 11                                  | <input type="text" value="34"/>   | 21     | <input type="text" value="7"/>  |
| 2                                 | <input type="text" value="61"/>     | 12                                  | <input type="text" value="39"/>   | 22     | <input type="text" value="8"/>  |
| 3                                 | <input type="text" value="45"/>     | 13                                  | <input type="text" value="46"/>   | 23     | <input type="text" value="25"/> |
| 4                                 | <input type="text" value="44"/>     | 14                                  | <input type="text" value="48"/>   | 24     | <input type="text" value="28"/> |
| 5                                 | <input type="text" value="12"/>     | 15                                  | <input type="text" value="49"/>   | 25     | <input type="text" value="46"/> |
| 6                                 | <input type="text" value="14"/>     | 16                                  | <input type="text" value="66"/>   | 26     | <input type="text" value="66"/> |
| 7                                 | <input type="text" value="67"/>     | 17                                  | <input type="text" value="91"/>   | 27     | <input type="text" value="72"/> |
| 8                                 | <input type="text" value="58"/>     | 18                                  | <input type="text" value="18"/>   | 28     | <input type="text" value="59"/> |
| 9                                 | <input type="text" value="26"/>     | 19                                  | <input type="text" value="10"/>   | 29     | <input type="text" value="60"/> |
| 10                                | <input type="text" value="28"/>     | 20                                  | <input type="text" value="8"/>    | 30     | <input type="text" value="41"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope outthroat trout (E=82.63e<sup>0.00055L</sup>), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \frac{\sum D_i \cdot P_i}{\sum P_i}$ ; where D<sub>i</sub> = the meometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve i.  
 \*\*\*\*\* The fredie index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 13.0  
 2 9.0  
 3 8.0  
 4 15.0  
 5 6.0  
 6 5.0  
 7 6.0  
 8 5.0  
 9 7.0  
 10 5.0  
 Avg. Width (ft.) = 8

**McNeil Core Sample**

| Date Processed:         | 09-Nov-98     | Total Dry Weight (g): | 2900.5     |        |     |
|-------------------------|---------------|-----------------------|------------|--------|-----|
| Particle                | Diameter (mm) | Weight (grams)        | %Total     | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 15                    | 1          | 1      | 1   |
| Fine Sand               | 0.250         | 39.5                  | 1          | 2      | 2   |
| Medium Sand             | 0.500         | 77                    | 3          | 5      | 5   |
| Coarse Sand             | 0.850         | 19.5                  | 1          | 5      | 5   |
| Vry Coarse Sand         | 1.0           | 110.5                 | 4          | 9      | 9   |
| Vry Fine Gravel         | 2.0           | 188                   | 6          | 15     | 15  |
| Fine Gravel             | 4.75          | 96.5                  | 3          | 19     | 19  |
| Fine Gravel             | 6.3           | 102.5                 | 4          | 22     | 22  |
| Medium Gravel           | 8.0           | 418                   | 14         | 37     | 37  |
| Coarse Gravel           | 16.0          | 1000                  | 34         | 71     | 71  |
| Vry Coarse Gravel       | 31.5          | 847                   | 29         | 100    | 100 |
| Small Cobble            | 63.5          | 0                     | 0          | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>2913.5</b>         | <b>100</b> |        |     |

