# LAKE ROOSEVELT FISHERIES AND LIMNOLOGICAL RESEARCH 

## 1995 ANNUAL REPORT

Prepared by:<br>Keith D. Underwood<br>and<br>John P. Shields<br>Department of Natural Resources<br>Spokane Tribe of Indians<br>Wellpinit, WA 99040

Prepared for:
U.S. Department of Energy

Bonneville Power Administration
Environment, Fish and Wildlife P. 0. Box 3621

Portland, OR 97208-3621

Project Number 88-63
Contract Number DE-8179-88DP91819

SEPTEMBER 1996

## EXECUTIVE SUMMARY

This document contains three 1995 annual reports, organized by sections. The first section contains the Lake Roosevelt Data Collection Project (BPA Contract No. 94BI32148; Project No. 94-043) report The second section contains the Lake Roosevelt Monitoring Program (BPA Contract No. DES 179-88DP918 19; Project No. 88-63) report, and the third section contains the Kokanee Salmon Imprinting (BPA Contract No. DE8179-88DP91819; Project No. 88-63) report.

These three projects were interdependent upon one another for data collection and interpretation. As a result, we combined the three annual reports in to one document in order to provide the reader with all aspects of the related analysis. Each section includes an abstract, introduction, methods, results, discussion, recommendations, references and appendices. In each discussion section, we have attempted to relate the results of the studies to existing knowledge, and speculate on management recommendations.

Section 1 discusses the Data Collection Project which was concerned with the effect of lake operations on the biota. This project was started in 1991 and was funded through the Systems Operation Review process that sought to develop an operational scenario of the Federal Columbia River Hydropower System which minimized the impacts of licensed river operations to all stakeholders of the Columbia River. The objective of the Data Collection Project was to build a biological model of the lake to predict how lake operations impact the biota of the lake. The goals of this project were to:

1. Quantify the impact of reservoir drawdowns and low water retention times on phytoplankton, zooplankton and fish;
2. Quantify the number, distribution, and use of fish food organisms in the reservoir by season;
3. Determine the seasonal growth of fish species as related to reservoir operations, prey abundance, and utilization; and
4. Quantification of entrainment levels of zooplankton and fish as related to reservoir operations and water retention times.

Limnological, reservoir operation, zooplankton, net-pen rainbow trout and kokanee tagging data were collected at up to eleven index stations in Lake Roosevelt, The 1995 results indicated that the reservoir reached a yearly low of 1,259 feet above sea level in March and a yearly high of 1,287 feet in December, with a mean yearly reservoir elevation of 1,277 feet. Water retention times ranged from 32 days in December to 69 days in September.

Duphnia spp. densities peaked during summer ranging from 6,381 to 403 \#/m $\mathrm{m}^{3}$ depending on sample site location. Minimum zooplankton densities occurred during fall ranging from 0 to 1 Daphnia spp/m ${ }^{3}$. In general, the zooplankton density was highest in the lower section of the lake.

A total of 12,984 net-pen rainbow trout were tagged at Kettle Falls and Seven Bays. Tag returns from anglers in Lake Roosevelt or below returned 200 of these tags yielding a recapture rate of 2\%. Entrainment of rainbow trout through Grand Coulee Dam was
considered low in 1995 (3\%) due to relatively high average water retention times in the Spring ( $40+$ days) and because rainbow trout were held in net pens until after June 1.

Results of the Data Collection Project found that reservoir elevations and storage capacity were similar in 1994 and 1995, but water retention times were reduced by an average of 12.9 days. This was due in part to a ten foot August drawdown implemented to benefit ESA listed stocks in the Snake River. Average zooplankton densities for 1995 were higher than 1994, but zooplankton biomass values for 1995 decreased due to significantly smaller Daphia spp. sizes. The temperature of Lake Roosevelt was positively correlated with zooplankton density.

Management recommendations for the Data Collection Project are as follows:

1. Continue to tag 10,000 rainbow trout at Kettle Falls and Seven Bays in order to increase the numbers of tag returns;
2. Continue to hold net pen rainbow trout until at least June 1 before release in order to reduce entrainment losses:
3. Increase the zooplankton sampling frequency to at least three times per month in the spring, summer and fall and twice per month in winter. In addition, begin to sample zooplankton in near shore areas along with mid channel tows. Also, due to the extreme variability of zooplankton data it is recommended that a minimum of three zooplankton tows be taken at each site and the densities averaged to attain a mean location density;
4. Continue to sample for zooplankton in Rufus Woods Reservoir in order to estimate entrainment losses of zooplankton from Lake Roosevelt;
5. Continue to collect zooplankton and water quality data at current sites;
6. Collect nutrient and $\mathrm{C}^{14}$ data to obtain nutrient abundance and assimilation rates for model development; and,
7. Detennine the depth of the euphotic zone via photometer to estimate the availability of phytoplankton habitat.

The Lake Roosevelt Monitoring Program is contained in section 2. This study was primarily concerned with the effect of stocking kokanee salmon and rainbow trout on the ecosystem. The Sherman Creek Hatchery (Washington Department of Fish and Wildlife) and the Spokane Tribal Hatchery (Spokane Tribe of Indians) were operational in 1991. To evaluate the effectiveness of stocking on the lake biota and fishery, baseline data was collected beginning in 1988. These data were compared to baseline fisheries data posthatchery stocking. The data generated from sampling was analyzed to determine food availability, utilization and preferences, growth rates, and angler use information (e.g. harvest). The objective of the Monitoring Program was to maximize angler harvest and maximize adult returns to egg collection sites.

The goals of the Monitoring Program were:

1. Determine angler pressure, number of fish harvested, average size of fish harvested and the economic value of the fishery;
2. Estimate the relative abundance of fish in the lake;
3. Determine the diet of kokanee salmon, rainbow trout and walleye; and

4 Estimate age and growth of kokanee salmon, rainbow trout and walleye.

Kokanee harvest increased from 11,906 fish in 1989 to 32,353 fish in 1995. Rainbow trout harvest increased from 65,515 fish in 1989 to 122,939 fish in 1995. Walleye harvest decreased from 80,626 fish in 1992 to 40,185 fish in 1995.

Relative abundance of kokanee salmon in fisheries surveys increased from $3 \%$ in past years to $20 \%$ in 1995. The relative abundance of rainbow trout remained at approximately $5 \%$. The relative abundance of yellow perch steadily decreased from $40 \%$ in 1989 to $7 \%$ in 1995.

The economic value of the lake increased from $\$ 2$ million in 1988 (before hatchery supplementation) to almost $\$ 9$ million in 1995 (after hatchery supplementation).

In 1995, mean lengths of kokanee were $219 \mathrm{~mm}, 385 \mathrm{~mm}$ and 472 mm for $1+, 2+$, and $3+$ fish. The mean condition factor for all age groups was 1.03. Rainbow trout mean lengths were $206,340 \mathrm{~mm}, 416 \mathrm{~mm}, 504 \mathrm{~mm}$, and 537 mm for age $1+, 2+, 3+, 4+$ and $5+$ fish. The mean condition factor for all age groups was 1.08 . Walleye lengths ranged from 149 mm for age $0+$ fish to 766 mm for age $11+$ fish.

Diet overlap predicted that kokanee and rainbow trout overlap was 0.80 (high overlap). Kokanee and walleye diet overlap was low (0.15) and rainbow and walleye overlap was moderate (0.45).

Results of the Monitoring Program suggests that the release of hatchery origin fish continued to improve harvest rates and the relative abundance of kokanee salmon and rainbow trout. Growth of kokanee and rainbow trout continued to exceed the mean growth per age class of fish in area lakes, but walleye growth and harvest rates have declined over the past 5 years. This may have been the result of food shortages since there has been a decrease in the relative abundance of yellow perch. Feeding habits of kokanee and walleye remain similar to past years with kokanee feeding mainly on Duphnia spp. and chironomids, and walleye feeding mainly on fish. However, food habits of rainbow trout were different than in past years. Rainbow fed mainly on Duphnia spp., and chironomids, but also fed on yellow perch. This apparent change in feeding habits may have been due to a change in sampling months from August to July in 1995.

Management recommendations for the Monitoring Program include:

1. Quantify the impact of walleye on newly stocked kokanee salmon;
2. Record the origin (fin clips) of every kokanee sampled to determine hatchery versus wild origin;
3. Evaluate the scientific design of the creel survey and methods used to compute indices;
4. Conduct hydroacoustic surveys monthly to identify spatial and temporal accumulations of fish along the length and width of the lake; and
5. Pursue new capture methods for fish assemblages to determine fish species, age structure, etc.

The third section of this report discusses the kokanee salmon imprinting and coded wire tagging program (Imprinting Program). This is a sub-contract to the Monitoring Program which began investigations to determine the critical period for thyroxine-induced olfactory imprinting in 1991. The objectives of this present study were to:

1. Determine the critical period(s) for olfactory imprinting; and
2. assess the best times and locations to release kokanee in order to prevent entrainment, and improve returns to creel and egg collection sites.

Field tests were conducted by exposing juvenile kokanee to the synthetic chemicals morpholine and phenethyl alcohol at different life stages. These artificially imprinted fish were coded wire tagged and stocked into Lake Roosevelt from 1992 through 1995. Adult kokanee salmon were collected during the spawning period and checked for tags to determine which life stage homed better to its release site.

Results continued to show that kokanee can be successfully imprinted to artificial odors morpholine and phenethyl alcohol - as juveniles from hatch to swimup and again at the smolt stage. Fish double imprinted at these stages exhibited the highest percentage homing $(86.5 \%)$. Also, fish exposed to chemicals showed greater homing ability than unexposed fish. Smolt releases continue to provide better adult recoveries than fry releases ( $99 \%$ of recoveries in 1995 came from fish released as smolts).

Based on the results of the Imprinting Program, the following recommendations were made:

1. Release more kokanee salmon into the reservoir,
2. Monitor entrainment of kokanee from Lake Roosevelt;
3. Modify hatchery ladder system at Sherman Creek Hatchery to attract more adult kokanee;
4. Set up an egg site at Hawk Creek since 286 fish were recovered at that site; and
5. Locate alternative stocks of kokanee with better genetic adaptations than Lake Whatcom stock for the Lake Roosevelt Program.

## TABLE OF CONTENTS

EXECUTIVE SUMMARY ..... ii
TABLE OF CONTENTS ..... iii
Section 1. Measurement of Lake Roosevelt Biota In Relation to Reservoir Operations. ..... 1
Section 2. Lake Roosevelt Monitoring Program ..... 159
Section 3. Imprinting Program ..... 281

## SECTION 1

# MEASUREMENT OF LAKE ROOSEVELT BIOTA IN RELATION TO RESERVOIR OPERATIONS <br> 1995 ANNUAL REPORT 

Prepared by:
John P. Shields
and
Keith D. Underwood
Department of Natural Resources
Spokane Tribe of Indians
Wellpinit, WA 99040

Prepared for:
U.S. Department of Energy

Bonneville Power Administration
Environment, Fish and Wildlife
P.O. Box 3621

Portland, OR 97208-3621

Project No. 94-043
Contract No. 94BI32148


#### Abstract

The purpose of the Lake Roosevelt Data Collection Project (Data Collection Project) was to collect data to develop a computer model that will predict biological responses to reservoir operations as part of the System Operation Review Program. The data collected by this project was gathered to quantify and qualify the impacts of reservoir operations on the ecology and limnology of the Lake. This project began 1991 and has since worked closely with the Lake Roosevelt Monitoring Project. In 1995, limnological, reservoir operation, zooplankton and net-pen rainbow trout tagging data were collected at up to eleven index stations throughout the reservoir. The monthly mean lake elevations ranged from 1,250 feet in March to 1,287 feet in December with a mean yearly reservoir elevation of 1,277 which was similar to 1994. However, for the first time, Lake Roosevelt was drawndown ten feet in August to benefit ESA listed stocks in the Snake River. This resulted in significantly lower water retention times of 47 days for August, 1995 compared to 59 days for August, 1994. Yearly mean water retention times was 47 days and ranged from 32 days in December to 69 days in September. Daphnia spp. densities peaked during July and August, reaching a maximum density of 6,380 organisms per $\mathrm{m}^{3}$ in August at the Confluence site. Minimum overall zooplankton densities of zero or near zero organisms per $\mathrm{m}^{3}$ were observed in March and April at most sites. Higher zooplankton densities were found at lower end of the reservoir. Daphnia and total zooplankton densities were much higher in 1995 when compared with 1994. However, 1995 biomass values were only a fraction of those reported in 1994 due to a dramatic reduction in zooplankton size. In 1995, 12,984 net-pen rainbow trout were tagged at Kettle Falls $(4,995)$ and Seven Bays $(7,989)$. Anglers captured and returned 200 of these tags, yielding a recapture rate of $2 \%$. Sixty-seven tags were returned from Kettle Falls releases and 133 were returned from Seven Bays releases. Two tags from each release location were recaptured in Rufus Woods reservoir and one additional Seven Bays fish was recaptured at Rocky Reach Dam yielding an entrainment rate of $3 \%$.


## ACKNOWLEDGMENTS

We gratefully acknowledge Dr. Allan (Eastern Washington University), and Dr. Ross Black (EWU), for advise and scientific input on this project. Creel clerks William Matt Sr. (Spokane Tribe of Indians), Jim Meskan (Washington Department of Fish and Wildlife), and Leroy Williams (Collville Confederated Tribes), for collection of creel data and assistance in tagging operations. John Hisata (Washington Department of Fish and Wildlife) and Jerry Marco (Colville Confederated Tribes), for advise and support in all areas. Dan Epstein and Dave Geist (Battelle Pacific Northwest Laboratories), for their oversight and assistance with this research. We give special thanks to Mary Beth Tilson, Amy Voeller, Mike Thatcher, Del Brown, Henry Etue, Bart Kieffer, William Matt Jr., Sam Abrahamson, Jason Wynecoop, Mitch Combs and Tim Peone for their assistance in the collection and analysis of field data. We further thank Hank Etue and Del Brown for their efficient maintenance and repair of all equipment used in this project.

We acknowledge Kelly Cash (National Park Service), and Gary Kuiper (NPS), for graciously providing camping at no cost to the project. We also thank Winn Self, Marty Blum, Roy Graffis, and students from Hunters High School for their assistance in the feeding and tagging of net-pen rainbow trout. Additional thanks goes to Charlie Craig (BPA, Division of Fish and Wildlife), who served as BPA project manager, Rob Pierson (Bonneville Power Administration), Larry Goodrow (Executive Director, Spokane Tribe of Indians and UCUT Administrator) Mary Vemer (Natural Resource Director, Spokane Tribe of Indians), Amy Voeller (previous project manager) and Carol Evans (Accountant, Spokane Tribe of Indians) for administration of contracts.

This project was supported by a contract from the U.S. Department of Energy, Bonneville Power Administration (BPA), Contact No.94BI32148, Modification No. A006, Project No. 94-043. Additional financial support for this project was provided by a grant from the U.S. Department of Interior, Bureau of Indian Affairs to the Upper Columbia United Tribes (UCUT), to fund the operation of the UCUT Fisheries Research Center at Eastern Washington University (Grant No. P12614208001). Capitol equipment for this project was supplied by the UCUT Fisheries Center.

## TABLE OF CONTENTS

ABSTRACT *.* ..... 2
ACKNOWLEDGMENTS ..... 3
LIST OF TABLES ..... 5
LIST OF FIGURES .....  8
1.0 INTRODUCTION ..... 11
1.1 Project History ..... 11
1.21994 Study Objectives ..... 11
2.0 MATERIALS AND METHODS ..... 13
2.1 Description of Study Area ..... 13
2.2 Reservoir Hydrology ..... 13
2.3 Zooplankton. ..... 15
2.4 Tagging Studies ..... 18
3.0 RESULTS ..... 19
3.1.. Hydrology ..... 19
3.2 Zooplankton ..... 22
3.3 Tagging ..... 49
4.0 DISCUSSION ..... 55
4.1 Reservoir Operations ..... 55
4.2 Hydrology ..... 55
4.3 Zooplankton Dynamics ..... 59
4.4 Rainbow Trout Tagging ..... 73
5.0 RECOMMENDATIONS ..... 75
LITERATURE CITED ..... 77
APPENDIX A. Hydrology ..... 81
APPENDIX B. Zooplankton ..... 95
APPENDIX C. Water Quality ..... 117

## LIST OF TABLES

Table
Page


$\begin{array}{ll}\text { Table 3.2 } & \begin{array}{l}\text { Synoptic list of zooplankton taxa historically identified in Lake } \\ \text { Roosevelt, including those identified during the } 1995 \text { study period. . }\end{array} 23\end{array}$
$\begin{array}{ll}\text { Table } 3.3 & \text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ & \text { Kettle Falls (Index Station 1), in 1995.................................................... }\end{array}$
$\begin{array}{ll}\text { Table 3.4 } & \text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ & \text { Gifford (Index Station 2), in 1995 . . . .................................. . } 25\end{array}$



$\begin{array}{ll}\text { Table } 3.8 & \begin{array}{l}\text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ \text { Seven Bays (Index Station 6), in } 1995 \ldots \ldots . . . . . . . . . . . . . . . . . . . . . . . . . . . . . . . ~\end{array} 29\end{array}$
$\begin{array}{ll}\text { Table } 3.9 & \text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ & \text { Keller Ferry (Index Station 7), in 1995.......................................................... }\end{array}$

$\begin{array}{ll}\text { Table 3.11 } & \begin{array}{l}\text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ \text { Spring Canyon (Index Station 9), in } 1995 \ldots \ldots \ldots \ldots . . . . . . . . . . . . . . . . . . . . . . . . ~\end{array} 32\end{array}$
$\begin{array}{ll}\text { Table 3.12 } & \begin{array}{l}\text { Mean monthly densities }\left(\# / \mathrm{m}^{3}\right) \text { for representative zooplankton at } \\ \\ \text { Rufus Woods (Index Station 10), in 1995.......................... } 33\end{array}\end{array}$
Table Page
Table 3.13 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Kettle Falls (Index Station 1), in 1995 ..... 35
Table 3.14 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Gifford (Index Station 2), in 1995 ..... 35
Table 3.15 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Hunters (Index Station 3), in 1995 ..... 35
Table 3.16 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Porcupine Bay (Index Station 4), in 1995 ..... 35
Table 3.17 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at the confluence of the Spokane River with the mainstem Columbia (Index Station Confluence), in 1995 ..... 36
Table 3.18 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Seven Bays (Index Station 6), in 1995 ..... 36
Table 3.19 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Keller Ferry (Index Station 7), in 1995 ..... 36
Table 3.20 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at San Poil (Index Station 8), in 1995 ..... 36
Table 3.21 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Spring Canyon (Index Station 9), in 1995 ..... 37
Table 3.22 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at Rufus Woods (Index Station 10), in 1995 ..... 37
Table 3.23 Monthly mean zooplankton lengths in mm for select cladocera at Kettle Falls (Index Station 1), in 1995 ..... 39
Table 3.24 Monthly mean zooplankton lengths in mm for select cladocera at Gifford (Index Station 2), in 1995. ..... 40
Table 3.25 Monthly mean zooplankton lengths in mm for select cladocera at Hunters (Index Station 3), in 1995 ..... 41
Table 3.26 Monthly mean zooplankton lengths in mm for select cladocera at Porcupine Bay (Index Station 4), in 1995 ..... 42
Table 3.27 Monthly mean zooplankton lengths in mm for select cladocera at the confluence of the Spokane River with the mainstem Columbia (Index Station Confluence), in 1995. ..... 43
Table 3.28 Monthly mean zooplankton lengths in mm for select cladocera at Seven Bays (Index Station 6), in 1995 ..... 44
Table Page
Table 3.29 Monthly mean zooplankton lengths in mm for select cladocera at Keller Ferry (Index Station 7), in 1995 ..... 45
Table 3.30 Monthly mean zooplankton lengths in mm for select cladocera at San Poil (Index Station 8), in 1995 ..... 46
Table 3.31 Monthly mean zooplankton lengths in mm for select cladocera at Spring Canyon (Index Station 9), in 1995 ..... 47
Table 3.32 Monthly mean zooplankton lengths in mm for select cladocera at Rufus Woods (Index Station 10), in 1995 ..... 48
Table 3.33 Summary of release dates, numbers, and subsequent capture locations of net-pen rainbow trout tagged and released from various net pen locations in 1995. ..... 50
Table 3.34 Summary of rainbow trout release times, water retention times and subsequent recapture numbers and percentages by year. ..... 51
Table 3.35 Number and percent of fish captured by location from the Kettle Falls tag releases in 1995 ..... 54
Table 3.36 Number and percent of fish captured by location from the Seven Bays tag releases in 1995 ..... 54
Table 4.1 Monthly and annual means for reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in 1994 and 1995. ..... 56

## LIST OF FIGURES

Figure
Figure 2.1 Map of Lake Roosevelt, WA showing the locations of sampling and tagging stations ..... 14
Figure 3.1 Mean monthly reservoir elevations in 1995 ..... 21
Figure 3.2 Mean monthly water retention times in 1995 ..... 21
Figure 3.3 Plot of length increases (mm) versus time since release (Days) for tagged rainbow trout at Kettle Falls in 1995 ..... 52
Figure 3.4 Plot of length increases (mm) versus time since release (Days) for tagged rainbow trout at Seven Bays in 1995 ..... 52
Figure 3.5 Plot of tagged rainbow trout weight increases (g) versus time since release (Days) at Kettle Falls in 1995 ..... 53
Figure 4.1 Mean monthly reservoir elevations from 1991 through 1995 ..... 58
Figure 4.2 Mean monthly water retention times from 1991 through 1995. ..... 58
Figure 4.3 Mean monthly Daphnia spp. densities (\#/m ${ }^{3}$ ) at Gifford, Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon in 1995. ..... 60
Figure 4.4 Mean monthly total zooplankton densities $\left(\# / \mathrm{m}^{3}\right)$ at Gifford, Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon in 1995. ..... 60
Figure 4.5 Mean Daphnia spp. densities ( $\# / \mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Gifford in 1995 ..... 61
Figure 4.6 Mean total zooplankton densities (\#/ $\mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Gifford in 1995 ..... 61
Figure 4.7 Mean Daphnia spp. densities (\#/m3) plotted against water temperature at 12 m depth at Porcupine Bay in 1995 ..... 62
Figure 4.8 Mean total zooplankton densites ( $\# / \mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Porcupine Bay in 1995. ..... 62
Figure 4.9 Mean Daphnia spp. densites (\#/m ${ }^{3}$ ) plotted against water temperature at 12 m depth at Seven Bays in 1995 . ..... 63
Figure 4.10 Mean total zooplankton densites ( $\# / \mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Seven Bays in 1995 ..... 63
Figure 4.11 Mean Daphnia spp. densites (\#/m ${ }^{3}$ ) plotted against water temperature at 12mdepthatKellerFerryin 1995 ..... 64
Figure Page
Figure 4.12 Mean total zooplankton densites ( $\# / \mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Keller Ferry in 1995. ..... 64
Figure 4.13 Mean Daphnia spp. densites (\#/m³) plotted against water temperature at 12 m depth at Spring Canyon in 1995 ..... 65
Figure 4.14 Mean total zooplankton densites ( $\# / \mathrm{m}^{3}$ ) plotted against water temperature at 12 m depth at Spring Canyon in 1995 ..... 65
Figure 4.15 Regression plot of Daphnia spp. densities versus water temperatures at 12 m for five locations in Lake Roosevelt in 1995 ..... 66
Figure 4.16 Regression plot of total zooplankton densities versus water temperatures at 12 m for five locations in Lake Roosevelt in 1995 ..... 66
Figure 4.17 Monthly mean total zooplankton densities and monthly average waterretention times at Gifford (Index Station 2), from 1991 through1995.68
Figure 4.18 Monthly mean total zooplankton densities and monthly average water retention times at Porcupine Bay (Index Station 4), from 1991 through 1995. ..... 69
Figure 4.19 Monthly mean total zooplankton densities and monthly average waterretention times at Seven Bays (Index Station 6), from 1991 through1995.70
Figure 4.20 Monthly mean total zooplankton densities and monthly average water retention times at Keller Ferry (Index Station 7), from 1991 through 1995.71
Figure 4.21 Monthly mean total zooplankton densities and monthly average water retention times at Spring Canyon (Index Station 9), from 1991 through 1995.

### 1.0 INTRODUCTION

### 1.1 Project History

The Data Collection Project began in July, 1991 as part of the Bonneville Power Administration (BPA), Bureau of Reclamation (BOR), and U.S. Army Corps of Engineer's (ACE) System Operation Review process. This process sought to develop an operational scenario for the Federal Columbia River Hydropower System which minimized impacts to all stakeholders of the Columbia River. The objective of the Data Collection Project was to collect data for the development of a biological model for Lake Roosevelt enabling researchers to identify which lake operation scenario best suites the biota of the lake. Major components of the Lake Roosevelt model will be: 1) quantification of impacts to phytoplankton, zooplankton and fish caused by reservoir drawdowns and low water retention times; 2) quantification of the number, distribution, and use of fish food organisms in the reservoir by season; 3) determination of seasonal growth of fish species as related to reservoir operations, prey abundance, and utilization; and 4) quantification of entrainment levels of zooplankton and fish as related to reservoir operations and water retention times. Upon completion in the year 1998, the model will predict biological responses to different reservoir operation strategies.

This chapter contains the results of the Data Collection Project for Lake Roosevelt from January through December 1995. Previous annual reports for the Data Collection Project were written by Griffith et al. (1991), Griffith and McDowell (1996), Voeller (1996) and Shields and Underwood (1996).

### 1.2 1995 Study Objectives

Objectives of the Lake Roosevelt Data Collection Project for 1995 were, as follows:

1. Collect zooplankton biomass and density data at eleven locations throughout Lake Roosevelt.
2. Tag rainbow trout in Lake Roosevelt and use tag return data to estimate entrainment, growth rates and habitat preferences.
3. Collect limnological data on the lake including: pH , temperature, dissolved oxygen, conductivity, oxidative reductive potential and secchii disk at ten sites throughout Lake Roosevelt.
4. Compare and contrast data collected during 1995 with previous years, to identify changes in the lake.
5. Participate in operational decisions on lake Roosevelt by providing technical input to the SOR through the resident fish work group.

### 2.0 MATERIALS AND METHODS

### 2.1 Description of Study Area

Lake Roosevelt is a mainstem Columbia River impoundment formed by the construction of Grand Coulee Dam in 1939 (Figure 2.1). Filled in 1941, the reservoir inundates 33,490 hectares at a full pool elevation of 393 m above mean sea level. It has a maximum width of 3.4 km and a maximum depth of 122 m (Stober er al. 1981). Grand Coulee Dam is a Bureau of Reclamation storage project operated primarily for power, flood control, and irrigation with secondary operations for recreation, fish, and wildlife.

### 2.2 Reservoir Hydrology

Water quality measurements of temperature, pH , dissolved oxygen, conductivity, and oxygen reduction potential were collected using a Hydrolab Surveyor II at ten sites in the reservoir. Samples were collected mid-channel to a depth of 33 M at Kettle Falls (Location 1), Gifford (Location 2), Hunters (Location 3), Porcupine Bay (Location 4), the Confluence of the Spokane River with the mainstem Columbia (Location confluence), Seven Bays (Location 6), Keller Ferry (Location 7), San Poil River (Location 8), Spring Canyon (Location 9), and at Rufus Woods Reservoir (Location 10), monthly in 1995 (Figure 2.1). Secchii disk readings were taken in conjunction with Hydrolab measurements at each of the above sites. This data collection continues investigations which began in 1991 (Appendix C).

Reservoir elevations and water retention times were calculated from daily midnight reservoir elevations ( ft ) and total outflows in thousand cubic feet per second per day (kcfs). Reservoir elevation and total outflow values were derived from summary reports for Grand Coulee Dam prepared monthly in 1995 by the U.S. Army Corps of Engineers, Reservoir Control Center in Portland, OR. Reservoir elevation was converted to volume of water stored (kcfsd) using a U.S. Army Corps of Engineers (198 1) reservoir water storage table. Water retention time was calculated using the formula:

$$
\text { Water retention time (days) } \quad=\quad \frac{\text { Reservoir volume }(\mathrm{kcfsd})}{\text { Outflow }(\mathrm{kcfs})}
$$

Daily values for each category were totaled and divided by the number of days in each month to attain mean reservoir elevations and water retention times (Appendix A).


Figure 2.1 Map of Lake Roosevelt, WA. " $\boldsymbol{\Delta}$ " denotes sampling locations and" " denotes tagging stations.

### 2.3 Zooplankton

Zooplankton samples were collected mid-channel at Seven Bays, Keller Ferry and Spring Canyon once every two weeks from January through October, 1995. Gifford, Porcupine Bay, Confluence, and Rufus Woods sites were sampled monthly from January through October, 1995. Additional samples were taken at Kettle Falls, Hunters, and the San Poil river three times per year in March, August and October as part of the Lake Roosevelt Monitoring Project's annual sampling sessions. Samples were taken using a Wisconsin vertical tow plankton net with an $80 \mu \mathrm{~m}$ silk mesh and bucket and a radius of 14.5 cm . Triplicate and some duplicate tows were made from 33 m depth to the surface at each location within Lake Roosevelt. At Rufus Woods, where main channel river depths average 20 m , two sets of three 15 m subsample tows were taken and combined into two samples. After all tows, collected organisms were washed into individual 253 ml bottles containing 10 ml of $37 \%$ formaldehyde and 0.5 g sugar Rigler (1978). Bottles were labeled with the date, location, time, tow number and tow depth. Organisms were then stained with 1.0 ml of five percent Lugol's solution and 1.0 ml of saturated eosin-y ethanol stain and brought to a volume of 200 mls .

In the laboratory, organisms were identified to species using taxonomic keys of Brandlova et al. (1972), Brooks (1957), Edmondson (1959), Pennak (1978;1989), Ruttner-Kolisko (1974), and Stemberger (1979). A Nikon SMZ-10 dissecting microscope with a ring illuminator system and Nikon Optiphot phase contrast microscope were used for identification. In cases where sample densities were high, three sub-samples were counted using a modified counting chamber Ward (1955), until 60 organisms or 25 ml of sample was counted (Edmondson and Winberg 1971, Downing and Rigler 1984). Volumes of sub-samples selected depended upon organism densities in the samples.

Zooplankton densities were calculated for each individual tow and the results of the three tows were averaged to arrive at a single location density. Zooplankton density (\# organisms $/ \mathrm{m}^{3}$ ) was calculated using the following sets of equations. First, the volume (L) of sample collected by the Wisconsin plankton sampler was calculated by the formula:

$$
V=\Pi r^{2} h
$$

where:

$$
\begin{aligned}
\mathrm{v} & =\text { volume of the sample (liters); } \\
\Pi & =\text { pi }(3.14) ; \\
\mathrm{r} & =\text { radius of sampler }(\mathrm{cm}) ; \text { and } \\
\mathrm{h} & =\text { depth of sample }(\mathrm{m}) .
\end{aligned}
$$

Next, microcrustacean zooplankton density (\# organisms/ m3) was calculated by the equation:

$$
\mathrm{D}=\frac{\left(\frac{T_{c}}{S_{n}} * \frac{S V}{S S V}\right)}{\mathrm{V}} \mathrm{DF} * 1000
$$

where:

$$
\begin{aligned}
\mathrm{D} & =\text { density (\# organisms/m3); } \\
\mathrm{Sn} & =\text { number of sub-samples; } \\
\text { S v } & =\text { sample volume; } \\
\text { s s v } & =\text { sub-sample volume; } \\
\mathrm{V} & =\text { volume of entire sample; } \\
\mathrm{DF} & =\text { dilution factor; and } \\
\mathrm{Tc} & =\text { total number counted of each species } \\
&
\end{aligned}
$$

Predominant cladocera biomass was determined using the length-weight regression equations summarized by Downing and Rigler (1984). Mean cladoceran length used in these equations was determined by measuring randomly chosen groups of up to twenty individual cladocera per species. Leptodora lengths were taken by direct measurement, while all other measurements were made by first calibrating a Nikon Optiphot scope so that 10 micrometer units equaled 10 mm . Next, individual cladocera were lined up to the micrometer and measured from the top of the head to the base of the carapace, excluding the spine. Observed length data was converted to actual lengths through the use of a
conversion factor of 0.02 . Actual length data was then averaged for each species and recorded. Next, dry weight estimates were made by the use of the equation:

$$
\ln w=\ln a+(b)(\ln L)
$$

Where:

$$
\begin{aligned}
\ln w & =\text { the natural log of the dry weight estimate }(\mu \mathrm{g}) \text { for the } \\
\ln a & =\text { Cladocera species; } \\
\mathrm{b} & =\text { the natural log of the intercept for the Cladocera species; } \\
\ln L & =\text { the natural log of the mean length value for the Cladocera } \\
& \text { species. }
\end{aligned}
$$

Slope (b) and intercept (ln a) values used for the dry weight estimate calculation (Downing and Rigler 1984) are found in table 3.1:

Table 2.1 Slope (b) and intercept (In a) values used for the dry weight estimate calculations.

| Cladocera Species | In $\boldsymbol{a}$ | b |
| :--- | :---: | :---: |
| Daphnia galeata mendota | 1.51 | 2.56 |
| Daphnia retrocurva | 1.4322 | 3.129 |
| Daphnia schødleri | 2.30 | 3.10 |
| Daphnia thorata | 2.64 | 2.54 |
| Leptodora kindtii | -0.822 | 2.67 |

Average cladocera biomass was calculated using the formula:

$$
B=(\ln w)(D)
$$

Where:

$$
\begin{aligned}
\mathrm{B} & =\text { biomass }\left(\mu \mathrm{g} / \mathrm{m}^{3}\right) ; \\
\text { lnw } & =\log \text { of the dry weight estimate for the Cladocera species } \\
& (\mu \mathrm{g}) ; \text { and } \\
\mathrm{D} & =\text { density }\left(\# \text { organisms } / \mathrm{m}^{3}\right) .
\end{aligned}
$$

### 2.4 Tagging Studies

Tagging studies were conducted on Lake Roosevelt using age one net-pen reared rainbow trout. Fish chosen for this study were randomly netted out of holding pens, measured to the nearest millimeter and tagged by placing individually numbered floy tags into the posterior base of the dorsal fin. Prior to tagging and length measurement, groups of up to 200 fish were anesthetized with carbon dioxide. This process involved placing 50 gallons of lake water into a large plastic holding tank and bubbling $\mathrm{CO}_{2}$ into it from a 750 psi main tank through two 32 inch oxygen stones at a rate of 30 psi for three minutes. pH levels in the holding tank were monitored with a Hydrolab II surveyor and buffered to a level of 6.5 to 7.0 with calcium bicarbonate (Post, 1979). Once acceptable pH ranges were attained, fish were netted from holding pens and placed in the $\mathrm{CO}_{2}$ water where they were rendered unconscious within one minute allowing for easy handling. Once measured and tagged, all fish were allowed to recuperate for up to 30 minutes in 20 gallon garbage cans prior to being returned to the net pens. Tagged fish were then held in net pens for three weeks, at which time mortality rates were calculated and fish released. Overall mortality rates for this process were less than $0.5 \%$. In 1995, 4,995 fish were tagged at Kettle Falls and 7,989 fish were tagged at Seven Bays. Tag colors were changed by year so that each age class of tagged fish could be easily differentiated. Orange colored tags were used in 1995.

In order to maximize angler tag returns, informational posters describing the Monitoring Program's tagging studies were distributed throughout Lake Roosevelt and Rufus Woods reservoir at locations frequented by anglers. These posters gave a visual description of floy tags and also requested that anglers return tags with recapture information which included: recapture date, location, fish length and fish weight. Any angler that returned tag information was sent a letter informing him or her of the fish release date, location, and length of fish at time of release. The angler was also provided with a brief summary of the tagging program.

Tag return data was compiled and analyzed to determine fish growth rates and movement within Lake Roosevelt and was also used to estimate entrainment rates through Grand Coulee Dam. Movement was analyzed by noting recapture location and plotting it against release location and date.

### 3.0 RESULTS

## $3.1 \quad$ Hydrology

Table 3.1 summarizes' mean monthly reservoir operations for Lake Roosevelt in 1995. Appendix A summarizes daily reservoir operations from January through December, 1995. Average reservoir elevations began at 1,278 feet above sea level in January and dropped to an elevation of 1,259 feet in April (Table 3.1 and Figure 3.1). The reservoir then began to refill to near full pool by June and remained stable within ten feet through the end of the year. Mean reservoir elevations ranged from a low of 1,259 feet in April to a maximum of 1,287 feet in July and December (Table and Figure 3.1). Mean yearly reservoir elevation was 1,277 feet. Mean outflows ranged from a low of 66 kcsf in September to 118 kcfs in June with a yearly mean of 96 kcfs . Mean inflows ranged from a low of 73 kcsf in January to 148 kcfs in June with a yearly mean of 100 kcfs (Table 3.1). Mean monthly water retention times ranged from a low of 32 days in December to a maximum of 69 days in September. Water retention times did not go below thirty days for any month (Table 3.1 and Figure 3.2).

Appendix C summarizes the water quality measurements taken by Hydrolab Surveyor II from January through October, 1995. Analysis of these data show that monthly temperatures ranged from a yearly low of $2.3^{\circ} \mathrm{C}$ in January at Gifford to a high of $24.1^{\circ} \mathrm{C}$ in July at Porcupine Bay. pH values ranged from a low of 6.9 in August at Porcupine Bay to a maximum of 8.7 at Spring Canyon in May. Higher pH readings of up to 11.2 were observed in January through March but these were due to a malfunctioning pH meter and as such are deemed unreliable. Dissolved oxygen readings ranged from a yearly minimum of $2.5 \mathrm{mg} / \mathrm{L}$ at Porcupine Bay in August to a maximum of $19.4 \mathrm{mg} / \mathrm{L}$ at Keller Ferry in March. Conductivity readings ranged from a low of $0.082 \mathrm{mmho} / \mathrm{cm}$ in March at Porcupine Bay to a high of $0.223 \mathrm{mmho} / \mathrm{cm}$ at Porcupine Bay in October. Oxidative reductive potential (ORP) in volts, ranged from a yearly low of 0.038 volts in June at Porcupine Bay to a yearly maximum of 0.302 volts in March at Spring Canyon. Appendix D also reports monthly secchii disk depths in meters for nine sampling stations within Lake Roosevelt. These measurements ranged from a yearly low of 0.3 meters in February at Porcupine Bay to a yearly high of 9.0 meters in July at Spring Canyon. Keller Ferry and Spring Canyon had the highest overall average secchii disk readings of 5.5 meters, while Porcupine Bay reported the lowest at 3.0 meters (Table D.64).

Table 3.1 Monthly and annual means for reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in 1995.

| Month | Inflow <br> (kcfs) | outflow <br> (kcfs) | Reservoir <br> Elevation <br> (Ft) | Storage <br> Capacity <br> (kcfsd) | Water <br> Retention <br> Time (Days) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Jan 1995 | 73.0 | 88.3 | $1,278.3$ | $4,127.8$ | 49.3 |
| Feb 1995 | 81.7 | 94.0 | $1,266.2$ | $3,688.9$ | 42.6 |
| Mar 1995 | 101.6 | 90.1 | $1,259.0$ | $3,434.3$ | 42.4 |
| Apr 1995 | 81.2 | 84.5 | $1,265.8$ | $3,669.5$ | 47.5 |
| May 1995 | 112.3 | 93.5 | $1,260-1$ | $3,460.4$ | 39.4 |
| Jun 1995 | 148.1 | 117.8 | $1,283.6$ | $4,335.3$ | 40.1 |
| Jul 1995 | 111.6 | 110.5 | $1,287.0$ | $4,467.4$ | 41.4 |
| Aug 1995 | $\mathbf{9 6 . 3}$ | 91.9 | $1,280.9$ | $4,227.8$ | 47.2 |
| Sep 1995 | 79.9 | 65.9 | $1,285.1$ | $4,392.8$ | 69.0 |
| Oct 1995 | 80.3 | 80.6 | $1,285.8$ | $4,420.3$ | 56.7 |
| Nov 1995 | 97.2 | 91.9 | $1,286.5$ | $4,448.3$ | 50.4 |
| Dec 1995 | 135.7 | 141.6 | $1,287.0$ | $4,466.7$ | 32.4 |
| Mean 1995 | $\mathbf{9 9 . 9}$ | $\mathbf{9 5 . 9}$ | $\mathbf{1 , 2 7 7 . 1}$ | $\mathbf{4 , 0 9 5 . 0}$ | $\mathbf{4 6 . 5}$ |



Figure 3.1 Mean monthly reservoir elevations in 1995.


Figure 3.2 Mean monthly water retention times in 1995.

### 3.2 Zooplankton

### 3.2.1 Zooplankton Densities

A total of 11 species of zooplankton were identified in Lake Roosevelt during 1995 (Table 3.2). Seven species were identified from Order Cladocera, the most diverse group, followed by 5 species from Order Eucopepoda.

Monthly mean densities (\# of organisms $/ \mathrm{m}^{3}$ ), of microcrustacean zooplankton collected at Kettle Falls, Gifford, Hunters, Porcupine Bay, Confluence, Seven Bays, Keller Ferry, San Poil, Spring Canyon and Rufus Woods are shown in Tables 3.3 through 3.12. Mean zooplankton densities by species for each location are found in Appendix B.

Mean total zooplankton densities at Kettle Falls ranged from a low of 24.3 organisms $/ \mathrm{m} 3$ in May to a yearly high of $2,105.3$ organisms/m3 in October yielding a mean of 742.7 organisms $/ \mathrm{m}^{3}$. Mean total zooplankton densities at Gifford ranged from 30.0 organisms/m3 in February to $8,382.3$ organisms $/ \mathrm{m} 3$ in August with an annual mean of $1,637.2$ organisms $/ \mathrm{m}^{3}$. Mean total zooplankton densities at Hunters ranged from 351.0 organisms/m3 in May to $2,844.5$ organisms $/ \mathrm{m} 3$ in July with a mean of $1,642.3$ organisms $/ \mathrm{m}^{3}$. Porcupine Bay values ranged from 128.7 organisms $/ \mathrm{m} 3$ in March to $7,597.3 / \mathrm{m}^{3}$ in June with an annual mean of $3,335.2 / \mathrm{m}^{3}$. At the confluence of the Spokane River with the mainstream Columbia, densities ranged from a yearly low of 710.3 organisms/m3 in February to a high of $11,291.3$ organisms/m3 in June yielding a yearly average of $3,977.9$ organisms $/ \mathrm{m}^{3}$. Seven Bays densities ranged from a low of 105.4 organisms/m3 in February to a maximum of $5,671.1$ organisms/m3 in September with an annual mean of $3,012.8$ organisms $/ \mathrm{m}^{3}$. Mean total densities at Keller Ferry ranged from 148.8 organisms/m3 in February to $9,492.9$ organisms $/ \mathrm{m}^{3}$ in June yielding an average of 3,590.8 organisms $/ \mathrm{m}^{3}$. Mean total densities at San Poil ranged from 4,336.4 organisms/m3 in October to $11,106.1$ organisms/m3 in July for a mean of 7.721.3 organisms $/ \mathrm{m}^{3}$. Mean total densities at Spring Canyon ranged from a yearly low of 788.2 organisms $/ \mathrm{m}^{3}$ in February to a maximum of 10,364.0 organisms/ $\mathrm{m}^{3}$ in June for a yearly mean of $4,427.8$ organisms $/ \mathrm{m}^{3}$. Finally, Rufus Woods mean total zooplankton densities ranged from 967,3 organisms/m3 in March to 2,557.0 organisms $/ \mathrm{m}^{3}$ in July with an annual mean of $1,485.8$ organisms $/ \mathrm{m}^{3}$.

# Table 3.2 Synoptic list of zooplankton taxa historically identified in Lake Roosevelt including those identified during the 1995 study period. 

\author{
Phylum Anthropoda <br> Class Crustacea <br> Subclass Brachiopoda <br> Order Cladocera <br> Family Daphnidae <br> 1. Ceriodaphnia quadranqula <br> 2. *Daphnia galeata mendotae <br> 3. *Daphnia retroarva <br> 4. 'Daphnia schdlen' <br> 5. *Daphniathorata <br> 6. Simocephalus serrulatus <br> Family Chydoridae <br> 7. Alona guttata <br> 8. Alona quadrangularis <br> 9. Chydorus sphaericus <br> Family Sididae <br> 10. *Diaphanosoma brachywwn <br> 11. Diaphanosoma birgi <br> 12 Sida crystallina <br> Family Bosminidae <br> 13. *Bosmina longirostris <br> Family Leptodoriidae <br> 14. *Leptodora kindtii <br> Subclass Copepoda <br> Order Eucopepoda <br> Suborder Calanoida <br> FamilyDiaptomidae <br> 15. *Leptodiaptomus ashlandi <br> 16. Skistodiaptomus oregonensis <br> Family Temoridae <br> 17. *Epischura nevadensis <br> Suborder Cyclopoida <br> Family Cyclopoidae <br> 18. *Diacyclops biaspidatus thomasi <br> 19. *Mesocyclop edax <br> Suborder Harpacticoida <br> Family Harpacticoidae <br> 20. Bryocamptus spp. <br> [^0]}

Table 3.3 Mean monthly densities $\left(\# / \mathrm{m}^{3}\right)$ for representative zooplankton at Kettle Falls (Index Station 1), in 1995.