$D_g =$   % < 6.3mm

$F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 2          | 2      | 2      |
| 2-4             | 6          | 6      | 8      |
| 4-8             | 5          | 5      | 13     |
| 8-16            | 22         | 22     | 35     |
| 16-32           | 25         | 25     | 60     |
| 32-64           | 31         | 31     | 91     |
| 64-128          | 9          | 9      | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                  |    |                                 |
|----|---------------------------------|----|----------------------------------|----|---------------------------------|
| 1  | <input type="text" value="45"/> | 11 | <input type="text" value="109"/> | 21 | <input type="text" value="18"/> |
| 2  | <input type="text" value="40"/> | 12 | <input type="text" value="12"/>  | 22 | <input type="text" value="12"/> |
| 3  | <input type="text" value="61"/> | 13 | <input type="text" value="72"/>  | 23 | <input type="text" value="94"/> |
| 4  | <input type="text" value="58"/> | 14 | <input type="text" value="60"/>  | 24 | <input type="text" value="80"/> |
| 5  | <input type="text" value="50"/> | 15 | <input type="text" value="62"/>  | 25 | <input type="text" value="62"/> |
| 6  | <input type="text" value="26"/> | 16 | <input type="text" value="58"/>  | 26 | <input type="text" value="64"/> |
| 7  | <input type="text" value="12"/> | 17 | <input type="text" value="44"/>  | 27 | <input type="text" value="26"/> |
| 8  | <input type="text" value="19"/> | 18 | <input type="text" value="50"/>  | 28 | <input type="text" value="33"/> |
| 9  | <input type="text" value="76"/> | 19 | <input type="text" value="26"/>  | 29 | <input type="text" value="19"/> |
| 10 | <input type="text" value="70"/> | 20 | <input type="text" value="34"/>  | 30 | <input type="text" value="16"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.

\*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)

\*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope cutthroat trout (E=82.63e<sup>1.00755L</sup>), where L is fork length in millimeters (Downs and Shepard).

\*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):

D<sub>g</sub> = D<sub>1</sub><sup>P<sub>1</sub></sup> x ... x D<sub>i</sub><sup>P<sub>i</sub></sup>; where D<sub>g</sub> = the geometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;

P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.

\*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula: F<sub>i</sub> = D<sub>i</sub>/S<sub>i</sub>; where S<sub>i</sub> = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