'. -' Indicates no data collected or no value.

Table 3.4 Mean monthly densities (\#/m³) for representative zooplankton at Gifford (Index Station 2), in 1995.


Table 3.5 Mean monthly densities $\left(\# / \mathbf{m}^{3}\right)$ for representative zooplankton at Hunters (Index Station 3), in 1995.


Table 3.6 Mean monthly densities (\#/m³) for representative zooplankton at Porcupine Bay (Index Station 4), in 1995.

'- -' Indicates no data collected or no value.

Table 3.7 Mean monthly densities $\left(\# / \mathrm{m}^{3}\right)$ for representative zooplankton at the confluence of the Spokane River with the main stem Columbia (Index Station Confluence), in 1995.

|  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total <br> Zooplankton |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | $338.8 \pm 56.8$ | $0.0 \pm-$ | $1,101.6 \pm 405.4$ | $1,007.0 \pm 279.8$ | $0.0 \pm-$ | $2,108.6 \pm 656.9$ |
| Feb | $78.3 \pm 30.2$ | $0.0 \pm-$ | $127.3 \pm 37.8$ | $508.2 \pm 3.5$ | $74.8 \pm 26.3$ | $710.3 \pm 19.3$ |
| Mar | $\mathbf{0 . 0} \pm-$ | $\mathbf{0 . 0} \pm-$ | $108.8 \pm 22.2$ | $2,178.4 \pm 16.4$ | $56.1 \pm 20.4$ | $2,343.2 \pm 12.8$ |
| Apr | $\mathbf{0 . 0} \pm-$ | $1.7 \pm 2.9$ | $90.1 \pm 17.9$ | $1,019.5 \pm 163.1$ | $137.6 \pm 27.0$ | $1,247.2 \pm 176.2$ |
| Jun | $822.4 \pm 132.6$ | $71.4 \pm 15.3$ | $1,075.6 \pm 145.7$ | $10,132.5 \pm 318.2$ | $83.3 \pm 15.6$ | $11,291.3 \pm 420.9$ |
| Aug | $4,049.2 \pm 700.0$ | $76.5 \pm 17.7$ | $4,144.4 \pm 705.5$ | $1,633.0 \pm 535.8$ | $1.7 \pm 2.9$ | $5,779.1 \pm 1,238.7$ |
| Sep | $\mathbf{1 , 1 7 4 . 2} \pm 44.0$ | $3.4 \pm 2.9$ | $1,323.7 \pm 16.4$ | $2,745.9 \pm 118.2$ | $294.0 \pm 29.4$ | $4,363.6 \pm 102.3$ |
| Mean | $923.3 \pm 192.7$ | $\mathbf{2 1 . 9} \pm 9.7$ | $\mathbf{1 , 1 3 8 . 8} \pm \mathbf{1 9 3 . 0}$ | $\mathbf{2 , 7 4 6 . 4} \pm \mathbf{2 0 5 . 0}$ | $\mathbf{9 2 . 5} \pm \mathbf{2 0 . 3} \mathbf{3 , 9 7 7 . 6} \pm \mathbf{3 7 5 . 3}$ |  |
| '. - Indicates no data collected or no value. |  |  |  |  |  |  |

Table 3.8 Mean monthly densities $\left(\# / \mathrm{m}^{3}\right)$ for representative zooplankton at Seven Bays (Index Station 6), in 1995.

|  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total Zooplankton |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | $66.1 \pm 6.1$ | $0.0 \pm \ldots$ | $107.3 \pm 9.2$ | $390.8 \pm 40.8$ | $50.9 \pm 16.8$ | $548.9 \pm 51.4$ |
| Feb | $12.7 \pm 3.7$ | $0.0 \pm \cdots$ | $23.0 \pm 4.0$ | $70.8 \pm 10.8$ | $11.6 \pm 4.0$ | $105.4 \pm 14.1$ |
| Mar | $0.9 \pm 1.5$ | $0.0 \pm \cdots$ | $121.5 \pm 33.7$ | $767.2 \pm 62.4$ | $147.8 \pm 5.1$ | $1,036.5 \pm 59.7$ |
| Apr | $3.4 \pm 1.5$ | $0.0 \pm \cdots$ | $752.8 \pm 97.1$ | $704.3 \pm 134.5$ | $338.1 \pm 74.7$ | 1,795.2 $\pm 80.9$ |
| May | 138.9 $\pm 25.5$ | $6.4 \pm 7.7$ | $296.9 \pm 42.3$ | $4,716.6 \pm 349.2$ | $114.7 \pm 83.7$ | $5,128.2 \pm 434.6$ |
| Jun | 920.1446 .5 | $79.9 \pm 2.9$ | $1,082.4 \pm 49.2$ | $3,222.6 \pm 233.4$ | $36.5 \pm 8.2$ | $4,341.5 \pm 288.7$ |
| Jul | 1,758.7 $\pm 27.0$ | $61.2 \pm 7.8$ | $1,852.2 \pm 15.6$ | $3,060.3 \pm 260.3$ | $13.6 \pm 5.9$ | $4,926.0 \pm 265.1$ |
| Aug | $1,855.6 \pm 238.8$ | $5.1 \pm 5.4$ | 1,900.6 $\pm 249.6$ | $1,559.9 \pm 158.9$ | $178.4 \pm 19.4$ | $3,638.9 \pm 373.7$ |
| Sep | $1,662.7 \pm 167.7$ | $5.1 \pm 5.4$ | $1,712.8 \pm 167.9$ | $3,683.1 \pm 183.9$ | $275.3 \pm 55.2$ | $5,671.1 \pm 342.1$ |
| Oct | $1,192.4 \pm 97.4$ | $0.0 \pm \cdots$ | $1,254.5 \pm 97.1$ | $1,644.9 \pm 124.6$ | $36.5 \pm 6.4$ | $2,935.8 \pm 89.7$ |
| Mean | $761.2 \pm 61.6$ | $15.8 \pm 5.8$ | $910.4 \pm 76.6$ | $1,982.1 \pm 155.9$ | $20.3 \pm 27.9$ | $012.8 \pm 200.0$ |

'- -' Indicates no data collected or no value.

Table 3.9 Mean monthly densities $\left(\# / \mathbf{m}^{3}\right)$ for representative zooplankton at Keller Ferry (Index Station 7), in 1995.

|  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total <br> Zooplankton |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | $266.9 \pm 14.3$ | $0.0 \pm-$ | $289.1 \pm 27.8$ | $1,054.9 \pm 115.9$ | $11.1 \pm 8.8$ | $1,355.1 \pm 113.1$ |
| Feb | $35.0 \pm 7.4$ | $0.0 \pm-$ | $39.8 \pm 6.8$ | $101.8 \pm 16.2$ | $7.2 \pm 1.7$ | $148.8 \pm 23.8$ |
| Mar | $0.2 \pm 0.2$ | $0.0 \pm-$ | $121.7 \pm 13.8$ | $496.1 \pm 67.2$ | $213.2 \pm 9.1$ | $831.0 \pm 75.4$ |
| Apr | $4.3 \pm 3.0$ | $0.0 \pm-$ | $876.0 \pm 59.0$ | $1,028.0 \pm 62.0$ | $232.8 \pm 41.0$ | $2,136.8 \pm 90.8$ |
| May | $83.3 \pm 10.6$ | $0.0 \pm-$ | $100.3 \pm 15.3$ | $2,421.4 \pm 367.4$ | $62.9 \pm 54.5$ | $2,584.5 \pm 336.2$ |
| Jun | $3,724.9 \pm 3,671$ | $42.5 \pm 11.5$ | $3,810.7 \pm 3,680$ | $5,680.5 \pm 379.4$ | $1.7 \pm 1.5$ | $9,492.9 \pm 3,686$ |
| Jul | $2,729.0 \pm 419.6$ | $30.6 \pm 19.2$ | $2,895.5 \pm 475.0$ | $4,579.4 \pm 413.5$ | $99.4 \pm 27.6$ | $7,574.3 \pm 716.7$ |
| Aug | $2,463.9 \pm 185.2$ | $18.7 \pm 10.3$ | $2,486.8 \pm 183.4$ | $2,969.4 \pm 255.4$ | $149.5 \pm 33.2$ | $5,605.7 \pm 292.8$ |
| Sep | $841.1 \pm 87.8$ | $2.6 \pm 4.4$ | $858.1 \pm 83.1$ | $3,994.9 \pm 235.2$ | $190.3 \pm 51.0$ | $5,043.3 \pm 222.3$ |
| Oct | $33.1 \pm 18.5$ | $0.0 \pm-$ | $57.8 \pm 18.8$ | $\mathbf{1 , 0 1 7 . 8} \pm 207.3$ | $60.3 \pm 15.6$ | $1,135.9 \pm 195.5$ |
| Mean | $\mathbf{1 , 0 1 8 . 2} \pm 441.7$ | $9.4 \pm \mathbf{1 1 . 4}$ | $\mathbf{1 , 1 5 3 . 6} \pm \mathbf{4 5 6 . 3} \mathbf{2 , 3 3 4 . 4} \pm \mathbf{2 1 2 . 0}$ | $\mathbf{1 0 2 . 8} \pm \mathbf{2 4 . 4}$ | $\mathbf{3 , 5 9 0 . 8} \pm \mathbf{5 7 5 . 3}$ |  |
| O - Indicates no data collected or no value. |  |  |  |  |  |  |

Table 3.10 Mean monthly densities (\#/m) for representative zooplankton at San Poil (Index Station 8), in 1995.

|  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total <br> Zooplankton |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Jul | $4,463.9 \pm 341.3$ | $25.5 \pm 22.2$ | $4,519.9 \pm 379.6$ | $6,542.0 \pm 76.5$ | $44.2 \pm 23.6$ | $11,106.1 \pm 299.3$ |
| Oct | $503.0 \pm 99.7$ | $0.0 \pm-$ | $520.0 \pm 91.9$ | $3,768.9 \pm 277.8$ | $47.6 \pm 12.8$ | $4,336.4 \pm 262.4$ |
| Mean | $\mathbf{2 , 4 8 3 . 5} \pm \mathbf{2 2 0 . 5}$ | $\mathbf{1 2 . 8} \pm \mathbf{2 2 . 2}$ | $2,520.0 \pm \mathbf{2 3 5 . 8}$ | $\mathbf{5 , 1 5 5 . 5} \pm \mathbf{1 7 7 . 2}$ | $45.9 \pm \mathbf{1 8 . 2}$ | $\mathbf{7 , 7 2 1 . 3} \pm \mathbf{2 8 0 . 9}$ |
| '.-' Indicates no data collected or no value. |  |  |  |  |  |  |

Table 3.11 Mean monthly densities $\left(\# / \mathbf{m}^{\mathbf{3}}\right)$ for representative zooplankton at Spring Canyon (Index Station 9), in 1995.

|  |  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total Zooplankton |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jan | $192.2 \pm 44.8$ | $0.0 \pm$ - | $224.9 \pm 34.8$ | $1,180.5 \pm 226.4$ | $51.4 \pm 57.0$ | $1,456.8 \pm 185.9$ |
|  | Feb | $138.2 \pm 33.0$ | $0.0 \pm$ - | $164.0 \pm 45.5$ | $573.6 \pm 253.3$ | $50.6 \pm 26.2$ | $788.2 \pm 318.3$ |
|  | Mar | $16.1 \pm 7.9$ | $0.0 \pm$ - | $123.2 \pm 43.8$ | $943.1 \pm 108.2$ | $124.0 \pm 24.7$ | $1,190.3 \pm 161.0$ |
|  | Apr | $63.7 \pm 20.3$ | $0.0 \pm$ - | $293.1 \pm 64.8$ | $1,900.6 \pm 129.6$ | $108.8 \pm 15.6$ | 2,302.4 $\pm 106.0$ |
| N | May | $69.3 \pm 0.0$ | $7.1 \pm 2.8$ | $161.1 \pm 19.1$ | 2,787.4 $\pm 599.1$ | $76.5 \pm 19.1$ | $3,025.0 \pm 610.7$ |
|  | Jun | $3,127.8 \pm 118.8$ | $64.6 \pm 8.9$ | $3,258.7 \pm 160.3$ | $7,087.4 \pm 570.6$ | $17.8 \pm 7.2$ | 10,364.0 $\pm 719.6$ |
|  | Jul | $3,806.3 \pm 462.3$ | $16.1 \pm 10.3$ | $3,942.9 \pm 456.1$ | $5,608.3 \pm 860.0$ | $113.0 \pm 41.1$ | $9,665.2 \pm 1,330.0$ |
|  | Aug | $1,562.4 \pm 281.7$ | $0.9 \pm 1.5$ | $1,579.4 \pm 283.2$ | 3,963.4 $\pm 434.7$ | $248.9 \pm 47.2$ | $5,791.8 \pm 601.7$ |
|  | Sep | $1,870.0 \pm 255.4$ | $1.7 \pm 2.9$ | $1,884.4 \pm 253.6$ | 5,214.9 $\pm 708.4$ | $270.2 \pm 93.9$ | $7,369.5 \pm 729.2$ |
|  | Oct | $186.9 \pm 49.5$ | $0.0 \pm$ - | $247.2 \pm 60.6$ | 2,065.4 $\pm 272.8$ | $11.9 \pm 1.5$ | $2,324.5 \pm 331.1$ |
|  | Mean | ,103.3 $\pm 127.4$ | $9.0 \pm 5.3$ | $187.9 \pm 142.23$ | 3,132.5 $\pm 416.3$ | $107.3 \pm 33.4$ | 4,427.8 $\pm 509.4$ |

[^1]Table 3.12 Mean monthly densities $\left(\# / \mathbf{m}^{3}\right)$ for representative zooplankton at Rufus Woods (Index Station 10), in 1995.

|  | Daphnia | Leptodora | Cladocera | Copepoda | Nauplii | Total <br> Zooplankton |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | $213.1 \pm 68.7$ | $0.0 \pm \ldots$ | $258.0 \pm 82.0$ | $1,523.4 \pm 105.7$ | $35.5 \pm 5.3$ | $1,819.5 \pm 179.8$ |
| Feb | $172.9 \pm 23.1$ | $0.0 \pm-$ | $239.5 \pm 61.1$ | $878.5 \pm 132.2$ | $67.8 \pm 13.2$ | $1,185.7 \pm 170.2$ |
| Mar | $25.2 \pm 1.3$ | $0.0 \pm \ldots$ | $67.3 \pm 10.6$ | $865.4 \pm 55.5$ | $34.6 \pm 17.2$ | $967.3 \pm 27.8$ |
| $\mathbf{w}$ | Apr | $9.8 \pm 9.9$ | $0.0 \pm \ldots$ | $84.1 \pm 11.9$ | $1,202.8 \pm 210.2$ | $67.3 \pm 7.9$ |
| May | $20.6 \pm 2.6$ | $0.0 \pm-$ | $44.9 \pm 10.6$ | $826.2 \pm 259.1$ | $24.3 \pm 23.8$ | $895.3 \pm 224.7$ |
| Jun | $856.1 \pm 195.6$ | $56.1 \pm 7.9$ | $946.7 \pm 212.8$ | $1,454.2 \pm 420.3$ | $0.0 \pm-$ | $2,400.9 \pm 633.1$ |
| Jul | $820.6 \pm 304.0$ | $0.0 \pm-$ | $859.8 \pm 317.2$ | $1,697.2 \pm 523.4$ | $0.0 \pm-$ | $2,557.0 \pm 840.6$ |
| Aug | $403.7 \pm 66.1$ | $0.9 \pm 1.3$ | $417.8 \pm 59.5$ | $716.8 \pm 75.3$ | $22.4 \pm 5.3$ | $1,159.7 \pm 23.8$ |
| Sep | $329.0 \pm 31.7$ | $0.0 \pm-$ | $336.4 \pm 31.7$ | $864.5 \pm 91.2$ | $0.0 \pm-$ | $1,200.9 \pm 85.9$ |
| Oct | $52.3 \pm 31.7$ | $0.0 \pm-$ | $76.6 \pm 34.4$ | $1,211.2 \pm 401.8$ | $29.9 \pm 26.4$ | $1,317.7 \pm 409.7$ |
| Mean | $290.3 \pm 73.5$ | $5.7 \pm 4.6$ | $333.1 \pm 83.2$ | $\mathbf{1 , 1 2 4 . 0} \pm \mathbf{2 2 7 . 5}$ | $\mathbf{2 8 . 2} \pm \mathbf{9 . 9}$ | $\mathbf{1 , 4 8 5 . 8} \pm \mathbf{2 8 2 . 2}$ |

'- -' Indicates no data collected or no value.

In 1995, the reservoir experienced one large peak in Daphnia spp. densities. This peak occurred between May and September and seemed to start at the lower end of the reservoir first and progress upstream with time (Tables 3.3-3.12). 1995 zooplankton population dynamics appear to be similar to those seen in 1994. There was a considerable difference in the densities of zooplankton between sample areas ranging from a yearly average low of $1,637.2$ organisms $/ \mathrm{m} 3$ at Gifford to a yearly average high of $4,427.8$ organisms $/ \mathrm{m} 3$ at Spring Canyon (Tables 3.3-3.12). The highest recorded Daphnia spp. density was 6380.6 organisms $/ \mathrm{m} 3$ at Gifford in August followed by a value of $4,049.2 / \mathrm{m}^{3}$ at the Confluence site in August. Total zooplankton values were highest at the Confluence site ( $11,291.3 / \mathrm{m}^{3}$ ) in June, followed by densities of $11,106.1 / \mathrm{m}^{3}$ in July at the San Poil River. Yearly low Daphnia spp. densities of $0.0 / \mathrm{m}^{3}$ were recorded at Gifford in January, Porcupine Bay in March and at the Confluence site in March and April. Total zooplankton densities were lowest at Gifford with a value of 30.0 organisms $/ \mathrm{m} 3$ recorded in March.

### 3.2.2 Zooplankton Biomass

Monthly mean biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ values for cladocera collected at Kettle Falls, Gifford, Hunters, Porcupine Bay, Confluence, Seven Bays, Keller Ferry, San Poil, Spring Canyon and Rufus Woods are shown in Tables 3.13 through 3.22. Mean biomass by species for each location can be found in Appendix B. Total cladocera biomass at Kettle Falls averaged $50.1 \mu \mathrm{~g} / \mathrm{m}^{3}$ and ranged from $0.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in May to $148.9 \mu \mathrm{~g} / \mathrm{m}^{3}$ (Table 3.13). Gifford biomass averaged $550.5 \mu \mathrm{~g} / \mathrm{m}^{3}$ for the year and ranged from $0.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in January to $4,005.8 \mu \mathrm{~g} / \mathrm{m}^{3}$ in August (Table 3.14). Total cladocera biomass at Hunters averaged $671.9 \mu \mathrm{~g} / \mathrm{m}^{3}$ and ranged from $0.1 \mu \mathrm{~g} / \mathrm{m}^{3}$ in May to $1,827.6 \mu \mathrm{~g} / \mathrm{m}^{3}$ in October (Table 3.15). Porcupine Bay averaged $494.2 \mu \mathrm{~g} / \mathrm{m}^{3}$ for the year and ranged from $0.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in March to $2,085 \mu \mathrm{~g} / \mathrm{m}^{3}$ in June (Table 3.16). The Confluence site averaged $1,719 \mu \mathrm{~g} / \mathrm{m}^{3}$ and ranged from $0.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in March and April to $8,690.6 \mu \mathrm{~g} / \mathrm{m}^{3}$ in August (Table 3.17). Total cladocera biomass at Seven Bays averaged $1,719.5 \mu \mathrm{~g} / \mathrm{m}^{3}$ for the year and ranged from $0.1 \mu \mathrm{~g} / \mathrm{m}^{3}$ in March to $3,422.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in July (Table 3.18). Keller Ferry total cladocera values ranged from $0.1 \mu \mathrm{~g} / \mathrm{m}^{3}$ in March to 1,781 in August with an average of $513.3 \mu \mathrm{~g} / \mathrm{m}^{3}$ (Table 3.19). San Poil total cladocera values ranged from $61.4 \mu \mathrm{~g} / \mathrm{m}^{3}$ in October to $2,414.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in July with a mean of $1,237.7 \mu \mathrm{~g} / \mathrm{m}^{3}$ (Table 3.20). Spring Canyon total cladocera biomass values ranged from $3.4 \mu \mathrm{~g} / \mathrm{m}^{3}$ in April to $6,085.8 \mu \mathrm{~g} / \mathrm{m}^{3}$ in June yielding an average of $839.8 \mu \mathrm{~g} / \mathrm{m}^{3}$ (Table 3.21). Finally, Rufus Woods total cladocera biomass values averaged $473.3 \mu \mathrm{~g} / \mathrm{m}^{3}$ and ranged from $1.0 \mu \mathrm{~g} / \mathrm{m}^{3}$ in April to $4,051.3 \mu \mathrm{~g} / \mathrm{m}^{3}$ in June (Table 3.22)

Table 3.13 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{\mathbf{3}}$ at Kettle Falls (Index Station 1), in 1995.

|  | May | Jul | 0ct | Mean |
| :--- | :---: | :---: | :---: | :---: |
| Daphnia Spp. | 0.0 | 0.2 | 148.9 | 49.7 |
| Leptodora kindtii | 0.0 | 1.1 | 0.0 | 0.4 |
| Total Cladocera | 0.0 | 1.3 | 148.9 | 50.1 |

Table 3.14 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathbf{m}^{\mathbf{3}}$ at Gifford (Index Station 2), in 1995.

Jan Feb Apr May Jun Jul Aug Sep 0ct Mean
$\begin{array}{lllllllllllll}\text { Daphnia } & \text { Spp. } & 0.0 & 0.1 & 0.0 & 0.1 & 0.1 & 111.9 & 903.2 & 47.3 & 86.1 & 127.6\end{array}$
L. $\begin{array}{lllllllllll} & k i n d t i i ~ & 0.0 & 0.0 & 0.0 & 2.3 & 13.0 & 338.0 & 3,102.5 & 0.0 & 349.5\end{array} 422.3$

| Total Cladocera | 0.0 | 0.1 | 0.0 | 2.4 | 13.1 | 449.9 | $4,005.8$ | 47.3 | 435.6 | 550.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 3.15 Monthiy mean zooplankton biomass values in $\mu \mathrm{g} / \mathbf{m}^{\mathbf{3}}$ at Hunters (Index Station 3), in 1995.

|  | May | Jul | 0ct | Mean |
| :--- | :---: | :---: | :---: | :---: |
| Daphnia Spp. | 0.1 | 77.5 | 187.9 | 88.5 |
| Leptodora kindtii | 0.0 | $1,750.2$ | 0.0 | 583.4 |
| Total Cladocera | 0.1 | $1,827.6$ | 187.9 | 671.9 |

Table 3.16 Monthly mean zooplankton biomass values in $\mu \mathbf{g} / \mathbf{m}^{\mathbf{3}}$ at Porcupine Bay (Index Station 4), in 1995.

Jan Feb Mar Apr May Jun Jul Aug Sep 0ct Mean

L. kindtii $0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 0.0 \quad 1,923 \quad 1,3730.0$

| Tot Cladocera | 2.4 | 13.2 | 0.0 | 0.1 | 4.6 | 2,085 | 1,467 | 262.1 | 770.1 | 337.9 | 494.2 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 3.17 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{3}$ at the confluence of the Spokane River with the mainstem Columbia (Index Station Confluence), in 1995.

|  | Jan | Feb | Mar | Apr | Jun | Aug | Sep | Mean |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Daphnia Spp. | 43.5 | 7.1 | 0.0 | 0.0 | 98.8 | $1,279.8$ | 327.0 | 250.9 |
| L. kindtii | 0.0 | 0.0 | 0.0 | 0.0 | $2,690.8$ | $7,410.8$ | 178.7 | $\mathbf{1 , 4 6 8 . 6}$ |
| Total Cladocera | 43.5 | 7.1 | 0.0 | 0.0 | $2,789.5$ | $8,690.6$ | 505.7 | $\mathbf{1 , 7 1 9 . 5}$ |

Table 3.18 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathbf{m}^{\mathbf{3}}$ at Seven Bays (Index Station 6), in 1995.

Jan Feb Mar Apr May Jun Jul Aug Sep 0ct Mean
Daphnia Spp. 18.41 .0
$\begin{array}{llllllllllll}\text { L. kindtii } & 0.0 & 0.0 & 0.0 & 0.0 & 69.7 & 2,799 & 3,293 & 459.7 & 156.1 & 0.0 & 677.7\end{array}$

| Tot Cladocera | 18.4 | 1.0 | 0.1 | 0.4 | 73.6 | 2,928 | 3,422 | 713.9 | 869.3292 .8 | 831.9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 3.19 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathrm{m}^{\mathbf{3}}$ at
Keller (Index Station 7), in 1995.

Jan Feb Mar Apr May Jun Jul Aug Sep Oct Mean
Daphnia Spp. $41.510 .0 \quad 0.1 \quad 0.3 \quad 3.4 \quad 555.6 \quad 126.4960 .5 \quad 209.8$
L. kindtii $\quad 0.0 \quad 0.0 \quad 0.0$

| Tot Cladocera | 41.5 | 10.0 | 0.1 | 0.3 | 3,4 | 1,274 | 1,531 | 1,781 | 490.1 | 1.9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 3.20 Monthly mean zooplankton biomass values in $\mu \mathbf{g} / \mathbf{m}^{\mathbf{3}}$ at San Poil (Index Station 8), in 1995.

|  | Jul | 0ct | Mean |
| :--- | :---: | :---: | :---: |
| Daphnia S pp. | 525.2 | 61.4 | 293.3 |
| Leptodora kindtii | $1,888.8$ | 0.0 | 944.4 |
| Total Cladocera | $2,414.0$ | 61.4 | $\mathbf{1 , 2 3 7 . 7}$ |


| Table 3.21 | Monthly <br> Spring | mean <br> Canyo | $\begin{aligned} & \text { in zoo } \\ & \text { on (In } \end{aligned}$ | ooplank Index | kton bio Station | aso 9), in | values $1995 .$ |  | $\mathrm{m}^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jan Feb | Mar | Apr | r May | Jun | Jul | Aug | Sep | Oct | Mean |
| Daphnia Spp. | 35.139 .3 | 7.0 | 3.4 | 2.8 | 1,013.4 | 541.5 | 597.4 | 605.1 | 17.5 | 286.3 |
| L. kindtii | $0.0 \quad 0.0$ | 0.0 | 0.0 | 28.5 | 5,072.4 | 397.7 | 7.0 | 30.3 |  | 553.6 |
| Tot Cladocera | 35.139 .3 | 7.0 | 3.43 | 31.36 | 6,085.8 9 | 939.2 | 604.4 | 635.3 | 17.5 | 839.8 |

Table 3.22 Monthly mean zooplankton biomass values in $\mu \mathrm{g} / \mathbf{m}^{\mathbf{3}}$ at Rufus Woods (Index Station 10), in 1995.

|  | Jan | Feb | Mar | Apr | M | Jun | Jul | Aug | Sep |  | n |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Daphnia Spp. | 88.1 | 32.9 | 1.9 | 1.0 | 1.7 | 194.0 | 124.3 | 159.8 | 77.3 | 2.7 | 68.4 |
| L. kindtii | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3,857.3 | 30.0 | 192.2 | 0.0 | 0.0 | 405.0 |
| Tot Cladocera | 88.1 | 32.9 | 1.9 | 1.0 | 1.7 | 4,051.3 | 124.3 | 352.1 | 77.3 | 2.7 | 473.3 |

### 3.2.3 Zooplankton Lengths

Lengths in millimeters were taken from randomly selected cladocera collected at Kettle Falls, Gifford, Hunters, Porcupine Bay, Confluence, Seven Bays, Keller Ferry, San Poil and Spring Canyon in 1995. The results of these measurements are shown in Tables 3.23 through 3.32. Length ranges and mean lengths by species at four sampling stations are located in Appendix B. Mean lengths of select cladocera at Kettle Falls were: Daphnia galeata mendotae - 0.20 mm ; Daphnia retrocurva - 0.20 mm ; Daphnia schødleri- 0.25 mm ; and Leptodora kindtii -6.10 mm . Mean lengths at Gifford were: Daphnia galeata mendotae -0.30 mm ; Daphnia retrocurva - 0.15 mm ; Daphnia sch $\phi$ dleri- 0.23 mm ; and Leptodora kindtii - 4.10 mm . Mean lengths at Hunters were: Daphnia galeata mendotae 0.25 mm ; Daphnia retrocurva - 0.20 mm ; Daphnia schødleri -0.27 mm ; and Leptodora kindtii -6.00 mm . Mean lengths at Porcupine Bay were: Daphnia galeata mendotae - 0.40 mm ; Daphnia retrocurva -0.20 mm ; Daphnia schodleri- 0.28 mm ; and Leptodora kindtii - 4.70 mm . Mean lengths at the Confluence site were: Daphnia galeata mendotae - 0.20 mm ; Daphnia retrocurva -0.20 mm ; Daphnia schødleri-0.28 mm; and Leptodora kindtii - 6.30 mm . Mean lengths at Seven Bays were: Daphnia galeata mendotae -0.20 mm ; Daphnia retrocurva-0.18 mm; Daphnia schødleri-0.25 mm; and Leptodora kindtii 5.55 mm . Mean lengths at Keller Ferry were: Daphnia galeata mendotae -0.30 mm ; Daphnia retrocurva - 0.23 mm ; Daphnia schødleri- 0.27 mm ; and Leptodora kindtii 5.50 mm . Mean lengths at San Poil were: Daphnia retrocurva - 0.10 mm ; Daphnia schødleri -0.20 mm ; and Leptodora kindtii -6.80 mm . Mean lengths at Spring Canyon were: Daphnia galeata mendotae -0.23 mm ; Daphnia retrocurva - 0.23 mm ; Daphnia schødleri-0.28 mm; and Leptodora kindtii -4.14 mm . Mean lengths at Rufus Woods were: Daphnia galeata mendotae - 0.50 mm ; Daphnia retrocurva - 0.23 mm ; Daphnia schødleri - 0.26 mm ; and Leptodora kindtii -6.90 mm . Overall, 1995 mean lengths were smaller than those reported in the 1994 report (Shields and Underwood, 1996).

Table 3.23 Monthly mean zooplankton lengths in mm for select cladocera at Kettle Falls (Index Station I), in 1995.

|  | May |  |  | Jul |  |  | Oct |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| D.g.mendota | - | $\pm$ | - | 0.2 | $\pm$ | 0.1 | 0.2 | $\pm$ | 0.1 |
| D.pulicaria | - | $\pm$ | - | - | $\pm$ | - | - | $\pm$ | - |
| D.retrocurva | - | $\pm$ | - | 0.2 | $\pm$ | 0.1 | 0.2 | $\pm$ | 0.1 |
| D.sch\#dleri | - | $\pm$ | - | 0.2 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 |
| D.thorata | - | $\pm$ | - | - | $\pm$ | - | - | $\pm$ | - |
| L. $k$ indtii | - | $\pm$ | - | 6.1 | $\pm$ | 2.4 | - | $\pm$ | - |
| Total Clad | ocera | - $\pm$ | - | 0.2 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 |

'-_' Indicates no data or no organisms found.

Table $3.24 \begin{aligned} & \text { Monthly mean zooplankton lengths in } \mathbf{m m} \text { for select cladocera at Gifford (Index Station 2), } \\ & \text { in 1995. }\end{aligned}$


Table 3.25 Monthly mean zooplankton lengths in mm for select cladocera at Hunters (Index Station 3), in 1995.

|  |  | May |  |  | Jul |  | Oct |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| D.g.mendota | - | $\pm$ | - | 0.2 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 |
| D.pulicaria | - | $\pm$ | - | - | $\pm$ | - | - | $\pm$ | - |
| D.retrocurva | - | $\pm$ | - | 0.2 | $\pm$ | 0.0 | 0.2 | $\pm$ | 0.0 |
| D.schødleri | 0.2 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 |
| D. thorata | - | $\pm$ | - | - | $\pm$ | - | - | $\pm$ | - |
| L.kindtii | - | $\pm$ | - | 6.0 | $\pm$ | 1.9 | - | $\pm$ | - |
| Total Cladocera | $\mathbf{0 . 2}$ | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 | 0.3 | $\pm$ | 0.1 |

'--' Indicates no data or no organisms found.

Table 3.26 Monthly mean zooplankton lengths in mm for select cladocera at Porcupine Bay (Index Station 4), in 1995.


Total Cladocera $0.2 \pm 0.10 .4 \pm 0.10 .2 \pm-0.3 \pm 0.10 .2 \pm 0.10 .2 \pm 0.10 .3 \pm 0.10 .3 \pm 0.10 .3 \pm 0.1$
-_- Indicates no data or no organisms found.

Table 3.27 Monthly mean zooplankton lengths in mm for select cladocera at the confluence of the Spokane River with the mainstem Columbia (Index Station Confluence), in 1995.

|  | Jan | Feb | Mar | Apr | Jun | Aug | Sep |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| D.g.mendota | $\pm$ | $\pm$ | $\pm$ |  | $\pm$ | - | $\pm$ - |
| D.pulicaria | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ |
| D.retrocurva | $\pm$ |  |  |  | $\pm 0$. | $\pm$ | $\pm 0.1$ |
| D.schødleri | $0.3 \pm 0$ | $\pm$ | $\pm 0$ |  | $\pm$ | $\pm$ | $\pm 0.1$ |
| D. thorata | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ |
| L. kindtii | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm \pm$ | $\pm 1.4$ |

$$
\begin{array}{cccccccccccccc}
\text { Total Cladocera } 0.3 \pm 0.1 & 0.2 \pm 0.1 & 0.3 \pm 0.1- \pm-0.2 \pm 0.1 & 0.3 \pm 0.1 & 0.3 \pm 0.1
\end{array}
$$

'-_' Indicates no data or no organisms found.

Table 3.28 Monthly mean zooplankton lengths in mm for select cladocera at Seven Rays (Index Station 6), in 1995.

|  |  | Jan | Feb | Mar | Apr | May | Jun | Jul | ul | Aug | Sep | O | ct |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | D.g.mendota | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ | $0.2 \pm 0$ | $0.2 \pm$ | $\pm 0.1$ | $0.2 \pm$ | $\pm$ |  | $\pm$ |
|  | D.pulicaria | $\pm$ | $\pm$ | $\pm$ | $\pm$ | - $\pm$ | $\pm$ | $\pm$ | $\pm$ | $\pm$ |  |  | $\pm$ |
| $\pm$ | D.retrocurva | - $\pm$ | $\pm$ | $\pm$ | $\pm$ | . $1 \pm 0$ | $0.2 \pm 0$ | $0.2 \pm$ | $\pm 0.1$ | $0.2 \pm$ | $\pm$ |  |  |
|  | D.schodleri | $0.3 \pm 0$ | $\pm 0$. | . $\pm 0$. | $2 \pm 0$. | $2 \pm 0$ | $0.3 \pm 0$ | $0.2 \pm$ | $\pm 0.1$ | $0.2 \pm$ | $\pm$ |  |  |
|  | D. thorata | - $\pm$ | $\pm$ | $\pm$ | $\pm$ | - $\pm$ | $\pm$ | - $\pm$ | $\pm$ | $\pm$ | $\pm$ |  | $\pm$ - |
|  | L. kindtii | -- $\pm$ | $\pm$ | $\pm$ | $\pm$ | - $\pm$ | . $2 \pm 2$ | $6.1 \pm$ | $\pm 2.4$ | $6.4 \pm 3$ | $\pm$ |  | $\pm$ |
|  | Total Cladoce | $0.3 \pm 0$ | $\pm 0$. | $3 \pm 0$. | $2 \pm 0$. | . $\pm 0$ | . $2 \pm 0$ | $0.2 \pm 0$ | $\pm 0.1$ | $0.3 \pm$ | $\pm$ |  |  |

'-_' Indicates no data or no organisms found,

Table 3.29
Monthly, mean zooplankton lengths in mm for select cladocera at Keller Ferry (Index Station 7), in 1995.


Table 3.30 Monthly mean zooplankton lengths in mm for select cladocera at San Poil (Index Station 8), in 1995.

|  |  | Jul |  |  | Oct |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| D.g.mendota | - | $\pm$ | - | - | $\pm$ | - |
| D.pulicaria | - | $\pm$ | - | - | $\pm$ | - |
| D.retrocurva | 0.1 | $\pm$ | 0.1 | - | $\pm$ | - |
| D.schødleri | 0.2 | $\pm$ | 0.1 | 0.2 | $\pm$ | 0.1 |
| D.thorata | - | $\pm$ | - | - | $\pm$ | - |
| L.kindtii | $\mathbf{6 . 8}$ | $\pm$ | $\mathbf{3 . 0}$ | - | $\pm$ | - |
| TotalCladocera | $\mathbf{0 . 2}$ | $\pm$ | $\mathbf{0 . 1}$ | 0.2 | $\pm$ | 0.1 |

'-I Indicates no data or no organisms found.

Table 3.31 $\quad \begin{aligned} & \text { Monthly mean zooplankton lengths in } \mathbf{~ m m} \text { for select cladocera at Spring Canyon (Index }\end{aligned}$ Station 9), in 1995.

$\overline{\text { Total Cladocera } 0.3 \pm 0.10 .3 \pm 0.10 .4 \pm 0.10 .2 \pm 0.10 .2 \pm 0.10 .3 \pm 0.10 .2 \pm 0.10 .3 \pm 0.10 .3 \pm 0.10 .2 \pm 0.1}$

- -' Indicates no data or no organisms found.

Table 3.32 Monthly mean zooplankton lengths in mm for select cladocera at Rufus Woods (Index Station 10), in 1995.

$\overline{\text { Total Cladocera } 0.3 \pm 0.10 .3 \pm 0.10 .2 \pm 0.10 .2 \pm 0.10 .2 \pm 0.10 .3 \pm 0.10 .2 \pm 0.10 .3 \pm 0.10 .3 \pm 0.10 .2 \pm 0.1}$
'-_ Indicates no data or no organisms found.

### 3.3 Rainbow Trout Tagging

In 1995, a total of 12,984 fish were tagged at the Kettle Falls and Seven Bays net-pens in May. Of these fish, 4,995 were released from Kettle Falls and 7,989 were released from Seven Bays. A total of 200 tags were returned from anglers fishing in Lake Roosevelt or below in 1995, yielding an overall recapture rate of $2.0 \%$ (Table 3.33). An analysis of the returns by location shows that 67 tags were returned from fish tagged at Kettle Falls, while 133 tags were returned from fish tagged at Seven Bays. Table 3.35 shows the capture locations and percentage breakdowns for all tags returned from 1995 Kettle Falls releases. Overall, the highest number of returns for Kettle Falls fish came from the Spring Canyon area $(42 \%, \mathrm{n}=25)$. Table 3.36 shows the capture locations and percentage breakdowns for all tags returned from the 1995 Seven Bays releases. The highest number of returns for Seven Bays area came from Seven Bays ( $33 \%, \mathrm{n}=36$ ). Table 3.33 summarizes fish tag recoveries from net-pen tagging efforts on Lake Roosevelt in 1995. Table 3.34 lists rainbow trout release times versus water retention times and their subsequent recapture rates. Based on this information, it can be seen that entrainment rates for rainbow trout appeared to be low in 1995. Total May releases show that only five fish out of a total of 200 returns came from below Lake Roosevelt, yielding a 3\% recapture rate below Grand Coulee Dam.

To determine growth rates of rainbow trout from the time of release to the time of recapture, we plotted days since release versus capture length and weight for the Kettle Falls and Seven Bays tag returns (Figures 3.3, 3.4 and 3.5).

Table 3.33 Summary of release dates, numbers, and subsequent capture locations of net-pen rainbow trout tagged and released from various net pen locations in 1995.

| Tag <br> Location | Release <br> Date | Total <br> \#Tagged | Total \# <br> Recoverad | Percent <br> Recovered | Number <br> Recovered <br> in FDR | Percent <br> Recovered <br> in FDR | \#Recovered <br> in Rufus <br> Woxls | Receries Below Grand Coulee <br> \# Recovered <br> at Rock Is. <br> or McNary | \% Recovered <br> Below <br> FDR |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | May-95 | 4,995 | 67 | $1 \%$ | 65 | $97 \%$ | 2 | 0 | $3 \%$ |
| 6 | May-95 | 7,989 | 133 | $2 \%$ | 131 | $98 \%$ | 2 | 1 | $2 \%$ |

Table $3.34 \quad \begin{aligned} & \text { Summary of rainbow trout release times, water retention times and subsequent } \\ & \text { recapture numbers and percentages by year. }\end{aligned}$

| Release Date | Water Retention Time | Total <br> \# Tagged | Total \# Recovered | Percent Recovered | Number Recovered in FDR | Percent Recovered in FDR | Recoveries Below Grand Coule <br> \# Recovered \# Recovered \% Recovered |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | in Rufus Woods | at Rock Is. or McNary | Below FDR |
|  | 36 |  |  |  |  |  | 0 | 5 |  |
| Mar. 89 | 32 | 768 | 8 | 1\% | 3 | 38\% | 0 | 3 | 63\% |
| Mar. 90 |  | 1,441 | 7 | 0\% | 4 | 57\% |  |  | 43\% |
|  | 48 |  |  |  |  |  |  | 0 | 2\% |
| Mar. 92 | 67 | 3,994 | 1167 | <1\% | 105 | 98\% | 2 | 0 | 0\% |
| Mar. 94 | 55 | 9,994 | 115 | 1\% | 113 | 98\% | 2 | 0 | 2\% |
| Apr. 89 | 33 | 985 | 20 | $2 \%$ | 38 | 55\% | 3 | 4 | 45\% |
| Apr. 90 | 31 | 1,470 | 52 | $3 \%$ | 52 | 73\% | 10 | 13 | 27\% |
| Apr. 91 | 18 | 2,300 | 78 | 3\% |  | 67\% | 13 | 0 | $33 \%$ $2 \%$ |
| Apr. 93 | 87 | 8,998 | 2108 | 1\% | 2484 | 100\% | 0 | 0 | 0\% |
| Apr. 94 | 55 | 7,998 | 123 | 2\% | 121 | 98\% | 2 | 0 | 2\% |
| May 88 | 40 | 1,171 | 99 | 9\% | 44 | 100\% | 0 | 2 | 0\% |
| May 90 | 29 | 1,450 | 54 | 4\% | 283 | 81\% | 8 | 0 | 19\% |
| May 92 | 34 39 | 6,090 | 2095 | 5\% | 64 | 96\% | 42 |  | 1\% |
|  | 44 |  |  |  |  |  |  | 0 | 1\% |
| May 94 | 47 | 182,9834 | 200 | 2\% | 195 | 98\% | 4 | 0 | 3\% |
| Jun. 91 |  | 296 | 32 | 11\% | 27 | 99\% | 5 |  | 1\% |
|  | 29 |  |  | 5\% |  |  | 0 | 0 | 0\% |
| Jun. 92 | 34 | 32090 | 1139 | 4\% | 1139 | 100\% | 0 | 0 | 0\% |
| Jul. 91 | 62 | 1.749 | 155 | 9\% | 148 | 97\% | 7 | 0 | 3\% |



Figure 3.3 Plot of length increases (mm) versus time since release (Days) for tagged rainbow trout at Kettle Falls in 1995.


Figure 3.4 Plot of length increases (mm) versus time since release (Days) for tagged rainbow trout at Seven Bays in 1995.


Figure $3.5 \quad$ Plot of tagged rainbow trout weight increases (g) versus time since release (Days) at Kettle Falls in 1995.

Table $3.35 \quad$ Number and percent of fish captured by location from the Kettle Falls tag releases in 1995.

| CAPTURE LOCATION | \# CAPTURED | \% CAPTURED |
| :---: | :---: | :---: |
| Kettle Falls | 3 | 5.0 |
| Gifford | 2 | 3.0 |
| Hunters | 3 | 5.0 |
| Seven Bays | 8 | 13.0 |
| Keller Ferry | 15 | 25.0 |
| San Poil | 2 | 3.0 |
| Spring Canyon | 25 | 42.0 |
| Rufus Woods | 2 | 3.0 |
| TOTAL | 60 | 100.0 |

Table $3.36 \quad$ Number and percent of fish captured by location from the Seven Bays tag releases in 1995.

| CAPTURE LOCATION | \# CAPTURED | \% CAPTURED |
| :---: | :---: | :---: |
| Porcupine Bay | 4 | 4.0 |
| Seven Bays | 31 | 28.0 |
| Keller Ferry | 36 | 32.0 |
| San Poil | 4 | 4.0 |
| Spring Canyon | 32 | 29.0 |
| Rufus Woods | 2 | $2 . \mathrm{b}$ |
| Rocky Reach Dam | 1 | 1.0 |
| TOTAL | 111 | 100.0 |

### 4.0 DISCUSSION

### 4.1 Reservoir operations

Grand Coulee Dam was commissioned by congress to operate for power, flood control, irrigation, recreation, fisheries and navigation. In 1995, Lake Roosevelt also provided additional water for anadromous fish as a result of the National Marine Fisheries Service's (NMFS) Biological Opinion. During the fall and winter, Grand Coulee was operated with the objective of providing water for anadromous fish by storing water above requirements needed by the power system. System wide flood control was not shifted from Dworshak to Grand Coulee as prescribed by the Biological Opinion due to the small April 1 runoff forecast of 8.6 maf at the Dalles. On October 1, 1994, Lake Roosevelt was at 1,287 feet and was operated above 1,275 feet through early February, 1995. The lake reached a low elevation of $1,253.3$ feet on March 8. The reservoir refilled to $1,269.3$ feet by April 16 before being drafted for flood control down to $1,253.4$ feet on May 5. Lake Roosevelt was then refilled back up to $1,289.0$ feet by early July. Lake operations continued flow augmentation, normal water budget, and Endangered Species Act (ESA), draft requirements until the end of August at which time lake elevations were at $1,281.7$ feet. The lowest elevation reached for ESA operations was $1,280.2$ feet on August 23 representing a draft of 9.8 feet below full pool. The reservoir was then filled to $1,287.3$ feet on September 30. The maximum daily inflow was 166.4 kcfs on June 9, and the highest daily outflow was 155.3 kcfs on June 26 (CRMG, 1995)

### 4.2 Hydrology

A comparison of monthly and annual means for reservoir inflow, outflow, elevation, storage capacity, and water retention time can be seen in Table 4.1 Overall, lake elevations and storage capacity were similar for 1994 and 1995 while reservoir inflows and outflows were greater in 1995 than 1994 due to a slightly above average water year. For 1995, higher water volumes flushing through the system resulted in a shorter yearly mean water retention time of 46.5 days when compared with a yearly average of 54.8 days for 1994. Flood control rule curves were not implemented in 1995 as the result of a near average predicted spring runoff and thus, Lake Roosevelt avoided an extreme spring draw down like that seen in 1991. The 1995 lake elevations from January through May averaged nearly 13 feet below those observed in 1994. This, was due to higher predicted spring runoffs for 1995 when compared with 1994 (Table 4.1). Figures 4.1 and 4.2 show that

Table 4.1 Monthly and annual means for reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in 1994 and 1995.

| Month | Inflow (kcfs) | outflow (kcfs) | Reservoir Elevation ( $\mathbf{F t}$ ) | Storage Capacity (kcfsd) | Water Retention Time (Days) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Jan. 1995 | 73.0 | 88.3 | 1,278.3 | 4,127.8 | 49.3 |
| Jan. 1994 | 81.0 | 77.2 | 1,285.4 | 4,403.6 | 61.8 |
| Feb. 1995 | 81.7 | 94.0 | 1,266.2 | 3,688.9 | 42.6 |
| Feb. 1994 | 97.5 | 103.6 | 1,281.8 | 4,261.2 | 42.5 |
| Mar. 1995 | 101.6 | -90.1 | 1,259.0 | 3,434.3 | 42.4 |
| Mar. 1994 | 67.9 | 77.7 | 1,276.5 | 4,061-1 | 54.9 |
| Apr. 1995 | 81.2 | 84.5 | 1,265.8 | 3,669.5 | 47.5 |
| Apr. 1994 | 89.5 | 73.0 | 1,268.1 | 3,754.4 | 55.0 |
| May 1995 | 112.3 | 93.5 | 1,260.1 | 3,460.4 | 39.4 |
| May 1994 | 112.4 | 99.6 | 1,280.6 | 4,215.0 | 44.0 |
| Jun. 1995 | 148.1 | 117.8 | 1,283.6 | 4,335.3 | 40.1 |
| Jun. 1994 | 133.1 | 135.9 | 1,276.0 | 4,041.3 | 30.1 |
| Jul. 1995 | 111.6 | 110.5 | 1,287.0 | 4,467.4 | 41.4 |
| Jul. 1994 | 101.7 | 95.8 | 1,274.9 | 3,996.1 | 43.5 |
| Aug. 1995 | 96.3 | 91.9 | 1,280.9 | 4,227.8 | 47.2 |
| Aug. 1994 | 82.5 | 73.3 | 1,277.1 | 4,080.0 | 58.7 |
| Sept. 1995 | 79.9 | 65.9 | 1,285.1 | 4,392.8 | 69.0 |
| Sept. 1994 | 67.6 | 55.9 | 1,281.3 | 4,244.6 | 78.4 |
| Oct. 1995 | 80.3 | 80.6 | 1,285.8 | 4,420.3 | 56.7 |
| Oct. 1994 | 61.6 | 64.0 | 1,287.2 | 4,474.9 | 72.6 |
| Nov. 1995 | 97.2 | 91.9 | 1,286.5 | 4,448.3 | 50.4 |
| Nov. 1994 | 75.5 | 75.7 | 1,284.7 | 4,374.9 | 60.1 |
| Dec. 1995 | 135.7 | 141.6 | 1,287.0 | 4,466.7 | 32.4 |
| Dec. 1994 | 85.0 | 83.5 | 1,284.2 | 4,356.8 | 56.3 |
| Annual 1995 | 99.9 | 95.9 | 1,277.1 | 4,095.0 | 46.5 |
| Annual 1994 | 87.9 | 84.6 | 1,279.8 | 4.188 .7 | 54.8 |

reservoir operations produced the lowest mean elevations and water retention times when compared to the years of 1991 through 1995, while 1995 elevations were similar to the elevations in 1992 through 1994.

In addition to lake elevation data, water temperature was measured monthly to correlate temperature with zooplankton densities (Appendix C). Overall trends for temperature data show that temperatures are at yearly minimums in January at all sites with up reservoir sites being the coldest. Yearly maximum temperatures were reached at all sites in July and August.


Figure 4.1 Mean monthly reservoir elevations from 1991 through 1995.


Figure 4.2 Mean monthly water retention times from 1991 through 1995.

### 4.3 Affect of Reservoir Operations on Zooplankton Dynamics

Figures 4.3 and 4.4 show the mean monthly Daphnia and total zooplankton density values at five locations in 1995. These figures indicated that Daphnia and total zooplankton densities remained low throughout the spring months but started to build in April and May in the lower river sections, reaching a peak density in June and July at Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon. After reaching yearly maximums, densities at the lower sites droped off through August and then incresased to a second smaller peak in September. Daphnia and total zooplankton densities at Gifford followed a much different pattern than the lower reservoir sites (Figures 4.5 and 4.6). At Gifford, densities started to build much later in the year, reaching a peak in August followed by a sharp decline by September. An analysis of zooplankton densities indicated that factors such as water temperature played a major role in zooplankton population dynamics. When water temperatures at 12 meters depth (chosen to avoid surface variations) were plotted against Daphnia spp. and total zooplankton densities, a positive correlation was evident (Figures 4.5 through 4.16). Results from a statistical analysis of Daphnia densities versus water temperatures yield a p value of $<0.0001$ and an $\mathrm{r}^{2}$ value of 0.441 (Figure 4.15). Statistical analysis of total zooplankton densities versus water temperatures were significant ( $\mathrm{p}<0.0001 ; \mathrm{r}^{2}=0.498$ ), meaning that, as temperature increased, zooplankton density also increased (Figure 4.16). The summer peak in zooplankton biomass and density values was probably the result of optimum temperatures coupled with a large quantity of nutrients and sunlight available for phytoplankton growth, which in turn increased the forage base for zooplankton. Warmer water temperatures increase the assimilation of nutrients by phytoplankton because they are poikilothermic organisms (Beckman et al. 1985). The drop in densities which occurs in August at the lower river sites may be due to the ten foot drafting of Lake Roosevelt resulting from ESA water demands, but at this time, were are unable to correlate zooplankton densities with drawdowns. August drawdowns have the greatest potential to impact zooplankton populations because they occur during a period


Figure 4.3 Mean monthly Daphnia spp. densities (\#/m³) at Gifford, Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon in 1995.


Figure 4.4 Mean monthly total zooplankton densities (\#/m ${ }^{\mathbf{3}}$ ) at Gifford, Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon in 1995.


Figure 4.5 Mean Daphnia spp. densities $\left(\# / \mathbf{m}^{\mathbf{3}}\right)$ plotted against water temperature at 12 m depth at Gifford in 1995.


Figure 4.6
Mean total zooplankton densities ( $\# / \mathrm{m}^{\mathbf{3}}$ ) plotted against water temperature at 12 m depth at Gifford in 1995.


Figure 4.7 Mean Daphnia spp. densities (\#/m³) plotted against water temperature at 12 m depth at Porcupine Bay in 1995.


Figure 4.8 Mean total zooplankton densities ( $\# / \mathbf{m}^{\mathbf{3}}$ ) plotted against water temperature at 12 m depth at Porcupine Bay in 1995.


Figure 4.9 Mean Daphnia spp. densities (\#/ $\mathbf{m}^{\mathbf{3}}$ ) plotted against water temperature at 12 m depth at Seven Bays in 1995.


Figure 4.10 Mean total zooplankton densities $\left(\# / \mathbf{m}^{\mathbf{3}}\right)$ plotted against water temperature at 12 m depth at Seven Bays in 1995.


Figure 4.11 Mean Daphnia spp. densities ( $\left({ }^{\left(\mathbf{m}^{\mathbf{3}}\right)}\right.$ ) plotted against water temperature at 12 mdepth at Keller Ferry in 1995.


Figure 4.12 Mean total zooplankton densities ( $\left(\# / \mathbf{m}^{3}\right)$ plotted against water temperature at 12 m depth at Keller Ferry in 1995.


Figure 4.13 Mean Daphnia spp. densities (\#/m³ ${ }^{\mathbf{3}}$ ) plotted against water temperature at 12 m depth at Spring Canyon in 1995.


Figure 4.14 Mean total zooplankton densities (\#/m³) plotted against water temperature at 12 m depth at Spring Canyon in 1995


Figure 4.15 Regression plot of Duphniu spp. densities versus water temperatures at 12 m for five locations in Lake Roosevelt in 1995.


Figure 4.16 Regression plot of total zooplankton densities versus water temperatures at 12 m for five locations in Lake Roosevelt in 1995.
when zooplankton populations are greatest. In order to investigate this possibility, it is recommended that zooplankton sampling frequency be increased to weekly samplings. August population reductions may also be due to over grazing of phytoplankton by zooplankton resulting in a population crash. As August zooplankton densities decline, the phytoplankton population has a chance to rebuild, thus allowing for an additional increase in zooplankton densitiy as seen in September (Figures 4.3, 4.4 and Appendix B). The lag in Daphnia and total zooplankton densities at Gifford is likely the result of lower water temperatures which inhibit phytoplankton and zooplankton growth until later in the year. Based on the above analysis, it is evident that water temperature will be a significant variable in the future modeling of zooplankton population dynamics. A comparison of yearly average total zooplankton densities for 1994 and 1995 shows a marked increase in zooplankton densities in 1995 (Shields and Underwood, 1996). This increase is not evident when yearly average biomass values are cornpaired. This lack of biomass increase, despite the increase in zooplankton densities, was the result of much smaller average lengths for individual zooplanktors in 1995. The reasons for this trend are not known, but may be due to increased predation by kokanee or a reduction of nutrient inputs into the system.

Water retention times may also indirectly affect zooplankton growth. For example, short water retention times may limit growth by not allowing enough time for assimilation of nutrients into phytoplankton and may also reduce overall water temperatures by not allowing the water in the lake as much time to warm up. Increased peaks in biomass and density values are thought to be related to increased reservoir elevations and water retention times but also appear to be affected by nutrient concentrations and water temperatures (Goldman and Home 1983).

Figures 4.17 through 4.21 give the monthly mean total zooplankton densities and water retention times for a five year period beginning in 1991. Figure 4.17 shows a clear relationship between water retention time and zooplankton densities at the Gifford area. During the growing season, as water retention time increases, zooplankton densities increase shortly afterward. Peaks in water retention time during the winter months are not followed by increased zooplankton densities. This may be due to the fact that since zooplankton are poikilotherms, their metabolism and reproduction rates are reduced in winter and their populations remain low regardless of water retention times. Figure 4.17 shows that the time period of the growing season increase was slightly different each year, corresponding to water retention time and not season. For example, graphs for
68


Figure 4.17 Monthly mean total zooplankton densities and monthly average water retention times at Gifford (Index Station 2), from 1991 through 1995.


Figure 4.18 Monthly mean total zooplankton oensities and monthly average water reten-ion times a Porcupine Bay (Index Station 4), from 1991 through 1995.


Figure 4.19 Monthly mean total zooplankton densities and monthly average water retention times at. Seven Bays (Index Station 6), from 1991 through 1995.


Figure 4.20
Monthly mean total zooplankton densities and monthly average water retention times at Keller Ferry (Index Station 7), from 1991 through 1995.


Figure 4.21 Monthly mean total zooplankton densities and monthly average water retention times at Spring Canyon (Index Station 9), from 1991 through 1995.
downstream areas of the reservoir do not show such a clear relationship between zooplankton densities and water retention time (Figures 4.18-4.21). The reasons for this are currently not known due to the fact that zooplankton dynamics are very complex and require a more in depth assessment before any conclusions can be drawn.

Overall, mean zooplankton densities and biomass were highest at the Confluence and Spring Canyon in 1995. The higher density values at the lower end of the reservoir may be a result of the flushing of water through the reservoir resulting in a "pileup" of organisms at lower reservoir sites. Declining pool elevation and large releases from the dam may further affect zooplankton by causing a downstream loss of zooplankton. Based on samples collected from Rufus Woods, it is estimated that a yearly average of $1,485.8$ zooplankton organisms are lost for every cubic meter of water run through the Coulee Dam turbines. This corresponds to a biomass loss of $4.7 \times 10^{8} \mu \mathrm{~g}$ per cubic meter of water.

### 4.4 Rainbow Trout Tagging

The percentage of tagged fish recovered below Grand Coulee Dam is a strong indicator of entrainment and has ranged from 0 to $63 \%$ of tag recoveries by month over the past eight years (Table 3.34). In 1995, entrainment was low, with only 5 fish recovered below Grand Coulee dam yielding a $3 \%$ recapture rate below Grand Coulee dam. This may be due to the fact that 1995 fish were held in net pens until June prior to release. Also, since water retention times averaged 46.5 days for 1995 , low entrainment rates should be expected. Table 3.34 shows that when water retention times average 40 or more days for a month, fish appear less likely to entrain from the reservoir. While entrainment through Grand Coulee dam appeared to be low for 1995, an analysis of tag release location versus fish capture location indicates a large down lake migration of net pen rainbow trout for both release sites. Of the tag returns sent in from the Kettle Falls releases, $95 \%$ ( $n=57$ ) were captured at down lake locations. The remaining $5 \%$ of fish ( $n=3$ ) were captured in the Kettle Falls area. No Kettle Falls fish were captured up lake from the release site. Down lake migrants traveled an average of 136.6 Km with a minimum migration of 41.9 Km and a maximum of 164.2 Km . An analysis of the Seven Bays releases finds that $68 \%(n=75)$ were captured down lake from the release location, $28 \%$ ( $n=31$ ) were recaptured in the Seven Bays area and $4 \%(n=4)$ were recaptured up lake from Seven Bays. The down lake migrants from the Seven Bays area traveled an average of 50.0 Km ranging from a minimum of 24.2 Km to a maximum of 264.8 Km . Up lake migrants traveled an average of 146.0 Km . This down lake migration may have been due to the fact that more food items were available in the lower ends of the reservoir and therefore were more attractive to
fish, or it may have been due to a smoltification process. The stock of rainbow trout used for the supplimentation program has been found to exhibit a smoltification process similar to that of steelhead trout and anadromous salmon (Muzi, 1984; Scholz et al., 1985; White et al., 1991). These rainbow trout have evidenced an increase in thyroxine, increased silvering, increased osmoregulatory capability and an increase in downstream migratory behavior during the spring (A. Scholz, personal communication). Therefore, if fish are released in early spring, they may exhibit partial smoltification and travel downstream.

When growth rates of fish released from Kettle Falls are compared to fish released from Seven Bays it can be seen that the Seven Bays fish grow at a much faster rate (Figures 3.3 and 3.4). This further indicates that the habitat in the lower reservoir is better for rainbow trout growth. When paired releases from Kettle Falls and Seven Bays areas are compared, the majority ( $81.7 \%$ ), of the Seven Bays fish are caught in the first year when compared to Kettle Falls releases where only $68.8 \%$ are caught in the first year. This indicates that Kettle Falls may be an important release site for carry over fish, while Seven Bays fish are more likely to contribute directly to the fishery.

### 5.0 Recommendations

1. Continue to tag 10,000 rainbow trout at Kettle Falls and Seven Bays in order to increase the numbers of tag returns.
2. Continue to hold net pen rainbow trout until at least June 1 before release in order to reduce entrainment losses.
3. Increase the zooplankton sampling frequency to at least three times per month in the spring, summer and fall and twice per month in winter. In addition, begin to sample zooplankton in near shore areas along with mid channel tows. Also, due to the extreme variability of zooplankton data it is recommended that a minimum of three zooplankton tows be taken at each site and the densities averaged to attain a mean location density.
4. Continue to sample for zooplankton in Rufus Woods Reservoir in order to estimate entrainment losses of zooplankton from Lake Roosevelt.
5. Continue to collect zooplankton and water quality data at current sites.
6. Collect nutrient and $C^{14}$ data to obtain nutrient abundance and assimilation rates for model development.
7. Determine the depth of the euphotic zone via photometer to estimate the availability of phytoplankton habitat.

## Literature Citations

APHA. 1976. Standard Methods for the Examination of Water and Wastewater, 14th Ed. American Public Health Association. Washington, D.C. 1192 pp.

Beckman, L.G., J.F. Novotny, W.R. Parsons, and T.T. Tarrell. 1985. Assessment of the fisheries and limnology in Lake F.D. Roosevelt 1980-1983. U.S. Fish and Wildlife Service. Final Report to U.S. Bureau of Reclamation. Contract No. WPRS-0-07-10-X02 16; FWS- 14-06-009-904, May 1985. 168 pp.

Brandlova, J., Z. Brandl and C.H. Fernando. 1972. The Cladocera of Ontario with remarks on some species and distribution. Can. J. of Zool. 50:1373-1403.

Brooks, J.L. 1957. The systematics of North America Daphnia. Conn. Acad. Arts and Sci. Vol. 13, New Haven, CT. 180 pp.

Carlander, K.D. 1943. Age, Growth, sexual maturity, and population fluctuations of the walleye, Stizostedian vitreum vitrewn (Mitchell), with references to the commercial fisheries, Lake of the Woods, Minnesota. Trans. Amer. Fish. Soc. 73:90-107.

Carlander, K.D. 1969. Handbook of Freshwater Fishery Biology. Vol. 1. Iowa State University Press. Ames, Iowa. 752 pp.

CRWMG (Columbia River Water Management Group). 1996. Columbia River Water Management Report for Water Year 1994. Columbia River Water Management Group. 132pp.

Downing, J.A. and F.H. Rigler. 1984. A Manual on Methods for the Assessment of Secondary Productivity in Fresh Waters. 2nd. Ed. IBP Handbook No. 17:500.

Edmondson, W.T. (ed). 1959. Fresh-water Biology. 2nd. ed. John Wiley and Sons. New York. 1248 pp.

Edmondson, W.T. and G.G. Winberg. 1971. A Manual for the Assessment of Secondary Productivity in Fresh Waters. IBP Handbook No. 17. 358 pp.

Fletcher, D.H. 1988. Phase management research, fiist year's work at Kitsap Lake, Kitsap County, Washington. Washington Department of Wildlife, Fisheries Management Division, Olympia, WA. Report No. 88-6:80.

Goldman, C.R. and A. J. Home. 1983. Limnology. McGraw-Hill, New York. 464pp.
Griffith, J.R. and A.C. McDowell. 1996. Measurement of Lake Roosevelt Biota in Relations to Reservoir Operations, Annual Report 1992. Bonneville Power Administration. Portland, OR. Project No. 88-63. 87 pp plus appendices

Griffith, J.R., A.C. McDowell and A.T.Scholz. 1992. Measurement of Lake Roosevelt Biota in Relation to Reservoir Operations, Annual Report 1991. Bonneville Power Administration. Portland, OR. Project No. 88-63. 138 plus appendices.

Griffith, J.R. and A.T. Scholz. 1991. Lake Roosevelt Fisheries Monitoring Program, Annual Report 1990. Bonneville Power Administration. Portland, OR. Project No. 88-63. 218 pp plus appendices

Jagielo, T. 1984. A comparison of nutrient loading, phytoplankton standing crop, and uophic state in two morphologically and hydraulically different reservoirs. MS thesis. University of Washington. Seattle, WA. 99 pp.

Lambou, V.W. 1961. Determination of fishing pressure from fishermen of party counts with a discussion of sampling problems. Proc. of the S.E. Association of Game and Fish Commissioners: 1961:380-401.

Lambou, V.W. 1966. Recommended method of reporting creel survey data for reservoirs. Oklahoma Fishery Research Laboratory. University of Oklahoma, Norman, OK. Bulletin No. 4. 39pp.

Lewis, S.L. 1975. Evaluation of the Kokanee fishery and hatchery releases at Ode11 Lake, 1964- 1975. Oregon Department of Fish and Wildlife. D.J. Rep. F-7 1-R.

Malvestuto, S.P., W.D. Davies, and W.C. Shelton. 1978. An evaluation of the roving creel survey with nonuniform probability sampling. Trans. Amer. Fish. Soc. 107:255-262.

Malvestuto, S.P. 1983. Sampling the Recreational Fishery. In: LA. Nielsen and P.L. Johnson (ed.). Fisheries Techniques. Amer. Fish. Soc. Bethesda, MD. 468 pp.

Mendel, G. and M. Schuck. 1987. Fall 1985 and Spring 1986 Snake River steelhead creel surveys. Washington Department of Wildlife. D.J. Rep. FRILSR- 87-8.

Muzi, M. 1984. Triiodothyronine binding sites in the brains of developing steelhead trout (Salmo gairdneri Richardson). M.S. thesis. Eastern Washington University, Cheney, WA. 123pp.

NWPPC. 1987. Columbia River Basin Fish and Wildlife Program. Section 900 Resident Fish. Northwest Power Planning Council, Portland, OR. 125-126 pp.

Pennak, R.W. 1978. Freshwater Invertebrates of the United States, 2nd ed. Wiley and sons, New York. 803 pp.

Pennak, R.W. 1989. Freshwater Invertebrates of. the United States, 3rd ed. Wiley and sons, New York. 628 pp.

Peone, T., A.T. Scholz, J.R. Griffith, S. Graves, and M.G. Thatcher. 1990. Lake Roosevelt fisheries monitoring program. Annual report 1989. Upper Columbia United Tribes Fisheries Center. Eastern Washington University. Cheney, WA. DE-8179-88B P91819

Post, G. 1979. Carbonic acid anesthesia for aquatic organisms. Progressive FishCulturist 41: 142-144.

Rigler, F. H. 1978. Sugar frosted Daphnia; An improved fixation technique for Cladocera. Limnol. Oceanogr. 23(3):557-559.

Ruttner-Kolisko, A. 1974. Plankton Rotifers Biology and Taxonomy. Die Binnengewasser, Stutgart. 26/1. 146 pp.

Scholz, A.T. 1989. Letter to Joe Keys, Director of Bureau of Reclamation, Boise, ID. Dated 8/14/89.

Scholz, A.T., K. O'Laughlin, T. Peone, J. Uehara. T. Kleist and J. Hisata. 1988. Environmental factors affecting Kokanee salmon, Oncorhynchus nerka (Walbaum) in Deer and Loon Lakes, Stevens County, Washington. Final Report submitted to Deer and Loon Lake Property Owners Association and Washington Department of Wildlife. Eastern Washington University, Department of Biology, Cheney, WA. 167 pp.

Scholz, A.T., J.K. Uehara, J. Hisata, and J. Marco. 1986. Feasibility report on restoration and enhancement of Lake Roosevelt Fisheries. In: Northwest Power Planning Council. Applications for Amendments. Vol. 3A: 1375-1489.

Scholz, A.T., K. O'Laughlin, D.R. Geist, D. Peone, J.K. Uehara, L. Fields, T. Kleist, I. Zozaya, T. Peone, and K. Teesatuskie. 1985. Compilation of information on salmon and steelhead total run size, catch and hydropower related losses in the Upper Columbia River Basin, above Grand Coulee Dam. Upper Columbia United Tribes Fisheries Center. Technical Report No. 2: 165.

Scholz, A.T., R.J. White, M. Muzi, and T. Smith. 1985. Uptake of radiolabelled triiodothyronine in the brain of steelhead trout (Salmo gairdneri) during pax-r-smolt transformation: Implications for the mechanism of thyroid activation of olfactory imprinting. Aquaculture 45: 199-214.

Shields, J.P., and K.D. Underwood. 1996. Measurement of Lake Roosevelt Biota in Relation to Reservoir Operations. Boneville Power Adminisuation. Portland, OR. Project No. 94-043. 236pp.

Stemberger, R.S. 1979. A guide to rotifers of the Laurentian Great Lakes. Environmental Monitoring and Support Laboratory, Office of Research and Development, U.S. Environmental Protection Agency, Cincinnati, OH. EPA-600/4-79-021. 1985 pp.

Stober, Q-J., M.E. Kopache and T.H. Jagielo. 1981. The limnology of Lake Roosevelt. Final Report Contract No. 14-16-0009-80-0004, to the U.S. Fish and Wildlife Service. National Fisheries Research Center, Seattle, WA. Fisheries Research Institute, University of Washington, Seattle, WA. FRI-VW-8106: 116 pp.

Stober, Q.J., R.W. Tyler, C.E. Peuosky, T.J. Carlson, D. Gaudet and R.E. Nakatani. 1977. Preliminary survey of fisheries resources in the forebay of FDR Reservoir. Annual report. College of Fisheries, Fisheries Research Institute. University of Washington, Seattle, WA. FRI-UW7701: 70 pp .

US. Army Corps of Engineers. 1981. Reservoir storage sables for Grand Coulee Reservoir. Prepared from table by U.S. Bureau of Reclamation and U.S. Geological Survey. October 1977.

Voeller, A.C. 1996. Measurements of Lake Roosevelt Biota in Relations to Reservoir Operations. Annual Report 1993. Boneville Power Administration. Portland, OR. Project No. 94-043. 109pp.

Ward, J. 1955. A description of a new zooplankton counter. Quart. J. Microscop. Scien. 96:371-373.

White, R.J., A.T. Scholz, M.V. Baker and D.E. Liljegren. 1991. Detection of triiodothyronine $\left(\mathrm{T}_{3}\right)$ receptors in isolated steelhead trout, Oncorhynchus mykiss, brain nuclei. Journal of Fish Biology 36: 783-785.

Wiggins, G.B. 1977. Larvae of the North American Caddisfly Genera (Trichoptera). University of Toronto. Toronto, ONT: 568 pp.

Williams, K. and L. Brown. 1984. Mid-Columbia walleye life history and management 1979-1982. Washington Department of Game, Fish Management Division. Olympia, WA. Internal report. 38 pp .

Willms, R.A., A.T. Scholz, and J. Whalen. 1989. The assessment of the developing mixed-species fishery in Sprague Lake, Adams and Lincoln Counties, Washington, following restoration with rotenone. Final report submitted to Washington Department of Wildlife, Olympia, WA. Department of Biology, Eastern Washington University, Cheney, WA. 160 pp.

Wonnacott, T.H. and R.J. Wonnacott. 1977. Introductory Statistics. Third edition. John Wiley and Sons, New York, N.Y. 650 pp.

APPENDIX A
HYDROLOGY

Table A. 1 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in January, 1995. Data from CORPs daily summary reports.

| JANUARY |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MOITH } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ \text { (KCFS) } \end{gathered}$ | OUTFLOW (KCFS) | $\begin{gathered} \hline \hline \text { RESERVOIR } \\ \text { ELEVATION } \\ \text { (FT) } \end{gathered}$ | STORAGE CAPACITY (KCFSD) |  <br> WATER <br> RETENTION <br> TIME (D) |
| 2 | 73.0660 .40 | 100.2089 .00 | 12843012885.30 | 4400.20 | 49.44 |
| 3 |  |  |  | 4360.40 | 43.52 |
| 4 | 73.8066 .00 | 119.2116 .80 | 1283.201281 .90 | 4317.00 | 36.22 |
|  |  |  |  | 4266.20 | 36.53 |
| 5 | 73.90 | 112.70 | 1280.90 | 4227.40 | 37.51 |
| 6 | 70.70 | 111.10 | 1279.80 | 4185.40 | 37.67 |
| 8 | 59.60 | 93.90 | 1278.90 | 4150.80 | 44.21 |
| 9 | 71.0073 .30 | 88.738280 | 1278.201278 .45 | 4131.80 | 46.57 |
|  |  |  |  | 4124.30 | 49.81 |
| 10 | 69.30 | 84.40 | 1277.80 | 4109.10 | 48.69 |
| 11 | 78.80 | 97.70 | 1277.30 | 4090.20 | 41.86 |
| 12 | 71.10 | 84.40 | 1277.00 | 4078.90 | 48.32 |
| 13 | 77.40 | 62.30 | 1277.40 | 4094.00 | 65.71 |
| 14 | 76.00 | 53.30 | 1278.00 | 4116.70 | 77.24 |
| 15 | 75.30 | 39.30 | 1278.90 | 4150.80 | 105.62 |
| 17 | 80.40 | 70.90 | 1279.20 | 4162.30 | 58.71 |
| 18 | 76.1080 .30 | 84.1083 .70 | 1278.901279 .10 | 4158.50 4150.80 | 49.44 |
|  |  |  |  |  | 49.59 |
| 20 | 78.40 | 89.90 | 1278.55 | 4135.60 | 46.00 |
| 21 | 74.1074 .80 | 77.9880 .50 | 1278.451278 .30 | 4131.80 4128.00 | 53.04 |
|  |  |  |  |  | 51.28 |
| 23 | 76.50 | 76.50 | 1278.30 | 4128.00 | 53.96 |
| 24 | 76.7080 .30 | 106.9010 .10 | 1277.701276 .90 | 4105.30 | 39.82 |
|  |  |  |  | 4075.10 | 38.12 |
| 25 | 76.00 | 94.80 | 1276.40 | 4056.30 | 42.79 |
| 26 | 75.20 | 92.10 | 1276.00 | 4041.30 | 43.88 |
| 27 | 72.40 | 102.30 | 1275.20 | 4011.40 | 39.21 |
| 28 | 62.70 | 92.50 | 1274.40 | 3981.60 | 43.04 |
| 29 | 66.20 | 60.60 | 1274.50 | 3985.30 | 65.76 |
| 30 | 72.20 | 94.50 | 1273.90 | 3963.00 | 41.94 |
| 31 | 71.20 | 91.60 | 1273.40 | 3944.50 | 43.06 |
| Average | 73.02 | 88.31 | 1278.28 | 4127.81 | 49.31 |

Table A. 2 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in February, 1995. Data from CORPS daily summary reports.

| FEBRRY |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DAY <br> OF <br> MONTH | INFLOW <br> (KCFS) | OUTFLOW <br> (KCFS) | 2ESERVOIR <br> ELEVATION <br> (FT) | STORAGE <br> CAPACITY <br> (KCFSD) | WATER <br> RETENTION <br> TIME (D) |  |
| 1 | 78.60 | 69.30 | 1273.60 | 3951.90 | 57.03 |  |
| 2 | 83.60 | 63.30 | 1274.20 | 3974.10 | 62.78 |  |
| 3 | 82.10 | 66.70 | 1274.50 | 3985.30 | 59.75 |  |
| 4 | 74.20 | 57.40 | 1274.90 | 4002.20 | 69.73 |  |
| 5 | 76.30 | 59.50 | 1275.20 | 4011.40 | 67.42 |  |
| 6 | 79.30 | 111.90 | 1274.20 | 3974.10 | 35.52 |  |
| 7 | 75.00 | 85.60 | 1273.60 | 3951.90 | 46.17 |  |
| 8 | 70.20 | 98.70 | 1272.50 | 3911.30 | 39.63 |  |
| 9 | 72.00 | 100.70 | 1271.35 | 3867.30 | 38.40 |  |
| 10 | 71.40 | 95.60 | 1270.30 | 3830.90 | 40.07 |  |
| 11 | 78.60 | 65.30 | 1270.30 | 3830.90 | 58.67 |  |
| 12 | 76.60 | 99.50 | 1269.20 | 3791.00 | 38.10 |  |
| 13 | 77.40 | 119.80 | 1268.00 | 3747.70 | 31.28 |  |
| 14 | 78.60 | 134.00 | 1266.40 | 3690.50 | 27.54 |  |
| 15 | 83.90 | 128.30 | 1265.15 | 3644.40 | 28.41 |  |
| 16 | 79.50 | 119.10 | 1263.80 | 3598.50 | 30.2 .1 |  |
| 17 | 76.60 | 104.20 | 1262.50 | 3553.00 | 34.10 |  |
| 18 | 81.60 | 79.50 | 1262.00 | 3535.60 | 44.47 |  |
| 19 | 71.90 | 57.70 | 1261.90 | 3532.10 | 61.22 |  |
| 20 | 86.70 | 89.90 | 1261.30 | 351.30 | 39.06 |  |
| 21 | 85.50 | 108.50 | 1260.15 | 3469.80 | 31.98 |  |
| 22 | 89.60 | 121.60 | 1258.90 | 3428.60 | 28.20 |  |
| 23 | 89.60 | 100.90 | 1258.50 | 3414.90 | 33.84 |  |
| 24 | 88.60 | 98.90 | 1258.20 | 3404.70 | 34.43 |  |
| 25 | 88.60 | 66.10 | 1258.40 | 3411.50 | 51.61 |  |
| 26 | 96.90 | 69.29 | 1258.80 | 3425.20 | 49.43 |  |
| 27 | 98.50 | 137.60 | 1257.60 | 3384.30 | 24.60 |  |
| 28 | 95.10 | 123.80 | 1259.70 | 3456.00 | 27.92 |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Average | 81.66 | 94.02 | 1266.15 | 36888.94 | 42.56 |  |

Table A. 3 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake
Roosevelt in March, 1995. Data from CORPs daily summary reports.

| MARCH |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DAY <br> OF <br> MONTH | INFLOW <br> (KCFS) | OUTFLOW <br> (KCFS) | RESERVOIR <br> ELEVATION <br> (FT) | STORAGE <br> CAPACITY <br> (KCFSD) | WATER <br> RTENTION <br> TIME (D) |  |
| $\mathbf{1}$ | 91.20 | 114.70 | 1256.00 | 3330.40 | 29.04 |  |
| $\mathbf{2}$ | 91.20 | 106.30 | 1255.60 | 3317.00 | 31.20 |  |
| 3 | 86.00 | 103.10 | 1255.00 | 3296.90 | 31.98 |  |
| 4 | 91.30 | 97.50 | 1254.40 | 3277.00 | 33.61 |  |
| 5 | 94.50 | 84.10 | 1254.30 | 3273.60 | 38.93 |  |
| 6 | 92.10 | 101.80 | 1253.90 | 3260.40 | 32.00 |  |
| 7 | 89.70 | 104.60 | 1253.45 | 3243.90 | 36.00 |  |
| 8 | 89.20 | 94.20 | 1253.30 | 3240.60 | 34.03 |  |
| 9 | 97.60 | 87.70 | 1253.60 | 3250.50 | 37.06 |  |
| 10 | 85.10 | 93.10 | 1253.40 | 3243.90 | 34.84 |  |
| 11 | 99.60 | 67.00 | 1254.20 | 3270.30 | 48.81 |  |
| 12 | 102.00 | 55.80 | 1255.20 | 3303.60 | 59.20 |  |
| 13 | 96.50 | 87.20 | 1255.40 | 3310.30 | 37.96 |  |
| 14 | 99.30 | 72.50 | 1256.20 | 3337.10 | 46.03 |  |
| 15 | 106.30 | 79.50 | 1257.00 | 3364.00 | 42.31 |  |
| 16 | 108.80 | 69.80 | 1258.20 | 3404.70 | 48.78 |  |
| 17 | 101.90 | 54.00 | 1259.60 | 3452.60 | 63.94 |  |
| 18 | 103.90 | 45.30 | 1261.30 | 3511.30 | 77.51 |  |
| 19 | 114.40 | 32.70 | 1263.60 | 3591.50 | 109.83 |  |
| 20 | 118.50 | 88.60 | 1264.50 | 3623.20 | 40.89 |  |
| 21 | 112.60 | 114.40 | 1264.40 | 3619.70 | 31.64 |  |
| 22 | 119.50 | 133.60 | 1264.00 | 3605.60 | 26.99 |  |
| 23 | 113.70 | 133.00 | 1263.45 | 3584.50 | 26.95 |  |
| 24 | 113.90 | 126.00 | 1263.10 | 3574.00 | 28.37 |  |
| 25 | 110.20 | 100.00 | 1263.30 | 3581.00 | 35.81 |  |
| 26 | 108.20 | 105.00 | 1263.40 | 3584.50 | 34.14 |  |
| 27 | 108.20 | 124.20 | 1262.90 | 3567.00 | 28.72 |  |
| 28 | 100.00 | 95.84 | 1263.63 | 3591.50 | 37.47 |  |
| 29 | 102.90 | 85.60 | 1263.50 | 3588.00 | 41.92 |  |
| 30 | 101.70 | 72.00 | 1264.30 | 3616.10 | 50.22 |  |
| 31 | 96.90 | 63.10 | 1265.20 | 3647.90 | 57.81 |  |
|  |  |  |  |  |  |  |
| Average | 101.55 | 90.07 | 1258.99 | 3434.28 | 42.39 |  |

Table A. 4 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in April, 1995. Data from CORPs daiiy summary reports.

| APRIL |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ \text { (KCFS) } \end{gathered}$ | $\begin{aligned} & \hline \hline \text { OUTFLOW } \\ & \text { (KCFS) } \end{aligned}$ | RESERVOIR <br> ELEVATION <br> (FT) | $\begin{gathered} \hline \hline \text { CTORAGE } \\ \text { CAPACITY } \\ \text { (KCFSD) } \end{gathered}$ | WATER <br> RETENTION <br> TIME (D) |
| 2 | 92.70 | 66.70 | 1265.80 | 3669.2 | 55.01 |
| 3 | 93.704 .00 | 22.0067 .00 | 1266.301266 .30 | 3667.03687 .0 | 40.0854 .50 |
| 5 | 83.30 | 115.60 | 1265.20 | 3647.9 | 31.56 |
| 6 | 86.6086 .40 | 110.1097 .00 | 1264.501266 .20 | 3633.23012 .6 | 329.37 .241 |
| 7 | 78.60 | 66.20 | 1264.50 | 3623.2 | 54.73 |
| 9 | 77.8081 .60 | 40.9067 .20 | 1266.801266 .00 | 3633.8367 .3 | 54.0789 .89 |
| 10 | 86.70 | 81.40 | 1266.10 | 3679.8 | 45.21 |
| 11 | 80.00 | 53.20 | 1266.41 | 3690.5 | 69.37 |
| 12 | 83.90 | 60.60 | 1267.50 | 3729.8 | 61.55 |
| 13 | 79.10 | 64.80 | 1267.90 | 3744.2 | 57.78 |
| 14 | 88.20 | 60.50 | 1268.70 | 3772.9 | 62.36 |
| 15 | 83.40 | 62.30 | 1269.00 | 3783.4 | 60.73 |
| 16 |  |  |  |  |  |
| 17 | 75.7086 .60 | 101.4052 .70 | 12693012688.80 | 3794.6377 .5 | 72.0037 .24 |
| 18 | 76.90 | 97.10 | 1268.10 | 3751.3 | 38.63 |
| 19 | 80.50 | 79.10 | 1268.10 | 3751.3 | 47.43 |
| 20 | 68.90 | 69.30 | 1268.00 | 3747.7 | 54.08 |
| 21 | 68.60 | 68.60 | 1268.00 | 3747.7 | 54.63 |
| 22 | 68.10 | 66.70 | 1267.90 | 3744.2 | 56.14 |
| 23 | 80.10 | 89.10 | 1267.50 | 3729.8 | 41.86 |
| 24 | 85.80 | 125.80 | 1266.20 | 3683.4 | 29.28 |
| 25 | 67.00 | 126.10 | 1264.35 | 3616.1 | 28.68 |
| 26 |  |  |  |  |  |
| 27 | 83.0075 .80 | 120.5011 .80 | 1261.801263 .10 | 3574.0328 .6 | 29.6630 .74 |
| 28 | 81.10 | 105.60 | 1260.90 | 3497.4 | 33.12 |
| 29 | 80.10 | 107.20 | 1259.80 | 3459.5 | 32.27 |
| 30 | 80.20 | 105.50 | 1258.70 | 3421.8 | 32.43 |
| Average | 81.15 | 84.52 | 1265.79 | 3669.49 | 47.51 |

Table A. 5 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in May, 1995. Data from CORPs daily summary reports.

| MAY |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ \text { (KCFS) } \end{gathered}$ | $\underset{\text { (KCFS) }}{\overline{\text { OUTFLOW }}}$ | RESERVOIR <br> ELEVATION <br> (FT) | $\begin{aligned} & \hline \text { STORAGE } \\ & \text { CAPACITY } \\ & \text { (KCFSD) } \\ & \hline \end{aligned}$ | WATER <br> RETENTION <br> TIME (D) |
| 2 | 83.70 | 124.30 | 1257.20 | 3370.8 | 27.12 |
| 3 | 87.5087 .50 | 113.20117 .9 | 1255.201256 .10 | 3333.7 | 28.28 |
| 4 |  |  |  | 3303.6 | 29.18 |
| 5 | 87.5087 .20 | 115.8510 .50 | 1253.411255 .20 | 3303.6 | 28.52 |
|  |  |  |  | 3277.0 | 31.03 |
| 6 | 91.10 | 62.10 | 1254.00 | 3263.7 | 52.56 |
| 8 | 91.70 | 65.80 | 1254.50 | 3280.3 | 49.85 |
| 9 | 96.2099 .80 | 97.0883 .90 | 1254.70 1254.65 | 3286.9 | 39.18 |
|  |  |  |  | 3285.3 | 33.87 |
| 10 | 102.90 | 86.40 | 1255.10 | 3300.3 | 38.20 |
| 11 | 108.80 | 80.30 | 1255.90 | 3327.0 | 41.43 |
| 12 | 104.80 | 90.30 | 1256.30 | 3340.4 | 36.99 |
| 13 | 108.40 | 60.88 | 1257.45 | 3379.3 | 55.51 |
| 14 | 109.60 | 67.33 | 1258.30 | 3408.1 | 50.62 |
| 15 | 112.70 | 118.70 | 1257.90 | 3394.5 | 28.60 |
| 16 | 125.10 | 115.80 | 1258.00 | 3397.9 | 29.34 |
| 17 | 127.00 | 116.00 | 1258.20 | 3404.7 | 29.35 |
| 19 | 128.40 | 99.66 | 1258.80 | 3425.2 | 34.37 |
| 20 | 132.30126 .60 | 93.8061 .90 | 1259.801266 .50 | 3459.5 | 36.88 |
|  |  |  |  | 3518.2 | 56.84 |
| 21 | 124.20 | 64.20 | 1263.00 | 3570.5 | 55.62 |
| 23 | 125.80 | 102.30 | 1263.50 | 3588.0 | 35.07 |
| 24 | 124.20120 .00 | 118.2097 .20 | 1264.00 | 3605.6 | 37.10 |
|  |  |  | 1264.00 | 3605.6 | 30.50 |
| 25 | 121.80 | 115.80 | 1264.00 | 3605.6 | 31.14 |
| 27 | 124.90 | 102.50 | 1264.50 | 3623.2 | 35.35 |
| 28 | 119.10 119.40 | 67.205 .20 | 1265.60 | 3623.1 | 53.92 |
|  |  |  | 1267.10 | 3715.5 | 72.57 |
| 30 | 128.50 | 101.90 | 1267.60 | 3733.4 | 36.64 |
| 31 | 140.41134 .0 | 117.3084 .80 | 1267.901269 .40 | 3744.2 | 31.92 |
|  |  |  |  | 3798.2 | 44.79 |
| Average | 112.34 | 93.53 | 1260.06 | 3460.40 | 39.43 |

iabie A.b Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in June, 1995. Data from CORPs daily summary reports.