| Spawning Gravel Analysis          |                                     |                                     |                                     |        |                                  |
|-----------------------------------|-------------------------------------|-------------------------------------|-------------------------------------|--------|----------------------------------|
| % Usable Spawning Area =          | <input type="text" value="2.0"/>    |                                     |                                     |        |                                  |
| Est. % Survival to Emergence:     | <input type="text" value="28.5"/>   |                                     |                                     |        |                                  |
| Redd Potential (#):               | <input type="text" value="25"/>     | Production Index:                   | <input type="text" value="1454"/>   |        |                                  |
| Bankfull Area (ft <sup>2</sup> ): | <input type="text" value="2020.0"/> | Gravel Quantity (ft <sup>2</sup> ): | <input type="text" value="40.0"/>   |        |                                  |
|                                   | Bankfull Width                      |                                     |                                     |        |                                  |
|                                   | 1                                   | 12.0                                |                                     |        |                                  |
|                                   | 2                                   | 9.0                                 |                                     |        |                                  |
|                                   | 3                                   | 10.0                                |                                     |        |                                  |
|                                   | 4                                   | 7.0                                 |                                     |        |                                  |
|                                   | 5                                   | 8.0                                 |                                     |        |                                  |
|                                   | 6                                   | 11.0                                |                                     |        |                                  |
|                                   | 7                                   | 10.0                                |                                     |        |                                  |
|                                   | 8                                   | 12.0                                |                                     |        |                                  |
|                                   | 9                                   | 10.0                                |                                     |        |                                  |
|                                   | 10                                  | <u>12.0</u>                         |                                     |        |                                  |
| Avg. Width (ft.) =                | 10.1                                |                                     |                                     |        |                                  |
| McNeil Core Sample                |                                     |                                     |                                     |        |                                  |
| Date Processed:                   | 02-Sep-98                           | Total Dry Weight (g):               | <input type="text" value="2567.5"/> |        |                                  |
| Particle                          | Diameter (mm)                       | Weight (grams)                      | %Total                              | %Cumm. |                                  |
| Vry Fine Sand                     | 0.125                               | 6                                   | 0                                   | 0      |                                  |
| Fine Sand                         | 0.250                               | 23                                  | 1                                   | 1      |                                  |
| Medium Sand                       | 0.500                               | 65                                  | 3                                   | 4      |                                  |
| Coarse Sand                       | 0.850                               | 17                                  | 1                                   | 4      |                                  |
| Vry Coarse Sand                   | 1.0                                 | 90                                  | 4                                   | 8      |                                  |
| Vry Fine Gravel                   | 2.0                                 | 183                                 | 7                                   | 15     |                                  |
| Fine Gravel                       | 4.75                                | 142                                 | 6                                   | 20     |                                  |
| Fine Gravel                       | 6.3                                 | 132                                 | 5                                   | 26     |                                  |
| Medium Gravel                     | 8.0                                 | 411                                 | 16                                  | 42     |                                  |
| Coarse Gravel                     | 16.0                                | 1036                                | 40                                  | 82     |                                  |
| Vry Coarse Gravel                 | 31.5                                | 464                                 | 18                                  | 100    |                                  |
| Small Cobble                      | 63.5                                | 0                                   | 0                                   | 100    |                                  |
| <b>Total Weight (g)</b>           |                                     | <b>2567</b>                         | <b>100</b>                          |        |                                  |
| $D_g =$                           | <input type="text" value="9.7"/>    | % < 6.3mm                           | <input type="text" value="26"/>     |        |                                  |
| $F_i =$                           | <input type="text" value="6.1"/>    | % < 0.88mm                          | <input type="text" value="4"/>      |        |                                  |
| Wolman Pebble Count               |                                     |                                     |                                     |        |                                  |
| Size Range (mm)                   | Item Count                          | %Total                              | %Cumm.                              |        |                                  |
| 0-2                               | 5                                   | 5                                   | 5                                   |        |                                  |
| 2-4                               | 8                                   | 8                                   | 13                                  |        |                                  |
| 4-8                               | 15                                  | 15                                  | 28                                  |        |                                  |
| 8-16                              | 16                                  | 16                                  | 44                                  |        |                                  |
| 16-32                             | 34                                  | 34                                  | 78                                  |        |                                  |
| 32-64                             | 20                                  | 20                                  | 98                                  |        |                                  |
| 64-128                            | 2                                   | 2                                   | 100                                 |        |                                  |
| 128-256                           | 0                                   | 0                                   | 100                                 |        |                                  |
| 256-512                           | 0                                   | 0                                   | 100                                 |        |                                  |
| 512-1024                          | 0                                   | 0                                   | 100                                 |        |                                  |
| 1024-2048                         | 0                                   | 0                                   | 100                                 |        |                                  |
| 2048-4096                         | 0                                   | 0                                   | 100                                 |        |                                  |
| Total                             | 100                                 | 100                                 |                                     |        |                                  |
| 30 Count Results                  |                                     |                                     |                                     |        |                                  |
| 1                                 | <input type="text" value="98"/>     | 11                                  | <input type="text" value="24"/>     | 21     | <input type="text" value="66"/>  |
| 2                                 | <input type="text" value="64"/>     | 12                                  | <input type="text" value="12"/>     | 22     | <input type="text" value="61"/>  |
| 3                                 | <input type="text" value="36"/>     | 13                                  | <input type="text" value="7"/>      | 23     | <input type="text" value="58"/>  |
| 4                                 | <input type="text" value="45"/>     | 14                                  | <input type="text" value="8"/>      | 24     | <input type="text" value="79"/>  |
| 5                                 | <input type="text" value="109"/>    | 15                                  | <input type="text" value="14"/>     | 25     | <input type="text" value="124"/> |
| 6                                 | <input type="text" value="70"/>     | 16                                  | <input type="text" value="18"/>     | 26     | <input type="text" value="110"/> |
| 7                                 | <input type="text" value="66"/>     | 17                                  | <input type="text" value="26"/>     | 27     | <input type="text" value="98"/>  |
| 8                                 | <input type="text" value="75"/>     | 18                                  | <input type="text" value="31"/>     | 28     | <input type="text" value="65"/>  |
| 9                                 | <input type="text" value="40"/>     | 19                                  | <input type="text" value="28"/>     | 29     | <input type="text" value="41"/>  |
| 10                                | <input type="text" value="45"/>     | 20                                  | <input type="text" value="70"/>     | 30     | <input type="text" value="36"/>  |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope outthroat trout ( $E = 82.63e^{0.0025L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = D_1 \cdot P_1 \cdot \dots \cdot D_i \cdot P_i$ ; where  $D_i$  = the meometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ , where  $S_i$  = a sorting coefficient