| JUNE |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ \text { (KCFS) } \end{gathered}$ | OUTFLOW (KCFS) | $\underset{\text { RESERVOIR }}{ }$ <br> (FT) | STORAGE CAPACITY (KCFSD) | WATER <br> RETENTION <br> TIME (D) |
| 1 | 132.70 | 100.80 | 1270.10 | 3823.6 | 37.93 |
| 2 | 134.40 | 79.80 | 1271.50 | 3874.6 | 48.55 |
| 4 | 129.50 | 49.70 | 1273.30 | 3940.8 | 72.29 |
| 5 | 130.8014 .50 | 42.8087 .20 | 1275.301276 .50 | 4015.14600 .1 | 46.5989 .81 |
| 6 | 158.10 | 88.80 | 1278.20 | 4124.3 | 46.45 |
| 7 | 158.00 | 76.20 | 1280.30 | 4204.3 | 55.18 |
| 8 |  |  |  |  |  |
| 9 | 154.00166 .40 | 127.1093 .00 | 1881.701282 .50 | 4258.44889 .5 | 45.7933 .75 |
| 10 | 165.00 | 133.90 | 1282.90 | 4305.2 | 32.15 |
| 11 | 165.10 | 121.30 | 1283.70 | 4336.7 | 35.75 |
| 12 | 163.70 | 139.00 | 1284.20 | 4356.4 | 31.34 |
| 13 | 161.00 | 134.40 | 1284.70 | 4376.3 | 32.56 |
| 15 | 150.50 | 130.10 | 1285.00 | 4388.2 | 33.73 |
| 16 | 133.50 135.80 | 118.90103 .80 | 1285.2012885 .80 | 4396.2420 .2 | 42.836 .97 |
| 17 | 139.10 | 85.90 | 1286.80 | 4460.5 | 51.93 |
| 18 | 141.50 | 88.70 | 1287.90 | 4505.2 | 50.79 |
| 19 |  |  |  |  |  |
| 20 | 140.08150 .30 | 129.71137 .60 | 1288.001288 .10 | 4509.3413 .4 | 34.7732 .80 |
| 21 | 146.60 | 141.30 | 1288.10 | 4513.4 | 31.94 |
| 22 |  |  |  |  |  |
| 23 | 148.7141 .90 | 153.4115 .00 | 1287.901287 .60 | 4505.2493 .0 | 29.329 .95 |
| 24 |  |  |  |  |  |
| 25 | 137.00 159.0 | 14.5118 .50 | 1287.111287 .30 | 472.7480 .8 | 30.1230 .00 |
| 26 |  |  |  |  |  |
| 27 | 160.11161 .30 | 14.30315 .30 | 1887.301287 .60 | 4880.8493 .0 | 28.8531 .57 |
| 28 | 153.70 | 145.20 | 1287.60 | 4493.0 | 30.94 |
| 29 | 144.80 | 144.60 | 1287.50 | 4488.9 | 31.04 |
| 30 | 135.80 | 138.20 | 1287.30 | 4480.8 | 32.42 |
| Average | 148.05 | 117.75 | 1283.57 | 4335.3 | 40.09 |

Table A. 7 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in July, 1995. Data from CORPs daily summary reports.

| JULY |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MOITH } \end{gathered}$ | $\begin{gathered} \hline \hline \text { (KCFS) } \\ \hline \text { (KLOW } \end{gathered}$ | $\underset{\text { (KCFS) }}{\overline{\text { OUTFLOW }}}$ | $\substack{\text { RESERVOIR } \\ \text { ELE VATION } \\ \text { (FT) }}$ | STORAGE CAPACITY (KCFSD) | WATER <br> RETENTION <br> TIME (D) |
| 2 | 129.10131 .30 | 92,30 73.90 | 1287,90 1289,00 | 4505.2 | 48.81 |
|  |  |  |  | 4550.3 | 61.57 |
| 3 | 137.30 | 104.30 | 1289.50 | 4570.9 | 43.82 |
| 5 | 124.80 | 108.30 | 1289.60 | 4575.0 | 42.24 |
| 6 | 119,30 114.10 | 129.801123 .90 | 1289.101288 .70 | $\begin{aligned} & 4554.4 \\ & 4537.9 \end{aligned}$ | 35.09 |
|  |  |  |  |  | 36.63 |
| 8 | $\begin{aligned} & 105.80 \\ & 115.50126 .50 \end{aligned}$ | $\begin{aligned} & 122.80 \\ & 107.3069 .80 \end{aligned}$ | 1288.10 <br> 1288.00 1889.10 | 4513.4 | 36.75 |
|  |  |  |  | 4509.3 | 42.03 |
| 9 |  |  |  | 4554.4 | 65.25 |
| 10 | 128.60 | 121.90 | 1289.10 | 4554.4 | 37.36 |
| 11 | 116.10 | 110.70 | 1289.10 1289.10 | 4554.4 | 41.14 |
| 12 | 119.50 | 105.90 | $\begin{aligned} & 1289.10 \\ & 1289.30 \end{aligned}$ | 4562.6 | 43.08 |
| 13 | 111.00 | 108.40 | $\begin{aligned} & 1289.30 \\ & 1289.20 \end{aligned}$ | 4558.5 | 42.05 |
| 14 | 104.80 | 114.10 | 1288.80 | 4542.0 | 39.81 |
| 15 | 96.90 | 112.90 | 1288.20 | 4517.5 | 40.01 |
| 16 | 99.60 | 95.10 | 1288.10 | 4513.4 | 47.46 |
| 17 | $\begin{gathered} 99.10 \\ 112.30 \end{gathered}$ | $\begin{aligned} & 131.30 \\ & 122.70 \end{aligned}$ | 1287.201286.90 | 4476.7 | 34.10 |
| 20 |  |  |  | 4464.6 | 36.39 |
|  | 107.80 10.90 | 116.70111 .30 | 12886011286.30 | 4452.4 | 38.15 |
|  |  |  |  |  | 39.9042.65 |
| 22 | $\begin{aligned} & 110.70 \\ & 105.5096 .30 \end{aligned}$ | $\begin{aligned} & 104.10 \\ & 105.3090 .00 \end{aligned}$ | $\begin{aligned} & 1286.30 \\ & 1286.001285 .90 \end{aligned}$ | 4440.3 |  |
| 23 |  |  |  | 428.3424 .2 | $\begin{aligned} & 42.05 \\ & 49.16 \end{aligned}$ |
|  |  |  |  |  |  |
| 2526 | $\begin{aligned} & 107.00 \\ & 109.00108 .80 \end{aligned}$ | $\begin{aligned} & 111.10 \\ & 123.70116 .00 \end{aligned}$ | 1285.60 12884.80 1285.30 | 4412.2 | 39.7137.93 |
|  |  |  |  | 4400.24380.3 |  |
|  |  |  |  |  | 35.41 |
| 28 | $\begin{aligned} & 103.60 \\ & 100.7094 .60 \end{aligned}$ | 134.10 | 1283.90 | 4317.0 4297.4 | 32.40 |
| 29 |  | 122.50100 .40. | 12883201882.70 |  | 35.24 |
| 30 |  |  |  | $\begin{aligned} & 4277.8 \\ & 4258.4 \\ & \hline \end{aligned}$ | $\begin{aligned} & 42.80 \\ & 37.04 \\ & 35.76 \end{aligned}$ |
| 31 | 109.80106 .10 | 199.1115 .50 | 1882.201281 .70 |  |  |
|  |  |  |  |  |  |
| Average | 111.63 | 110.49 | 1286.95 | 4467.4 | 41.35 |

Table A. 8 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in August, 1995. Data from CORPs daily summary reports.

| AUGUST |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ \text { (KCFS) } \end{gathered}$ | $\underset{(\text { KCFS })}{\overline{\text { OUTFLOW }}}$ | RESERVOIR <br> ELEVATION <br> (FT) | STORAGE CAPACITY (KCFSD) | WATER <br> RETENTION <br> TIME (D) |
| 1 |  |  |  |  |  |
| 2 | 103.2086 .60 | 105.9094.80 | 1281.151281 .50 | 4250.64237 .0 | 40.144 .69 |
| 3 | 86.90 | 94.90 | 1280.80 | 4223.6 | 44.51 |
| 4 | 83.70 | 84.80 | 1280.60 | 4215.9 | 49.72 |
| 5 |  |  |  |  |  |
| 6 | 104.2092 .90 | 73.3069 .20 | 1281.301188080 | 4233.6442 .9 | 57.6261 .31 |
| 7 | 109.90 | 96.70 | 1281.50 | 4250.6 | 43.96 |
| 8 |  |  |  |  |  |
| 9 | 101.5095 .60 | 102.3099 .70 | 1281281.301 .20 | 4239.0442 .9 | 42.5641 .4 |
| 10 | 105.80 | 100.80 | 1281.20 | 4239.0 | 42.05 |
| 11 | 103.20 | 119.30 | 1280.60 | 4215.9 | 35.34 |
| 12 | 100.30 | 87.90 | 1280.70 | 4219.7 | 48.01 |
| 13 | 96.40 | 72.40 | 1281.00 | 4231.3 | 58.44 |
| 14 | 106.10 | 108.30 | 1280.80 | 4223.6 | 39.00 |
| 15 | 108.40 | 97.60 | 1281.00 | 4231.3 | 43.35 |
| 16 | 96.70 | 104.60 | 1280.60 | 4215.9 | 40.30 |
| 17 | 93.50 | 92.00 | 1280.50 | 4212.0 | 45.78 |
| 18 |  |  |  |  |  |
| 19 | 101.9092 .10 | 78.8075 .40 | 1281.1011880 .70 | 4219.74235 .2 | 53.5556 .17 |
| 20 | 91.50 | 71.80 | 1281.30 | 4242.9 | 59.09 |
| 22 | 96.00 | 109.00 | 1280.90 | 4227.4 | 38.78 |
| 23 | 101.6095 .40 | 111.70103 .20 | 1280:50 1280.20 | 4212.04200 .5 | 40.7037 .71 |
| 24 | 103.70 | 87.30 | 1280.60 | 4215.9 | 48.29 |
| 25 |  |  |  |  |  |
| 26 | 101.7087 .80 | 73.0075 .30 | 1281.201281 .20 | 4239.0423390 | 58.756 .29 |
| 27 | 88.70 | 68.50 | 1281.40 | 4246.8 | 62.00 |
| 28 | 83.90 | 109.20 | 1281.40 | 4246.8 | 38.89 |
| 29 | 87.40 | 87.40 | 1280.60 | 4215.9 | 48.24 |
| 30 |  |  |  |  |  |
| 31 | 22.80 87.10 | 96.709 .70 | 1280.30 1280.20 | 4204.34200 .5 | 43.4843 .89 |
| Average | 96.34 | 91.86 | 1280.91 | 4227.77 | 47.21 |

Table A. 9 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in September, 1995. Data from CORPs daily summary reports.

| SEPTEMBER |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \end{gathered}$ | $\begin{gathered} \hline \hline \text { INFLOW } \\ (\mathrm{KCFS}) \end{gathered}$ | $\begin{gathered} \hline \hline \text { OUTFLOW } \\ \text { (KCFS) } \end{gathered}$ | RESERVOIR ELEVATION (FT) | $\begin{aligned} & \hline \hline \text { STORAGE } \\ & \text { CAPACITY } \\ & \text { (KCFSD) } \end{aligned}$ | $\begin{aligned} & \hline \text { WATER } \\ & \text { 2ETENTION } \\ & \text { TIME (D) } \end{aligned}$ |
| 1 | 90.20 | 75.40 | 1280.45 | 4210.1 | 55.84 |
| 2 | 82.40 | 55.80 | 1280.90 | 4227.4 | 75.76 |
| 3 | 67.40 | 43.10 | 1281.20 | 4239.0 | 98.35 |
| 4 | 71.40 | 52.80 | 1281.40 | 4246.8 | 80.43 |
| 5 | 84.30 | 67.70 | 1281.70 | 4258.4 | 62.90 |
| 6 | 83.90 | 77.30 | 1281.80 | 4262.3 | 55.14 |
| 7 | 86.90 | 63.00 | 1282.40 | 4285.6 | 68.03 |
| 8 | 90.30 | 60.00 | 1283.10 | 4313.1 | 71.89 |
| 9 | 89.70 | 46.90 | 1283.90 | 4344.6 | 92.64 |
| 10 | 100.90 | 44.50 | 1285.10 | 4392.2 | 98.70 |
| 11 | 101.40 | 69.70 | 1285.90 | 4424.2 | 63.48 |
| 12 | 93.60 | 67.60 | 1286.50 | 4448.4 | 65.81 |
| 13 | 85.80 | 77.90 | 1286.60 | 4452.4 | 57.16 |
| 14 | 87.40 | 73.40 | 1286.90 | 4464.6 | 60.83 |
| 15 | 73.30 | 79.30 | 1286.70 | 4456.5 | 56.20 |
| 16 | 64.50 | 51.80 | 1286.80 | 4460.5 | 86.11 |
| 17 | 55.90 | 62.10 | 1286.40 | 4444.4 | 71.57 |
| 18 | 60.90 | 88.00 | 1285.70 | 4416.2 | 50.18 |
| 19 | 61.50 | 67.50 | 1285.50 | 4408.2 | 65.31 |
| 20 | 69.00 | 68.70 | 1285.60 | 4412.2 | 64.22 |
| 21 | 81.40 | 70.10 | 1285.70 | 4416.2 | 63.00 |
| 22 | 82.90 | 69.10 | 1286.00 | 4428.3 | 64.09 |
| 23 | 88.20 | 57.40 | 1286.50 | 4448.4 | 77.50 |
| 24 | 80.00 | 51.20 | 1286.95 | 4466.6 | 87.24 |
| 25 | 81.70 | 72.20 | 1287.10 | 4472.7 | 61.95 |
| 26 | 81.50 | 69.20 | 1287.40 | 4484.8 | 64.81 |
| 27 | 79.10 | 87.20 | 1287.20 | 4476.7 | 51.34 |
| 28 | 69.60 | 73.60 | 1287.10 | 4472.7 | 60.77 |
| 29 | 79.10 | 80.40 | 1287.05 | 4470.7 | 55.61 |
| 30 | 71.20 | 54.20 | 1287.25 | 4478.8 | 82.64 |
| Average | 79.85 | 65.90 | 1285.09 | 4392.8 | 68.98 |

Table A. 10 Daily midnight reservoir inflow, outflow, elevation, storage capacity, and water retention time for Lake Roosevelt in October, 1995. Data from CORPs daily summary reports.

| OCTOBER |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MONTH } \\ \hline \end{gathered}$ | INFLOW (KCFS) | $\underset{\substack{\text { OUTFLOW } \\ \hline \hline \text { (KCFS) }}}{ }$ | RESERVOIR <br> ELEVATION <br> (FT) | STORAGE CAPACITY (KCFSD) | WATER <br> RETENTION <br> TIME (D) |
| I | 75.30 | 51.70 | 1287.60 | 4493.0 | 86.91 |
| 2 | 75.00 | 95.70 | 1287.00 | 4468.6 | 46.69 |
| 3 | 73.30 | 82.10 | 1286.80 | 4460.5 | 54.33 |
| 4 | 76.70 | 67.70 | 1287.00 | 4468.6 | 66.01 |
| 5 | 81.20 | 69.00 | 1287.30 | 4480.8 | 64.94 |
| 7 | 76.1080 .40 | 75.5073 .10 | 1287.301287 .25 | 4880.64778 .8 | 59.3561 .27 |
| 9 | 73.5085 .60 | 66.5084 .69 | 1287.301287 .10 | 4880.8472 .7 | 67.352 .81 |
| 10 | 75.90 | 82.30 | 1286.90 | 4464.6 | 54.25 |
| 11 | 82.80 | 108.90 | 1286.30 | 4440.3 | 40.77 |
| 12 | 76.50 | 112.60 | 1285.40 | 4404.2 | 39.11 |
| 13 | 74.40 | 83.50 | 1285.10 | 4392.2 | 52.60 |
| 14 | 72.20 | 57.20 | 1285.30 | 4400.2 | 76.93 |
| 15 | 72.00 | 57.00 | 1285.50 | 4408.2 | 77.34 |
| 16 | 76.80 | 91.10 | 1285.10 | 4392.2 | 48.21 |
| 17 | 77.00 | 88.90 | 1284.80 | 4380.3 | 49.27 |
| 18 | 72.60 | 64.70 | 1285.00 | 4388.2 | 67.82 |
| 19 | 86.80 | 78.80 | 1285.20 | 4396.2 | 55.79 |
| 20 | 75.20 | 84.80 | 1285.00 | 4388.2 | 51.75 |
| 21 | 86.00 | 72.30 | 1285.20 | 4396.2 | 60.81 |
| 22 | 94.00 | 86.30 | 1285.20 | 4396.2 | 50.94 |
| 23 |  |  |  |  |  |
| 24 | 84.8080 .00 | 90.708 .100 | 1285,00 1884,90 | 4388.24384 .2 | 48.3852 .19 |
| 25 | 82.30 | 96.20 | 1284.60 | 4372.3 | 45.45 |
| 26 | 82.90 | 78.90 | 1284.70 | 4376.3 | 55.47 |
| 27 | 79.40 | 73.20 | 1284.80 | 4380.3 | 59.84 |
| 28 | 90.00 | 66.20 | 1285.30 | 4400.2 | 66.47 |
| 29 | 100.30 | 88.60 | 1285.50 | 4408.2 | 49.75 |
| 30 |  |  |  |  |  |
| 31 | 86.0082 .00 | 98.0080 .90 | 1288.201283 .10 | 439692 | 44.8650 .54 |
| Average | 80.25 | 80.55 | 1285.80 | 4420.31 | 56.72 |

Table A. 11 Daily midnight reservoir inflow, outflow elevation, storage capacity, and water retention time for Lake Roosevelt in November, 1995. Data from CORPs daily summary reports.

| NOVEMBER |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \hline \text { DAY } \\ \text { OF } \\ \text { MOIITH } \\ \hline \end{gathered}$ | INFLOW (KCFS) | $\begin{aligned} & \hline \hline \text { OUTFLOW } \\ & \text { (KCFS) } \end{aligned}$ | $\begin{array}{\|c\|} \hline \text { RESERVOIR } \\ \text { ELEVATION } \\ \text { (FT) } \end{array}$ | $\begin{aligned} & \hline \hline \text { STORAGE } \\ & \text { CAPACITY } \\ & \text { (KCFSD) } \end{aligned}$ | WATER <br> RETENTION <br> TIME (D) |
| 2 | 75.6062 .40 | 87.5088 .30 | 1288.2012884 .80 | 4380.30 | 50.06 |
| 3 |  |  |  | 4356.40 | 49.34 |
| 4 | 63.804 .000 | 73.9883 .60 | 1283.70 | 4336.70 | 51.87 |
| 5 |  |  | 1283.40 | 4324.80 | 58.52 |
| 6 | 81.2081 .20 | 65.4097 .00 | 1283.8012883 .40 | 4340.60 | 66.37 |
|  |  |  |  | 4324.80 | 44.59 |
| 8 | 81.90 | 81.90 | 1283.40 | 4324.80 | 52.81 |
| 9 | 75.2090 .70 | 79.2080 .90 | 1288,60 1283,30 | 4320.90 | 54.56 |
|  |  |  |  | 4332.70 | 53.56 |
| 10 | 79.00 | 81.00 | 1283.50 | 4328.80 | 53.44 |
| 11 | 87.50 | 59.80 | 1284.20 | 4356.40 | 72.85 |
| 12 | 90.20 | 58.40 | 1285.00 | 4388.2 | 75.14 |
| 13 | 105.40 | 83.40 | 1285.60 | 4412.20 | 52.90 |
| 14 | 99.80 | 73.70 | 1286.20 | 4436.30 | 60.19 |
| 15 | 102.20 | 90.10 | 1286.50 | 4448.40 | 49.37 |
| 17 | 98.60 | 90.50 | 1286.70 | 4456.50 | 49.24 |
| 18 | 108.10109 .9 | 93.7075 .60 | 1287.901287 .10 | 4472.70 | 47.73 |
| 19 |  |  |  | 4505.20 | 59.59 |
| 20 | 119.00115 .0 | 120.0090 .90 | 1288.60 1288.50 | 4533.80 | 49.88 |
| 21 |  |  |  | 4529.70 | 37.75 |
| 22 | 110.10107 .50 | 113.7099 .80 | 1288.40 | 4525.60 | 39.80 |
|  |  |  | 1288.60 | 4533.80 | 45.43 |
| 23 | 97.20 | 76.60 | 1289.10 | 4554.40 | 59.46 |
| 25 | 117.40 | 100.90 | 1289.50 | 4570.90 | 45.30 |
| 26 | 105.60112 .30 | 117.90108 .10 | 1289,20 1289,30 | 4558.50 | 38.66 |
|  |  |  |  | 4562.60 | 42.21 |
| 28 | 118.80 | 122.90 | 1289.20 | 4558.5 | 37.09 |
| 29 | 123.50128 .50 | 114.8013 .50 | 1289,40 1289.20 | 4566.80 | 39.78 |
|  |  |  |  | 4558.50 | 33.64 |
| 30 | 101.70 | 110.80 | 1289.00 | 4550.30 | 41.07 |
| Average | 97.16 | 91.86 | 1286.48 | 4448.34 | 50.41 |

Table A. 12 Daily midnight reservoir inflow, outflow elevation, storage capacity, and water retention time for Lake Roosevelt in December, 1995. Data from CORPs daily summary reports.

| DECEMBER |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DAY <br> OF <br> MONTH | INFLOW <br> (KCFS) | OUTFLOW <br> (KCFS) | RESERVOIR <br> ZLEVATION <br> (FT) | STORAGE <br> CAPACITY <br> (KCFSD) | WATER <br> RETENTION <br> TIME (D) |  |
| 1 | 117.10 | 93.10 | 1289.50 | 4570.90 | 49.10 |  |
| 2 | 119.30 | 117.50 | 1289.50 | 4570.90 | 38.90 |  |
| 3 | 134.30 | 130.50 | 1289.50 | 4570.90 | 35.03 |  |
| 4 | 123.60 | 120.60 | 1289.50 | 4570.90 | 37.90 |  |
| 5 | 133.20 | 137.30 | 1289.40 | 4566.80 | 33.26 |  |
| 6 | 143.10 | 147.30 | 1289.30 | 4562.60 | 30.97 |  |
| 7 | 158.80 | 152.50 | 1289.50 | 4570.90 | 29.97 |  |
| 8 | 154.60 | 169.00 | 1289.10 | 4554.40 | 26.95 |  |
| 9 | 141.90 | 160.10 | 1288.70 | 4537.90 | 28.34 |  |
| 10 | 145.80 | 145.80 | 1288.60 | 4533.80 | 31.10 |  |
| 11 | 147.80 | 146.90 | 1288.55 | 4531.00 | 30.84 |  |
| 12 | 136.10 | 138.10 | 1288.50 | 4529.70 | 32.80 |  |
| 13 | 155.90 | 169.40 | 1288.20 | 4517.50 | 26.67 |  |
| 14 | 144.60 | 140.50 | 1288.25 | 4519.00 | 32.16 |  |
| 15 | 148.70 | 156.40 | 1288.05 | 4511.00 | 28.84 |  |
| 16 | 142.10 | 136.50 | 1288.00 | 4509.30 | 33.04 |  |
| 17 | 152.80 | 145.20 | 1288.00 | 4509.30 | 31.06 |  |
| 18 | 152.10 | 142.70 | 1288.20 | 4517.50 | 31.66 |  |
| 19 | 146.10 | 158.40 | 1287.90 | 4505.20 | 28.44 |  |
| 20 | 131.70 | 158.10 | 1287.30 | 4480.80 | 28.34 |  |
| 21 | 138.70 | 158.90 | 1286.80 | 4460.50 | 28.07 |  |
| 22 | 130.00 | 160.20 | 1286.00 | 4428.30 | 27.64 |  |
| 23 | 126.80 | 164.80 | 1285.10 | 4392.20 | 26.65 |  |
| 24 | 125.60 | 161.40 | 1284.20 | 4356.40 | 26.99 |  |
| 25 | 132.00 | 151.70 | 1283.70 | 4336.70 | 28.59 |  |
| 26 | 125.20 | 156.60 | 1282.90 | 4305.20 | 27.49 |  |
| 27 | 117.60 | 139.10 | 1282.30 | 4281.70 | 30.78 |  |
| 28 | 115.20 | 122.90 | 1282.10 | 4273.90 | 34.78 |  |
| 29 | 113.60 | 119.50 | 1282.30 | 4288.70 | 35.83 |  |
| 30 | 126.30 | 100.90 | 1282.90 | 4305.20 | 42.67 |  |
| 31 | 126.20 | 86.80 | 1283.90 | 4305.20 | 49.60 |  |
|  |  |  |  |  |  |  |
| Average | 135.70 | 141.57 | 1286.96 | 4466.69 | 32.40 |  |
|  |  |  |  |  |  |  |

APPENDIX B ZOOPLANKTON

Table B. $1 \quad$ Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in January, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density $\left(\# / \mathrm{m}^{3}\right)$ | Porcupine Bay Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring Canyon Mean Density $\left(\# / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae |  |  |  |  |
| Daphnia retrocurva |  |  |  |  |
| Daphnia schødleri |  | 21.55 | 66.05 | 192.18 |
| Daphnia thorata |  |  |  |  |
| Daphnia pulex |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangularis |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 0.21 |  |  |  |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 0.42 | 5.99 | 41.20 | 32.71 |
| Leptodora kindtii |  |  |  |  |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 16.72 | 7.68 | 155.02 | 433.99 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis |  |  | 0.59 |  |
| Diacyclops bicuspidatus thomasi | 32.49 | 93.47 | 235.21 | 745.91 |
| Mesocyclop edax |  |  |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 0.42 |  | 50.87 | 51.41 |
| Total Daphnia spp. | 0.00 | 21.55 | 66.10 | 192.20 |
| Total Cladocera | 0.63 | 27.55 | 107.30 | 224.90 |
| Total Copepoda | 49.21 | 101.14 | 390.80 | 1,180.50 |
| Total Nauplii | 0.42 | 0.00 | 50.90 | 51.41 |
| Grand Total | 50.26 | 128.69 | 548.90 | 1,456.80 |

Table B. 2 Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in February, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density (\#/m ${ }^{3}$ ) | $\begin{gathered} \text { Porcupine } \\ \text { Bay } \\ \text { Mean } \\ \text { Density } \\ \left(\# / \mathbf{m}^{3}\right) \\ \hline \end{gathered}$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring Canyon Mean Density (\#/m3) |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae |  |  |  |  |
| Daphnia retrocwva |  |  |  |  |
| Daphnia schødleri | 0.46 | 30.39 | 12.73 | 138.21 |
| Daphnia thorata |  |  |  |  |
| Daphnia pulex |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangular-is |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum |  |  | 0.38 |  |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 1.22 | 76.96 | 9.85 | 25.83 |
| Leptodora kindtii |  |  |  |  |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 6.88 | 16.72 | 25.81 | 269.06 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis |  |  | 0.11 |  |
| Diacyclops bicuspidatus thomasi | 20.49 | 63.61 | 44.90 | 304.51 |
| Mesocyclop edax |  |  |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 0.92 | 3.15 | 11.59 | 50.59 |
| Total Daphnia spp. | 0.46 | 30.39 | 12.70 | 138.20 |
| Total Cladocera | 1.68 | 107.35 | 23.00 | 164.00 |
| Total Copepoda | 27.37 | so. 33 | 70.80 | 573.60 |
| Total Nauplii | 0.92 | 3.15 | 11.60 | 50.59 |
| Grand Total | 29.97 | 190.83 | 105.40 | 788.20 |

Table B. 3 Mean density (\#/m $\mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in March, 1995 at three sampling locations on Lake Roosevelt, WA

|  | Porcupine Bay Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring Canyon Mean Density $\left(\# / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |
| Daphnia galeata mendotae |  |  |  |
| Daphnia retrocwva |  |  |  |
| Daphnia schødleri |  | 0.85 | 16.15 |
| Daphnia thorata |  |  |  |
| Megafenestra aurita |  |  |  |
| Simocephalus serrulatus |  |  |  |
| Alona guttata |  |  |  |
| Alona quadrangular-is |  |  |  |
| Chydorus sphaericus |  |  |  |
| Eurycerus lamellatus |  |  |  |
| Pleuroxus denticulatus |  |  |  |
| Diaphanosoma brachyurum |  |  |  |
| Diapharwsoma birgei |  |  |  |
| Sida crystallina |  |  |  |
| Macrothrix laticornis |  |  |  |
| Streblocerus serricaudatus |  |  |  |
| Bosmina longirostris | 161.43 | 120.64 | 107.05 |
| Leptodora kindtii |  |  |  |
| Eucopepoda |  |  |  |
| Leptodiaptomus ashlandi | 22.09 | 56.93 | 158.03 |
| Skistodiaptomus oregonensis |  |  |  |
| Epischura nevadensis |  |  |  |
| Diacyclops bicuspidatus |  |  |  |
| thomasi | 1,474.92 | 710.28 | 785.04 |
| Mesocyclop edax |  |  |  |
| Bryocamptus spp. |  |  |  |
| Nauplii | 54.38 | 147.83 | 124.04 |
| Total Daphnia spp. | 0.00 | 0.85 | 16.10 |
| Total Cladocera | 161.43 | 121.49 | 123.20 |
| Total Copepoda | 1,497.01 | 767.20 | 943.10 |
| Total Nauplii | 54.38 | 147.83 | 124.04 |
| Grand Total | 1,712.81 | 1,036.52 | 1,190.30 |

Table B. 4 Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in April, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density $\left(\# / \mathrm{m}^{3}\right)$ | $\begin{gathered} \hline \text { Porcupine } \\ \text { Bay } \\ \text { Mean } \\ \text { Density } \\ \left(\# / \mathbf{m}^{3}\right) \end{gathered}$ | Seven Bays Mean Density $\left(\# / \mathrm{m}^{3}\right)$ | Spring Canyon Mean Density $\left(\# / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae <br> Daphnia retrocwva |  |  |  |  |
| Daphnia schødleri | 0.76 | 1.70 | 3.40 | 63.72 |
| Daphnia thorata |  |  |  |  |
| Megajenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangular-is |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum |  |  |  | 0.85 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 19.12 | 259.98 | 749.36 | 228.55 |
| Leptodora kindtii |  |  |  |  |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 24.62 | 5.10 | 101.96 | 382.33 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis |  |  |  |  |
| Diacyclops bicuspidatus thomasi | 60.10 | 880.20 | 602.37 | 1,518.25 |
| Mesocyclop edax |  |  |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 38.54 | 76.46 | 338.15 | 108.76 |
| Total Daphnia spp. | 0.76 | 1.70 | 3.40 | 63.70 |
| Total Cladocera | 19.88 | 261.68 | 752.83 | 293.10 |
| Total Copepoda | 34.72 | 885.29 | 704.32 | 1,900.60 |
| Total Nauplii | 38.54 | 76.46 | 338.11 | 108.76 |
| Grand Total | 143.14 | 1,223.44 | 1,795.24 | 2,302.40 |

Table B. 5 Mean density (\#/m $\mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in May, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density (\#/m ${ }^{3}$ ) | Porcupine <br> Bay <br> Mean <br> Density <br> (\#/m ${ }^{3}$ ) | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring <br> Canyon <br> Mean <br> Density <br> (\#/m ${ }^{3}$ ) |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula Daphnia galeata mendotae |  |  |  |  |
| Daphnia retrocurva |  |  | 29.31 |  |
| Daphnia schodleri | 0.76 | 22.43 | 109.60 | 69.33 |
| Daphnia thorata |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Aiona quadrangular-is |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 0.61 |  | 14.02 | 17.33 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 1.53 | 362.61 | 137.64 | 67.29 |
| Leptodora kindtii | 0.31 |  | 6.37 | 7.14 |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 6.27 | 7.48 | 188.61 | 360.91 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis | 1.53 |  | 21.67 | 35.68 |
| Diacyclops bicuspidatus |  |  |  |  |
| thomasi | 113.02 | 3,835.48 | 4,494.86 | 2,390.80 |
| Mesocyclop edax |  | 9.35 | 11.47 |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 25.23 | 31.78 | 114.70 | 76.46 |
| Total Daphnia spp. | 0.76 | 22.43 | 138.91 | 69.33 |
| Total Cladocera | 3.21 | 385.04 | 296.94 | 161.09 |
| Total Copepoda | 120.81 | 3,852.30 | 4,716.61 | 2,787.40 |
| Total Nauplii | 25.23 | 31.78 | 114.70 | 76.46 |
| Grand Total | 149.26 | 4,269.12 | 5,128.24 | 3,024.95 |

Table B. 6 Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in June, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density $\left(\# / m^{3}\right)$ | $\begin{gathered} \text { Porcupine } \\ \text { Bay } \\ \text { Mean } \\ \text { Density } \\ \left(\# / \mathbf{m}^{\mathbf{3}}\right) \\ \hline \end{gathered}$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring <br> Canyon <br> Mean <br> Density <br> (\#/m³) |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae |  | 11.89 | 50.13 |  |
| Daphnia retrocurva |  | 586.23 | 79.02 | 92.19 |
| Daphnia schødleri | 4.13 | 764.65 | 790.99 | 3,035.66 |
| Daphnia thorata |  |  |  |  |
| Daphnia thorata |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangular-is |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 1.07 | 67.97 | 36.54 | 59.05 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 17.59 | 326. 25 | 45.88 | 7.23 |
| Leptodora kindtii | 0.46 | 69.67 | 79.87 | 64.57 |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 10, 55 | 256.58 | 341.54 | 1,075.19 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis | 2.14 | 365.33 | 15.30 | 20.82 |
| Diacyclops bicuspidatus |  |  |  |  |
| thomasi | 264.87 | 4,978.71 | 2,865.74 | 5,991.45 |
| Mesocyclop edax |  | 33.98 |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nau plii | 64.23 | 66.27 | 36.54 | 17.84 |
| Total Daphnia spp. | 4.13 | 1,362.77 | 920.10 | 3,127.80 |
| Total Cladocera | 23.25 | 1,826.66 | 1,082.40 | 3,258.70 |
| Total Copepoda | 277.57 | 5,634.61 | 3,222.60 | 7,087.40 |
| Total Nauplii | 64.23 | 66.27 | 36.50 | 17.84 |
| Grand Total | 365.04 | 7,527.54 | 4,341.50 | 10,364.00 |

Table B. 7 Mean density (\#/m $\mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in July, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density $\left(\# / m^{3}\right)$ | Porcupine Bay Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring <br> Canyon <br> Mean <br> Density <br> (\#/m ${ }^{3}$ ) |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae | 11.05 | 5.10 | 169.92 | 4.25 |
| Daphnia retrocwva | 216.81 | 40,78 | 52,68 | 118.95 |
| Daphnia schødleri | 1,598.41 | 2,380.61 | 1,536.09 | 3,683.06 |
| Daphnia thorata 2,380.61 1,536.09 3,683.06 |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Aiona quadrangularis |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 26. 57 | 59.47 | 5.10 | 49.28 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 34.06 | 8.50 | 27.19 | 72.22 |
| Leptodora kindtii | 12.59 | 22.09 | 61.17 | 16.15 |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 266.05 | 312.66 | 987.25 | 1,959.20 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis | 40.78 | 42.48 | 37.38 | 54.38 |
| Diacyclops bicuspidatus |  |  |  |  |
| thomasi | 968.47 | 1,092.60 | 2,035.66 | 3,594.70 |
| Mesocyclop edax |  |  |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 43.01 | 13.59 | 13.59 | 113.00 |
| Total Daphnia spp. | 1,826.30 | 2,426.49 | 1,758.69 | 3,806.30 |
| Total Cladocera | 1,899.50 | 2,516.54 | 1,852.15 | 3,942.90 |
| Total Copepoda | 1,275.30 | 1,447.73 | 3.060.29 | 5,608.30 |
| Total Nauplii | 42.30 | 13.59 | 13.59 | 113.00 |
| Grand Total | 3,217-1 | 3,977.87 | 4,926.04 | 9,665.20 |

Table B. 8 Mean density $\left(\# / \mathbf{m}^{\mathbf{3}}\right.$ ) values for zooplankton samples collected in August, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford <br> Mean <br> Density $\left(\# / \mathrm{m}^{3}\right)$ | Porcupine <br> Bay <br> Mean <br> Density <br> (\#/m ${ }^{3}$ ) | Seven Bays Mean Density $\left(\# / m^{3}\right)$ | Spring Canyon Mean Density $\left(\# / m^{3}\right)$ |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae | 700.08 |  | 56.93 | 45.03 |
| Daphnia retrocurva | 1,102.79 | 30.59 | 132.54 | 25.49 |
| Daphnia schødleri | 4,577.70 | 788.44 | 1,666.09 | 1,487.67 |
| Daphnia thorata |  |  |  | 4.25 |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangularis |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 35.68 | 137.64 | 14.45 | 8.50 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 49.28 |  | 25.49 | 7.65 |
| Leptodora kindtii | 74.77 |  | 5.10 | 0.85 |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 688.18 | 1,469.82 | 784.19 | 2,561.58 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis | 336.45 | 158.03 | 79.01 | 108.75 |
| Diacyclops bicuspidatus thomasi | 555.64 | 1,432.44 | 696.68 | 1,293.11 |
| Mesocyclop edax |  |  |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 186.91 | 96.86 | 179.42 | 248.92 |
| Total Daphnia spp. | 6,380.57 | 819.02 | 1,855.60 | 1,562.40 |
| Total Cladocera | 6,540.30 | 956.66 | 1,900.60 | 1,579.40 |
| Total Copepoda | 1,580.27 | 3,060.29 | 1,559.90 | 3,963.40 |
| Total Nauplii | 186.91 | 96.86 | 178.4 | 248.92 |
| Grand Total | 8,307.48 | 4,113.81 | 3,638.90 | 5,791.80 |

Table B. 9 Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in September, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density $\left(\# / m^{3}\right)$ | Porcupine Bay Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Spring <br> Canyon <br> Mean <br> Density <br> (\#/m ${ }^{3}$ ) |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphniaquadranqula |  |  |  |  |
| Daphnia galeata mendotae |  |  |  | 2.55 |
| Daphnia retrocw-va | 156.33 |  |  | 3.40 |
| Daphnia schødleri | 436.70 | 2,276.95 | 1,662.69 | 1,864.05 |
| Daphnia thorata |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangular-is |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum | 5.10 | 231.09 | 25.49 | 5.10 |
| D iaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 1.70 |  | 19.54 | 7.65 |
| Leptodora kindtii |  |  | 5.10 | 1.70 |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 317.75 | 1,870.84 | 2,678.82 | 4,167.34 |
| Skistodiaptomusoregonensis |  |  |  |  |
| Epischura nevadensis | 1.70 | 69.67 | 26.34 | 100. 25 |
| Diacyclops bicuspidatus |  |  |  |  |
| thomasi | 95.16 | 1,172.46 | 977.90 | 947.32 |
| Mesocyclop edax |  | 5.10 |  |  |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 326.25 | 47.58 | 275.28 | 270.18 |
| Total Daphnia spp. | 593.03 | 2,276.95 | 1,662.70 | 1,870.00 |
| Total Cladocera | 599.82 | 2,508.05 | 1,712.80 | 1,884.40 |
| Total Copepoda | 414.61 | 3,118.07 | 3,683.10 | 5,214.90 |
| Total Nauplii | 326.25 | 47.58 | 275.30 | 270.18 |
| Grand Total | 1,340.68 | 5,673.69 | 5,671.10 | 7,369.50 |

Table B. 10 Mean density ( $\# / \mathbf{m}^{\mathbf{3}}$ ) values for zooplankton samples collected in October, 1995 at four sampling locations on Lake Roosevelt, WA

|  | Gifford Mean Density (\#/m³) | Porcupine Bay Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | Seven Bays Mean Density $\left(\# / \mathbf{m}^{3}\right)$ | $\begin{gathered} \text { Spring } \\ \text { Canyon } \\ \text { Mean } \\ \text { Density } \\ \left(\# / \mathrm{m}^{3}\right) \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Ceriodaphnia quadranqula |  |  |  |  |
| Daphnia galeata mendotae | 56.07 |  |  |  |
| Daphnia retrocwva | 61.17 |  |  |  |
| Daphnia schødleri | 457.09 | 1,673.73 | 1,192.43 | 186.92 |
| Daphnia thorata |  |  |  |  |
| Megafenestra aurita |  |  |  |  |
| Simocephalus serrulatus |  |  |  |  |
| Alona guttata |  |  |  |  |
| Alona quadrangularis |  |  |  |  |
| Chydorus sphaericus |  |  |  |  |
| Eurycerus lamellatus |  |  |  |  |
| Pleuroxus denticulatus |  |  |  |  |
| Diaphanosoma brachyurum |  | 33.98 | 5.95 | 22.94 |
| Diaphanosoma birgei |  |  |  |  |
| Sida crystallina |  |  |  |  |
| Macrothrix laticornis |  |  |  |  |
| Streblocerus serricaudatus |  |  |  |  |
| Bosmina longirostris | 54.38 | 22.09 | 56.08 | 37.39 |
| Leptodora kindtii | 1.70 | 1.70 |  |  |
| Eucopepoda |  |  |  |  |
| Leptodiaptomus ashlandi | 86.66 | 971.95 | 1,040.77 | 1,334.74 |
| Skistodiaptomus oregonensis |  |  |  |  |
| Epischura nevadensis | 1.70 |  | 11.47 | 14.45 |
| Diacyclops bicuspidatus |  |  |  |  |
| thomasi | 188.61 | 1,700.92 | 592.61 | 716.22 |
| Bryocamptus spp. |  |  |  |  |
| Nauplii | 134.24 | 35.68 | 36.54 | 11.90 |
| Total Daphnia spp. | 574.34 | 1,673.73 | 1,192.43 | 186.90 |
| Total Cladocera | 630.41 | 1,731.50 | 1,254.50 | 247.20 |
| Total Copepoda | 276.97 | 2,672.87 | 1,644.90 | 2,065.40 |
| Total Nauplii | 134.24 | 35.68 | 36.50 | 11.90 |
| Grand Total | 1,041.62 | 4,440.06 | 2,935.80 | 2,324.50 |

Table B.ll Representative zooplankton size ranges (mm), mean lengths (mm) and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ for samples collected at four locations in January, 1995 on Lake Roosevelt, WA

|  | $\begin{gathered} \text { Size } \\ \text { range } \\ \text { (mm) } \end{gathered}$ | $\begin{gathered} \text { Mean } \\ \text { length } \\ (\mathrm{mm}) \end{gathered}$ | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata menabtae | -- | -- | 0.00 |
| Daphnia retrocurva | . | . | 0.00 |
| Daphnia schødleri | -- | -- | 0.00 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 2 Biomass |  |  | 0.00 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocw-va |  |  | 0.00 |
| Daphnia schødleri | 0.06-0.44 | 0.23 | 2.36 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 2.36 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva | --- |  | 0.00 |
| Daphnia schødleri | 0.12-0.52 | 0.32 | 18.38 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 6 Biomass |  |  | 18.38 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  |  | 0.00 |
| Daphnia schødleri | 0.10-0.64 | 0.28 | 35.10 |
| Daphnia pulex | -- | -- | 0.00 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 9 Biomass |  |  | 35.10 |

Table B. 12 Representative zooplankton size ranges (mm), mean lengths ( mm ) and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right.$ ) for samples collected at four locations in February, 1995 on Lake Roosevelt, WA

|  | $\begin{gathered} \text { Size } \\ \text { range } \\ \text { (mm) } \end{gathered}$ | Mean length (mm) | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schødleri | 0.20-0.34 | 0.26 | 0.07 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 2 Biomass |  |  | 0.07 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva | -- |  | 0.00 |
| Daphnia schødleri | 0.10-0.54 | 0.36 | 13.21 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptoa'ora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 13.21 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva | $\stackrel{--}{0}$ | $\stackrel{-}{-}$ | 0.00 |
| Daphnia schødleri | 0.06-0.60 | 0.22 | 1.03 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptohra kindtii | -- | -- | 0.00 |
| Total Loc 6 Biomass |  |  | 1.03 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schødleri | 0.08-0.52 | 0.31 | 39.27 |
| Daphnia pulex | -- | -- | 0.00 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 9 Biomass |  |  | 39.27 |

Table B. 13 Representative zooplankton size ranges (mm), mean lengths (mm) and biomass values ( $\mu \mathrm{g} / \mathbf{m}^{3}$ ) for samples collected at three locations in March, 1995 on Lake Roosevelt, WA

|  | $\begin{gathered} \text { Size } \\ \text { range } \\ \text { (mm) } \end{gathered}$ | Mean length <br> (mm) | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schødleri | 0.22-0.38 | 0.29 | 0.00 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 0.00 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schødleri | 0.16-0.38 | 0.27 | 0.05 |
| Daphnia thorata | -- | -- | 0.00 |
| - Leptodora_kindtii | -- | -- | 0.00 |
| Total Loc 6 Biomass |  |  | 0.05 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva | --- |  | 0.00 |
| Daphnia schødleri | 0.22-0.60 | 0.36 | 6.98 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 9 Biomass |  |  | 6.98 |

Table B. 14 Representative zooplankton 'size ranges (mm), mean lengths $(\mathrm{mm})$ and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ for samples collected at four locations in April, 1995 on Lake Roosevelt, WA

|  | $\begin{gathered} \hline \text { Size } \\ \text { range } \\ \text { (mm) } \end{gathered}$ | Mean length (mm) | $\begin{aligned} & \text { Biomass } \\ & \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva | -- | -- | 0.00 |
| Daphnia schødleri | 0.06-0.18 | 0.13 | 0.01 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 2 Biomass |  |  | 0.01 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva | -- | -- | 0.00 |
| Daphnia schødleri | -- | 0.16 | 0.06 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 0.06, |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocuwa |  |  | 0.00 |
| Daphnia schødleri | 0.20-0.30 | 0.24 | 0.41 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 6 Biomass |  |  | 0.41 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  | $\stackrel{-}{-}$ | 0.00 |
| Daphnia schødleri | 0.08-0.34 | 0.18 | 3.38 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 9 Biomass |  |  | 3.38 |

Table B. 15 Representative zooplankton size ranges (mm), mean lengths (mm) and biomass values ( $\mu \mathrm{g} / \mathrm{m}^{3}$ ) for samples collected at four locations in May, 1995 on Lake Roosevelt, WA

|  | $\begin{gathered} \text { Size } \\ \text { range } \\ \text { (mm) } \end{gathered}$ | Mean length | $\begin{aligned} & \text { Biomass } \\ & \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  |  | 0.00 |
| Danhnia schodleri | 0.14-0.26 | 0.21 | 0.06 |
| Daphnia thorata | -- ${ }^{-0}$ | 150 | 0.00 |
| Leptodora kindtii | 1 .00-2.00 | 1.50 | 2.32 |
| Total Loc 2 Biomass |  |  | 2.38 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  | $\stackrel{-}{29}$ | 0.00 |
| Daphnia schødleri | 0.14-0.44 | 0.29 | 4.62 |
| Daphnia thorata | - | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 4.62 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae |  | -- | 0.00 |
| Daphnia retrocurva | 0.08-0.24 | 0.12 | 0.17 |
| Daphnia schødleri | O-06-0.38 | 0.16 | 3.69 |
| Davhnia thorata | - | -- | 0.00 |
| Leptodorakindtii | 2.00-5.00 | 3.33 | 69.70 |
| Total Loc 6 Biomass |  |  | 73.56 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  | $\stackrel{-}{-}$ | 0.00 |
| Daphnia schødleri | 0.06-0.36 | 0.17 | 2.80 |
| Daphnia thorata |  |  | 0.00 |
| Leptodora kindtii | 2.00-3.00 | 2.29 | 28.53 |
| Total Loc 9 Biomass |  |  | 31.33 |

Table B. 16 Representative zoopiankton size ranges (mm), mean lengths (mm) and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ for samples collected at four locations in June, 1995 on Lake Roosevelt, WA

|  | Size range (mm) | Mean length (mm) | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva | -- | -- | 0.00 |
| Daphnia schødleri | 0.06-0.40 | 0.13 | 0.08 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 1.00-10.00 | 4.75 | 12.96 |
| Total Loc 2 Biomass |  |  | 13.04 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | 0.20-0.58 | 0.37 | 4.29 |
| Daphnia retrocwva | 0.06-0.54 | 0.19 | 12.65 |
| Daphnia schødleri | 0.10-0.52 | 0.28 | 144.98 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 2.00-7.00 | 4.71 | 1,923.43 |
| Total Loc 4 Biomass |  |  | 2.085 .34 |
| Location 6 Seven Bays $0.0{ }^{\text {a }}$ |  |  |  |
| Daphnia galeata mendotae | 0.04-0.40 | 0.15 | 3.04 |
| Daphnia retrocwva | 0.08-0.42 | 0.26 | 4.19 |
| Daphnia schødleri | 0.06-0.48 | 0.26 | 123.61 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 1.00-13.00 | 4.45 | 2,798.50 |
| Total Loc 6 Biomass |  |  | 2.927 .50 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- |  | 0.00 |
| Daphnia retrocwva | 0.10-0.48 | 0.25 | 4.03 |
| Daphnia schødleri | O-12-0.66 | 0.32 | 1,009.38 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 3.00-15.00 | 6.98 | \$072.41 |
| Total Loc 9 Biomass |  |  | 6.085 .82 |

Table B. 17 Representative zooplankton size ranges (mm), mean lengths (mm) and biomass values ( $\mu \mathrm{g} / \mathrm{m}^{\mathbf{3}}$ ) for samples collected at four locations in July, 1995 on Lake Roosevelt, WA

|  | $\begin{aligned} & \text { Size } \\ & \text { range } \\ & \text { (mm) } \end{aligned}$ | Mean length (mm) | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | O-10-0.38 | 0.25 | 1.49 |
| Daphnia retrocurva | 0.08-0.30 | 0.14 | 2.27 |
| Daphnia schodleri | 0.06-0.50 | 0.18 | 108.12 |
| Daphnia thorata | -- |  | 0.00 |
| Leptodora kindtii | 2.00-13.00 | 451 | 338.00 |
| Total Loc 2 Biomass |  |  | 449.9 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | 0.26-0.48 | 0.37 | 3.02 |
| Daphnia retrocurva | 0.06-0.40 | 0.15 | 0.71 |
| Daphnia schødleri | 0.06-0.58 | 0.17 | 91.33 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 2.00-12.00. | 6.39 | 1,901.41 |
| Total Loc 4 Biomass |  |  |  |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | 0.06-0.34 | 0.20 | 12.71 |
| Daphnia retrocwva | 0.1 O-0.28 | 0.18 | 1.06 |
| Daphnia schødleri | 0.1 o-0.40 | 0.21 | 114.94 |
| Danhnia thorata | -0-1 | -- | 0.00 |
| Leptodora kindtii | 2.00-12.00 | 6.05 | 3,293.15 |
| Total Loc 6 Biomass |  |  | R.-421.85 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | 0.18-0.20 | 0.19 | 0.11 |
| Daphnia retrocurva | 0.06-0.40 | 0.21 | 3.20 |
| Daphnia sch dlleri | 0.1 o-0.50 | 0.26 | 538.15 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 2.00-8.00 | 4.42 | 467.57 |
| Total Loc 9 Biomass |  |  | 1,009.03 |

Table B. 18 Representative zooplankton size ranges (mm), mean lengths ( mm ) and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ for samples collected at four locations in August, 1995 on Lake Roosevelt, WA

|  | $\begin{aligned} & \text { Size } \\ & \text { range } \\ & \text { (mm) } \end{aligned}$ | $\begin{gathered} \text { Mean } \\ \text { length } \\ \text { (mm) } \end{gathered}$ | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | 0.10 .0 .44 | 0.25 | 91.13 |
| Daphnia retrocurva | 0.1 o-0.40 | 0.22 | 38.19 |
| Daphnia schødleri | 0.16-0.50 | 0.27 | 773.90 |
| Daphnia thorata |  |  | 0.00 |
| Leptodora kindtii | 3.00-9.00 | 5.49 | 3,102.54 |
| Total Loc 2 Biomass |  |  | 4,005.77 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | $\cdots$ | 0.00 |
| Daphnia retrocurva | 0.10-0.48 | 0.22 | 1.11 |
| Daphnia schodleri | 0.10-0.56 | 0.33 | 260.96 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 262.07 |
| Location 6 Seven Bays 262.07 |  |  |  |
| Daphnia galeata mendotae | 0.10-0.42 | 0.24 | 6.77 |
| Daphnia retrocwva | 0.10-0.46 | 0.21 | 5.61 |
| Daphnia schødleri | 0.1 O-0.60 | 0.24 | 241.86 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | 3.00-14.00 | 6.38 | 459.69 |
| Total Loc 6 Biomass |  |  | 713.91 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | 0.20-0.54 | 0.31 | 10.08 |
| Daphnia retrocwva | 0.12-0.28 | 0.21 | 0.81 |
| Daphnia schodleri | 0,08.0,52 | 0.31 | 586.35 |
| Daphnia thorata | $0.18 \cdot 0.48$ | 0.40 | 0.18 |
| Leptodora kindtii | -- | 3.00 | 7.02 |
| Total Loc 9 Biomass |  |  | 604.40 |

Table B. 19 Representative zooplankton size ranges (mm), mean Iengths (mm) and biomass values ( $\mu \mathrm{g} / \mathrm{m}^{3}$ ) for samples collected at four locations in September, 1995 on Lake Roosevelt, WA

|  | $\begin{aligned} & \text { Size } \\ & \text { range } \\ & \text { (mm) } \end{aligned}$ | Mean length (mm) | $\begin{aligned} & \text { Biomass } \\ & \left(\mu \mathrm{g} / \mathrm{m}^{3}\right) \end{aligned}$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae |  | -- | 0.00 |
| Daphnia retrocwva | 0.08-0.40 | 0.17 | 2.79 |
| Daphnia schodleri | 0.10-0.52 | 0.23 | 44.53 |
| Daphnia thorata |  | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 2 Biomass |  |  | 47.32 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  |  | 0.00 |
| Daphnia schodleri | 0.14-0.54 | 0.34 | 770.10 |
| Daphnia thorata |  |  | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 4 Biomass |  |  | 770.10 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | $0.00{ }^{\text { }}$ |
| Daphnia retrocurva | -- |  | 0.00 |
| Daphnia schødleri | 0.12-0.58 | 0.31 | 713.17 |
| Daphnia thorata | ---7 | -- | 0.00 |
| Leptodora kindtii | 3.00-7.00 | 4.45 | 156.08 |
| Total Loc 6 Biomass |  |  | 869.25 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia gaieata mendotae | 0.18-0.26 | 0.22 | 0.23 |
| Daphnia retrocurva | 0.16-0.24 | 0.19 | 0.09 |
| Daphnia schodleri | 0.10-0.56 | 0.32 | 604.75 |
| Daphnia thorata |  |  | 0.00 |
| Leptodora kindtii | -- | 4.00 | 30.27 |
| Total Loc 9 Biomass |  |  | 635.30 |

Table B. 20 Representative zooplankton size ranges (mm), mean lengths $(\mathrm{mm})$ and biomass values $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ for samples collected at four locations in October, 1995 on Lake Roosevelt, WA

|  | $\begin{aligned} & \text { Size } \\ & \text { range } \\ & \text { (mm) } \end{aligned}$ | $\begin{aligned} & \text { Mean } \\ & \text { length } \\ & \text { (mm) } \end{aligned}$ | Biomass $\left(\mu \mathrm{g} / \mathrm{m}^{3}\right)$ |
| :---: | :---: | :---: | :---: |
| Location 2 Gifford |  |  |  |
| Daphnia galeata mendotae | 0.10-0.46 | 0.27 | 9.00 |
| Daphnia retrocwva | 0.06-0.32 | 0.22 | 0.41 |
| Daphnia schødleri | 0.12-0.52 | 0.27 | 76.63 |
| Daphnia thorata | -- | -- |  |
| Leptodora kindtii | -- | 10.00 | 349.51 |
| Total Loc 2 Biomass |  |  | 435.56 |
| Location 4 Porcupine Bay |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocurva |  | -- | 0.00 |
| Daphnia schødleri | 0.10-0.50 | 0.28 | 323.86 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | 3.00 | 14.04 |
| Total Loc 4 Biomass |  |  | 337.90 |
| Location 6 Seven Bays |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schodleri | 0.10-0.52 | 0.29 | 292.84 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 6 Biomass |  |  | 292.84 |
| Location 9 Spring Canyon |  |  |  |
| Daphnia galeata mendotae | -- | -- | 0.00 |
| Daphnia retrocwva |  |  | 0.00 |
| Daphnia schødleri | 0.10-0.50 | 0.24 | 17.45 |
| Daphnia thorata | -- | -- | 0.00 |
| Leptodora kindtii | -- | -- | 0.00 |
| Total Loc 9 Biomass |  |  | 17.45 |

## APPENDIX C

WATER QUALITY

Tables C.1-C. 4 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Seven Bays, Keller Ferry and Spring Canyon in January, 1995.
Table C.l;

## GIFFORD

| $\begin{gathered} \text { Depth } \\ (\mathrm{m}) \end{gathered}$ | $\begin{gathered} \text { Temp. } \\ \left({ }^{\circ} \mathrm{C}\right) \end{gathered}$ | pH | $\begin{gathered} \mathrm{D} . \mathrm{O} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Conduct. mmho/cm | $\begin{aligned} & \text { ORP } \\ & \text { (V) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 2.30 | 9.37* | 13.33 | 0.171 | 0.247 |
| 6 | 2.332 .32 | 9.38* | 13.711 .79 | 0.1500 .166 | 0.249 |
|  |  |  |  |  | 0.250 |
| 9 | 2.32 | 9.37* | 11.53 | 0.162 | 0.251 |
| 12 | 2.32 | 9.38 " | 11.66 | 0.146 | 0.252 |
| 15 | 2.33 | 9.37* | 11.57 | 0.182 | 0.253 |
| 18 | 2.30 | 9.37* | 11.46 | 0.153 | 0.254 |
| 21 |  |  |  |  |  |
| 24 | 2.322 .30 | 9.34* | 11.4911 .50 | 0.1860 .17 | 0.2540 .255 |
| 27 | 2.32 | 9.33* | 11.50 | 0.186 | 0.256 |
| 30 |  |  |  |  |  |
| 33 | 2.332 .33 | 9.31* | 11.4911 .43 | 0.1560 .134 | 0.2570 .258 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.2;

## SEVEN BAYS

| $\begin{gathered} \text { Depth } \\ (\mathbf{m}) \end{gathered}$ | $\begin{aligned} & \text { Temp. } \\ & \left({ }^{\circ} \mathrm{C}\right) \end{aligned}$ | pH | $\begin{gathered} \mathrm{D} . \mathrm{O} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | conduct. mmho/cm | $\begin{aligned} & \hline \text { ORP } \\ & (\mathrm{V}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.52 | 9.66" | 11.70 | 0.164 | 0.214 |
| 3 |  |  |  |  |  |
| 6 | 2.22 .52 | 9.74* | 13.95120 | 0.1520 .163 | 02140.20 |
| 9 | 2.57 | 9.78 " | 11.82 | 0.167 | 0.208 |
| 12 | 2.60 | 9.75 " | 11.79 | 0.160 | 0.210 |
| 15 |  |  |  |  |  |
| 18 | 2.64 .60 | 9.791\% | 11.111 .16 | 0.184 .1 .19 | 02120214 |
| 21 | 2.64 | 9.67 " | 11.67 | 0.148 | 0.215 |
| 24 | 2.64 | 9.64* | 11.76 | 0.178 | 0.217 |
| 27 | 2.67 | 9.61 " | 11.64 | 0.160 | 0.218 |
| 30 33 | 2722.76 | 9.56* | 11.511 .58 | 0.1480 .137 | 02280.21 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.3;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.82 | $9.79^{*}$ | 13.02 | 0.160 | 0.221 |
| 3 | 2.84 | $9.76^{*}$ | 12.00 | 0.165 | 0.219 |
| 6 |  |  |  |  |  |
| 9 | 2.82 .84 | $9.80^{*}$ | 12.0412 .38 | 0.1710 .158 | 0.2170 .212 |
| 12 | 2.86 | $9.90^{*}$ | 12.45 | 0.152 | 0.214 |
| 15 |  |  |  |  |  |
| 18 | 2.892 .96 | $9.85^{*}$ | 12.5112 .63 | 0.1730 .124 | 0.2150 .216 |
| 21 | 3.02 | $9.82^{*}$ | 12.79 | 0.146 | 0.217 |
| 24 | 3.01 | $9.80^{*}$ | 12.81 | 0.155 | 0.218 |
| 27 | 2.99 | $9.78^{*}$ | 12.70 | 0.151 | 0.219 |
| 30 | 3.01 | $9.77^{*}$ | 12.57 | 0.171 | 0.221 |
| 33 | 3.03 | $9.86^{*}$ | 12.49 | 0.194 | 0.221 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.4;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.27 | $9.46^{*}$ | 13.28 | 0.165 | 0.217 |
| 3 |  |  |  |  |  |
| 6 | 3.273 .24 | $9.63^{*}$ | 11.2611 .59 | 0.1660 .171 | 0.2000 .207 |
| 9 | 3.22 | $9.67^{*}$ | 11.72 | 0.155 | 0.209 |
| 12 | 3.26 | $9.66^{*}$ | 11.74 | 0.159 | 0.210 |
| 15 | 3.24 | $9.67^{\prime \prime}$ | 12.04 | 0.166 | 0.211 |
| 18 | 3.27 | $9.62^{*}$ | 12.13 | 0.174 | 0.213 |
| 21 |  |  |  |  |  |
| 24 | 3.263 .24 | $9.60^{*}$ | 12.1112 .18 | 0.200 |  |
| 27 | 3.33 | $9.56^{*}$ | 12.21 | 0.155 | 0.2140 .215 |
| 30 | 3.34 | $9.55^{*}$ | 12.21 | 0.166 | 0.217 |
| 33 | 3.44 | $9.54^{*}$ | 12.25 | 0.104 | 0.218 |

* Indicates suspect pH readings due to sensor malfunction.

Tables C.5-C. 10 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Porcupine Bay, Confluence, Seven Bays, Keller Ferry and Spring Canyon in February, 1995.
Table C. 5
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> mmho/ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.55 | $8.56^{*}$ | 16.10 | 0.173 | 0.297 |
| 3 | 2.54 | $8.69^{*}$ | 12.60 | 0.172 | 0.288 |
| 6 | 2.54 | $8.4^{*}$ | 12.31 | 0.175 | 0.285 |
| 9 | 2.54 | $9.18^{*}$ | 12.28 | 0.175 | 0.273 |
| 12 | 2.57 | $9.30^{*}$ | 12.76 | 0.170 | 0.272 |
| 15 | 2.55 | $9.34^{*}$ | 12.21 | 0.182 | 0.271 |
| 18 | 2.55 | $9.37^{*}$ | 12.10 | 0.193 | 0.272 |
| 21 | 2.55 | $9.41^{*}$ | 12.06 | 0.158 | 0.271 |

* Indicates suspect pH readings due to sensor malfunction.

Table C. 6
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.26 | $9.05^{*}$ | 18.20 | 0.131 | 0.263 |
| 3 | 3.33 | $9.1^{*}$ | 12.86 | 0.129 | 0.259 |
| 6 | 3.41 | $9.8^{*}$ | 12.00 | 0.128 | 0.256 |
| 9 | 3.43 | $9.59^{*}$ | 11.88 | 0.129 | 0.234 |
| 12 | 3.48 | $9.57^{*}$ | 11.62 | 0.125 | 0.234 |
| 15 | 3.54 | $9.56^{*}$ | 11.54 | 0.127 | 0.235 |
| 18 | 3.59 | $9.56^{*}$ | 11.32 | 0.127 | 0.235 |
| 21 | 3.61 | $9.54^{*}$ | 11.43 | 0.124 | 0.236 |
| 24 | 3.63 | $9.53^{*}$ | 11.55 | 0.125 | 0.237 |
| 27 | 3.63 | $9.52^{*}$ | 11.44 | 0.122 | 0.237 |
| 30 | 3.63 | $9.52^{*}$ | 11.50 | 0.121 | 0.237 |
| 33 | 3.61 | $9.51^{*}$ | 11.51 | 0.122 | 0.238 |

* Indicates suspect pH readings due to sensor malfunction.

Table C. 7

## CONFLUENCE

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.67 | $8.82^{*}$ | 14.08 | 0.159 | 0.251 |
| 3 | 2.62 | $9.4^{*}$ | 11.85 | 0.155 | 0.241 |
| 6 | 2.70 | $9.34^{*}$ | 11.41 | 0.147 | 0.227 |
| 9 | 2.77 | $9.5^{*}$ | 11.34 | 0.156 | 0.226 |
| 12 | 3.06 | $9.46^{*}$ | 11.32 | 0.148 | 0.226 |
| 15 | 3.09 | $9.50^{*}$ | 11.21 | 0.149 | 0.229 |
| 18 | 3.09 | $9.50^{*}$ | 11.26 | 0.147 | 0.227 |
| 21 | 3.07 | $9.51^{*}$ | 11.26 | 0.146 | 0.227 |
| 24 | 3.09 | $9.52^{*}$ | 11.31 | 0.148 | 0.228 |
| 27 | 3.06 | $9.52^{*}$ | 11.31 | 0.147 | 0.229 |
| 30 | 3.07 | $9.52^{*}$ | 11.31 | 0.143 | 0.230 |
| 33 | 3.07 | $9.52^{*}$ | 11.41 | 0.144 | 0.230 |

* Indicates suspect pH readings due to sensor malfunction.

Table C. 8
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{0}$ <br> $(\&$ i i $)$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.67 | -- | 12.49 | 0.164 | 0.181 |
| 3 | 2.64 | - | 12.09 | 0.166 | 0.182 |
| 6 | 2.59 | -- | 11.99 | 0.178 | 0.179 |
| 9 | 2.57 | -- | 11.96 | 0.156 | 0.180 |
| 12 | 2.60 | -- | 11.94 | 0.148 | 0.182 |
| 15 | 2.59 | -- | 11.94 | 0.185 | 0.183 |
| 18 | 2.59 | -- | 11.92 | 0.173 | 0.184 |
| 21 | 2.62 | -- | 1.93 | 0.153 | 0.185 |
| 24 | 2.64 | -- | 11.92 | 0.161 | 0.185 |
| 27 | 2.64 | -- | 11.92 | 0.162 | 0.186 |
| 30 | 2.64 | -- | 11.95 | 0.166 | 0.187 |
| 33 | 2.64 | -- | 11.49 | 0.140 | 0.187 |

Table C. 9
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.55 | - | 17.12 | 0.164 | 0.289 |
| 3 | 2.55 | - | 11.64 | 0.161 | 0.288 |
| 6 | 2.55 | - | 11.40 | 0.155 | 0.286 |
| 9 | 2.55 | - | 11.30 | 0.152 | 0.268 |
| 12 | 2.55 | - | 11.29 | 0.153 | 0.271 |
| 15 | 2.57 | - | 11.29 | 0.166 | 0.273 |
| 18 | 2.59 | -- | 11.36 | 0.146 | 0.274 |
| 21 | 2.59 | - | 11.22 | 0.164 | 0.276 |
| 24 | 2.59 | -- | 11.24 | 0.181 | 0.277 |
| 27 | 2.59 | -- | 11.23 | 0.146 | 0.279 |
| 30 | 2.60 | -- | 11.24 | 0.155 | 0.279 |
| 33 | 2.60 | -- | 11.15 | 0.170 | 0.281 |

Table C. 10
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2.47 | -- | 12.75 | 0.162 | 0.183 |
| 3 | 2.45 | - | 12.06 | 0.167 | 0.183 |
| 6 | 2.45 | -- | 11.96 | 0.170 | 0.182 |
| 9 | 2.45 | - | 11.93 | 0.168 | 0.184 |
| 12 | 2.44 | - | 11.86 | 0.198 | 0.186 |
| 15 | 2.44 | - | 12.05 | 0.166 | 0.188 |
| 18 | 2.42 | - | 12.59 | 0.170 | 0.189 |
| 21 | 2.40 | - | 11.80 | 0.159 | 0.191 |
| 24 | 2.39 | - | 11.76 | 0.184 | 0.192 |
| 27 | 2.39 | - | 11.75 | 0.172 | 0.193 |
| 30 | 2.39 | - | 11.75 | 0.191 | 0.193 |
| 33 | 2.39 | -- | 11.76 | 0.196 | 0.194 |

Tables C.11-C. 15 Water quality measurements taken with a Hydrolab Surveyor II at Porcupine Bay, Confluence, Seven Bays, Keller Ferry and Spring Canyon in March, 1995.
Table C.11;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.87 | $10.79^{*}$ | 15.41 | 0.084 | 0.179 |
| 3 | 3.83 | $10.71^{*}$ | 17.94 | 0.084 | 0.181 |
| 6 | 3.84 | $10.67^{*}$ | 17.40 | 0.084 | 0.182 |
| 9 | 3.81 | $10.69^{*}$ | 16.23 | 0.082 | 0.180 |
| 12 | 3.83 | $10.71^{*}$ | 13.67 | 0.085 | 0.182 |
| 15 | 3.84 | $10.60^{*}$ | 13.27 | 0.082 | 0.188 |
| 18 | 3.83 | $10.55^{*}$ | 19.63 | 0.085 | 0.190 |
| 21 | 3.84 | $10.55^{*}$ | 21.10 | 0.085 | 0.192 |
| 24 | 3.89 | $10.50^{*}$ | 13.73 | 0.086 | 0.193 |
| 27 | 3.89 | $10.46^{*}$ | 13.05 | 0.081 | 0.196 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.12;
CONFLUENCE

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D. $\mathbf{O}$. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 4.24 | $10.63^{*}$ | 14.65 | 0.094 | 0.198 |
| 3 | 4.11 | $10.60^{*}$ | 12.87 | 0.095 | 0.199 |
| 6 | 4.06 | $10.54^{*}$ | 12.75 | 0.095 | 0.199 |
| 9 | 4.03 | $10.65^{*}$ | 12.65 | 0.099 | 0.197 |
| 12 | 3.98 | $10.59^{*}$ | 12.58 | 0.106 | 0.200 |
| 15 | 3.99 | $10.58^{*}$ | 12.58 | 0.105 | 0.202 |
| 18 | 3.98 | $10.58^{*}$ | 12.54 | 0.107 | 0.204 |
| 21 | 3.98 | $10.55^{*}$ | 12.55 | 0.109 | 0.205 |
| 24 | 3.90 | $10.54^{*}$ | 12.55 | 0.109 | 0.206 |
| 27 | 3.98 | $10.53^{*}$ | 12.60 | 0.106 | 0.207 |
| 30 | 3.98 | $10.52^{*}$ | 12.59 | 0.103 | 0.208 |
| 33 | 3.94 | $10.51^{*}$ | 12.57 | 0.110 | 0.209 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.13;

## SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.49 | $11.00^{*}$ | 16.31 | 0.154 | 0.216 |
| 3 | 3.44 | $11.02^{*}$ | 13.00 | 0.157 | 0.225 |
| 6 | 3.48 | $10.86^{*}$ | 14.56 | 0.160 | 0.226 |
| 9 | 3.38 | $10.85^{*}$ | 18.44 | 0.159 | 0.222 |
| 12 | 3.38 | $10.94^{*}$ | 12.50 | 0.172 | 0.224 |
| 15 | 3.58 | $10.93^{*}$ | 12.22 | 0.173 | 0.226 |
| 18 | 3.26 | $10.89^{*}$ | 12.15 | 0.179 | 0.228 |
| 21 | 3.24 | $10.89^{*}$ | 11.96 | 0.185 | 0.229 |
| 24 | 3.24 | $10.91^{*}$ | 13.73 | 0.169 | 0.229 |
| 27 | 3.26 | $10.91^{*}$ | 15.64 | 0.166 | 0.232 |
| 30 | 3.22 | $10.90^{*}$ | 12.66 | 0.186 | 0.233 |
| 33 | 3.22 | $10.91^{*}$ | 13.18 | 0.170 | 0.237 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.14;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.93 | $11.18^{*}$ | 19.40 | 0.160 | 0.166 |
| 3 | 3.94 | $11.12^{*}$ | 12.13 | 0.161 | 0.170 |
| 6 | 3.81 | $11.08^{*}$ | 11.63 | 0.146 | 0.170 |
| 9. | 3.81 | $11.14^{*}$ | 11.49 | 0.174 | 0.167 |
| 12 | 3.81 | $11.12^{*}$ | 11.45 | 0.154 | 0.168 |
| 15 | 3.79 | $11.11^{*}$ | 11.45 | 0.154 | 0.169 |
| 18 | 3.79 | $11.08^{*}$ | 11.50 | 0.157 | 0.169 |
| 21 | 3.79 | $11.07^{*}$ | 11.50 | 0.145 | 0.170 |
| 24 | 3.79 | $11.06^{*}$ | 11.50 | 0.167 | 0.170 |
| 27 | 3.79 | $11.05^{*}$ | 11.43 | 0.147 | 0.171 |
| 30 | 3.81 | $11.05^{*}$ | 11.44 | 0.167 | 0.171 |
| 33 | 3.81 | $11.05^{*}$ | 11.44 | 0.164 | 0.171 |

[^2]Table C.15;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 3.58 | $10.84^{*}$ | 15.98 | 0.158 | 0.253 |
| 3 | 3.56 | $10.80^{*}$ | 12.23 | 0.154 | 0.263 |
| 6 | 3.54 | $10.92^{*}$ | 11.85 | 0.158 | 0.265 |
| 9 | 3.56 | $10.87^{*}$ | 11.78 | 0.140 | 0.273 |
| 12 | 3.56 | $10.85^{*}$ | 11.71 | 0.158 | 0.279 |
| 15 | 3.54 | $10.83^{*}$ | 11.69 | 0.155 | 0.285 |
| 18 | 3.54 | $10.81^{*}$ | 11.66 | 0.172 | 0.288 |
| 21 | 3.54 | $10.81^{*}$ | 11.63 | 0.157 | 0.292 |
| 24 | 3.54 | $10.80^{*}$ | 11.65 | 0.152 | 0.294 |
| 27 | 3.54 | $10.80^{*}$ | 11.63 | 0.155 | 0.297 |
| 30 | 3.56 | $10.79^{*}$ | 11.62 | 0.153 | 0.300 |
| 33 | 3.54 | $10.79^{*}$ | 11.65 | 0.154 | 0.302 |

* Indicates suspect pH readings due to sensor malfunction.

Tables C.16-C. 21 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Porcupine Bay, Confluence, Seven Bays, Keller Ferry and Spring Canyon in April, 1995.

Table C.16;
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 6.14 | - | 11.78 | 0.178 | 0.200 |
| 3 | 5.93 | - | 11.81 | 0.174 | 0.199 |
| 6 | 5.81 | - | 11.80 | 0.172 | 0.200 |
| 9 | 5.60 | - | 11.71 | 0.165 | 0.202 |
| 12 | 5.57 | - | 11.59 | 0.168 | 0.203 |
| 15 | 5.57 | - | 11.53 | 0.166 | 0.204 |
| 18 | 5.53 | -- | 11.42 | 0.154 | 0.205 |
| 21 | 5.52 | - | 11.43 | 0.172 | 0.205 |
| 24 | 5.50 | - | 11.53 | 0.183 | 0.205 |
| 27 | 5.45 | -- | 11.49 | 0.190 | 0.206 |

Table C.17;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 7.69 | $11.19^{*}$ | 11.70 | 0.092 | 0.205 |
| 3 | 6.78 | $11.19^{*}$ | 12.00 | 0.091 | 0.207 |
| 6 | 6.69 | $11.16^{*}$ | 11.82 | 0.091 | 0.208 |
| 9 | 6.58 | $11.07^{*}$ | 11.35 | 0.092 | 0.214 |
| 12 | 6.52 | $11.05^{*}$ | 11.50 | 0.092 | 0.216 |
| 15 | 6.51 | $11.01^{*}$ | 11.73 | 0.089 | 0.219 |
| 18 | 6.42 | $10.98^{*}$ | 11.21 | 0.089 | 0.223 |
| 21 | 6.36 | $10.97^{*}$ | 12.17 | 0.090 | 0.225 |
| 24 | 6.32 | $10.94^{*}$ | 11.60 | 0.091 | 0.227 |
| 27 | 6.13 | $10.91^{*}$ | 11.85 | 0.090 | 0.229 |
| 30 | 5.88 | $10.90^{*}$ | 11.71 | 0.091 | 0.232 |
| 33 | 5.88 | $10.94^{*}$ | 11.70 | 0.091 | 0.232 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.18;
CONFLUENCE

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 6.78 | -- | 11.75 | 0.156 | 0.230 |
| 3 | 6.26 | - | 11.86 | 0.161 | 0.230 |
| 6 | 6.18 | - | 12.76 | 0.183 | 0.231 |
| 9 | 6.04 | - | 12.83 | 0.145 | 0.233 |
| 12 | 5.96 | - | 12.30 | 0.145 | 0.236 |
| 15 | 5.95 | - | 11.93 | 0.121 | 0.237 |
| 18 | 5.96 | - | 11.81 | 0.140 | 0.239 |
| 21 | 5.95 | - | 11.81 | 0.195 | 0.240 |
| 24 | 5.91 | - | 11.85 | 0.165 | 0.241 |
| 27 | 5.80 | -- | 11.78 | 0.169 | 0.242 |
| 30 | 5.14 | -- | 11.75 | 0.161 | 0.243 |
| 33 | 4.79 | -- | 11.56 | 0.172 | 0.243 |

Table C.19;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 7.06 | $10.95^{*}$ | 16.09 | 0.139 | 0.195 |
| 3 | 6.04 | $11.63^{*}$ | 16.04 | 0.138 | 0.195 |
| 6 | 6.03 | $11.09^{*}$ | 11.40 | 0.142 | 0.195 |
| 9 | 5.86 | $11.11^{*}$ | 11.45 | 0.142 | 0.195 |
| 12 | 5.73 | $11.08^{*}$ | 11.53 | 0.149 | 0.197 |
| 15 | 5.68 | $11.08^{*}$ | 11.50 | 0.149 | 0.198 |
| 18 | 5.63 | $11.07^{*}$ | 11.37 | 0.146 | 0.198 |
| 21 | 5.58 | $11.09^{*}$ | 11.35 | 0.145 | 0.199 |
| 24 | 5.42 | $11.08^{*}$ | 11.40 | 0.150 | 0.199 |
| 27 | 5.30 | $11.09^{*}$ | 11.41 | 0.152 | 0.200 |
| 30 | 5.10 | $11.10^{*}$ | 11.26 | 0.146 | 0.200 |
| 33 | 4.92 | $11.11^{*}$ | 11.27 | 0.153 | 0.201 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.20;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 4.76 | $10.98^{*}$ | 18.08 | 0.164 | 0.250 |
| 3 | 4.76 | $11.01^{*}$ | 12.55 | 0.162 | 0.248 |
| 6 | 4.74 | $11.05^{*}$ | 11.58 | 0.157 | 0.248 |
| 9 | 4.74 | $11.07^{*}$ | 11.53 | 0.167 | 0.248 |
| 12 | 4.72 | $11.06^{*}$ | 11.49 | 0.158 | 0.249 |
| 15 | 4.72 | $11.03^{*}$ | 11.49 | 0.171 | 0.251 |
| 18 | 4.72 | $11.03^{*}$ | 11.49 | 0.155 | 0.252 |
| 21 | 4.72 | $11.03^{*}$ | 11.46 | 0.166 | 0.253 |
| 24 | 4.72 | $11.03^{*}$ | 11.41 | 0.166 | 0.253 |
| 27 | 4.67 | $11.03^{*}$ | 11.45 | 0.179 | 0.254 |
| 30 | 4.66 | $11.04^{*}$ | 11.47 | 0.166 | 0.254 |
| 33 | 4.64 | $11.04^{*}$ | 11.46 | 0.130 | 0.255 |

* Indicates suspect pH readings due to sensor malfunction.

Table C.21;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 6.08 | - | 11.56 | 0.163 | 0.235 |
| 3 | 6.01 | -- | 11.57 | 0.172 | 0.234 |
| 6 | 6.01 | -- | 11.51 | 0.149 | 0.230 |
| 9 | 6.00 | - | 11.51 | 0.151 | 0.232 |
| 12 | 6.00 | - | 11.45 | 0.151 | 0.232 |
| 15 | 5.95 | - | 11.45 | 0.168 | 0.233 |
| 18 | 5.95 | - | 11.46 | 0.165 | 0.233 |
| 21 | 5.90 | - | 11.42 | 0.147 | 0.235 |
| 24 | 5.91 | -- | 11.40 | 0.143 | 0.236 |
| 27 | 5.86 | -- | 11.43 | 0.151 | 0.236 |
| 30 | 5.73 | - | 11.39 | 0.164 | 0.237 |
| 33 | 5.73 | - | 11.32 | 0.153 | 0.237 |

Tables C.22-C. 29 Water quality measurements taken with a Hydrolab Surveyor II at Kettle Falls, Gifford, Hunters, Porcupine Bay, Seven Bays, Keller Ferry, San Poil and Spring Canyon in May, 1995.
Table C.22;
KETTLE FALLS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 9.82 | 7.83 | 11.69 | 0.134 | 0.144 |
| 3 | 9.72 | 7.82 | 11.76 | 0.134 | 0.148 |
| 6 | 9.71 | 7.81 | 11.77 | 0.133 | 0.150 |
| 9 | 9.67 | 7.82 | 11.86 | 0.135 | 0.152 |
| 12 | 9.66 | 7.82 | 11.86 | 0.134 | 0.154 |
| 15 | 9.66 | 7.82 | 11.87 | 0.132 | 0.155 |
| 18 | 9.61 | 7.83 | 11.89 | 0.134 | 0.156 |
| 21 | 9.60 | 7.84 | 11.90 | 0.134 | 0.158 |
| 24 | 9.60 | 7.84 | 11.85 | 0.133 | 0.159 |
| 27 | 9.60 | 7.84 | 11.89 | 0.133 | 0.161 |
| 30 | 9.62 | 7.85 | 11.89 | 0.133 | 0.162 |

Table C.23;

## GIFFORD

| Depth <br> $(\mathrm{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathrm{mg} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 11.57 | 8.05 | 11.34 | 0.142 | 0.180 |
| 3 | 11.41 | 8.03 | 11.38 | 0.143 | 0.181 |
| 6 | 11.21 | 7.98 | 11.36 | 0.142 | 0.183 |
| 9 | 11.20 | 7.95 | 11.32 | 0.142 | 0.184 |
| 12 | 11.18 | 7.94 | 11.32 | 0.140 | 0.185 |
| 15 | 11.05 | 7.92 | 11.31 | 0.140 | 0.186 |
| 18 | 10.70 | 7.89 | 11.28 | 0.142 | 0.187 |
| 21 | 10.56 | 7.87 | 11.29 | 0.140 | 0.187 |
| 24 | 10.48 | 7.85 | 1.29 | 0.139 | 0.189 |
| 27 | 10.42 | 7.84 | 11.28 | 0.140 | 0.190 |
| 30 | 10.37 | 7.83 | 11.26 | 0.142 | 0.190 |
| 33 | 10.37 | 7.83 | 11.26 | 0.141 | 0.191 |

Table C.24;
HUNTERS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 11.75 | 8.00 | 11.36 | 0.146 | 0.130 |
| 3 | 11.13 | 7.99 | 11.36 | 0.146 | 0.136 |
| 6 | 10.82 | 7.94 | 11.34 | 0.146 | 0.140 |
| 9 | 10.76 | 7.93 | 11.35 | 0.145 | 0.142 |
| 12 | 10.86 | 7.94 | 1.33 | 0.145 | 0.144 |
| 15 | 10.70 | 7.92 | 11.29 | 0.145 | 0.145 |
| 18 | 10.57 | 7.90 | 11.28 | 0.143 | 0.147 |
| 21 | 10.53 | 7.89 | 11.24 | 0.144 | 0.148 |
| 24 | 10.59 | 7.90 | 11.27 | 0.144 | 0.150 |
| 27 | 10.56 | 7.90 | 11.28 | 0.144 | 0.151 |
| 30 | 10.56 | 7.90 | 11.30 | 0.146 | 0.154 |
| 33 | 10.48 | 7.90 | 11.27 | 0.146 | 0.155 |

Table C.25;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 13.54 | 7.37 | 12.19 | 0.104 | 0.158 |
| 3 | 12.35 | 7.57 | 11.69 | 0.103 | 0.158 |
| 6 | 12.29 | 7.57 | 11.56 | 0.103 | 0.159 |
| 9 | 12.27 | 7.57 | 1.51 | 0.102 | 0.160 |
| 12 | 12.25 | 7.57 | 11.53 | 0.102 | 0.162 |
| 15 | 12.17 | 7.57 | 11.46 | 0.105 | 0.163 |
| 18 | 11.90 | 7.56 | 11.49 | 0.104 | 0.164 |
| 21 | 10.91 | 7.51 | 11.42 | 0.107 | 0.167 |
| 24 | 10.70 | 7.47 | 11.21 | 0.109 | 0.169 |
| 27 | 10.01 | 7.38 | 10.71 | 0.111 | 0.173 |
| 30 | 9.84 | 7.30 | 10.33 | 0.111 | 0.175 |

Table C.26;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathrm{mg} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 11.47 | 8.50 | 12.05 | 0.127 | 0.158 |
| 3 | 13.00 | 8.32 | 12.04 | 0.136 | 0.166 |
| 6 | 12.62 | 8.23 | 11.93 | 0.136 | 0.169 |
| 9 | 11.75 | 8.02 | 11.79 | 0.146 | 0.176 |
| 12 | 11.26 | 7.90 | 11.62 | 0.144 | 0.180 |
| 15 | 10.83 | 7.83 | 11.51 | 0.145 | 0.182 |
| 18 | 10.60 | 7.81 | 11.41 | 0.147 | 0.184 |
| 21 | 10.50 | 7.81 | 11.37 | 0.149 | 0.185 |
| 24 | 10.31 | 7.80 | 11.30 | 0.153 | 0.186 |
| 27 | 10.23 | 7.79 | 11.29 | 0.150 | 0.187 |
| 30 | 10.16 | 7.78 | 11.29 | 0.151 | 0.187 |
| 33 | 10.05 | 7.79 | 11.30 | 0.151 | 0.187 |

Table C.27;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.06 | 8.64 | 11.92 | 0.150 | 0.165 |
| 3 | 13.24 | 8.57 | 12.10 | 0.151 | 0.167 |
| 6 | 12.83 | 8.47 | 12.08 | 0.151 | 0.171 |
| 9 | 11.14 | 8.19 | 12.02 | 0.154 | 0.179 |
| 12 | 10.41 | 8.02 | 11.06 | 0.156 | 0.185 |
| 15 | 10.30 | 7.98 | 11.78 | 0.156 | 0.187 |
| 18 | 10.16 | 7.95 | 11.76 | 0.156 | 0.187 |
| 21 | 10.14 | 7.93 | 11.71 | 0.155 | 0.190 |
| 24 | 10.16 | 7.92 | 11.66 | 0.155 | 0.191 |
| 27 | 9.46 | 7.90 | 11.72 | 0.158 | 0.192 |
| 30 | 9.43 | 7.89 | 11.70 | 0.157 | 0.192 |
| 33 | 9.38 | 7.89 | 11.67 | 0.157 | 0.193 |

Table C.28;
SAN POIL RIVER

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.14 | 7.98 | 10.31 | 0.136 | 0.138 |
| 3 | 12.67 | 7.97 | 10.74 | 0.144 | 0.141 |
| 6 | 11.33 | 8.08 | 11.48 | 0.153 | 0.141 |
| 9 | 10.90 | 8.08 | 11.68 | 0.153 | 0.143 |
| 12 | 10.66 | 8.03 | 11.69 | 0.151 | 0.145 |
| 15 | 10.44 | 7.97 | 11.68 | 0.151 | 0.147 |
| 18 | 10.18 | 7.90 | 11.60 | 0.148 | 0.150 |
| 21 | 9.93 | 7.86 | 11.56 | 0.151 | 0.151 |
| 24 | 9.75 | 7.83 | 11.52 | 0.149 | 0.153 |
| 27 | 9.55 | 7.79 | 11.43 | 0.149 | 0.155 |
| 30 | 9.22 | 7.65 | 10.93 | 0.150 | 0.158 |

Table C.29;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.12 | 8.60 | 13.21 | 0.154 | 0.168 |
| 3 | 12.94 | 8.70 | 12.80 | 0.153 | 0.165 |
| 6 | 11.56 | 8.46 | 12.50 | 0.153 | 0.172 |
| 9 | 10.71 | 8.28 | 123 | 0.155 | 0.177 |
| 12 | 10.50 | 8.22 | 12.16 | 0.155 | 0.180 |
| 15 | 10.17 | 8.16 | 12.10 | 0.156 | 0.182 |
| 18 | 10.10 | 8.13 | 11.97 | 0.156 | 0.184 |
| 21 | 9.46 | 8.07 | 12.16 | 0.157 | 0.186 |
| 24 | 9.24 | 8.02 | 11.94 | 0.157 | 0.187 |
| 27 | 9.21 | 8.01 | 11.90 | 0.156 | 0.188 |
| 30 | 9.12 | 8.01 | 11.92 | 0.155 | 0.189 |
| 33 | 8.87 | 7.99 | 11.92 | 0.157 | 0.189 |

Tables C.30-C. 35 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Porcupine Bay, Seven Bays, Keller Ferry and Spring Canyon in June, 1995.
Table C. $\mathbf{3 0}$;
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 13.79 | 7.89 | 12.02 | 0.137 | 0.118 |
| 3 | 13.78 | 7.88 | 11.99 | 0.137 | 0.120 |
| 6 | 13.71 | 7.86 | 11.93 | 0.136 | 0.123 |
| 9 | 13.63 | 7.84 | 11.91 | 0.135 | 0.125 |
| 12 | 13.62 | 7.84 | 11.88 | 0.136 | 0.127 |
| 15 | 13.62 | 7.84 | 11.84 | 0.136 | 0.129 |
| 18 | 13.60 | 7.85 | 11.80 | 0.134 | 0.130 |
| 21 | 13.56 | 7.84 | 11.79 | 0.137 | 0.132 |
| 24 | 13.50 | 7.82 | 11.82 | 0.135 | 0.134 |
| 27 | 13.41 | 7.82 | 11.78 | 0.134 | 0.136 |
| 30 | 1335 | 7.81 | 11.75 | 0.134 | 0.138 |
| 33 | 13.43 | 7.80 | 11.80 | 0.135 | 0.139 |

Table C. 31 ;

## PORCUPINE BAY

| Depth <br> $(\mathrm{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathrm{mg} / \mathrm{L})$ | Conduct. <br> $\mathrm{mmho} / \mathrm{cm}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 17.28 | 8.66 | 10.94 | 0.101 | 0.038 |
| 3 | 17.03 | 8.68 | 11.08 | 0.101 | 0.043 |
| 6 | 16.84 | 8.66 | 11.06 | 0.101 | 0.048 |
| 9 | 16.14 | 8.17 | 10.63 | 0.105 | 0.064 |
| 12 | 15.05 | 7.79 | 10.35 | 0.105 | 0.075 |
| 15 | 14.27 | 7.60 | 10.21 | 0.101 | 0.082 |
| 18 | 13.20 | 7.50 | 10.35 | 0.101 | 0.087 |
| 21 | 11.79 | 7.40 | 10.36 | 0.106 | 0.094 |
| 24 | 11.38 | 7.34 | 10.15 | 0.108 | 0.098 |
| 27 | 11.00 | 7.28 | 9.65 | 0.110 | 0.102 |