**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 4.0  
 2 6.0  
 3 4.0  
 4 8.0  
 5 5.0  
 6 3.0  
 7 6.0  
 8 6.0  
 9 8.0  
 10 7.0  
 Avg. Width (ft.) = 5.7

**McNeil Core Sample**

| Particle   | Diameter (mm) | Weight (grams) | %Total     | %Cumm. |
|--|---------------|----------------|------------|--------|
| Date Processed: 09-Nov-98 Total Dry Weight (g): 2129.5 |               |                |            |        |
| Vry Fine Sand  | 0.125         | 4              | 0          | 0      |
| Fine Sand  | 0.250         | 16.5           | 1          | 1      |
| Medium Sand  | 0.500         | 34             | 2          | 3      |
| Coarse Sand  | 0.850         | 11.5           | 1          | 3      |
| Vry Coarse Sand  | 1.0           | 58             | 3          | 6      |
| Vry Fine Gravel  | 2.0           | 128.5          | 6          | 13     |
| Fine Gravel  | 4.75          | 76             | 4          | 16     |
| Fine Gravel  | 6.3           | 78.5           | 4          | 20     |
| Medium Gravel  | 8.0           | 312            | 15         | 36     |
| Coarse Gravel  | 16.0          | 789.5          | 39         | 75     |
| Vry Coarse Gravel                                      | 31.5          | 506.5          | 25         | 100    |
| Small Cobble   | 63.5          | 0              | 0          | 100    |
| <b>Total Weight (g)</b>                                |               | <b>2015</b>    | <b>100</b> |        |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 9          | 10     | 10     |
| 2-4             | 7          | 8      | 18     |
| 4-8             | 15         | 17     | 34     |
| 8-16            | 16         | 18     | 52     |
| 16-32           | 18         | 20     | 72     |
| 32-64           | 15         | 17     | 89     |
| 64-128          | 10         | 11     | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                 |    |                                 |    |                                 |
|----|---------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="40"/> | 11 | <input type="text" value="66"/> | 21 | <input type="text" value="28"/> |
| 2  | <input type="text" value="37"/> | 12 | <input type="text" value="78"/> | 22 | <input type="text" value="40"/> |
| 3  | <input type="text" value="29"/> | 13 | <input type="text" value="85"/> | 23 | <input type="text" value="62"/> |
| 4  | <input type="text" value="31"/> | 14 | <input type="text" value="64"/> | 24 | <input type="text" value="80"/> |
| 5  | <input type="text" value="49"/> | 15 | <input type="text" value="51"/> | 25 | <input type="text" value="65"/> |
| 6  | <input type="text" value="55"/> | 16 | <input type="text" value="40"/> | 26 | <input type="text" value="35"/> |
| 7  | <input type="text" value="52"/> | 17 | <input type="text" value="42"/> | 27 | <input type="text" value="41"/> |
| 8  | <input type="text" value="71"/> | 18 | <input type="text" value="31"/> | 28 | <input type="text" value="40"/> |
| 9  | <input type="text" value="79"/> | 19 | <input type="text" value="20"/> | 29 | <input type="text" value="24"/> |
| 10 | <input type="text" value="64"/> | 20 | <input type="text" value="22"/> | 30 | <input type="text" value="24"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope outthroat trout ( $E = 82.63e^{0.00025L}$ ), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter ( $D_g$ ) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = D_1 \cdot P_1 + \dots + D_n \cdot P_n$ ; where  $D_i$  = the geometric mean (mm);  $D_i$  = the mean diameter (mm) of material retained on sieve  $i$ ;  
 $P_i$  = the proportion by weight of the entire sample made up of material retained on sieve  $i$ .  
 \*\*\*\*\* The fredle index ( $F_i$ ) was calculated using the formula:  $F_i = D_i/S_i$ ; where  $S_i$  = a sorting coefficient



**Spawning Gravel Survey, 1998 - Results and Analysis**

Date Collected:  Location:   
 Initials:  Reach:  Site Number:

**Spawning Gravel Analysis**

% Usable Spawning Area =   
 Est. % Survival to Emergence:   
 Redd Potential (#):  Production Index:   
 Bankfull Area (ft<sup>2</sup>):  Gravel Quantity (ft<sup>2</sup>):   
 Bankfull Width  
 1 10.0  
 2 8.0  
 3 6.0  
 4 8.0  
 5 15.0  
 6 24.0  
 7 12.0  
 8 10.0  
 9 8.0  
 10 9.0  
 Avg. Width (ft.) = 11

**McNeil Core Sample**

| Date Processed:         | 10-Nov-98     | Total Dry Weight (g): | <input type="text" value="2811.5"/> |        |     |
|-------------------------|---------------|-----------------------|-------------------------------------|--------|-----|
| Particle                | Diameter (mm) | Weight (grams)        | %Total                              | %Cumm. |     |
| Vry Fine Sand           | 0.125         | 13                    | 0                                   | 0      | 0   |
| Fine Sand               | 0.250         | 59.5                  | 2                                   | 3      | 3   |
| Medium Sand             | 0.500         | 151.5                 | 5                                   | 8      | 8   |
| Coarse Sand             | 0.850         | 36.5                  | 1                                   | 9      | 9   |
| Vry Coarse Sand         | 1.0           | 122.5                 | 4                                   | 14     | 14  |
| Vry Fine Gravel         | 2.0           | 109                   | 4                                   | 17     | 17  |
| Fine Gravel             | 4.75          | 57                    | 2                                   | 19     | 19  |
| Fine Gravel             | 6.3           | 62.5                  | 2                                   | 22     | 22  |
| Medium Gravel           | 8.0           | 380.5                 | 14                                  | 35     | 35  |
| Coarse Gravel           | 16.0          | 1253.5                | 44                                  | 80     | 80  |
| Vry Coarse Gravel       | 31.5          | 572                   | 20                                  | 100    | 100 |
| Small Cobble            | 63.5          | 0                     | 0                                   | 100    | 100 |
| <b>Total Weight (g)</b> |               | <b>2817.5</b>         | <b>100</b>                          |        |     |

$D_g =$   % < 6.3mm   
 $F_i =$   % < 0.88mm

**Wolman Pebble Count**

| Size Range (mm) | Item Count | %Total | %Cumm. |
|-----------------|------------|--------|--------|
| 0-2             | 1          | 1      | 1      |
| 2-4             | 5          | 5      | 6      |
| 4-8             | 11         | 11     | 17     |
| 8-16            | 10         | 10     | 27     |
| 16-32           | 30         | 30     | 57     |
| 32-64           | 22         | 22     | 79     |
| 64-128          | 21         | 21     | 100    |
| 128-256         |            | 0      | 100    |
| 256-512         |            | 0      | 100    |
| 512-1024        |            | 0      | 100    |
| 1024-2048       |            | 0      | 100    |
| 2048-4096       |            | 0      | 100    |

Total

**30 Count Results**

|    |                                  |    |                                 |    |                                 |
|----|----------------------------------|----|---------------------------------|----|---------------------------------|
| 1  | <input type="text" value="46"/>  | 11 | <input type="text" value="40"/> | 21 | <input type="text" value="58"/> |
| 2  | <input type="text" value="121"/> | 12 | <input type="text" value="5"/>  | 22 | <input type="text" value="50"/> |
| 3  | <input type="text" value="17"/>  | 13 | <input type="text" value="4"/>  | 23 | <input type="text" value="36"/> |
| 4  | <input type="text" value="24"/>  | 14 | <input type="text" value="29"/> | 24 | <input type="text" value="47"/> |
| 5  | <input type="text" value="26"/>  | 15 | <input type="text" value="36"/> | 25 | <input type="text" value="48"/> |
| 6  | <input type="text" value="24"/>  | 16 | <input type="text" value="31"/> | 26 | <input type="text" value="65"/> |
| 7  | <input type="text" value="10"/>  | 17 | <input type="text" value="29"/> | 27 | <input type="text" value="34"/> |
| 8  | <input type="text" value="18"/>  | 18 | <input type="text" value="74"/> | 28 | <input type="text" value="39"/> |
| 9  | <input type="text" value="16"/>  | 19 | <input type="text" value="81"/> | 29 | <input type="text" value="6"/>  |
| 10 | <input type="text" value="96"/>  | 20 | <input type="text" value="86"/> | 30 | <input type="text" value="12"/> |