Table C.32;
CONFLUENCE

| Depth <br> $(\mathrm{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathrm{mg} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.15 | 8.49 | 11.92 | 0.110 | 0.099 |
| 3 | 15.90 | 8.39 | 11.43 | 0.115 | 0.104 |
| 6 | 15.72 | 8.28 | 11.22 | 0.123 | 0.108 |
| 9 | 15.57 | 8.15 | 11.09 | 0.127 | 0.116 |
| 12 | 14.71 | 7.96 | 11.16 | 0.128 | 0.124 |
| 15 | 13.87 | 7.88 | 11.29 | 0.132 | 0.127 |
| 18 | 13.10 | 7.83 | 11.45 | 0.135 | 0.130 |
| 21 | 12.78 | 7.81 | 11.47 | 0.138 | 0.132 |
| 24 | 12.48 | 7.79 | 11.49 | 0.139 | 0.133 |
| 27 | 12.08 | 7.77 | 11.49 | 0.139 | 0.135 |
| 30 | 11.80 | 7.74 | 11.46 | 0.136 | 0.137 |
| 33 | 11.70 | 7.72 | 11.38 | 0.140 | 0.140 |

Table C.33;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 15.64 | 8.28 | 11.52 | 0.134 | 0.077 |
| 3 | 15.53 | 8.27 | 11.50 | 0.135 | 0.079 |
| 6 | 15.01 | 8.19 | 11.54 | 0.140 | 0.082 |
| 9 | 14.42 | 8.08 | 11.53 | 0.143 | 0.086 |
| 12 | 14.04 | 7.99 | 11.56 | 0.143 | 0.090 |
| 15 | 13.44 | 7.92 | 11.54 | 0.143 | 0.093 |
| 18 | 13.21 | 7.88 | 11.50 | 0.143 | 0.096 |
| 21 | 13.05 | 7.85 | 11.56 | 0.143 | 0.097 |
| 24 | 12.90 | 7.84 | 11.53 | 0.140 | 0.099 |
| 27 | 12.67 | 7.82 | 11.54 | 0.143 | 0.101 |
| 30 | 12.55 | 7.80 | 11.48 | 0.140 | 0.102 |
| 33 | 12.34 | 7.79 | 11.52 | 0.141 | 0.103 |

Table C.34;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.56 | 8.32 | 11.19 | 0.133 | 0.135 |
| 3 | 16.52 | 8.31 | 11.03 | 0.134 | 0.136 |
| 6 | 16.37 | 8.29 | 10.96 | 0.133 | 0.137 |
| 9 | 16.35 | 8.29 | 10.91 | 0.134 | 0.138 |
| 12 | 16.30 | 8.26 | 10.90 | 0.133 | 0.139 |
| 15 | 13.54 | 7.94 | 11.34 | 0.134 | 0.149 |
| 18 | 12.97 | 7.84 | 11.46 | 0.138 | 0.153 |
| 21 | 12.71 | 7.80 | 11.51 | 0.136 | 0.155 |
| 24 | 12.32 | 7.77 | 11.45 | 0.136 | 0.157 |
| 27 | 12.16 | 7.75 | 11.42 | 0.137 | 0.158 |
| 30 | 12.06 | 7.74 | 11.43 | 0.136 | 0.159 |
| 33 | 11.98 | 7.73 | 11.38 | 0.136 | 0.160 |

Table C.35;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 17.06 | 8.28 | 10.74 | 0.121 | 0.112 |
| 3 | 17.00 | 8.29 | 10.60 | 0.143 | 0.112 |
| 6 | 16.52 | 8.26 | 10.62 | 0.141 | 0.114 |
| 9 | 16.06 | 8.23 | 10.71 | 0.142 | 0.115 |
| 12 | 14.95 | 8.23 | 10.90 | 0.142 | 0.120 |
| 15 | 13.96 | 7.98 | 10.97 | 0.141 | 0.125 |
| 18 | 13.60 | 7.91 | 11.00 | 0.140 | 0.127 |
| 21 | 12.77 | 7.83 | 11.06 | 0.139 | 0.131 |
| 24 | 12.53 | 7.79 | 11.03 | 0.140 | 0.132 |
| 27 | 12.41 | 7.77 | 11.01 | 0.139 | 0.134 |
| 30 | 12.23 | 7.76 | 11.09 | 0.138 | 0.135 |
| 33 | 12.08 | 7.74 | 11.13 | 0.137 | 0.136 |

Tables C.36-C. 43 Water quality measurements taken with a Hydrolab Surveyor II at Kettle Falls, Gifford, Hunters, Porcupine Bay, Seven Bays, Keller Ferry, San Poil and Spring Canyon in July, 1995.
Table C.36;
KETTLE FALLS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 17.79 | 8.03 | 11.70 | 0.153 | 0.138 |
| 3 | 18.16 | 8.08 | 11.61 | 0.153 | 0.137 |
| 6 | 18.08 | 8.10 | 11.31 | 0.152 | 0.138 |
| 9 | 18.07 | 8.11 | 11.22 | 0.152 | 0.139 |
| 12 | 18.06 | 8.11 | 11.16 | 0.152 | 0.141 |
| 15 | 18.06 | 8.12 | 11.13 | 0.152 | 0.141 |
| 18 | 18.04 | 8.12 | 11.13 | 0.152 | 0.142 |
| 21 | 18.03 | 8.13 | 11.11 | 0.152 | 0.143 |
| 24 | 18.01 | 8.13 | 11.12 | 0.152 | 0.143 |
| 27 | 17.99 | 8.13 | 11.14 | 0.151 | 0.145 |
| 30 | 18.00 | 8.14 | 11.12 | 0.152 | 0.145 |
| 33 | 17.98 | 8.15 | 11.13 | 0.151 | 0.145 |

Table C.37;

## GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/ $\mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 19.41 | 8.31 | 12.47 | 0.152 | 0.113 |
| 3 | 18.37 | 8.37 | 11.74 | 0.151 | 0.114 |
| 6 | 18.16 | 8.35 | 11.24 | 0.152 | 0.117 |
| 9 | 17.71 | 8.24 | 10.96 | 0.151 | 0.122 |
| 12 | 17.29 | 8.14 | 10.86 | 0.151 | 0.124 |
| 15 | 17.05 | 8.08 | 1.01 | 0.151 | 0.127 |
| 18 | 16.84 | 8.03 | 11.14 | 0.151 | 0.129 |
| 21 | 16.75 | 8.00 | 11.05 | 0.151 | 0.131 |
| 24 | 16.68 | 7.98 | 11.00 | 0.150 | 0.132 |
| 27 | 16.68 | 7.97 | 10.93 | 0.150 | 0.133 |
| 30 | 16.61 | 7.96 | 10.89 | 0.150 | 0.135 |
| 33 | 16.44 | 7.94 | 10.86 | 0.150 | 0.136 |

Table C. 38
HUNTERS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 19.41 | 8.13 | 11.54 | 0.150 | 0.099 |
| 3 | 18.86 | 8.31 | 11.36 | 0.151 | 0.101 |
| 6 | 18.87 | 8.37 | 11.29 | 0.151 | 0.102 |
| 9 | 17.62 | 8.22 | 11.14 | 0.150 | 0.107 |
| 12 | 17.25 | 8.13 | 10.95 | 0.152 | 0.111 |
| 15 | 17.03 | 8.05 | 10.83 | 0.150 | 0.114 |
| 18 | 17.02 | 8.03 | 10.79 | 0.150 | 0.115 |
| 21 | 17.01 | 8.02 | 10.79 | 0.151 | 0.117 |
| 24 | 16.63 | 7.98 | 10.71 | 0.149 | 0.119 |
| 27 | 16.55 | 7.94 | 10.75 | 0.150 | 0.121 |
| 30 | 16.41 | 7.92 | 10.76 | 0.149 | 0.122 |
| 33 | 16.14 | 7.89 | 10.73 | 0.149 | 0.125 |

Table C.39;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 24.12 | 8.20 | 9.64 | 0.140 | 0.123 |
| 3 | 23.63 | 8.21 | 9.31 | 0.142 | 0.125 |
| 6 | 19.95 | 7.93 | 9.10 | 0.146 | 0.137 |
| 9 | 18.70 | 7.68 | 8.45 | 0.145 | 0.145 |
| 12 | 18.20 | 7.55 | 7.82 | 0.141 | 0.149 |
| 15 | 17.21 | 7.45 | 7.60 | 0.131 | 0.152 |
| 18 | 16.17 | 7.38 | 7.55 | 0.123 | 0.156 |
| 21 | 15.49 | 7.31 | 7.62 | 0.118 | 0.159 |
| 24 | 14.71 | 7.27 | 7.68 | 0.113 | 0.161 |
| 27 | 14.05 | 7.23 | 7.69 | 0.111 | 0.163 |
| 30 | 13.37 | 7.17 | 7.08 | 0.111 | 0.166 |

Table C.40;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 20.64 | 8.12 | 10.18 | 0.151 | 0.105 |
| 3 | 20.39 | 8.30 | 10.23 | 0.150 | 0.102 |
| 6 | 18.33 | 8.19 | 10.64 | 0.151 | 0.108 |
| 9 | 17.40 | 8.06 | 10.67 | 0.151 | 0.113 |
| 12 | 16.94 | 7.99 | 10.71 | 0.151 | 0.115 |
| 15 | 16.47 | 7.94 | 10.80 | 0.151 | 0.118 |
| 18 | 16.23 | 7.91 | 10.84 | 0.151 | 0.119 |
| 21 | 15.93 | 7.89 | 10.89 | 0.151 | 0.121 |
| 24 | 15.82 | 7.88 | 10.93 | 0.151 | 0.122 |
| 27 | 15.62 | 7.86 | 10.91 | 0.150 | 0.123 |
| 30 | 15.43 | 7.85 | 11.00 | 0.151 | 0.125 |
| 33 | 15.34 | 7.84 | 11.02 | 0.150 | 0.125 |

Table C.41;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} \mathbf{L}$ ) | Conduct. <br> mmho/cm | ORP <br> (V) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 21.42 | 8.19 | 10.08 | 0.154 | 0.104 |
| 3 | 21.00 | 8.17 | 9.16 | 0.153 | 0.106 |
| 6 | 20.78 | 8.13 | 9.08 | 0.153 | 0.109 |
| 9 | 20.55 | 8.10 | 9.04 | 0.152 | 0.111 |
| 12 | 20.00 | 8.02 | 9.14 | 0.153 | 0.115 |
| 15 | 19.50 | 7.96 | 9.25 | 0.152 | 0.118 |
| 18 | 19.23 | 7.91 | 9.33 | 0.152 | 0.121 |
| 21 | 18.70 | 7.87 | 9.42 | 0.153 | 0.125 |
| 24 | 18.45 | 7.84 | 9.46 | 0.152 | 0.126 |
| 27 | 18.36 | 7.82 | 9.51 | 0.154 | 0.127 |
| 30 | 18.20 | 7.81 | 9.53 | 0.153 | 0.129 |
| 33 | 17.85 | 7.78 | 9.60 | 0.151 | 0.131 |

Table C.42;
SAN POIL RIVER

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | $\mathbf{D . O}$ <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 22.57 | 7.90 | 9.74 | 0.157 | 0.107 |
| 3 | 19.67 | 8.07 | 10.15 | 0.151 | 0.107 |
| 6 | 17.78 | 7.99 | 10.50 | 0.149 | 0.112 |
| 9 | 17.28 | 7.93 | 10.65 | 0.150 | 0.114 |
| 12 | 17.06 | 7.90 | 10.60 | 0.150 | 0.116 |
| 15 | 16.83 | 7.88 | 10.73 | 0.150 | 0.118 |
| 18 | 16.52 | 7.87 | 10.80 | 0.150 | 0.119 |
| 21 | 16.46 | 7.87 | 10.66 | 0.149 | 0.120 |
| 24 | 16.28 | 7.79 | 10.32 | 0.147 | 0.123 |
| 27 | 16.08 | 7.73 | 9.97 | 0.147 | 0.125 |
| 30 | 15.89 | 7.69 | 9.66 | 0.147 | 0.126 |
| 33 | 15.64 | 7.66 | 9.53 | 0.144 | 0.128 |

Table C.43;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 22.78 | 7.77 | 10.25 | 0.060 | 0.162 |
| 3 | 21.57 | 8.33 | 9.67 | 0.145 | 0.145 |
| 6 | 21.12 | 8.33 | 9.73 | 0.146 | 0.146 |
| 9 | 19.66 | 8.15 | 9.79 | 0.147 | 0.153 |
| 12 | 17.36 | 7.97 | 10.22 | 0.149 | 0.160 |
| 15 | 17.00 | 7.91 | 10.34 | 0.150 | 0.163 |
| 18 | 16.72 | 7.89 | 10.40 | 0.150 | 0.164 |
| 21 | 16.50 | 7.87 | 10.48 | 0.151 | 0.165 |
| 24 | 16.00 | 7.85 | 10.56 | 0.150 | 0.167 |
| 27 | 15.80 | 7.83 | 10.63 | 0.150 | 0.168 |
| 30 | 15.74 | 7.82 | 10.62 | 0.150 | 0.168 |
| 33 | 15.64 | 7.81 | 10.64 | 0.150 | 0.169 |

Tables C.44-C. 49 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Porcupine Bay, Confluence, Seven Bays, and Spring Canyon in August, 1995.
Table C.44;
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 19.91 | 8.07 | 9.12 | 0.144 | 0.092 |
| 3 | 19.92 | 8.06 | 9.12 | 0.145 | 0.094 |
| 6 | 19.91 | 8.04 | 9.13 | 0.145 | 0.095 |
| 9 | 19.90 | 8.03 | 9.16 | 0.145 | 0.096 |
| 12 | 19.90 | 8.02 | 9.16 | 0.145 | 0.096 |
| 15 | 19.90 | 8.02 | 9.16 | 0.145 | 0.097 |
| 18 | 19.90 | 8.02 | 9.16 | 0.144 | 0.098 |
| 21 | 19.89 | 8.02 | 9.20 | 0.144 | 0.099 |
| 24 | 19.86 | 7.99 | 9.18 | 0.145 | 0.100 |
| 27 | 19.83 | 7.95 | 9.16 | 0.143 | 0.102 |
| 30 | 19.72 | 7.88 | 9.10 | 0.144 | 0.105 |
| 33 | 19.72 | 7.85 | 9.07 | 0.145 | 0.106 |

Table C.45;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{( } \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 21.97 | 7.97 | 9.91 | 0.153 | 0.104 |
| 3 | 21.60 | 7.99 | 8.57 | 0.152 | 0.105 |
| 6 | 21.44 | 7.97 | 8.43 | 0.153 | 0.106 |
| 9 | 21.41 | 7.96 | 8.35 | 0.152 | 0.107 |
| 12 | 21.31 | 7.89 | 8.20 | 0.157 | 0.110 |
| 15 | 20.87 | 7.69 | 7.60 | 0.165 | 0.117 |
| 18 | 20.07 | 7.48 | 6.75 | 0.169 | 0.123 |
| 21 | 19.17 | 7.34 | 6.03 | 0.159 | 0.128 |
| 24 | 18.36 | 7.25 | 5.67 | 0.151 | 0.132 |
| 27 | 16.04 | 7.13 | 4.53 | 0.131 | 0.138 |
| 30 | 14.65 | 6.97 | 3.55 | 0.125 | 0.143 |
| 33 | 13.43 | 6.88 | 2.45 | 0.123 | 0.147 |

Table C.46;
CONFLUENCE

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathrm{cm}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 20.46 | 8.01 | 8.80 | 0.149 | 0.128 |
| 3 | 20.39 | 8.00 | 8.81 | 0.149 | 0.128 |
| 6 | 20.30 | 7.98 | 8.82 | 0.149 | 0.129 |
| 9 | 20.27 | 7.97 | 8.83 | 0.149 | 0.130 |
| 12 | 20.22 | 7.96 | 8.86 | 0.149 | 0.130 |
| 15 | 20.20 | 7.95 | 8.86 | 0.148 | 0.130 |
| 18 | 20.15 | 7.94 | 8.84 | 0.148 | 0.131 |
| 21 | 19.84 | 7.86 | 8.85 | 0.148 | 0.134 |
| 24 | 19.53 | 7.79 | 8.84 | 0.147 | 0.136 |
| 27 | 19.49 | 7.75 | 8.82 | 0.147 | 0.137 |
| 30 | 19.31 | 7.72 | 8.81 | 0.147 | 0.138 |
| 33 | 19.15 | 7.67 | 8.74 | 0.147 | 0.140 |

Table C.47;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 19.37 | 7.57 | 9.40 | 0.152 | 0.105 |
| 3 | 20.45 | 7.86 | 9.00 | 0.148 | 0.102 |
| 6 | 20.34 | 7.91 | 8.96 | 0.148 | 0.102 |
| 9 | 20.28 | 7.91 | 8.92 | 0.148 | 0.103 |
| 12 | 20.24 | 7.91 | 8.91 | 0.148 | 0.103 |
| 15 | 20.19 | 7.91 | 8.90 | 0.147 | 0.104 |
| 18 | 20.08 | 7.89 | 8.89 | 0.148 | 0.105 |
| 21 | 19.81 | 7.83 | 8.90 | 0.147 | 0.108 |
| 24 | 19.54 | 7.77 | 8.85 | 0.148 | 0.117 |
| 27 | 19.42 | 7.74 | 8.83 | 0.148 | 0.114 |
| 30 | 19.25 | 7.71 | 8.82 | 0.147 | 0.116 |
| 33 | 19.05 | 7.67 | 8.84 | 0.149 | 0.118 |

Table C.48;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 19.57 | 7.67 | 8.95 | 0.148 | 0.163 |
| 3 | 19.57 | 7.70 | 8.95 | 0.149 | 0.164 |
| 6 | 19.54 | 7.70 | 8.93 | 0.150 | 0.165 |
| 9 | 19.54 | 7.70 | 8.90 | 0.148 | 0.166 |
| 12 | 19.54 | 7.71 | 8.90 | 0.148 | 0.166 |
| 15 | 19.53 | 7.72 | 8.91 | 0.149 | 0.167 |
| 18 | 19.51 | 7.72 | 8.91 | 0.148 | 0.168 |
| 21 | 19.44 | 7.71 | 8.89 | 0.149 | 0.169 |
| 24 | 19.40 | 7.68 | 8.84 | 0.147 | 0.170 |
| 27 | 19.32 | 7.63 | 8.73 | 0.148 | 0.173 |
| 30 | 19.18 | 7.57 | 8.62 | 0.146 | 0.175 |
| 33 | 19.12 | 7.54 | 8.62 | 0.145 | 0.177 |

Table C.49;
SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 21.05 | 8.05 | 10.26 | 0.152 | 0.097 |
| 3 | 21.10 | 8.07 | 9.24 | 0.153 | 0.097 |
| 6 | 21.06 | 8.07 | 9.01 | 0.153 | 0.098 |
| 9 | 20.95 | 8.06 | 8.91 | 0.154 | 0.100 |
| 12 | 20.86 | 8.05 | 8.83 | 0.153 | 0.101 |
| 15 | 20.68 | 8.01 | 8.82 | 0.153 | 0.104 |
| 18 | 20.51 | 7.96 | 8.84 | 0.152 | 0.106 |
| 21 | 20.28 | 7.92 | 8.83 | 0.154 | 0.103 |
| 24 | 19.99 | 7.83 | 8.86 | 0.152 | 0.114 |
| 27 | 19.66 | 7.78 | 8.91 | 0.151 | 0.117 |
| 30 | 19.43 | 7.76 | 8.96 | 0.151 | 0.119 |
| 33 | 19.25 | 7.73 | 9.00 | 0.151 | 0.123 |

Tables C.50-C.55 Water quality measurements taken with a Hydrolab Surveyor II at Gifford, Porcupine Bay, Confluence, Seven Bays, Keller Ferry, and Spring Canyon in September, 1995.
Table C.50;
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 18.21 | 7.80 | 9.86 | 0.144 | 0.170 |
| 3 | 17.81 | 7.85 | 10.02 | 0.142 | 0.170 |
| 6 | 17.58 | 7.84 | 10.07 | 0.142 | 0.172 |
| 9 | 17.43 | 7.82 | 10.08 | 0.142 | 0.173 |
| 12 | 17.36 | 7.80 | 10.11 | 0.142 | 0.175 |
| 15 | 17.33 | 7.78 | 10.10 | 0.142 | 0.175 |
| 18 | 17.28 | 7.75 | 10.05 | 0.142 | 0.177 |
| 21 | 17.24 | 7.73 | 10.04 | 0.141 | 0.179 |
| 24 | 17.23 | 7.72 | 10.04 | 0.141 | 0.180 |
| 27 | 17.21 | 7.71 | 10.05 | 0.140 | 0.180 |
| 30 | 17.17 | 7.69 | 10.00 | 0.142 | 0.182 |
| 33 | 17.12 | 7.67 | 9.99 | 0.140 | 0.183 |

Table C.51;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 20.49 | 7.76 | 8.63 | 0.169 | 0.144 |
| 3 | 20.27 | 7.78 | 8.68 | 0.168 | 0.145 |
| 6 | 20.80 | 7.78 | 8.75 | 0.167 | 0.146 |
| 9 | 19.91 | 7.76 | 8.67 | 0.168 | 0.148 |
| 12 | 19.46 | 7.57 | 7.81 | 0.189 | 0.157 |
| 15 | 19.04 | 7.43 | 7.15 | 0.201 | 0.161 |
| 18 | 18.09 | 7.37 | 6.84 | 0.206 | 0.163 |
| 21 | 18.06 | 7.36 | 6.79 | 0.213 | 0.164 |
| 24 | 18.70 | 7.34 | 6.62 | 0.217 | 0.166 |
| 27 | 18.57 | 7.32 | 6.34 | 0.215 | 0.167 |
| 30 | 18.27 | 7.27 | 5.84 | 0.216 | 0.170 |

Table C.52;

| CONFLUENCE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | pH | D.O. <br> $(\mathrm{mg} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |  |
| 0 | 20.41 | 7.77 | 8.92 | 0.149 | 0.162 |  |
| 3 | 20.11 | 7.81 | 8.95 | 0.148 | 0.161 |  |
| 6 | 19.79 | 7.81 | 8.97 | 0.148 | 0.162 |  |
| 9 | 19.55 | 7.79 | 8.96 | 0.147 | 0.163 |  |
| 12 | 19.38 | 7.77 | 8.91 | 0.147 | 0.165 |  |
| 15 | 19.21 | 7.70 | 8.75 | 0.147 | 0.169 |  |
| 18 | 19.07 | 7.63 | 8.61 | 0.147 | 0.171 |  |
| 21 | 18.87 | 7.62 | 8.72 | 0.146 | 0.172 |  |
| 24 | 18.59 | 7.61 | 8.78 | 0.146 | 0.173 |  |
| 27 | 18.44 | 7.59 | 8.82 | 0.145 | 0.175 |  |
| 30 | 18.67 | 7.59 | 8.93 | 0.145 | 0.176 |  |
| 33 | 17.63 | 7.59 | 9.14 | 0.144 | 0.178 |  |

Table C.53;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/ $\mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 21.83 | 7.58 | 8.82 | 0.149 | 0.135 |
| 3 | 19.85 | 7.90 | 9.15 | 0.148 | 0.136 |
| 6 | 19.51 | 7.92 | 9.17 | 0.147 | 0.136 |
| 9 | 19.42 | 7.90 | 9.11 | 0.147 | 0.138 |
| 12 | 19.39 | 7.87 | 9.06 | 0.147 | 0.140 |
| 15 | 19.23 | 7.80 | 8.93 | 0.148 | 0.144 |
| 18 | 18.65 | 7.71 | 8.85 | 0.145 | 0.147 |
| 21 | 18.29 | 7.68 | 8.91 | 0.146 | 0.148 |
| 24 | 18.14 | 7.66 | 8.99 | 0.146 | 0.150 |
| 27 | 17.91 | 7.65 | 9.06 | 0.146 | 0.151 |
| 30 | 17.84 | 7.63 | 9.10 | 0.143 | 0.152 |
| 33 | 17.78 | 7.62 | 9.12 | 0.146 | 0.153 |

Table C.54;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 18.47 | 7.33 | 9.07 | 0.149 | 0.159 |
| 3 | 19.05 | 7.72 | 8.91 | 0.149 | 0.159 |
| 6 | 19.07 | 7.72 | 8.94 | 0.148 | 0.160 |
| 9 | 19.06 | 7.72 | 8.91 | 0.147 | 0.161 |
| 12 | 19.06 | 7.70 | 8.91 | 0.147 | 0.162 |
| 15 | 19.00 | 7.79 | 8.66 | 0.148 | 0.166 |
| 18 | 18.85 | 7.50 | 8.53 | 0.147 | 0.169 |
| 21 | 18.50 | 7.43 | 8.47 | 0.146 | 0.171 |
| 24 | 18.09 | 7.43 | 8.64 | 0.146 | 0.172 |
| 27 | 18.00 | 7.44 | 8.75 | 0.145 | 0.172 |
| 30 | 17.90 | 7.44 | 8.80 | 0.145 | 0.173 |
| 33 | 17.70 | 7.44 | 8.87 | 0.144 | 0.174 |

Table C.55;
SPRING CANYON

| Depth <br> $(\mathrm{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathrm{mg} / \mathrm{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 20.80 | 7.93 | 8.97 | 0.150 | 0.139 |
| 3 | 20.72 | 7.93 | 9.01 | 0.150 | 0.141 |
| 6 | 20.40 | 7.94 | 9.05 | 0.149 | 0.142 |
| 9 | 20.02 | 7.91 | 9.05 | 0.149 | 0.144 |
| 12 | 19.83 | 7.82 | 8.92 | 0.149 | 0.148 |
| 15 | 19.67 | 7.76 | 8.78 | 0.149 | 0.151 |
| 18 | 19.57 | 7.69 | 8.64 | 0.148 | 0.154 |
| 21 | 19.26 | 7.61 | 8.47 | 0.148 | 0.158 |
| 24 | 18.84 | 7.52 | 8.33 | 0.146 | 0.161 |
| 27 | 18.55 | 7.48 | 8.29 | 0.145 | 0.163 |
| 30 | 18.36 | 7.46 | 8.35 | 0.146 | 0.165 |
| 33 | 18.19 | 7.44 | 8.38 | 0.144 | 0.166 |

Tables C.56-C.63 Water quality measurements taken with a Hydrolab Surveyor II at Kettle Falls, Gifford, Hunters, Porcupine Bay, Seven Bays, Keller Ferry, San Poil and Spring Canyon in October, 1995.
Table C.56;
KETTLE FALLS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | PH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.09 | 7.67 | 10.05 | 0.155 | 0.207 |
| 3 | 14.08 | 7.71 | 9.93 | 0.154 | 0.204 |
| 6 | 14.08 | 7.74 | 9.88 | 0.153 | 0.203 |
| 9 | 14.07 | 7.75 | 9.84 | 0.153 | 0.203 |
| 12 | 14.06 | 7.76 | 9.85 | 0.154 | 0.203 |
| 15 | 14.07 | 7.76 | 9.84 | 0.153 | 0.203 |
| 18 | 14.06 | 7.76 | 9.85 | 0.152 | 0.203 |
| 21 | 14.06 | 7.76 | 9.89 | 0.154 | 0.204 |
| 24 | 14.06 | 7.77 | 9.89 | 0.153 | 0.204 |
| 27 | 14.06 | 7.77 | 9.89 | 0.153 | 0.204 |
| 30 | 14.06 | 7.77 | 9.93 | 0.154 | 0.204 |
| 33 | 14.06 | 7.77 | 9.93 | 0.151 | 0.204 |

Table C.57;
GIFFORD

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | PH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.96 | 7.64 | 10.01 | 0.152 | 0.163 |
| 3 | 14.98 | 7.22 | 10.24 | 0.151 | 0.187 |
| 6 | 14.98 | 7.39 | 9.63 | 0.151 | 0.178 |
| 9 | 14.98 | 7.47 | 9.51 | 0.151 | 0.173 |
| 12 | 14.98 | 7.53 | 9.47 | 0.151 | 0.170 |
| 15 | 14.98 | 7.57 | 9.48 | 0.152 | 0.168 |
| 18 | 14.94 | 7.59 | 9.46 | 0.151 | 0.167 |
| 21 | 14.88 | 7.61 | 9.49 | 0.152 | 0.167 |
| 24 | 14.85 | 7.62 | 9.50 | 0.153 | 0.166 |
| 27 | 14.81 | 7.63 | 9.49 | 0.151 | 0.166 |
| 30 | 14.97 | 7.64 | 9.50 | 0.153 | 0.166 |
| 33 | 14.78 | 7.64 | 9.50 | 0.150 | 0.167 |

Table C.58;
HUNTERS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 15.73 | 7.37 | 9.63 | 0.148 | 0.201 |
| 3 | 15.73 | 7.42 | 9.36 | 0.148 | 0.196 |
| 6 | 15.73 | 7.45 | 9.24 | 0.148 | 0.193 |
| 9 | 15.73 | 7.50 | 9.21 | 0.147 | 0.190 |
| 12 | 15.71 | 7.54 | 9.17 | 0.146 | 0.188 |
| 15 | 15.69 | 7.56 | 9.15 | 0.148 | 0.187 |
| 18 | 15.64 | 7.57 | 9.22 | 0.148 | 0.187 |
| 21 | 1562 | 7.61 | 9.21 | 0.149 | 0.187 |
| 24 | 15.69 | 7.61 | 9.19 | 0.147 | 0.187 |
| 27 | 15.69 | 7.62 | 9.19 | 0.148 | 0.187 |
| 30 | 15.67 | 7.62 | 9.20 | 0.146 | 0.187 |
| 33 | 15.65 | 7.62 | 9.21 | 0.147 | 0.188 |

Table C.59;
PORCUPINE BAY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 15.70 | 7.55 | 8.76 | 0.211 | 0.203 |
| 3 | 15.70 | 7.56 | 8.75 | 0.210 | 0.203 |
| 6 | 15.78 | 7.56 | 8.70 | 0.209 | 0.203 |
| 9 | 15.67 | 7.56 | 8.69 | 0.210 | 0.203 |
| 12 | 15.66 | 7.55 | 8.70 | 0.210 | 0.203 |
| 15 | 15.64 | 7.56 | 8.67 | 0.209 | 0.204 |
| 18 | 15.45 | 7.53 | 8.63 | 0.211 | 0.205 |
| 21 | 15.15 | 7.49 | 8.56 | 0.218 | 0.207 |
| 24 | 14.47 | 7.46 | 8.47 | 0.221 | 0.208 |
| 27 | 14.33 | 7.44 | 8.45 | 0.221 | 0.208 |
| 30 | 14.24 | 7.43 | 8.41 | 0.220 | 0.209 |
| 33 | 14.19 | 7.42 | 8.39 | 0.223 | 0.209 |

Table C.60;
SEVEN BAYS

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | pH | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> mmho/cm | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.46 | 7.50 | 9.61 | 0.138 | 0.184 |
| 3 | 16.46 | 7.51 | 9.25 | 0.146 | 0.183 |
| 6 | 16.45 | 7.51 | 9.11 | 0.147 | 0.183 |
| 9 | 16.45 | 7.52 | 9.03 | 0.146 | 0.183 |
| 12 | 16.43 | 7.52 | 9.00 | 0.146 | 0.183 |
| 15 | 16.42 | 7.52 | 8.97 | 0.147 | 0.183 |
| 18 | 16.42 | 7.52 | 8.97 | 0.145 | 0.184 |
| 21 | 16.41 | 7.53 | 8.95 | 0.146 | 0.184 |
| 24 | 16.40 | 7.52 | 8.95 | 0.146 | 0.184 |
| 27 | 16.39 | 7.53 | 8.95 | 0.145 | 0.185 |
| 30 | 16.39 | 7.53 | 8.95 | 0.144 | 0.185 |
| 33 | 16.37 | 7.53 | 8.96 | 0.145 | 0.185 |

Table C.61;
KELLER FERRY

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathrm{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathrm{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 14.79 | 7.46 | 9.12 | 0.149 | 0.189 |
| 3 | 14.82 | 7.46 | 9.12 | 0.148 | 0.189 |
| 6 | 14.83 | 7.47 | 9.07 | 0.148 | 0.189 |
| 9 | 14.83 | 7.46 | 9.07 | 0.148 | 0.189 |
| 12 | 14.82 | 7.46 | 9.07 | 0.147 | 0.190 |
| 15 | 14.82 | 7.46 | 9.08 | 0.147 | 0.190 |
| 18 | 14.83 | 7.46 | 9.07 | 0.148 | 0.190 |
| 21 | 14.83 | 7.47 | 9.07 | 0.147 | 0.191 |
| 24 | 14.82 | 7.46 | 9.07 | 0.147 | 0.191 |
| 27 | 14.82 | 7.46 | 9.07 | 0.147 | 0.191 |
| 30 | 14.82 | 7.47 | 9.07 | 0.146 | 0.191 |
| 33 | 14.83 | 7.47 | 9.08 | 0.145 | 0.191 |

Table C.62;
SAN POIL RIVER

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathrm{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.54 | 7.51 | 9.28 | 0.148 | 0.120 |
| 3 | 16.49 | 7.53 | 9.12 | 0.147 | 0.120 |
| 6 | 16.48 | 7.54 | 9.06 | 0.148 | 0.121 |
| 9 | 16.44 | 7.53 | 8.96 | 0.146 | 0.123 |
| 12 | 16.44 | 7.52 | 8.93 | 0.148 | 0.124 |
| 15 | 16.43 | 7.52 | 8.93 | 0.147 | 0.125 |
| 18 | 16.43 | 7.52 | 8.93 | 0.147 | 0.126 |
| 21 | 16.42 | 7.52 | 8.93 | 0.146 | 0.127 |
| 24 | 16.41 | 7.52 | 8.94 | 0.145 | 0.129 |
| 27 | 16.41 | 7.52 | 8.94 | 0.145 | 0.130 |
| 30 | 16.40 | 7.52 | 8.94 | 0.146 | 0.131 |
| 33 | 16.40 | 7.52 | 8.94 | 0.144 | 0.132 |

Table C.63;

## SPRING CANYON

| Depth <br> $(\mathbf{m})$ | Temp. <br> $\left({ }^{\circ} \mathbf{C}\right)$ | $\mathbf{p H}$ | D.O. <br> $(\mathbf{m g} / \mathbf{L})$ | Conduct. <br> $\mathbf{m m h o} / \mathbf{c m}$ | ORP <br> $(\mathbf{V})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 16.61 | 7.48 | 8.99 | 0.148 | 0.169 |
| 3 | 16.62 | 7.50 | 8.95 | 0.146 | 0.168 |
| 6 | 16.62 | 7.50 | 8.95 | 0.146 | 0.168 |
| 9 | 16.61 | 7.49 | 8.92 | 0.146 | 0.168 |
| 12 | 16.60 | 7.49 | 8.92 | 0.145 | 0.169 |
| 15 | 16.59 | 7.48 | 8.92 | 0.146 | 0.169 |
| 18 | 16.59 | 7.48 | 8.89 | 0.145 | 0.170 |
| 21 | 16.59 | 7.48 | 8.99 | 0.145 | 0.170 |
| 24 | 16.59 | 7.48 | 8.89 | 0.146 | 0.170 |
| 27 | 16.59 | 7.48 | 8.89 | 0.145 | 0.171 |
| 30 | 16.59 | 7.48 | 8.89 | 0.144 | 0.171 |
| 33 | 16.59 | 7.48 | 8.89 | 0.143 | 0.171 |

Table C. 64 Monthly secchii disk depths in meters for nine index stations on Lake Roosevelt in 1995.

## LOCATION

|  | 1 | 2 | 3 | 4 | Confluence | 6 | 7 | 8 | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Jan |  | 3.5 |  | 0.4 | 4.8 | 5.3 | 5.9 |  | 8.1 |
| Feb |  | 7.0 |  | 0.3 | 3.0 | 3.8 | 4.8 |  | 3.5 |
| M a r |  |  |  | 1.3 | 1.5 | 1.5 | 3.1 |  | 2.5 |
| Apr |  | 3.5 |  | 2.0 | 0.5 | 1.5 | 2.5 |  | 2.6 |
| May |  |  |  |  |  | 2.7 | 2.4 | 1.3 | 2.5 |
| Jun |  |  |  |  |  |  |  | 4.5 |  |
| Jul | 4.8 | 3.3 | 3.0 | 4.3 |  | 6.3 | 7.1 | 4.5 | 9.0 |
| Aug |  | 4.0 |  | 3.5 | 5.5 | 6.5 | 9.3 |  | 7.5 |
| Sep |  | 7.0 |  | 5.0 | 6.5 | 6.0 | 8.5 |  | 8.5 |
| Oct | 3.5 | 7.3 | 5.5 | 7.0 |  | 6.0 | 7.0 | 7.0 | 7.0 |
| Mean | 4.2 | 5.1 | 4.3 | 3.0 | 3.6 | 4.4 | 5.5 | 4.3 | 5.5 |

## SECTION 2

# LAKE ROOSEVELT FISHERIES MONITORING PROGRAM 1995 ANNUAL REPORT 

Prepared by:<br>Keith D. Underwood and<br>John P. Shields<br>Department of Natural Resources<br>Spokane Tribe of Indians<br>Wellpinit, WA 99040<br>and<br>Mary Beth Tilson<br>Upper Columbia United Tribes Fisheries Research Center<br>Biology Department Eastern Washington University<br>Cheney, WA 99004

Prepared for:
Charlie Craig, Project Manager
U.S. Department of Energy

Bonneville-Power Administration
Division of Fish and Wildlife
P.O. Box 3621

Portland, OR. 97208-3621

Project Number 88-63
Contract Number DE-8179-88D P91819

September 1996


#### Abstract

The purpose of the Lake Roosevelt Monitoring Program was to evaluate the effects of releasing hatchery origin kokanee salmon and rainbow trout on the fishery and to determine stocking strategies which increases fish harvest while maximizing the return of spawning kokanee to egg collection facilities. Two hatcheries stock kokanee and rainbow trout into lake Roosevelt, the Spokane Tribal Hatchery and the Sherman Creek Hatchery. The Spokane Tribal Hatchery began stocking fish in 1991 and Sherman Creek Hatchery began stocking in 1992. Approximately, 2.5 million kokanee salmon and 400,000 rainbow trout have been released annually since both hatcheries went on line. We estimated that 231,202 angler trips were taken to Lake Roosevelt spending a total of $\$ 8,697,819$. The harvest of kokanee increased steadily from 8,021 fish in 1992 to 32,353 fish in 1995. Rainbow trout harvest was estimated to be 122,939 fish in 1995. Walleye harvest was down slightly from 53,589 fish in 1994 to 40,185 fish in 1995. The relative abundance of kokanee increased from $3 \%$ in past years to $20 \%$ in 1995. On the other hand, the relative abundance of yellow perch has steadily decreased from $40 \%$ in 1989 to $7 \%$ in 1995. Growth of kokanee salmon, rainbow trout and walleye appeared to be similar to previous years. Kokanee salmon and rainbow trout exceed the mean growth per age class of fish in area lakes, but walleye growth was significantly less than fish in area lakes. The feeding habits of kokanee and walleye in 1995 were similar to previous years, but rainbow trout feeding habits differed. Kokanee fed mainly on Daphnia spp. and chironomids, and walleye mainly feed on fish. Rainbow trout fed primarily on Daphnia spp., chironomids and a new food item yellow perch. Food habits and growth suggest that kokanee and rainbow had ample food, but the reduced walleye growth may be the result of food shortages. The continued decline of yellow perch may be contributing to the reduced growth rates of walleye.


## ACKNOWLEDGMENTS

We gratefully acknowledge Ray Duff (WDFW) and Jerry Marco (Colville Confederated Tribes), for advise and coordination in all aspects of this study. Creel clerks William Matt Sr. (Spokane Tribe of Indians), Jim Meskan (WDFW), and Leroy Williams (Colville Confederated Tribes), for collection of creel data. We thank Amy Voeller, Hank Etue, Bart Kieffer, and William Matt Jr. for collection of field data and laboratory analysis. A special thanks to Dr. Allan Scholz (Eastern Washington University Biology Department) for his advice and technical support

This project was supported by a contract from the U.S. Department of Energy, Bonneville Power Administration (BPA), Contract number DE-8 179-88BP9 18 19; BPA project number 88-63. Additional financial support for this project was provided by a grant from the U.S. Department of Interior, Bureau of Indian Affairs to the Upper Columbia United Tribes (UCUT.) (Grant No. P12614208001). Capital equipment for this project was supplied by the Spokane Tribe of Indians through a grant by the Bureau of Indian Affairs.

Special thanks goes to Charlie Craig (BPA, Division of Fish and Wildlife), who served as BPA Contracting Offices Technical Representative, Larry Goodrow (Executive Director, Spokane Tribe of Indians), Carol Evans (Accountant, Spokane Tribe of Indians), and Mary Vemer (Natural Resource Director, Spokane Tribe of Indians) for administration of this contract.

## TABLE OF CONTENTS

ABSTRACT ..... 152
ACKNOWLEDGMENTS ..... 153
LIST OF TABLES ..... 156
LIST OF FIGURES ..... 158
1.0 INTRODUCTION ..... 159
1.1 Project History ..... 159
1.2 Kokanee and Rainbow Stocking History ..... 159
1.31995 Study Objectives ..... 160
2.0 MATERIALS AND METHODS ..... 163
2.1 Study Area ..... 163
2.2 Creel Survey Design and Procedures ..... 163
2.3 Fisheries Surveys and Relative Abundance ..... 171
2.4 Age, Back Calculations and Condition Factors ..... 172
2.5 Feeding Habits ..... 173
3.0 RESULTS ..... 175
3.1 Creel Survey ..... 175
3.2 Fisheries Surveys and Relative Abundance ..... 186
3.3 Age, Back Calculations and Condition Factors ..... 189
3.4 Feeding Habits ..... 193
4.0 DISCUSSION ..... 195
4.1 Historical Stocking and Lake Operations ..... 195
4.2 Creel Survey Trends ..... 201
4.2.1 Kokanee salmon ..... 202
4.2.2 Rainbow trout ..... 203
4.2.3 Walleye ..... 203
4.2.4 Accuracy and Precision ..... 205
4.3 Fisheries Surveys and Relative Abundance ..... 206
4.3 Growth and Feeding ..... 209
5.0 RECOMMENDATIONS AND RESEARCH NEEDS ..... 211
LITERATURE CITED ..... 212
APPENDIX A Creel Data ..... 217
APPENDIX B Fisheries Surveys ..... 249
APPENDIX C Feeding Habits ..... 267

## LIST OF TABLES

Table Page
Table 3.1 Total monthly angler pressure estimates in hours ( $\pm 95 \% \mathrm{CI}$ ), by creel section on Lake Roosevelt from December 1994 through November 1995 ..... 176
Table 3.2 Angler trip estimates by section based on angler hours (hrs) and average trip length for Lake Roosevelt from December 1994 through November 1995 ..... 177
177
Table 3.3 Harvest (kept fish) per unit effort (HPUE) by section from December 1994 through November 1995 at Lake Roosevelt. ..... 179
Table 3.4 Catch (kept and released fish) per unit effort (CPUE) by section from December 1994 through November 1995 at Lake Roosevelt ..... 179
Table 3.5 Number of fish harvest (kept), with $\pm 95 \%$ confidence intervals, for Lake Roosevelt during December 1994 through November 1995.. ..... 180
Table 3.6 Number of fish caught (kept and released), with $\pm 95 \%$ confidence intervals, for Lake Roosevelt during December 1994 through November 1995 ..... 181
Table 3.7 Annual numbers ( n ), mean lengths ( mm ) and weights ( g ) with standard deviations for all fish harvested on Lake Roosevelt from December 1994 through November 1995 ..... 182
Table 3.8 Percent of anglers that were satisfied with the fishery by species, section and season from December 1994 through November 1995 ..... 184
Table 3.9 Percent of anglers targeting various fish species by section and season on Lake Roosevelt from December 1994 through November 1995. ..... 185
Table 3.10 Economic value of the sport fishery at Lake Roosevelt during December 1994 through November 1995 ..... 186
Table 3.11 Relative abundance of fish collected by electrofishing boat and gillnets in Lake Roosevelt during 1995 ..... 187
Table 3.12 Catch per unit effort based on time (min) for fish captured by electrofishing boat or gillnets during 1995 ..... 188
Table 3.13 Lengths, weights, and condition factors (mean $\pm$ standard deviation) of kokanee salmon collected during 1995 ..... 189
Table 3.14 Back calculated total length (mean $\pm$ standard deviation) of kokanee salmon sampled during 1995 . ..... 189
Table 3.15 Lengths, weights, and condition factors (mean $\pm$ standard deviation) of rainbow trout collected during 1995 ..... 190

## LIST OF TABLES (cont.)

Table 3.16 Back calculated total length (mean $\pm$ standard deviation) of rainbow trout sampled during 1995 ..... 190
Table 3.17 Lengths, weights, and condition factors (mean $\pm$ standard deviation) of walleye collected during 1995........I. ..... 191
Table 3.18 Back calculated total length (mean $\pm$ standard deviation) of walleye sampled during 1995 ..... 192
Table 3.19 Index of relative importance for kokanee, rainbow trout, and walleye from fish collected during 1995 ..... 194
Table 4.1 Summary of hatchery origin rainbow trout released into Lake Roosevelt from 1988 through 1995 ..... 196
Table 4.2 Summary of hatchery origin kokanee released into Lake Roosevelt from 1988 through 1995 ..... 197
Table 4.3 Summary of angler trips, number of fish caught and harvested. Catch and harvest per unit effort and mean lengths of kokanee, rainbow trout and walleye from 1989 through 1995 ..... 204
Table 4.4 Comparison of relative abundance (\%) of fish collected during the 1989 through 1995 sampling periods via electroshocking and gillnettings. ..... 207
Table 4.5 Comparison of catch per unit effort (No. fish per hour) for fish collected during the 1998 through 1995 sampling periods viaelectrofishing and gillnetting. ..... 208

## LIST OF FIGURES

Figure Page
Figure 2.1 Map of Lake Roosevelt, Washington ..... 165
Figure 4.1 Lake elevations for Lake Roosevelt from 1990 through 1995 ..... 199
Figure 4.2 Water retention times for Lake Roosevelt from 1990 through 1995 ..... 200

### 1.0 INTRODUCTION

### 1.1 Project History

The primary objective of this project was to determine stocking strategies of hatchery origin kokanee salmon (Oncorhynchus nerka) and rainbow trout (Oncorhynchus mykiss) which maximized angler harvest and return of kokanee to egg collection facilities; and collect baseline data on the fishery to evaluated effects of stocking kokanee and the rainbow trout on the environment. Another responsibility of this program was to assess the effectiveness of kokanee hatcheries and rainbow trout program funded by BPA. Tasks of the Monitoring Program were: to conduct a year round reservoir wide creel survey; sample the fishery by electroshocking boat during spring, summer and fall; and collect information on diet, length, weight and age information. The data generated from sampling was analyzed to determine food availability, utilization, growth rates, and angler use information (e.g. harvest). This 1995 annual report marks the sixth report produced by the Lake Roosevelt Monitoring Program (Monitoring Program). Peone et al. (1990), Griffith and Scholz (1991), Thatcher et al. (1993), Thatcher et al. (1994) and Underwood and Shields (1996), and Underwood et al. (1996) wrote previous Monitoring Program reports.

### 1.2 History of Kokanee and Rainbow Trout Stocking

From 1988 to 1990, kokanee reared at the Ford Hatchery by the Washington Department of Fish and Wildlife were stocked into Lake Roosevelt. Approximately 750,000 kokanee fry were stocked into Sherman Creek and 100,000 kokanee fiy were stocked into the Spokane River at Little Falls Dam each year during July or May. Rainbow trout fry were provided to the Lake Roosevelt Net Pen Program by the Spokane Hatchery (WDFW operated) from 1986 to 1990. The number of rainbow trout provided by the Spokane Hatchery began at 50,000 increasing to 276,500 by 1990 . The rainbow trout were stocked in net pens during October. Rainbow trout were held in net pens until May or June and then released as yearlings. The Net Pen Program was operated by the Lake Roosevelt Development Association, a nonprofit volunteer group.

The Spokane Tribal Hatchery went on line in 1990 and began stocking kokanee and rainbow trout into Lake Roosevelt in 199 1. The Sherman Creek Hatchery began rearing and releasing kokanee in 1992. Construction and operation of these hatcheries was funded
by the Bonneville Power Administration as partial mitigation for the loss of anadromous salmon and steelhead. The loss occurred in the 1939 when the Grand Coulee Dam was installed The dam was not equipped with a fish ladder, thus blocking the migration path of anadromous salmon and steelhead. The blockage caused the permanent loss of anadromous stocks upstream from the dam.

The Spokane Tribal Hatchery was a full production facility operated by the Spokane Tribe and located on their reservation. The Sherman Creek Hatchery was a part time (spring to fall) rearing facility operated by the Washington Department of Fish and Wildlife and located near Kettle Falls, Washington. The Sherman Creek Hatchery imprinted juvenile kokanee to the creek water, then released the juveniles and collected eggs from returning adults. The collected eggs were transferred to the Spokane Tribal Hatchery for incubation and rearing. To initiate production a majority of the kokanee eggs have come from Lake Whatcom Hatchery near Bellingham, WA (operated by WDFW). Also, due to limited returning adults to egg collection sites in Lake Roosevelt, kokanee eggs continue to be supplemented by the Lake Whatcom Hatchery. A portion of the kokanee reared in the Spokane Tribal Hatchery were transferred to Sherman Creek Hatchery in early Spring for imprinting and later released. The hatcheries original production goals were 8 million kokanee salmon fry for release into Lake Roosevelt and 500,000 rainbow trout fry for the Lake Roosevelt Net Pen Program. However, due to a limited water supply at the Spokane Tribal Hatchery, approximately 2.5 million kokanee and 250,000 rainbow trout fiy have been released annually.

### 1.31995 Study Goals

Goals of the Lake Roosevelt Monitoring Program for 1995 were as follows:

1) Determine angler pressure, number of fish harvested, average size of fish harvested and economic value of the fishery;
2) Estimate the relative abundance of fish in the lake;
3) Determine diet of kokanee, rainbow trout and walleye;
4) Back calculate length at age using scales from kokanee, rainbow trout and walleye; and
5) Compare and contrast data collected during 1995 with previous years to identify changes in the fishery.

### 2.0 MATERIALS AND METHODS

### 2.1 Study Area

Lake Roosevelt is a mainstem Columbia River impoundment formed by the installation of Grand Coulee Dam in 1939 (Figure 2.1). Filled in 1941, the reservoir inundated 33,490 hectares at a full pool elevation of $393 \mathrm{~m}(1,290 \mathrm{ft})$ above mean sea level. It has a maximum width of 3.4 km and a maximum depth of 122 m (Stober et al. 198 1).

### 2.2 Creel Design and Procedures

A two-stage probability sampling scheme was used to determine annual fishing pressure, catch-per-unit-effort (CPUE), and sport fish harvest by species on Lake Roosevelt (Lambou 1961;1966, Malvestuto 1983). Creel surveys were conducted at Spokane and Colville tribal campgrounds and National Park Service boat launches for a total of 48 survey locations (Figure 2.1).

Three creel clerks were employed to interview anglers at access points along Lake Roosevelt. The lake was split into three sections an upper, middle and lower section (Figure 2.1). One creel clerk was permanently assigned to a section. Each creel clerk was scheduled approximately 21 days per month to make roving instantaneous pressure and effort counts at access points.

Schedules were constructed by dividing each month into weekday and weekend/holiday stratum and days were stratified into a.m. (sunrise to $12: 00$ ) and p.m. (12:00 to sunset) time periods. Eighteen weekdays and four weekend/holidays with were randomly selected to schedule roving instantaneous pressure counts with half of the surveys conducted during the a.m. and the other half were conducted during the p.m.. The remaining a.m. or p.m. time slots over the 21 day time period were used to conduct five hour access point surveys. The schedules were developed monthly by randomly pulling index cards from a hat that specified the time, day, survey type (roving instantaneous pressure count or access point survey) and, if an access type of survey, the location. Roving instantaneous pressure counts and access point survey schedules differed among creel clerks both spatially and temporally.

During roving instantaneous pressure counts, each creel clerk recorded the number of boat trailers and shore anglers at the access points in their section. The creel clerk reached the access points by road. No angler interviews were performed during roving instantaneous pressure counts.

During each access point survey creel clerks interviewed anglers. The following data was collected from the anglers during the interviews: angler type, hours fished, completed trip, satisfaction, zip code of origin, target species, and number of fish caught and released. Fish harvested were identified to species, measured in millimeters, weighed in grams and examined for floy tags, fin clips, and physical markings such as eroded pectoral and pelvic fins, and stubbed dorsal fins. Physical marks were used to differentiate rainbow trout of net-pen or hatchery origin from wild fish. Scale samples were collected from representative kokanee, rainbow trout, and walleye, and stomach samples were collected from kokanee. Heads were taken from fin clipped kokanee for coded wire tag analysis. Additionally, incoming boaters (angler or non angler) were surveyed to determine the number of boats angling and the number of anglers per boat.

During 1990 through 1993, four air flights (one flight per stratum) were scheduled to coincide with roving instantaneous pressure counts monthly. The three creel clerks recorded the number of boat trailers and shore anglers in their section while concurrently the surveyor in the airplane recorded the number of boats on the water and the number of shore anglers. This information was used to compute a correction factor for the number of boats on the water versus the number of boat trailers at access points:

$$
C F_{b}=\left(\frac{B_{a}}{B_{c}}\right)
$$

Where:

$$
\begin{aligned}
C F_{b}= & \text { boat trailer correction factor for each stratum per } \\
& \text { month; } \\
B_{a}= & \text { boat count from air survey for each stratum; and } \\
B_{C}= & \text { number of boat trailers counted by creel clerks during } \\
& \text { air flights for each stratum. }
\end{aligned}
$$

The correction factor for boat trailers versus boats on the water that was determined during 1990 through 1993 was averaged among years and then applied to 1995 due to the facts that a limitation of funds negated our ability to conduct regularly scheduled air flights.


Figure 2.1. Map of Lake Roosevelt, Washington. " $\boldsymbol{*}$ " indicates fish sampling stations.

Table A. 1 and A. 2 of Appendix A summarize the correction factors determined in past years and the correction factor applied to 1995 data

The number of boats on the reservoir was determined for the stratum weekday/weekend, section and month by completing the following calculation:

$$
T_{b}=\left(C_{b t}\right)\left(C F_{b}\right)
$$

Where:

$$
\begin{aligned}
T_{b} & =\begin{array}{l}
\text { number of boats on the water for each stratum per } \\
\text { month; }
\end{array} \\
C_{b t} & =\begin{array}{l}
\text { mean boat trailer count from pressure counts for each } \\
\text { stratum per month; and }
\end{array} \\
C F_{b} & =\begin{array}{l}
\text { boat trailer correction factor for each stratum per } \\
\text { month. }
\end{array}
\end{aligned}
$$

The number of boats fishing for the strata weekday/weekend, section and month was calculated by using the formula:

$$
B_{f}=\left(T_{b}\right)\left(\% B_{f}\right)
$$

Where:

$$
\begin{aligned}
& B_{f}=\text { number of boats fishing for each stratum per month; } \\
& T_{b}=\text { number of boats on the water for each stratum per } \\
& \text { month; and } \\
& \% B f=\begin{array}{l}
\text { percent of boats fishing for each stratum per month } \\
\text { (number is in decimal form). }
\end{array}
\end{aligned}
$$

The adjusted mean number of boat anglers per day for the strata weekday/weekend, section and month was estimated using the formula:

$$
X_{d}=(A d)\left(B_{f}\right)
$$

Where:
$X_{d}=$ adjusted mean number of anglers per boat per day for each stratum per month;
$\mathbf{A d}=$ mean number of anglers per boat from effort counts for each stratum per month; and
$B_{f}=$ number of boats fishing for each stratum per month.

The above calculations used to estimate the instantaneous number of boat anglers were estimated separately by section then summed to obtain a full lake estimate.

The number of hours available for fishing (sunrise to sunset) was estimated using the following formula:

$$
N_{s}=\left(D_{s}\right)\left(H_{d}\right)
$$

Where:

$$
\begin{aligned}
N_{s}= & \text { number of hours per weekend, weekday per month; } \\
D_{s}= & \text { number of days per month a weekday or weekend, and } \\
H_{d}= & \text { average number of hours per day for each } \\
& \text { stratum per month. }
\end{aligned}
$$

The number of hours sampled for each stratum per month was estimated using the formula:

$$
n=\sum_{i=1}^{D_{s}}\left(H_{c i}\right)
$$

Where:

$$
\begin{aligned}
n= & \text { number of hours sampled for each stratum per month; } \\
D_{S}= & \text { number of days per month within each stratum; and } \\
H_{C i}= & \text { mean number of hours creeled per day for each stratum } \\
& \text { per month. }
\end{aligned}
$$

The number of shore anglers per day for each stratum per month was estimated using the formula:

$$
\mathbf{X d}=\sum_{i=1}^{P_{d}}(S p i)
$$

Where:
Xd = mean number of shore anglers per day for each stratum per month from pressure counts;
$P_{d}=$ number of pressure counts conducted for each stratum per month; and
$S_{p i}=$ total number of shore anglers counted during pressure counts for each stratum per month.

The mean number of anglers (boat or shore) for each stratum per month was estimated using the formula:

$$
X_{s}=\left(X_{d}\right)\left(D_{s}\right)
$$

Where:
$X_{S}=$ mean number of anglers for each stratum per month;
$X_{d}=$ mean number of anglers for each stratum per day; and
$D_{S}=$ number of days per month within the stratum.

The standard deviation of anglers (boat or shore) for each stratum per month was estimated using the formula:

$$
S_{s}=\left(S_{d}\right)\left(D_{s}\right)
$$

Where:

$$
\begin{aligned}
& S_{S}=\begin{array}{l}
\text { standard deviation of anglers for each stratum per } \\
\text { month; }
\end{array} \\
& S_{d}=\begin{array}{l}
\text { standard deviation of anglers per day for each stratum } \\
\text { per month; and }
\end{array} \\
& D_{S}=\text { number of days per month for each stratum per month. }
\end{aligned}
$$

The mean number of angler hours per angler for each stratum was estimated using the formula:

$$
H_{a}=\left(\frac{T_{h}}{A_{i}}\right)
$$

Where:

$$
\begin{aligned}
H_{a}= & \text { mean number of angler hours per angler for each } \\
T_{h}= & \text { stratum per month; } \\
& \text { and hours spent fishing for each stratum per month; } \\
A_{i}= & \begin{array}{l}
\text { total number of anglers interviewed for each stratum } \\
\\
\\
\text { per month. }
\end{array}
\end{aligned}
$$

Pressure (hours fished) was estimated for day stratum (week day or weekend/holiday) for boat and shore anglers for each month by section by the formula:

$$
P E_{s}=\left(\frac{N_{s}}{n}\right)\left(X_{s}\right)\left(H_{a}\right)
$$

where:

$$
\begin{aligned}
& P E_{S}=\text { pressure estimate for each stratum per month; } \\
& N_{S}=\text { number of hours for each stratum per month; } \\
& \mathbf{n}=\text { number of hours sampled for each stratum per month; } \\
& X_{S}=\text { mean number of anglers for each stratum per month; } \\
& H_{a}=\begin{array}{l}
\text { and } \\
\end{array} \\
& \text { mean number of angler hours per angler for each } \\
& \text { suath. }
\end{aligned}
$$

The variance of the pressure (hours fished) estimate for each stratum per month was calculated by:

$$
V P E_{s}=\left(\frac{N_{s}}{n}\right) S_{s}^{2}
$$

where:

$$
\begin{aligned}
V P E_{S} & =\text { variance of pressure estimate for each stratum per } \\
& \text { month; } \\
N_{S} & =\text { number of hours for each stratum per month; } \\
\mathbf{n} & =\text { number of hours sampled for each stratum per month; } \\
S_{S}= & \begin{array}{l}
\text { and } \\
\\
\\
\text { standard deviation of mean number of angler hours for } \\
\text { each stratum per month. }
\end{array}
\end{aligned}
$$

Ninety-five percent confidence intervals for each stratum per month were calculated by:

$$
\text { C.I. }=P E \pm \sqrt{\left(V P E_{s}\right) 1.96}
$$

where:

$$
\begin{aligned}
C . I .= & 95 \% \text { confidence intervals for each stratum per month; } \\
\text { PEE }= & \text { pressure estimate for each stratum per month; and } \\
V P E_{S}= & \text { variance of the pressure estimate for each stratum } \\
& \text { per month. }
\end{aligned}
$$

Monthly angler pressure and 95\% C.I. was determined by calculating by weekend/weekday, boat/shore anglers, per month by section. If data gaps existed in any suata the quarterly averages were used to fill the gaps. Annual angler pressure and 95\% C.I. estimates were calculated by summing monthly angler pressure estimates and $95 \%$ C.I. estimates for that section. Each section was added together to get full lake estimates.

Studies by Fletcher (1988) and Malvestuto et al. (1978) have shown that CPUE values calculated independently from complete and incomplete trip data are not statistically different. Therefore, complete and incomplete angler trips were used to compute CPUE for fish species in each stratum. CPUE was calculated independently for fish captured (kept and released) and fish harvested (kept) for each stratum for the month by the formula:

$$
\text { CPUE }=\left(\frac{F}{T_{h}}\right)
$$

where:

$$
\begin{aligned}
& \text { CPUE }=\begin{array}{l}
\text { Catch per unit effort of a particular fish species for } \\
\text { each stratum per month; }
\end{array} \\
& F=\begin{array}{l}
\text { number of fish captured (harvested) for each stratum }
\end{array} \\
& T_{h}=\text { per month; and } \\
& \text { total hours spent fishing for each stratum per month. }
\end{aligned}
$$

Monthly CPUE of a particular fish species was calculated by dividing the total catch for the entire month (all stratum) by the total angler hours (all stratum) for each section. Annual CPUE values of a particular fish species were calculated by dividing the total catch for the year by the total number of angler hours for the year.

Harvest of fish species was determined for each stratum per month by the formula:

$$
\text { Harvest }=\left(H_{c p u e}\right)\left(P E_{s}\right)
$$

where:

$$
\begin{aligned}
\text { Harvest }= & \begin{array}{l}
\text { harvest of a particular fish species for each stratum per } \\
\text { month; }
\end{array} \\
H_{\text {cpue }}= & \begin{array}{l}
\text { number of fish harvested of a particular fish species for } \\
\text { each stratum per month for each stratum per month; }
\end{array} \\
P E_{S} & =\begin{array}{l}
\text { and } \\
\text { pressure (hours fished estimate for each stratum per } \\
\text { month. }
\end{array}
\end{aligned}
$$

Monthly harvest estimates for a particular fish species by stratum were combined to calculate a total monthly harvest estimate by section. Monthly harvest estimates were combined to calculate annual estimates for each fish species by section. Section harvest estirnates were added by month to obtain full lake monthly harvest.

Data compiled by the U.S. Fish and Wildlife Service in 1980 and 1985, showed a typical angler spent $\$ 23.00$ /fishing trip in 1980 and $\$ 26.00$ /fishing trip in 1985 in inland waters of Washington State (USFWS 1989). To calculate current dollar amount spent by anglers per trip, the 1985 cost per fishing trip was adjusted for inflation using the regional consumer price index (CPI). The following formula was used:

$$
D_{95}=\left(\frac{D_{85} x C_{95}}{C_{85}}\right)
$$

where:

$$
\begin{aligned}
\text { D95 }= & \text { dollar value per fishing trip for the Lake Roosevelt } \\
& \text { Fishery in 1995; } \\
C 85= & \text { regional CPI for 1985; } \\
\text { C95 }= & \text { regional CPI for 1995; and } \\
D 85= & \text { dollar value per fishing trip for the Lake Roosevelt } \\
& \text { Fishery in } 1985(\$ 26.00) .
\end{aligned}
$$

The 1995 dollar value was multiplied by total number of angler trips in 1995 to provide an estimate of the economic value of the fishery. The number of angler trips per month was determined by dividing the total number of angler hours per month by the average length of
a completed fishing trip for the month. Annual angler trips were calculated by summing monthly angler trip values.

### 2.3 Fisheries Surveys and Relative Abundance

Fishery samples were collected at nine index stations in the reservoir, which included: Kettle Falls, Gifford, Hunters, Porcupine Bay, Little Falls Dam, Seven Bays, Keller Ferry, Sanpoil, and Spring Canyon (Figure 2.1). Fishery data was collected at each index station over 24 hour periods. Principle target species included kokanee salmon, rainbow trout, and walleye, although it was assumed that all fish captured were caught in proportion to their relative abundance in the lake.

Relative abundance surveys were performed in littoral areas and tributaries by electrofishing 10 minute transects along 0.5 km of shoreline using SR- 180 and SR-23 electrofishing boats (Smith Root, Inc., Vancouver, WA) according to procedures outlined by Reynolds (1983) and Novotany and Prigel(1974). Voltage was adjusted to produce a pulsating DC current of approximately 5 amperes. Fish were collected using dip nets and placed into live wells on the boat for examination and data collection. A minimum of six 10 minute transects were performed at each sample station.

Additional relative abundance surveys were performed in pelagic zones with bottom and surface monofilament gillnets using methodologies described by Hubert (1983). The following gillnets were used: two horizontal surface set gillnets measuring 61 m in length by 6.1 m deep, with four 15.2 m long panels graded from 1.3 to 7.6 cm stretch mesh; and two horizontal bottom set gillnets measuring 61 m in length by 6.1 m deep, with four 15.2 m long panels graded from 1.3 to 8.9 cm stretch mesh. Gillnets were set in early afternoon (2:00 p.m.) and pulled at 10:00 a.m. the next morning.

Fish captured were identified by species using the taxonomic key of Wydoski and Whimey (1979). Total lengths were measured to the nearest millimeter using a metric measuring board and scale samples were removed from target fish species to determine age and growth. Target species were weighed to the nearest gram using an spring scales. Sexes were determined when possible. Stomach samples were collected from representative sizes of target species. The heads of kokanee were cut off and sent to the Fisheries Research Center at Eastern Washington University, where coded wire tags (cwt) were dissected out
and examined to determine the lot code. These results are discussed in chapter 3 of fish report.

### 2.4 Age, Back Calculations and Condition Factor

In the field, scales were taken from appropriate locations for each species as described by Jearld (1983) and placed in coin envelopes labeled with fish number, length, weight, location, date, and species for later analysis. In the laboratory, back-calculation measurements and age class of each fish were determined simultaneously. To obtain data, scales were removed from the envelope and placed between two microscope slides. Slides were then placed in a Realist Vantage 5, Model 3315 microfiche reader. Scale images were projected onto the screen and a non-regenerated, uniform scale was selected to determine age and back calculate length at age. Age was determined by counting the number of annuli (Jearld 1983). For back calculations, the annulus distance was measured from the origin of the scale to the last circuli of each respective annulus. Each measurement was made under constant magnification to the nearest millimeter.

Lee's back-calculation method was used to determine the length of the fish at the formation of each annulus (Carlander 1950, 1981; Hile 1970). However, due to a small number of samples the " y " intercept was assumed to be zero.

Back-calculations were computed using the formula:

$$
L_{i}=a+\left(\frac{L_{c}-a}{S_{c}}\right) S_{i}
$$

where:

$$
\begin{aligned}
L_{i} & =\text { length of fish (in } \mathrm{mm}) \text { at each annulus formation; } \\
\mathbf{a} & =\text { intercept of the body-scale regression line; } \\
L_{c} & =\text { length of fish (in } \mathrm{mm} \text { ) at time of capture; } \\
S_{c} & =\text { distance (in mm) from the focus to the edge of the } \\
S_{i} & =\text { scale; and }
\end{aligned}
$$

Condition factors were determined for each fish to serve as an indicator of fish condition (Hile 1970, Everhart and Youngs 1981). Condition factor describes how a fish adds weight in relation to incremental changes in length. The relationship is shown by the
formula:

$$
K_{\pi Z}=\left(\frac{w}{l^{3}}\right) 10^{s}
$$

where:

$$
\begin{aligned}
K_{T L} & =\text { condition factor; } \\
\mathrm{w} & =\text { weight of fish }(\mathrm{g}) ; \text { and } \\
l & =\text { total length of fish }(\mathrm{mm}) .
\end{aligned}
$$

### 2.5 Feeding Habits

Fish stomachs were collected from kokanee, rainbow, and walleye at each index station. Additional kokanee stomachs were obtained by creel clerks from anglers throughout the year. Stomachs from representative sizes of fish were collected by making an incision into the body cavity, cutting the esophagus, and pinching the pyloric sphincter. The esophagus was clamped to keep prey items from being expelled and the stomach was placed in $10 \%$ formalin.

In the laboratory, stomachs were transferred to a $70 \%$ isopropyl alcohol solution. Contents were identified and enumerated by taxa using the taxonomic keys of Brooks (1957), Ward and Whipple (1966), Bon-or et al. (1976), Ruttner-Kolisko (1974), Edmonds et al. (1976), Wiggins (1977), Pennak (1978, 1989), and Merritt and Cummins (1984). Food organisms were identified using a Nikon SMZ- 1B dissecting microscope equipped with a fiber optics illumination system and 5 mm ocular micrometer.

Dry weights were obtained by drying sorted stomach contents in an oven at $105^{\circ}$ for 24 hours on a stainless steel wire screen and weighing them on a Sartorius Model H51 analytical balance to the nearest 0.0001 g (Weber 1973, APHA 1976). Weight values were combined for each age class, annual mean and standard deviation.

Index of relative importance values were used to compensate for numerical estimate biases that tend to over-emphasize small prey groups consumed in large numbers and weight estimate biases that overemphasize large prey items consumed in small numbers (Bowen 1983). The index of relative importance (George and Hadley 1979) was calculated using the formula:

## R $\boldsymbol{I}^{100}{ }^{10} l_{+}$ <br> $\sum_{a=1} A l_{a}$

where:

$$
\begin{aligned}
& R I_{a}= \text { relative importance of food item a; } \\
& A I_{a}= \text { absolute importance of food item a (i.e., frequency of } \\
& \text { occurrence }+ \text { numerical frequency }+ \text { weight frequency } \\
& \text { of food item a); and } \\
& n= \text { number of different food types. }
\end{aligned}
$$

Relative importance values range from zero to $100 \%$ with prey items near zero being relatively less important than those prey items near one hundred percent.

Diet overlap was calculated to determine the degree to which intra and inter species competition exists in Lake Roosevelt. Fish diet overlaps were computed by using the overlap formula of Morisita (1959) as modified by Horn (1966). Overlap values were based upon indices obtained from IRI calculations. Overlap index was expressed in the equation:

$$
C_{x}=\frac{2 \sum_{i=1}^{n}\left(P_{x i} x P_{y i}\right)}{\sum_{i=1}^{n} P_{x i} 2+\sum_{i=1}^{n} p_{y i} 2}
$$

where: $\quad C_{X}=$ overlap coefficient;
$n=$ number of food categories;
$P_{x i}=$ proportion of food category (i) in the diet of species x; and
$P_{y i}=$ proportion of food category (i) in the diet of species $y$.
Overlap coefficients were computed using IRI values in the equation for the variables Pxi and Pyi. Overlap coefficients range from 0 (no overlap) to 1 (complete overlap). Values of less than 0.3 are considered low and values greater than 0.7 indicate high overlap (Peterson and Martin-Robichaud 1982). High diet overlap indices may indicate competition if food items utilized by the species are limited (MacArthur 1968), or there may be an abundant food supply and therefore competition does not exist.

### 3.0 RESULTS

### 3.1 Creel Data

The angler pressure (hours fished) estimates for Lake Roosevelt are reported by section and month in Table 3.1. Appendix A reports the pressure estimates by the lowest stratification levels.

The results of the creel analyses are reported for the time period December 1994 through November 1995. December 1994 was included in this report so that quarterly averages could be used to fill data gaps at the lowest stratification level. Quarters were split into December 1993 through February 1994 (winter), March 1994 through May 1994 (spring), June 1994 through August 1994 (summer), and September 1994 through November 1994 (fall). Quarters were established based on historic weather trends and angler use of the fishery. For example, a quarterly average was used if no boat anglers were surveyed during the month of January on a weekend in Section 1, but boat trailers were counted at the access points during the weekends. Since no boat anglers were surveyed, we were unable to estimate the average number of hours fished by boat anglers on a weekend in January without using some other means to estimate the number of boat angler hours. As a result, the quarterly average was used to fill the data hole, "weekend boat angler".

Fishing pressure (angler hours) was greatest in Section 3 ( $684,026 \mathrm{hrs}$ ) followed by Section 2 ( $512,730 \mathrm{hrs}$ ) and then Section 1 ( $152,773 \mathrm{hrs}$ ). The annual pressure for the lake was estimated to be 1.3 million hours. Total pressure was greatest during June ( $243,852 \mathrm{hrs}$ ) and August (203,724 hrs). Pressure was least during November $(12,299)$ and January ( $39,018 \mathrm{hrs}$ ).

The number of angler trips to Lake Roosevelt were estimated by dividing the estimated number of angler hours fished by the mean trip length for each section and month (Table 3.2). The total number of trips estimated from the period December 1994 through November 1995 was 231,202 angler trips. A total of 24,223 trips ( $11 \%$ ) were made in Section 1; 109,709 angler trips (47\%) in Section 2 and 97,270 angler trips (42\%) were in Section 3. The greatest number of trips was during February (33,962 trips), June ( 33,387 trips) and May (29,015 trips).

Table 3.1. Total monthly angler pressure estimates in hours ( $\pm 95 \% \mathrm{CI}$ ), by creel section on Lake Roosevelt from December 1994 through November 1995.

Section

|  | 1 | 2 | 3 | Total |
| :--- | :---: | :---: | :---: | :---: |
| Dec | $\mathbf{2 , 0 0 9} \pm \mathbf{8 7}$ | $36,923 \pm 1,854$ | $6,966 \pm 388$ | $\mathbf{4 5 , 8 9 8} \pm \mathbf{2 , 3 2 9}$ |
| Jan | $\mathbf{1 , 5 8 9} \pm \mathbf{1 4 4}$ | $20,152 \pm 1,146$ | $17,277 \pm 564$ | $\mathbf{3 9 , 0 1 8} \pm \mathbf{1 , 8 5 4}$ |
| Feb | $\mathbf{1 , 5 2 8} \pm \mathbf{1 9 5}$ | $94,308 \pm 1,738$ | $72,490 \pm 3,025$ | $\mathbf{1 6 8 , 3 2 6} \pm \mathbf{4 , 9 5 8}$ |
| Mar | $5,913 \pm 199$ | $44,784 \pm 2,672$ | $35,282 \pm 2,007$ | $\mathbf{8 5 , 9 7 9} \pm \mathbf{4 , 8 7 8}$ |
| Apr | $10,807 \pm 612$ | $52,297 \pm 1,729$ | $56,248 \pm 1,687$ | $\mathbf{1 1 9 , 3 5 2} \pm \mathbf{4 , 0 2 8}$ |
| May | $20,514 \pm 690$ | $62,152 \pm 2,012$ | $94,690 \pm 2,969$ | $\mathbf{1 7 7 , 3 5 6} \pm \mathbf{5 , 6 7 1}$ |
| Jun | $19,948 \pm 520$ | $47,427 \pm 684$ | $176,477 \pm 2,893$ | $\mathbf{2 4 3 , 8 5 2} \pm \mathbf{4 , 0 9 7}$ |
| Jul | $55,896 \pm 2,275$ | $31,883 \pm 1,150$ | $22,579 \pm 1,249$ | $\mathbf{1 1 0 , 3 5 8} \pm \mathbf{4 , 6 7 4}$ |
| Aug | $5,772 \pm 983$ | $60,269 \pm 3,116$ | $137,683 \pm 1,715$ | $\mathbf{2 0 3 , 7 2 4} \pm \mathbf{5 , 8 1 4}$ |
| Sep | $23,215 \pm 807$ | $36,464 \pm \mathbf{1 , 6 1 9}$ | $\mathbf{3 2 , 9 9 1} \pm 1,232$ | $\mathbf{9 2 , 6 7 0} \pm \mathbf{3 , 6 5 8}$ |
| Oct | $3,456 \pm 135$ | $22,188 \pm 655$ | $25,053 \pm 553$ | $\mathbf{5 0 , 6 9 7} \pm \mathbf{1 , 3 4 3}$ |
| Nov | $2,126 \pm 207$ | $3,883 \pm 339$ | $6,290 \pm 23$ | $\mathbf{1 2 , 2 9 9} \pm \mathbf{5 6 9}$ |
| Total | $152,773 \pm 6,854$ | $512,730 \pm 18,714$ | $684,026 \pm 18,305$ | $\mathbf{1 , 3 4 9 , 5 2 9} \pm \mathbf{4 3 , 8 7 3}$ |

Table 3.2. Angler trip estimates by section based on angler hours (hrs) and average trip length for Lake Roosevelt from December 1994 through November 1995.

|  | Section | Mean Trip Length | No. Angler Hours | No. Angler Trips |
| :---: | :---: | :---: | :---: | :---: |
| December | 1 | 3.2 | 2,009 | 628 |
|  | 2 | 3.3 | 36,923 | 11,189 |
|  | 3 | 4.5 | 6,966 | 1,548 |
| January | 1 | 3.7 | 1,589 | 429 |
|  | 2 | 5.0 | 20,152 | 4,030 |
|  | 3 | 5.0 | 17,277 | 3,455 |
| February | 1 | 5.0 | 1,528 | 306 |
|  | 2 | 5.0 | 94,308 | 18,862 |
|  | 3 | 4.9 | 72,490 | 14,794 |
| March | 1 | 5.1 | 5,913 | 1,159 |
|  | 2 | 4.8 | 44,784 | 9,330 |
|  | 3 | 5.2 | 35,282 | 6,785 |
| April | 1 | 5.3 | 10,807 | 2,039 |
|  | 2 | 4.7 | 52,297 | 11,127 |
|  | 3 | 7.7 | 56,248 | 7,305 |
| May | 1 | 6.1 | 20,514 | 3,363 |
|  | $2$ | 5.3 | 62,152 | $11,727$ |
|  | 3 | 6.8 | 94,690 | 13,925 |
| June | 1 | 5.9 | 19,948 | 3,381 |
|  | 2 | 5.0 | 47,427 | 9,485 |
|  | 3 | 8.6 | 176,477 | 20,521 |
| July | 1 | 6.8 | 55,896 | 8,220 |
|  | 2 | 4.2 | 31,883 | 7,591 |
|  | 3 | 6.0 | 22,579 | 3,763 |
| August | 1 | 6.6 | 5,772 | 875 |
|  | 2 | 5.3 | 60,269 | 11,372 |
|  | 3 | 8.9 | 137,683 | 15,470 |
| September | 1 | 8.0 | 23,215 | 2,902 |
|  | 2 | 3.9 | 36,464 | 9,350 |
|  | 3 | 6.6 | 32,991 | 4,997 |
| October |  | $6.1$ | $3,456$ |  |
|  | $2$ | 4.7 6.7 | $22,188$ | $4,721$ |
|  | 3 | 6.7 | 25.053 | 3,739 |
| November | 1 | 6.0 | 2,126 | 354 |
|  | 2 | 4.2 | 3,883 | 925 |
|  | 3 | 6.5 | 6,290 | 968 |
| Total |  | 5.62 | 1,349,529 | 231,202 |

Table 3.3 reports the harvest rates by catch per unit effort (CPUE) for fish harvested during 1995. The annual mean harvest rate for rainbow trout was 0.089 HPUE ( 11 angler hrs/fish); 0.060 HPUE ( 17 angler hrs/fish) for walleye; 0.018 HPUE ( 55 angler hrs/fish) for kokanee and 0.004 ( 250 angler hrs/fish) for smallmouth bass. Section 3 had the quickest harvest rate for rainbow trout ( 0.154 HPUE ( 6 angler hrs/fish) ) and smallmouth bass ( 0.010 HPUE ( 100 angler hrs/fish)). Section 1 had the highest mean annual harvest rate for walleye ( 0.154 HPUE ( 6 angler hrs/fish)) and Section 2 had the highest for kokanee ( 0.042 HPUE ( 23 angler hrs/fish)). The HPUE was 0.000 for sturgeon due to the fact that Washington Department of Fish and Wildlife closed harvest of sturgeon in 1995 as a means to conserve the remnant Lake Roosevelt population.

The 1995 catch (kept and released fish) estimates were similar to harvest estimates for most species (Table 3.4). However, walleye harvest was 0.060 HPUE ( 17 angler hrs/fish), and the catch was 0.131 CPUE ( 8 angler hrs/fish). The difference between the harvest and catch rates for walleye were due to a slot limit enforced by the Washington Department of Fish and Wildlife. The walleye slot limit specified that no more than one fish over 20 inches and only walleye less than 16 inches may be kept. This meant that it took approximately 17 hours for each angler to harvest a legal sized walleye, but 8 angler hours for each fish to catch a walleye within or outside of the slot limit.

The largest contribution to the fishery in terms of harvested (kept only) was rainbow trout $(122,939$ fish) followed by walleye $(40,185)$ and kokanee $(32,353)$. Smallmouth bass harvest was estimated to be 9,558 fish. Among lake sections, a majority of the rainbow trout $(87,428)$ were harvested in Section 3 and a majority of walleye $(28,743)$ were harvested in Section 1. Of the walleye observed in the creel, $6 \%$ of the fish were outside of the legal size limit, which was between $406 \mathrm{~mm}(16 \mathrm{in})$ and 508 mm ( 20 in ). The number of walleye harvested outside of the legal size was estimated to be 2,411 fish. Section 1 was the only area where sturgeon (297) were harvested. However, the estimated number of sturgeon captured must be viewed with caution. One sturgeon were seen by creel clerks in 1995. The small sample size may have caused significant errors in the estimate. The estimated number of fish harvested with $95 \%$ confidence intervals are reported in Table 3.5. Appendix A also reports harvest by section, month and species.

Table 3.3. Harvest (kept fish) catch per unit effort (HPUE) by section from December 1994 through November 1995 at Lake Roosevelt.

|  | Section |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | Annual |
| kokanee | $<0.001$ | 0.004 | 0.042 | $\mathbf{0 . 0 1 8}$ |
| rainbow trout | 0.033 | 0.064 | 0.143 | 0.083 |
| walleye | 0.154 | 0.008 | 0.009 | 0.060 |
| smallmouth bass | 0.000 | 0.000 | 0.010 | 0.004 |
| sturgeon | 0.000 | 0.000 | 0.000 | 0.000 |
| other species | 0.003 | 0.000 | 0.000 | 0.001 |
| Annual HPUE | 0.190 | 0.076 | 0.204 | 0.173 |

Table 3.4. Catch (kept and released fish) catch per unit effort (CPUE) by section from December 1994 through November 1995 at Lake Roosevelt.

|  | Section |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | Annual |
| kokanee | 0.000 | 0.004 | 0.042 | 0.018 |
| rainbow trout | 0.034 | 0.065 | 0.145 | 0.084 |
| walleye | 0.312 | 0.014 | 0.010 | 0.131 |
| smallmouth bass | 0.000 | 0.001 | 0.100 | 0.040 |
| sturgeon | $\mathrm{CO.001}$ | 0.000 | 0.000 | $<\mathbf{0 . 0 0 1}$ |
| other species | 0.003 | 0.000 | 0.000 | 0.001 |
| Annual CPUE | 0.350 | 0.084 | 0.298 | 0.276 |

Table 3.5. Number of fish harvested (kept), with $\pm 95 \%$ confidence intervals, for Lake Roosevelt during December 1994 through November 1995.

|  | Section |  |  | Total |
| :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 |  |
| kokanee | $\begin{array}{r} 175 \\ ( \pm 7) \end{array}$ | $\begin{array}{r} 3,076 \\ ( \pm 140) \end{array}$ | $\begin{array}{r} 29,102 \\ ( \pm 938) \end{array}$ | $\begin{array}{r} 32,353 \\ ( \pm 1,085) \end{array}$ |
| rainbow trout | $\begin{array}{r} 4,497 \\ ( \pm 251) \end{array}$ | $\begin{array}{r} 31,014 \\ ( \pm 1,143) \end{array}$ | $\begin{array}{r} 87,428 \\ ( \pm 2,505) \end{array}$ | $\begin{array}{r} 122,939 \\ ( \pm 3,899) \end{array}$ |
| walleye | $\begin{array}{r} 28,743 \\ ( \pm 1,132) \end{array}$ | $\begin{array}{r} 6,008 \\ ( \pm 146) \end{array}$ | $\begin{array}{r} 5,434 \\ ( \pm 116) \end{array}$ | $\begin{array}{r} 40,185 \\ ( \pm 1,394) \end{array}$ |
| smallmouth bass | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 9,558 \\ ( \pm 162) \end{array}$ | $\begin{array}{r} 9,558 \\ ( \pm 162) \end{array}$ |
| sturgeon | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ |
| other species | $\begin{array}{r} 537 \\ ( \pm 25) \\ \hline \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \\ \hline \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \\ \hline \end{array}$ | $\begin{array}{r} 537 \\ +\quad 25) \\ \hline \end{array}$ |
| Annual Harvest | $\begin{gathered} 33,952 \\ ( \pm 1,417) \end{gathered}$ | $\begin{array}{r} 40,098 \\ \pm \quad 1,429) \end{array}$ | $\begin{array}{r} 131,759 \\ +\quad 3,734) \end{array}$ | $\begin{array}{r} 205,809 \\ ( \pm 6,580) \end{array}$ |

Table 3.6. Number of fish caught (kept and released), with $\pm 95 \%$ confidence intervals, for Lake Roosevelt during December 1994 through November 1995.

|  | Section |  |  | Total |
| :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 |  |
| kokanee | $\begin{array}{r} 175 \\ ( \pm 7) \end{array}$ | $\begin{array}{r} 3,076 \\ ( \pm 140) \end{array}$ | $\begin{aligned} & 29,102 \\ & ( \pm 938) \end{aligned}$ | $\begin{array}{r} 32,353 \\ ( \pm 1,085) \end{array}$ |
| rainbow trout | $\begin{array}{r} 4,609 \\ ( \pm 255) \end{array}$ | $\begin{array}{r} 32,305 \\ ( \pm 1,186) \end{array}$ | $\begin{gathered} 89,044 \\ ( \pm 2,537) \end{gathered}$ | $\begin{array}{r} 125,958 \\ ( \pm 3,978) \end{array}$ |
| walleye | $\begin{array}{r} 58,602 \\ ( \pm 2,234) \end{array}$ | $\begin{array}{r} 9,002 \\ ( \pm 239) \end{array}$ | $\begin{gathered} 6,063 \\ ( \pm 136) \end{gathered}$ | $\begin{array}{r} 73,667 \\ ( \pm 2,609) \end{array}$ |
| smallmouth bass | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 681 \\ ( \pm 10) \end{array}$ | $\begin{gathered} 75,543 \\ ( \pm 2,120) \end{gathered}$ | $\begin{array}{r} 76,224 \\ ( \pm 2,130) \end{array}$ |
| sturgeon | $\begin{array}{r} 13 \\ ( \pm 2) \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \end{array}$ | $\begin{array}{r} 0 \\ \pm \\ \pm \end{array}$ | $\begin{array}{r} 13 \\ ( \pm \quad 2) \end{array}$ |
| other species | $\begin{array}{r} 624 \\ ( \pm 28) \\ \hline \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \\ \hline \end{array}$ | $\begin{array}{r} 0 \\ ( \pm 0) \\ \hline \end{array}$ | $\begin{array}{r} 624 \\ \pm \quad 28) \\ \hline \end{array}$ |
| Annual Catch | $\begin{array}{r} 64,024 \\ ( \pm 2,526) \end{array}$ | $\begin{gathered} 45,065 \\ 1,575) \end{gathered}$ | $\begin{array}{r} 199,990 \\ \pm \quad 5,740) \end{array}$ | $\begin{array}{r} 309,079 \\ ( \pm 9,841) \end{array}$ |

Table 3.7. Annual numbers ( n ), mean lengths (mm) and weights (g) with standard deviations for all fish harvested on Lake Roosevelt from December 1994 through November 1995.

Small-

Kokanee Rainbow Walleye \begin{tabular}{c}
mouth <br>
Bass

 Burbot 

Yellow <br>
Perch
\end{tabular}

See 1

| $n$ | 1 | 87 | 409 | - | 4 | 4 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{L n}$ | $221 \pm-$ | $384 \pm 66$ | 372.3 | - | $563 \pm 47$ | $267 \pm 53$ |
| $\mathbf{W t}$ | $85 \pm-$ | $643 \pm 359$ | $388 \pm 258$ | - | $1,054 \pm 224$ | $245 \pm 129$ |

Sec 2

| $\mathbf{n}$ |  | 5 | 83 | 11 |
| :--- | :--- | :---: | :---: | :---: |
| $\mathbf{L n}$ | $507 \pm 87$ | $431 \pm 55$ | $396 \pm 90$ | - |
| W | t | $1,380 \pm 657$ | $928 \pm 322$ | $451 \pm 64$ |

Sec 3

| n | 112 | 375 | 24 | 26 | - |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ln | $468 \pm 48$ | 41 lf 84 | $321 \pm 83$ | $265 \pm 57$ | - |  |
| Wt | $1,219 \pm 327$ | $1,017 \pm 474$ | $418 \pm 238$ | $625 \pm 519$ | - |  |
| Totis $\mathbf{n}$ (118 |  |  |  |  |  |  |
| Ln | 118 | $545$ | 444 | 26 | + | 4 |
|  | $467 \pm 55$ | $410 \pm 79$ | $370 \pm 58$ | $265 \pm 57$ | $56.3 \pm 47$ | $267 \pm 53$ |
| W | t | 1,216 $\pm 3629$ | $948 \pm 457$ | $0 \pm 25562$ | $\pm 5191,05$ | $\pm 224245 \pm 129$ |

The number of fish in the catch (kept and released) was similar to the harvest numbers. However, the walleye harvest deviated from catch due to the slot limit. Table 3.6. identifies the catch numbers by section and species with $95 \%$ confidence intervals. Appendix A reports catch by section, month and species.