\* Estimated % survival to emergence was calculated using the equation published by Weaver and Fraley, 1993.  
 \*\* Redd potential was determined by dividing area of suitable spawning gravel by average redd size (Thurow and King, 1994)  
 \*\*\* Production Index is an estimate of fry production calculated by multiplying redd potential by the estimated % survival to emergence and the calculated egg deposition (E) for the average sized female (310mm), using a length-fecundity relation for westslope outthroat trout (E=82.63e<sup>0.00259L</sup>), where L is fork length in millimeters (Downs and Shepard).  
 \*\*\*\* Geometric mean diameter (D<sub>g</sub>) was calculated by a method of moments using the formula of Lotspeich and Everest (1981):  
 $D_g = \sum D_i P_i / \sum P_i$ ; where D<sub>g</sub> = the geometric mean (mm); D<sub>i</sub> = the mean diameter (mm) of material retained on sieve i;  
 P<sub>i</sub> = the proportion by weight of the entire sample made up of material retained on sieve i.  
 \*\*\*\*\* The fredle index (F<sub>i</sub>) was calculated using the formula:  $F_i = D_i / S_i$ ; where S<sub>i</sub> = a sorting coefficient

## **Appendix G**

Cohort analysis of growth for cutthroat trout and brook trout in  
select tributaries of the Coeur d'Alene Reservation, 1998.

**Mean back-calculated lengths at annulus formation by age class and cohort - 1998.**

| SPECIES            |            | Age Class  |           |            |            |            |            |            |            |            |     |
|--------------------|------------|------------|-----------|------------|------------|------------|------------|------------|------------|------------|-----|
| Cutthroat Trout    |            |            |           |            |            |            |            |            |            |            |     |
| LOCATION           | Cohort     | Number     | 1         | 2          | 3          | 4          | 5          | 6          | 7          | 8          |     |
| ALDER              | 1996       | 3          | 77        | 112        |            |            |            |            |            |            |     |
|                    | 1995       | 8          | 82        | 117        | 154        |            |            |            |            |            |     |
|                    | 1994       | 5          | 86        | 127        | 165        | 203        |            |            |            |            |     |
|                    | 1993       | 1          | 82        | 111        | 147        | 183        | 220        |            |            |            |     |
| ALDER Total        |            | 17         | 82        | 119        | 158        | 200        | 220        |            |            |            |     |
| BENEWAH            | 1996       | 17         | 45        | 104        |            |            |            |            |            |            |     |
|                    | 1995       | 55         | 43        | 97         | 157        |            |            |            |            |            |     |
|                    | 1994       | 24         | 41        | 95         | 151        | 208        |            |            |            |            |     |
|                    | 1993       | 18         | 45        | 103        | 166        | 238        | 297        |            |            |            |     |
|                    | 1992       | 24         | 40        | 98         | 163        | 224        | 281        | 340        |            |            |     |
|                    | 1991       | 3          | 32        | 76         | 124        | 182        | 237        | 305        | 367        |            |     |
| BENEWAH Total      |            | 141        | 43        | 98         | 157        | 220        | 285        | 336        | 367        |            |     |
| EVANS              | 1996       | 1          | 38        | 92         |            |            |            |            |            |            |     |
|                    | 1995       | 27         | 61        | 108        | 155        |            |            |            |            |            |     |
|                    | 1994       | 15         | 64        | 112        | 161        | 209        |            |            |            |            |     |
|                    | 1993       | 2          | 67        | 116        | 159        | 209        | 258        |            |            |            |     |
| EVANS Total        |            | 45         | 62        | 109        | 157        | 209        | 258        |            |            |            |     |
| LAKE               | 1997       | 1          | 43        |            |            |            |            |            |            |            |     |
|                    | 1996       | 11         | 58        | 105        |            |            |            |            |            |            |     |
|                    | 1995       | 98         | 61        | 110        | 156        |            |            |            |            |            |     |
|                    | 1994       | 32         | 60        | 107        | 156        | 203        |            |            |            |            |     |
|                    | 1993       | 14         | 61        | 116        | 162        | 212        | 261        |            |            |            |     |
|                    | 1992       | 20         | 65        | 119        | 165        | 216        | 267        | 325        |            |            |     |
|                    | 1991       | 16         | 61        | 104        | 158        | 206        | 256        | 304        | 350        |            |     |
|                    | 1990       | 1          | 61        | 108        | 155        | 202        | 258        | 305        | 352        | 390        |     |
|                    | LAKE Total |            | 193       | 61         | 110        | 158        | 208        | 262        | 315        | 350        | 390 |
| <b>Grand Total</b> |            | <b>396</b> | <b>56</b> | <b>106</b> | <b>157</b> | <b>213</b> | <b>272</b> | <b>324</b> | <b>352</b> | <b>390</b> |     |

**Standard deviation of mean back-calculated lengths at annulus formation - 1998.**

| SPECIES            |            | Age Class  |           |           |           |           |           |           |           |             |      |
|--------------------|------------|------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-------------|------|
| Cutthroat Trout    |            |            |           |           |           |           |           |           |           |             |      |
| LOCATION           | Cohort     | Number     | 1         | 2         | 3         | 4         | 5         | 6         | 7         | 8           |      |
| ALDER              | 1996       | 3          | 2         | 4         |           |           |           |           |           |             |      |
|                    | 1995       | 8          | 8         | 6         | 10        |           |           |           |           |             |      |
|                    | 1994       | 5          | 4         | 11        | 8         | 11        |           |           |           |             |      |
|                    | 1993       | 1          | ####      | ####      | ####      | ####      | ####      |           |           |             |      |
| ALDER Total        |            | 17         | 7         | 9         | 11        | 13        | ####      |           |           |             |      |
| BENEWAH            | 1996       | 17         | 8         | 12        |           |           |           |           |           |             |      |
|                    | 1995       | 55         | 10        | 14        | 15        |           |           |           |           |             |      |
|                    | 1994       | 24         | 9         | 15        | 12        | 18        |           |           |           |             |      |
|                    | 1993       | 18         | 10        | 15        | 17        | 30        | 32        |           |           |             |      |
|                    | 1992       | 24         | 7         | 13        | 16        | 22        | 21        | 22        |           |             |      |
|                    | 1991       | 3          | 12        | 34        | 29        | 28        | 32        | 32        | 18        |             |      |
| BENEWAH Total      |            | 141        | 9         | 15        | 16        | 27        | 30        | 25        | 18        |             |      |
| EVANS              | 1996       | 1          | ####      | ####      |           |           |           |           |           |             |      |
|                    | 1995       | 27         | 15        | 17        | 14        |           |           |           |           |             |      |
|                    | 1994       | 15         | 15        | 16        | 18        | 18        |           |           |           |             |      |
|                    | 1993       | 2          | 20        | 15        | 19        | 30        | 26        |           |           |             |      |
| EVANS Total        |            | 45         | 15        | 17        | 16        | 18        | 26        |           |           |             |      |
| LAKE               | 1997       | 1          | ####      |           |           |           |           |           |           |             |      |
|                    | 1996       | 11         | 11        | 18        |           |           |           |           |           |             |      |
|                    | 1995       | 98         | 10        | 16        | 20        |           |           |           |           |             |      |
|                    | 1994       | 32         | 7         | 11        | 14        | 14        |           |           |           |             |      |
|                    | 1993       | 14         | 7         | 28        | 27        | 37        | 36        |           |           |             |      |
|                    | 1992       | 20         | 9         | 13        | 17        | 19        | 35        | 31        |           |             |      |
|                    | 1991       | 16         | 7         | 22        | 9         | 15        | 17        | 20        | 24        |             |      |
|                    | 1990       | 1          | ####      | ####      | ####      | ####      | ####      | ####      | ####      | ####        |      |
|                    | LAKE Total |            | 193       | 9         | 17        | 19        | 21        | 30        | 28        | 24          | #### |
| <b>Grand Total</b> |            | <b>396</b> | <b>15</b> | <b>17</b> | <b>17</b> | <b>24</b> | <b>32</b> | <b>28</b> | <b>23</b> | <b>####</b> |      |

**Mean back-calculated lengths at annulus formation by age class and cohort - 1998.**

| SPECIES       |        | Age Class |    |     |     |     |     |     |     |   |  |
|---------------|--------|-----------|----|-----|-----|-----|-----|-----|-----|---|--|
| Brook Trout   |        |           |    |     |     |     |     |     |     |   |  |
| LOCATION      | Cohort | Number    | 1  | 2   | 3   | 4   | 5   | 6   | 7   | 8 |  |
| ALDER         | 1996   | 1         | 70 | 124 |     |     |     |     |     |   |  |
|               | 1995   | 36        | 71 | 107 | 145 |     |     |     |     |   |  |
|               | 1994   | 20        | 73 | 112 | 150 | 186 |     |     |     |   |  |
|               | 1993   | 3         | 74 | 122 | 170 | 207 | 239 |     |     |   |  |
|               | 1992   | 5         | 72 | 115 | 156 | 202 | 241 | 281 |     |   |  |
|               | 1991   | 1         | 76 | 108 | 150 | 182 | 225 | 267 | 299 |   |  |
| ALDER Total   |        | 66        | 72 | 110 | 148 | 191 | 238 | 279 | 299 |   |  |
| BENEWAH       | 1997   | 1         | 55 |     |     |     |     |     |     |   |  |
|               | 1996   | 6         | 79 | 105 |     |     |     |     |     |   |  |
|               | 1995   | 14        | 84 | 121 | 153 |     |     |     |     |   |  |
|               | 1994   | 7         | 84 | 114 | 144 | 172 |     |     |     |   |  |
|               | 1993   | 2         | 81 | 118 | 154 | 182 | 209 |     |     |   |  |
| BENEWAH Total |        | 30        | 82 | 116 | 150 | 175 | 209 |     |     |   |  |
| Grand Total   |        | 96        | 75 | 112 | 149 | 187 | 233 | 279 | 299 |   |  |

**Standard deviation of mean back-calculated lengths at annulus formation - 1998.**

| SPECIES       |             | Age Class |      |      |      |      |      |      |      |      |  |
|---------------|-------------|-----------|------|------|------|------|------|------|------|------|--|
| Brook Trout   |             |           |      |      |      |      |      |      |      |      |  |
| LOCATION      | Cohort      | Number    | 1    | 2    | 3    | 4    | 5    | 6    | 7    | 8    |  |
| ALDER         | 2           | 1         | #### | #### |      |      |      |      |      |      |  |
|               | 3           | 36        | 7    | 11   | 15   |      |      |      |      |      |  |
|               | 4           | 20        | 6    | 10   | 9    | 13   |      |      |      |      |  |
|               | 5           | 3         | 9    | 12   | 12   | 13   | 9    |      |      |      |  |
|               | 6           | 5         | 7    | 6    | 17   | 25   | 29   | 32   |      |      |  |
|               | 7           | 1         | #### | #### | #### | #### | #### | #### | #### | #### |  |
|               | ALDER Total |           | 66   | 7    | 11   | 15   | 17   | 22   | 29   | #### |  |
| BENEWAH       | 1           | 1         | #### |      |      |      |      |      |      |      |  |
|               | 2           | 6         | 6    | 7    |      |      |      |      |      |      |  |
|               | 3           | 14        | 8    | 16   | 24   |      |      |      |      |      |  |
|               | 4           | 7         | 5    | 8    | 10   | 11   |      |      |      |      |  |
|               | 5           | 2         | 8    | 18   | 28   | 36   | 44   |      |      |      |  |
| BENEWAH Total |             | 30        | 9    | 14   | 20   | 17   | 44   |      |      |      |  |
| Grand Total   |             | 96        | 9    | 12   | 16   | 18   | 27   | 29   | #### |      |  |