The average length and weight of fish observed in the creel are reported in Table 3.7 by section and species. Section 2 contained the largest rainbow trout (length 431 mm and weight 928 g ). Walleye were larger in Section 2 than in Section 1 or 3. The mean length of walleye in Section 1 was $360 \mathrm{~mm}(\mathrm{n}=370)$ for fish classified in the smaller than 16 inch slot limit. Fish harvested in the 20 inches slot limit had a mean length of $561 \mathrm{~mm}(\mathrm{n}=13)$. In Section 2 walleye smaller than 16 inches had a mean length of $360 \mathrm{~mm}(\mathrm{n}=9)$ and walleye larger than 20 inches had a mean length of $559 \mathrm{~mm}(\mathrm{n}=2)$. In Section 3 walleye smaller than 16 inches had a mean length of $319 \mathrm{~mm}(\mathrm{n}=22)$ and walleye larger than 20 inches had a mean length of $508 \mathrm{~mm}(\mathrm{n}=1)$. There were 26 illegally sized walleye measured in Section 1, averaging 458 mm long.

Table 3.8 identifies the percent of anglers satisfied with the fishery by species, section and season. Based on annual time step, a majority of anglers ( $100 \%$ ) were satisfied with the smallmouth bass fishery. However, fewer anglers were satisfied with the walleye (12\%), kokanee (40\%), rainbow trout (36\%) and sturgeon (20\%) fisheries. Anglers fishing during the summer and fall appeared to have a higher satisfaction then anglers fishing in the winter and spring.

Of all the anglers who fished on Lake Roosevelt during 1995, 49\% targeted rainbow, 26\% targeted walleye, $19 \%$ targeted kokanee and the rest targeted other species (Table 3.9). The Other species were mainly smallmouth bass. The winter fishery consisted of mainly of rainbow trout anglers ( $82 \%$ ) with fewer kokanee anglers ( $10 \%$ ) and very few walleye anglers (3\%). In spring, the percent of rainbow trout anglers began tailing off in Sections 1 and 2 while the percent of walleye anglers increased. The majority of anglers in Section 3 during spring targeted kokanee (68\%). In the summer period, walleye anglers made up the majority in Sections 1 ( $83 \%$ ) and $2(53 \%)$ while ( $53 \%$ ) of the anglers were fishing for kokanee in Section 3. During the fall, $60 \%$ of the anglers were fishing for rainbow trout ( $60 \%$ ), compared to $27 \%$ fishing for walleye and $11 \%$ fishing for kokanee.

Table 3.10 shows the economic value of the sport fishery based on total number of angler trips of 231,202 at $\$ 37.62$ for each trip. The economic value was $\$ 8,697,819$.

Table 3.8. Percent of anglers that were satisfied with the fishery by species, section and season from December 1994 through November 1995.

| Quarter Section | Kokanee | $\begin{gathered} \text { Rainbow } \\ \text { Trout } \\ \hline \end{gathered}$ | Walleye | Sturgeon | Smallmouth Bass |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Winter |  |  |  |  |  |
| One |  | 23\% | 33\% | 0\% |  |
| Two |  | 12\% |  |  |  |
| Three | 27\% | 43\% |  |  |  |
| Spring |  |  |  |  |  |
| One |  | 45\% | 56\% | 20\% |  |
| Two | - | 19\% | 5\% |  |  |
| Three | 33\% | 72\% |  |  |  |
| Summer |  |  |  |  |  |
| One |  | 50\% | 80\% | 0\% |  |
| Two | - | 33\% | 25\% | - |  |
| Three | 62\% | 74\% | 83\% | 0\% | 100\% |
| Fall |  |  |  |  |  |
| One |  | 42\% | 71\% | 0\% |  |
| Two | 100\% | 17\% | 0\% |  |  |
| Three | 12\% | 86\% |  |  |  |
| Qrtly Totals |  |  |  |  |  |
| Winter | 27\% | 22\% | 14\% | 0\% | 0\% |
| S pring | 33\% | 44\% | 41\% | 20\% | 0\% |
| Summer | 62\% | 59\% | 77\% | 0\% | 100\% |
| Fall | 22\% | 56\% | 56\% | 0\% | 0\% |
| Annual Total | 40\% | 36\% | 12\% | 20\% | 100\% |

Table 3.9. Percent of anglers targeting various fish species by section and season on Lake Roosevelt from December 1994 through November 1995.

| Quarter |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Winter |  |  |  |  |
| One | 0\% | 95\% | 4\% | 1\% |
| Two | 0\% | 91\% | 3\% | 6\% |
| Three | 53\% | 31\% | 0\% | 6\% |
| Spring |  |  |  |  |
| One | 0\% | 30\% | 62\% | 8\% |
| Two | 0\% | 51\% | 41\% | 8\% |
| Three | 68\% | 25\% | 0\% | 6\% |
| Summer |  |  |  |  |
| One | 0\% | 16\% | 83\% | 1\% |
| Two | 0\% | 37\% | 53\% | 10\% |
| Three | 58\% | 26\% | 8\% | 8\% |
| Fall |  |  |  |  |
| One | 0\% | 56\% | 41\% | 3\% |
| Two | 10\% | 54\% | 34\% | 2\% |
| Three | 27\% | 73\% | 0\% | 0\% |
| Qrtly Totals |  |  |  |  |
| Winter | 10\% | 82\% | 3\% | 5\% |
| Spring | 29\% | 36\% | 28\% | 7\% |
| Summer | 22\% | 22\% | 52\% | 4\% |
| Fall | 11\% | 60\% | 27\% | $2 \%$ |
| Annual Total | 19\% | 49\% | 26\% | 6\% |

Table 3.10 Economic value of the sport fishery at Lake Roosevelt during December 1994 through November 1995.

|  | 1985 | 1995 |
| :--- | ---: | :---: |
| Consumer Price Index | $\$ 167.87$ | $\$ 242.90$ |
| Dollars Spent per Angler Trip | $\$ 26.00$ | $\$ 37.62$ |
| Number of Angler Trips |  | 231,202 |

Economic .Value of Fishery
$\$ 8,697,819$

### 3.2 Fisheries Surveys

Electrofishing and gillnet sets were used to estimate the relative abundance of each fish species in Lake Roosevelt. The most common fish species was the largescale sucker (Catostomus macrocheilus) at $27 \%$ based on all fish sampled (Table 3.11). The second most abundant fish was kokanee salmon ( $20 \%$ ), followed by walleye ( $12 \%$ ), smallmouth bass (10\%), yellow perch (7\%) and rainbow trout (5\%).

The catch per unit effort, based on duration of effort only, was determined for electrofishing and gillnet surveys (Table 3.12). These efforts were from all sampling during 1995. The annual sampling effort was 118 hrs electrofishing and $2,099 \mathrm{hrs}$ of gillnetting totaling $2,217 \mathrm{hrs}$ of sampling effort. Appendix B lists the number of fish captured, relative abundance, and CPUE by site, month and species.

Table 3.11 Relative abundance of fish collected by electrofishing boat and gillnets in Lake Roosevelt during 1995.

| Family species | Common Name | Electro- | Gillnet | Total <br> Annual |
| :---: | :---: | :---: | :---: | :---: |
| Catostomidae |  |  |  |  |
| Catostomusmacrocheilus | largescale sucker | $30 \%$ | 4 \% | $27 \%$ |
| Catostomus catostomus | longnose sucker | $<1 \%$ | $2 \%$ | $<1 \%$ |
| Catostomus columbianus | bridgelip sucker | <1\% | 3\% | 1\% |
| Catostomus spp. | sucker spp. | <1\% | 0\% | $<1 \%$ |
| Centrarchidae |  |  |  |  |
| Micropterus dolomiai | smallmouth bass | 10\% | $3 \%$ | 10\% |
| Lepomis gibloosus | pumpkinseed | $<1 \%$ | 0\% | <1\% |
| Pomoxis sp. | crappie sp. | c1\% | 0\% | <1\% |
| Cottidae |  |  |  |  |
| Cottus beldingi | piute sculpin | 6 \% | 0\% | 6 \% |
| Cyprinidae |  |  |  |  |
| Cyprinus carpio | carp | 2 \% | $2 \%$ | 2 \% |
| Richardsonius balteatus | redside shiner | <1\% | 0\% | < $1 \%$ |
| Tinca tinca | tench | 0\% | < $1 \%$ | <1\% |
| Ptychocheilus oregonensis | squawfish | 2 \% | 2 \% | 2 \% |
| Gadidae |  |  |  |  |
| Lotaiota | burbot | <1\% | $4 \%$ | < $1 \%$ |
| Ictaluridae |  |  |  |  |
| Ictalurus nebulosus | brown bullhead | $<1 \%$ | 0\% | $<1 \%$ |
| Percidae |  |  |  |  |
| Stizostedion vitreum vitram | walleye | 11\% | $18 \%$ | 12\% |
| Percaflavescens | yellow perch | 7 \% | 7 \% | 7 \% |
| Salmonidae |  |  |  |  |
| Savelinus fontinalis | brown trout | <1\% | 0\% | <1\% |
| Salvelinus confluentus | bull trout | c 1\% | 0\% | < $1 \%$ |
| Salmotrutta | brook trout | $<1 \%$ | 0\% | <1\% |
| Oncorhynchus tshawytscha | chinook salmon | <1\% | 0\% | <1\% |
| Oncorhyndhus nerka | kokanee salmon | 22 \% | 5 \% | $20 \%$ |
| Coregonus clupeaformis | lake whitefish | <1\% | $46 \%$ | 5 \% |
| Prosopium williamsoni | mt . whitefish | $<1 \%$ | <1\% | <1\% |
| Oncorhynduusmykiss | rainbow trout | 5\% | 4 \% | 5\% |

Table 3.12. Catch per unit effort based on time (hours) for fish captured by electrofishing boat or gillnets during 1995.

|  | Electrofish <br> CPUENo. |  | Gillnet |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| largescale sucker | 22.73 | 2,680 | 0.06 | 41 | 1.30 | 2,721 |
| Cottus spp. | 4.73 | 558 | 0.00 | 0 | 0.27 | 558 |
| walleye | 8.74 | 1,030 | 0.26 | 182 | 0.58 | 1,212 |
| smallmouth bass | 7.96 | 939 | 0.04 | 30 | 0.46 | 969 |
| rainbow trout | 3.88 | 458 | 0.06 | 44 | 0.24 | 502 |
| squawfish | 1.81 | 213 | 0.03 | 20 | 0.11 | 233 |
| carp | 1.75 | 206 | 0.03 | 20 | 0.11 | 226 |
| yellow perch | 5.45 | 643 | 0.10 | 71 | 0.34 | 714 |
| brown trout | 0.31 | 37 | $<0.01$ | 1 | 0.02 | 38 |
| kokanee salmon | 17.01 | 2,006 | 0.07 | 47 | 0.98 | 2,053 |
| chinook salmon | 0.02 | 2 | 0.00 | 0 | co. 01 | 2 |
| crappie | 0.08 | 10 | 0.00 | 0 | co. 01 | 10 |
| bull trout | 0.01 | 1 | 0.00 | 0 | co. 01 | 1 |
| brook trout | 0.18 | 21 | 0.00 | 0 | 0.01 | 21 |
| burbot | 0.26 | 31 | 0.07 | 46 | 0.04 | 71 |
| lake whitefish | 0.34 | 40 | 0.68 | 482 | 0.25 | 52 |
| mountain whitefish | 0.08 | 10 | CO. 01 | 1 | 0.01 | 11 |
| brown bullhead | 0.05 | 6 | 0.00 | 0 | co. 01 | 6 |
| longnose sucker | 0.21 | 25 | 0.03 | 19 | 0.02 | 44 |
| bridgelip sucker | 0.70 | 83 | 0.05 | 33 | 0.06 | 116 |
| redside shiner | 0.01 | 1 | 0.00 | 0 | co. 01 | 1 |
| tench | 0.00 | 0 | CO. 01 | 1 | co. 01 | 1 |
| Totals | 76.39 | 9,006 | 1.47 | 1,038 | 4.79 | 10,044 |

### 3.3 Age, Back Calculations and Condition Factor

Length, weight and condition of kokanee collected by electrofishing or gillnet surveys are reported in Table 3.13. The length, weight, and scales of 180 kokanee were collected during 1995. The condition factor of the kokanee was greater than 1.00 for age classes two through three, age one fish was slightly lower than 1.00. The back calculated growth of kokanee indicated an average annual growth of 134 mm for the first year of life, 151 mm for the second year and 126 mm for the third year (Table 3.14). This translated into a mean total length of 134 mm for age $1+$ fish, 285 mm for age $2+$ fish and 411 mm for age 3+ fish.

Table 3.13. Lengths, weights, and condition factors (mean $\pm$ standard deviation) of kokanee salmon collected during 1995.

| Age | n | Length $(\mathrm{mm})$ | Weight $(\mathrm{g})$ | Condition | Factor |  |
| :---: | :---: | :---: | ---: | ---: | ---: | ---: |
| $1+$ | 5 | 219 | $\pm 35$ | $92 \pm 38$ | 0.96 | $\pm$ |
| $2+$ | 62 | 385 | $\pm 81$ | 437 | $\pm 207$ | 1.01 |
|  | 0.15 |  |  |  |  |  |
| $3+$ | 113 | 472 | $\pm 60$ | 1,180 | $\pm$ | 559 |

Table 3.14. Back calculated total length (mean $\pm$ standard deviation) of kokanee salmon sampled during 1995.

|  | Back Calculated Total Length (mm) at Annulus |  |  |  |  |
| :---: | ---: | :---: | :---: | :---: | :---: |
| Cohort | n | 1 | 2 | 3 |  |
| 1994 | 5 | $132 \pm 20$ |  |  |  |
| 1993 | 62 | $159 \pm 51$ | $291 \pm 83$ | $411 \pm 70$ |  |
| 1992 | 113 | 121436 | $283 \pm 56$ |  |  |
| Grand |  |  | $\mathbf{2 8 5} \pm 67$ | $\mathbf{4 1 1} \pm 70$ |  |
| Mean | 180 | 134 f 45 |  |  |  |

Annual
Growth
134
151

The lengths, weights and condition factors of rainbow trout collected during 1995 are identified in Table 3.15. The condition factor of rainbow trout ages one through four was greater than 1.00 with ages $0+$ and $5+$ slightly under 1.00 . A condition factor close to 1.00 suggests that the rainbow trout population in Lake Roosevelt is healthy. The back calculated lengths indicate no significant growth differences among cohorts (Table 3.16). The incremental growth for age one through four year old rainbow ranged from 96 mm to 124 mm . The growth rates of rainbow between ages one through four appears to be relatively constant, however, growth seems to slow dramatically at age 5 . The incremental growth of age 5 rainbow was 37 mm .

Table 3.15. Lengths, weights, and condition factors (mean $\pm$ standard deviation) of rainbow trout collected during 1995.

| Age | n | Length $(\mathrm{mm})$ | Weight $(\mathrm{g})$ | Condition | Factor |
| :---: | :---: | :---: | ---: | ---: | ---: | :--- |
| $0+$ | 7 | $169 \pm 21$ | $55 \pm 7$ | $0.93 \pm 0.15$ |  |
| $1+$ | 44 | $206 \pm 52$ | $102 \pm 79$ | $1.15 \pm 0.30$ |  |
| $2+$ | 51 | $340 \pm 68$ | $490 \pm 237$ | $1.17 \pm$ | 0.18 |
| $3+$ | 58 | $416 \pm 56$ | $922 \pm 254$ | $1.21 \pm$ | 0.16 |
| $4+$ | 13 | $504 \pm 32$ | $1,325 \pm 203$ | $1.08 \pm$ | 0.18 |
| $5+$ | 4 | $537 \pm 20$ | $1,572 \pm 117$ | $0.93 \pm$ | $\pm .08$ |

Weight for $5+$ fish based on 2 fish.

Table 3.16. Back calculated total length (mean $\pm$ standard deviation) of rainbow trout sampled during 1995.

Back Calculated Total Length (mm) at Annulus

| Cohort n | $\mathbf{1}$ | 2 | 3 | 4 | 5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 44 | $116 \pm 36$ |  |  |  |  |
| 1993 | 51 | $121 \pm 38$ | $256 \pm 60$ |  |  |  |
| 1992 | 58 | $122 \pm 35$ | $239 \pm 52$ | $354 \pm 48$ |  |  |
| 1991 | 13 | $116 \pm 33$ | $248 \pm 49$ | $369 \pm 39$ | $459 \pm 34$ |  |
| 1990 | 4 | $106 \pm 44$ | $190 \pm 27$ | $272 \pm 57$ | $385 \pm 59$ | $479 \pm 31$ |
| Grand |  |  |  |  |  |  |
| Mean | 170 | $\mathbf{1 1 9} \pm 36$ | $\mathbf{2 4 3} \pm 59$ | $\mathbf{3 3 9} \pm 84$ | $\mathbf{4 4 2} \pm 51$ | $479 \pm 31$ |

Annual

| Growth | $\mathbf{1 1 9}$ | 124 | 96 | 103 | 37 |
| :--- | :--- | :--- | :--- | :--- | :--- |

The length, weight and condition factor of walleye sampled by electrofishing and gillnet sets are summarized in Table 3.17. The length, weight and scales of 372 walleye were measured in 1995. The walleye condition factor ranged from 0.67 to 1.01 depending on the age of the fish. Mean lengths ranged from 149 mm for walleye age $0+$ to 766 mm for age $11+$ fish. The back calculated length by cohort is reported in Table 3.18. The mean annual growth was 66 mm per year for cohorts covering the last eleven years. Growth was quickest for the first three years of ranging from 121 mm to 91 mm a year. In general, the older the fish the slower the growth.

Table 3.17. Lengths, weights, and condition factors (mean $\pm$ standard deviation) of walleye collected during 1995.

| Age | n | Length $(\mathrm{mm})$ | Weight $(\mathrm{g})$ |  | Condition Factor |  |
| :---: | :---: | :---: | ---: | ---: | ---: | ---: |
| $\mathrm{o}+$ | 13 | $1^{\prime} 49 \pm 43$ | $28 \pm 22$ | $0.67 \pm 0.12$ |  |  |
| $1+$ | 82 | $211 \pm 55$ | $85 \pm 58$ | $0.76 \pm 0.22$ |  |  |
| $2+$ | 101 | $277 \pm 53$ | $192 \pm 103$ | $0.83 \pm 0.21$ |  |  |
| $3+$ | 76 | $359 \pm 41$ | $411 \pm 140$ | $0.87 \pm 0.20$ |  |  |
| $4+$ | 45 | $440 \pm 44$ | $811 \pm 356$ | $0.90 \pm 0.18$ |  |  |
| $5+$ | 29 | $513 \pm 67$ | $1,267 \pm 511$ | $0.88 \pm 0.17$ |  |  |
| $6+$ | 12 | $507 \pm 55$ | $1,346 \pm 444$ | $0.92 \pm 0.18$ |  |  |
| $7+$ | 7 | $621 \pm 114$ | $2,374 \pm 365$ | $1.01 \pm 0.40$ |  |  |
| $8+$ | 3 | $600 \pm 79$ | $2,310 \pm$ | 0 | $0.90 \pm$ | 0.00 |
| $9+$ | 2 | 633 | $\pm 74$ | $2,840 \pm$ | 1,640 | $1.05 \pm$ |
| $10+$ | 1 | 723 | $\pm 0$ | -- | $\pm$ | -- |
| $11+$ | 1 | $766 \pm 0$ | $3,358 \pm 0$ | 0.27 |  |  |

Table 3.18. Back calculated total length (mean $\pm$ standard deviation) of walleye sampled during 1995.

|  | Back Calculated Total Length (mm) at Annulus |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Cohort $n$ | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
|  | 199482 | $115 \pm 30$ |  |  |  |  |  |  |  |  |  |  |
|  | 1993102 | $120 \pm 78$ | $227 \pm 100$ |  |  |  |  |  |  |  |  |  |
|  |  | $126 \pm 67$ | $230 \pm 150$ | $329 \pm 227$ |  |  |  |  |  |  |  |  |
| $\stackrel{\rightharpoonup}{0}$ | 199199476 | $121 \pm 25$ | $219 \pm 36$ | $310 \pm 44$ | $379 \pm 69$ |  |  |  |  |  |  |  |
|  | 199029 | $124 \pm 26$ | $225 \pm 53$ | $312 \pm 65$ | $398 \pm 71$ | $460 \pm 73$ |  |  |  |  |  |  |
|  | $1989 \quad 12$ | $122 \pm 20$ | $210 \pm 41$ | $289 \pm 53$ | $369 \pm 52$ | $423 \pm 58$ | $474 \pm 58$ |  |  |  |  |  |
|  | 19887 | $130 \pm 20$ | $239 \pm 38$ | $319 \pm 53$ | $393 \pm 69$ | $462 \pm 84$ | $528 \pm 97$ | $576 \pm 114$ |  |  |  |  |
|  | 19873 | $114 \pm 11$ | $194 \pm 43$ | $295 \pm 17$ | $366 \pm 26$ | $428 \pm 22$ | $472 \pm 32$ | $523 \pm 45$ | $575 \pm 70$ |  |  |  |
|  | 19862 | $74 \pm 18$ | $142 \pm 4$ | $220 \pm 10$ | $268 \pm 23$ | $343 \pm 17$ | $394 \pm 40$ | $450 \pm 51$ | $512 \pm 46$ | $589 \pm 100$ |  |  |
|  | 19851 | $114 \pm 0$ | $202 \pm 0$ | $264 \pm 0$ | $341 \pm 0$ | $422 \pm 0$ | $481 \pm 0$ | $536 \pm 0$ | $580 \pm 0$ | $668 \pm 0$ | $697 \pm 0$ |  |
|  | 19841 | $73 \pm 0$ | $125 \pm 0$ | $198 \pm 0$ | $290 \pm 0$ | $351 \pm 0$ | $435 \pm 0$ | $504 \pm 0$ | $564 \pm 0$ | $597 \pm 0$ | $693 \pm 0$ | $726 \pm 0$ |
|  | Grand |  |  |  |  |  |  |  |  |  |  |  |
|  | Mean | 121f 55 | $224 \pm 102$ | $315 \pm 155$ | $5380 \pm 68$ | $444 \pm 71$ | $481 \pm 72$ | $539 \pm 93$ | $556 \pm 54$ | $611 \pm 69$ | $695 \pm 3$ | $726 \pm$ |
|  | Annual Growth | 121 | 103 | 91 | 65 | 64 | 37 | 58 | 17 | 55 | 84 | 31 |

### 3.4 Feeding Habits

Feeding habits were based on fish sampled during electrofishing and gillnet sets. A total of 50 kokanee, 90 rainbow trout and 153 walleye stomachs were collected and the contents of the stomachs were enumerated by taxa. The annual index of relative importance (IRI) is reported in Table 3.19 for each species regardless of age by food item. Appendix C lists the index of relative importance, percent of food items by number and weight and the frequency of food item occurring for each fish species and age.

According to the IRI, kokanee's primary food item was Daphnia spp. (70.21). The two most important food items for rainbow trout were Daphnia spp (36.69) and percidae (13.83). The most important food items for walleye were fish (totalling 51.90) and Leptodora kindtii (2226).

Diet overlap analysis predicted that kokanee and rainbow trout overlap was 0.80 (high overlap). Kokanee and walleye diet overlap was 0.15 (low overlap) and rainbow and walleye diet overlap was 0.45 (moderate overlap).

Table 3.19. Index of relative importance for kokanee $(\mathrm{n}=\mathbf{5 0})$, rainbow trout ( $\mathrm{n}=90$ ), and walleye ( $\mathrm{n}=153$ ) from fish collected during 1995. '--' indicates no organisms found.

| PREY ITEM | Index of Relative Importance |  |  |
| :---: | :---: | :---: | :---: |
|  | Kokanee | Rainbow | Walleye |
| Osteichthyes |  |  |  |
| Catostomidae | $\cdots$ | - | 1.17 |
| Cottidae | -- | $\cdots$ | 15.62 |
| Cyprinidae | - | 2.30 | 4.02 |
| Percidae | -- | 13.83 | 16.4 |
| Sahnonidae |  | -- | 11.19 |
| Unidentified fish | 3.06 | 7.65 | 18.98 |
| Fish eggs |  | 0.93 | -- |
| Amphipoda |  |  |  |
| Gammeras sp. | -- | -- | 0.38 |
| Cladocera |  |  |  |
| Daphnia spp. | 70.21 | 36.69 | 4.49 |
| Eucopepoda |  |  |  |
| E. nevadensis | -- | 0.29 | -- |
| L. ashlandi | 0.62 | 0.29 | 0.94 |
| Copepoda sp. | 0.59 | -- | -- |
| Basommatophora |  |  |  |
| Planordidae | -- | 0.29 | -- |
| Physidae | -- | 0.51 | -- |
| Diptera |  |  |  |
| Chironomidae pupa | 13.10 | 10.60 | -- |
| Chironomidae larvae | 5.31 | 4.37 | -- |
| Sciomyzidae | - | 0.29 | -- |
| Simulidae larvae | 0.57 |  |  |
| Trichoptera |  |  |  |
| Hydroptilidae | -- | 0.60 | -- |
| Lepidostomatidae |  | 0.58 |  |
| Hydropyschidae | 0.64 | 0.29 | -- |
| Decapoda |  |  |  |
| Hemiptera |  |  |  |
| Corixidae | 0.58 | 1.78 | 0.26 |
| Plecoptera |  |  |  |
| Perlodidae | -- | 0.29 | 0.23 |
| Nemouridae | -- | 0.29 | -- |
| Ephemeroptera |  |  |  |
| Baetidae | -- | 0.29 |  |
| Leptophlebiidae | -- | 0.69 | 0.23 |
| Odonata |  |  |  |
| Anisoptera | -- | 1.16 | 0.48 |
| Zygoptera | -- | 0.33 | 0.46 |
| Coleoptera |  |  |  |
| Elmidae | -- | 0.29 | -- |
| Lepidoptera |  |  |  |
| Pyralidae | -- | -- | -- |
| Oligochaeta |  |  |  |
| Lumbriculidae | -- | -- | 1.15 |
| Hydrachnellae |  |  |  |
| Hydracharina | -- | 0.88 | $\cdots$ |
| Terrestrial | \%.33 | 10.14 | 3.74 |

### 4.0 DISCUSSION

The main objective of this study was to monitor and evaluate the effects of stocking hatchery reared kokanee salmon and rainbow trout into Lake Roosevelt on the ecosystem and the fishery. Sub-objectives were to identify stocking strategies which: maximize the numb\& of hatchery kokanee and rainbow trout harvested or captured by anglers; maximize the collection of kokanee eggs at egg collection facilities and maximize the quality of fish harvested (large size and good condition). We evaluated the effects of the stocking program on the fishery by comparing data collected prior to stocking Spokane Tribal and Sherman Creek Hatcheries fish (pre hatchery) with data collected after stocking began (post hatchery).

### 4.1 Historical Stocking and Lake Operations

There were two general factors effecting the recruitment of hatchery origin rainbow trout and kokanee salmon into the fishery. The first was stocking strategies conuolled by the Hatchery Coordination Team (Team). One member each from the Washington Department of Fish and Wildlife (WDFW), the Colville Confederated Tribes (CCT) and the Spokane Tribe of Indians (STOI) made up the Team. The Team's job was to determine: the number of fish stocked; the size of fish stocked; the time of year to stock the fish; location in the lake to stock the fish and method of stocking (e.g.. by truck). The other variables, not under Lake Roosevelt Fish Managers control, were lake operations. Lake operations were controlled by mother nature (i.e.. rainfall, snowmelt), economics and politics (i.e., power demand, irrigation).

Stocking of rainbow trout began in 1986 when the WDFW started supplying rainbow trout fry to the Lake Roosevelt Net Pen Program (operated by a volunteer organization, Lake Roosevelt Development Association). Table 4.1 indicates the number and the source of rainbow trout provided to the net pen operators. By July of 1988, the WDFW began stocking kokanee into the lake (Table 4.2). The kokanee were stocked at Sherman Creek (760,000 fry) and at Little Falls Dam on the Spokane River (141,000 fry). WDFW continued stocking approximately the same number of kokanee at Sherman Creek and Spokane River in 1989 and 1990. The Spokane Tribal Hatchery went on line in 1990 and began releasing rainbow trout and kokanee in 1991. Sherman Creek went on line

Table 4.1 Summary of hatchery origin rainbow trout released into Lake Roosevelt from 1988 though 1995.

| Year | Hatchery | Number |
| :---: | :---: | :---: |
| 1986 | Spokane (WDFW) | 50,000 |
| 1987 | Spokane (WDFW) | 80,000 |
| 1988 | Spokane (WDFW) | 150,00 |
| 1989 | Spokane (WDFW) | 175,00 |
| 1990 | Spokane (WDFW) | 276,500 |
| 1991 | Spokane Tribal | 326,461 |
| 1992 | Spokane Tribal | 424,395 |
| 1993 | Spokane Tribal | 446,798 |
| 1994 | Spokane Tribal | 448,992 |
| 1995 | Spokane Tribal | 415,844 |

Table 4.2 Summary of hatchery origin kokanee released into Lake Roosevelt from 1988 though 1995.

| Year | Hatchery | Number | Life Stage | Size <br> $(\# / L B)$ |
| :---: | :---: | ---: | :---: | ---: |
| 1988 | Ford | 872,150 | fry | 500 |
| 1989 | F o r d | 861,442 | fry | 280 |
| 1990 | Ford | $1,025,400$ | fry | 247 |
| 1991 | Spokane Tribal | $1,674,577$ | fry | 119 |
| 1992 | Spokane Tribal | 71,256 | yearling | 9 |
| 1992 | Spokane Tribal | 819,220 | fry | 158 |
| 1992 | Sherman Creek | 68,552 | yearling | 22 |
| 1992 | Sherman Creek | $1,099,000$ | fry | $616^{a}$ |
| 1993 | Spokane Tribal | 21,190 | yearling | 7 |
| 1993 | Spokane Tribal | $1,024,293$ | fry | 225 |
| 1993 | Sherman Creek | 72,508 | yearling | 15 |
| 1993 | Sherman Creek | 675,572 | fry | 228 |
| 1994 | Spokane Tribal | 29,111 | yearling | 8 |
| 1994 | Spokane Tribal | 540,220 | fry | 425 |
| 1994 | S herman Creek | 90,881 | yearlings | $11^{\mathrm{a}}$ |
| 1994 | Sherman Creek | $1,087,161$ | fry | $372^{\mathrm{a}}$ |
| 1995 | Spokane Tribal | 1,401 | brood | 1 |
| 1995 | Spokane Tribal | 59,825 | yearling | 10 |
| 1995 | Spokane Tribal | 515,425 | fry | 202 |
| 1995 | S herman Creek | 210,643 | yearlings | $15^{\mathrm{a}}$ |
| 19 | 5 5herman Creek | 164,328 | yearlings | $28^{\mathrm{a}}$ |

a size transferred from Spokane Tribal Hatchery not at release.
and began releasing kokanee in 1992. Once the new hatcheries were operational, close to 450,000 rainbow trout and 2 million kokanee were released annually. Of the kokanee released approximately 100,000 were yearlings and the rest were fry. In 1994, Tilson et al. (1995) recommended that we discontinue fry releases and instead release the kokanee as yearlings. This recommendation was made due to the fact that only $1 \%$ of all tagged fish recovered as adults from 1993 through 1994 were fish releases as fry and the remaining $99 \%$ were released as yearlings. As a result, the hatcheries released 434,796 yearlings and 515,425 fry in 1995. Shifting to a yearling release program caused a reduction in the total number of kokanee being stocked from the hatcheries because the yearlings require more hatchery space for rearing.

Lake operations for the time period 1990 through 1995 are depicted in Figures 4.1 and 4.2. Figure 4.1 identifies the monthly mean lake elevation above sea level in feet. Figure 4.2 shows the monthly mean water retention time of Lake Roosevelt in days. Generally, when the elevation of the lake fell below 1,240 feet elevation the water retention time fell below thirty days, however, this was dependent on the volume of water flowing into and out of the lake. The year 1991 was considered to be extraordinarily bad for the fishery (Thatcher et al. 1993, 1994). The lake elevation fell below 1,240 feet and the water retention time was below 30 days. Griffith and Scholz (1991) and Thatcher et al.'s (in press) 1991 and 1992 annual reports identified that water retention times less than thirty days had dramatic effects on the biota of Lake Roosevelt. The zooplankton population decreased and the entrainment of fish out of the lake through Grand Coulee Dam increased. The resulting decrease of fish food (zooplankton) and decrease of fish (fish entrainment) negatively impacted the fishery.

In contrast to 1991 operations, overall lake operations in 1995 appeared to be the best for fish over the life of this project. The mean lake elevation during 1995 was $1,277 \mathrm{ft}$ with the lowest elevation occurring in March at $1,259 \mathrm{ft}$. The mean water retention time during 1995 was 47 days with the fewest days water retention in December at 32 days. However, in August 1995 a ten foot drawdown was employed to benefited Snake River salmon which had not occurred in previous years. We were unable and unequipped to make a determination as to whether a summer drawdown had a significant effect on the fishery. Lake operations during the summer may significantly impact fish growth. The summer period (June - August) is considered the most important growing period for fish in the lake due to the rise in water temperature and food availability. The rise in temperature increases the metabolism of the fish and the high availability of food provides the fish with nutrient


Figure 4.1 Lake elevations for Lake Roosevelt from 1990 through 1995.


Figure 4.2 Water retention times for Lake Roosevelt from 1990 through 1995.
input-for growth. The combination of increased metabolism and nutrient input results in fast growth. Thus, we believe that lake operations have the greatest potential for negatively impacting fish growth during the summer months because of decreased food availability. Lake operation in 1994 was similar to 1995 with the exception that 1994 did not have an August drawdown. The water retention time in August, 1994 was 59 days and in Augusts, 1995 was 47 days. The reduced water retention time may have an effect on food availability and therefore fish growth. The abundance of fish food (i.e.. zooplankton, fry fish) in comparison to fish growth may help determine the effect of August drawdowns on the fishery.

### 4.2 Creel Survey Trends

The number of angler trips, economic worth of the fishery and the number of fish harvested in 1995 was down slightly from 1992, but the number of hours fished, the mean length of an angler trip and the catch per unit of effort was up (Table 4.3). In 1995, the total number of angler trips was 231,202 which was down slightly to the 291,380 angler trips made in 1992 (Table 4.3). The number of kokanee harvested in $1995(32,323)$ was four times greater than the number of kokanee harvested $(8,021)$ in 1992. The number of rainbow trout harvested $(122,939)$ in 1995 was down slightly to the number harvested $(140,609)$ 1992. The most dramatic change was the number of walleye harvested. The harvest went from 118,863 fish in 1992 to 40,185 fish in 1995. The 1995 economic worth of the fishery was close to $\$ 8.7$ million which was down in comparison to the 1992 economic worth of $\$ 9.7$ million. The total number of fish harvested was 205,809 fish with a total annual harvest rate of 0.173 harvest per unit effort (HPUE). In 1992, the number of fish harvested was 291,886 with a total annual harvest rate of 0.377 HPUE. The mean annual angler trip length had increased steadily from 3.3 hours per trip in 1992 to 5.1 hour in 1995.

Overall, the number of angler trips appear to be down due to the increased trip length. Angler trips are determined by dividing average trip lengths by total estimated hours fished. Angler trips may be increasing due to the lower catch rate resulting in more angler hours in order to obtain bag limits or at least accustomed number of fish harvest per trip.

### 4.2.1 Kokanee Salmon

Table 4.3 indicates that the kokanee fishery is building. Since 1992, the number of kokanee harvested per year has steadily increased from 8,021 kokanee in 1992 to 32,353 kokanee in 1995 and the harvest rate has increased from <0.01 HPUE in 1994 to 0.02 HPUE in 1995. The increased harvest rate suggests that anglers were catching more kokanee with less effort due to either an increase in the abundance of kokanee causing greater availability or kokanee anglers had become more proficient at catching kokanee which would reduce the amount of time spent by anglers between catches. Lake operations during 1993 though 1995 were relatively fish friendly. The drawdowns were limited and water retention time did not fall below thirty days. This led us to believe that kokanee were more abundant in 1995 than in years past. However, angler proficiency could not be mled out. The information collected by the survey does not provide for determining whether anglers were becoming better fisherman thereby requiring less time to catch the same number of fish.

Another factor contributing to the kokanee harvest was the stocking of hatchery reared fish. Of the 112 kokanee observed by the creel clerks in 1995, one kokanee ( $0.8 \%$ ) had a mark identifying it as a hatchery origin fish. In addition, Eastern Washington University (EWU) conducted an augmented creel during 1995 targeting kokanee anglers. EWU's creel clerks observed 282 kokanee of which $19(6.7 \%)$ had marks identifying them as hatchery origin fish (Tilson et al 1996). Approximately $12 \%$ of the kokanee stocked by the hatcheries from 1992 through 1994 bore adipose fin clips. If we assume that all fish released during 1992 through 1994 had an equal probability of being recruited into the fishery during 1995, and that all of the kokanee harvested by anglers were of hatchery origin, then we would expect $12 \%$ of the kokanee observed by the creel clerks to bear marks identifying them as hatchery origin fish. Instead the creel clerks observed between $0.08 \%$ and $6.7 \%$ of the kokanee bearing hatchery origin markings. the estimated percentages of hatchery fish caught could be calculated by dividing the percentage of hatchery fish (marked) seen by the creel clerks by the percentage of marked fish stocked from the hatcheries. Therefore, the estimated percentage of hatchery fish harvested could be as low as $7 \%$ ( $0.8 \% / 12 \%$ ) or as high as $56 \% ~(6.7 \% / 12 \%)$ depending on the survey. Thus, from $7 \%$ to $56 \%$ of the estimated harvest were hatchery origin fish which suggests that of the 32,353 harvested kokanee 2,265 to 18,118 kokanee were of hatchery origin. In order to obtain better composition estimates, beginning in 1995 all hatchery origin fish released into the lake have been adipose clipped. The marked fish will enable us to make positive identification of hatchery
versus wild origin fish. This will allow for more precise estimates of the hatchery origin contribution to the fishery.

### 4.2.2 Rainbow trout

The number of rainbow trout harvested $(122,929)$ in 1995 appeared to be significantly lower than that estimated for $1993(398,943)$ and for $1994(499,460)$ (Table 4.3). We identified two possible reasons for the large change among years. The 1993 and 1994 harvest may have been overestimated and/or the rainbow trout were overharvested. During 1992, 1993 and 1994 lake operations were relatively fish friendly in comparison to previous years which led us to believe that harvest had increased in those years. However, to achieve the harvest estimated for 1993 and 1994 virtually every rainbow stocked in those years would have to have been harvested. We do not believe this was true due to the abundance of fish predators in the lake, know to feed on newly stocked rainbow trout. On the other hand, if the 1993 and 1994 estimates were accurate, than the rainbow trout population was probably overharvested. The overharvest would have caused a reduction in the total number rainbow available for harvest, thus, resulting in a decreased harvest. However, according to 1995 rainbow trout tag results (see chapter 1) approximately 56\% of the fish released in 1995 recruited into the fishery and the remaining $44 \%$ were rainbow trout released one to three years prior to 1995. If overharvest had occurred, we would not expect to see almost half of the tag rainbow trout to be from the years in which overharvest occurred. We do not recommend altering our current management strategy of rainbow. Instead, we plan to reanalyze the 1993 and 1994 creel data to determine if the rainbow harvest was overestimated. Furthermore, another year of data collection will help to establish the harvest trend for rainbow trout.

### 4.2.3. Walleye

The number of walleye harvested in 1994 and 1995 appears to be significantly lower than the number harvested in previous years (Table 4.3). As a result, we have hypothesized that either the walleye population was overharvested in previous years or factors unknown to us have impeded the recruitment of walleye to the fishery over the past two years. The estimated number of walleye harvested in 1991 was 168,736 the number of walleye harvested fell to 118,863 in 1992. In 1993, the number of walleye harvested grew to 307,663 fish. We question the magnitude of the harvest, but accept that the harvest was

Table 4.3 Summary of angler trips, number of fish caught and harvested, catch and harvest per unit of effort and mean lengths of kokanee, rainbow trout and walleye from 1990 through 1995.

|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Angler Trips | 17 | 1,725 | 398,408 | 29 | 1,380 | 594,508 |
|  |  |  | 469,998 | 23 | 1,202 |  |
| No. Caught |  |  |  |  |  |  |
| kokanee | 17,756 | 31,651 | 8,146 | 13,986 | 16,567 | 32,353 |
| rainbow | 81,560 | 81,529 | 167,156 | 402,277 | 499,460 | 125,958 |
| walleye | 116,473 | 231,813 | 163,995 | 337,413 | 123,612 | 73,667 |
|  |  |  |  |  |  |  |
| No. Harvested |  |  |  |  |  |  |
| kokanee | 17,515 | 31,651 | 8,021 | 13,960 | 16,567 | 32,353 |
| rainbow | 79,683 | 73,777 | 140,609 | 398,943 | 499,293 | 122,939 |
| walleye | 82,284 | 168,736 | 118,863 | 307,663 | 53,589 | 40,185 |
|  |  |  |  |  |  |  |
| CPUE |  |  |  |  |  |  |
| kokanee | 0.03 | 0.06 | 0.03 | 0.01 | co.01 | 0.02 |
| rainbow | 0.13 | 0.20 | 0.22 | 0.17 | 0.21 | 0.08 |
| walleye | 0.11 | 0.11 | 0.15 | 0.12 | 0.08 | 0.13 |
|  |  |  |  |  |  |  |
| HPUE |  |  |  |  |  |  |
| kokanee | 0.02 | 0.06 | 0.03 | 0.01 | co.01 | 0.02 |
| rainbow | 0.12 | 0.20 | 0.18 | 0.16 | 0.21 | 0.08 |
| walleye | 0.08 | 0.08 | 0.11 | 0.08 | 0.05 | 0.06 |
|  |  |  |  |  |  |  |
| Mean Length |  |  |  |  |  |  |
| kokanee | 391 | 361 | 436 | 486 | 481 | 467 |
| rainbow | 346 | 348 | 422 | 471 | 473 | 410 |
| walleye | 376 | 397 | 361 | 382 | 385 | 370 |
|  |  |  |  |  |  |  |

likely greater than previous years (see section 4.2.4.). At any rate, the increased harvest may have caused the decrease of the walleye fishery in 1994 and 1995. The harvest per unit of effort has decreased from 0.08 to 0.11 in 1991 and 1992 respectively and decreased further to 0.05 in 1994 and 0.06 in 1995. The reduced harvest rate indicates that the number of hours required to harvest a walleye has grown from 9 hours per fish in 1992 to 16 hours per fish in 1995.

The reduced number of walleye harvested over the past two years indicates that the walleye fishery is possibly in a long term slump. The exact reason for the reduced number of walleye is not known, however, overharvest is conceivable. Another year of data is necessary to attain more accurate trend data before making any management decisions, but reducing the bag limits of walleye may be necessary in future years.

### 4.2.4 Accuracy and Precision

The estimated harvest for rainbow trout and walleye in 1993 and 1994 has come into question. The number harvested seem to be greater than the number of fish available for harvest. Less than $10 \%$ of the rainbow trout harvest was composed of wild origin fish and the rest were from net pens according to accounts by the creel clerks. In 1993 and 1994, we estimated that 446,798 and 449,183 net pen origin rainbow were released, respectively. The harvest estimates for 1993 and 1994 were 398,943 and 499,294 trout respectively. If $90 \%$ of the rainbow trout harvested were of net pen origin then $79 \%$ of the released rainbow trout were harvested in 1993 and over 100\% in 1994. We do not believe that all of the rainbow trout released were harvested. We have documented walleye prey on released rainbow, and entrainment of rainbow trout over the dam. These factors would reduce the available rainbow trout for harvest. As a result the harvested number of fish appears to be overestimated. There are also indications that a harvest in excess of 200,000 walleye was not probable. Beckman et al. (1985) estimated that Lake Roosevelt walleye fishery could sustain approximately 100,000 fish annually. For the duration of this project the relative abundance of walleye has decreased suggesting that the prevalence of walleye has decreased. Thus, the harvest of walleye in 1993 at three times the rate indicated by Beckman et al. (1985) is not unlikely.

In addition, the number of angler hours (pressure) seem to be overestimated. The National Park Service tracked the number of boats launched at access points along Lake Roosevelt.

The number of boats launched did not change appreciably from 1992 to 1993 (Scott Hebner, personal communication, NPS). This suggests that the number of anglers fishing did not change from 1992 to 1993. The overestimate of angler hours in 1993 may have been caused from the method of calculation. The Lake Roosevelt creel survey uses a count of the number of boat trailers as a method for estimating the number of boats fishing, but not all boat trailers equal boats fishing. Many boats during the summer contained persons not fishing, instead they were boating, sun bathing, water skiing, swimming, etc. The creel survey was designed to correct the number of boat anglers by comparing the number of boat anglers to non boat anglers. The creel clerks attempted to contact all boat at the access point to ask them whether they were fishing or not. When the number of angler hours were estimated a ratio was determined by comparing the number of boaters fishing to non fishing boaters as recorded by the creel clerks. However, the creel clerks contacted mainly anglers thereby biasing toward contacting anglers instead of non anglers. As a result, the percent of boaters angling versus not angling may be biased towards angling.

A large majority of the boats that were non anglers occurred during the summer in the mid to lower river. A substantial number of rainbow and walleye were harvested during the summer. Thus, if the estimate of angler hours during summer months was skewed, then the number of fish harvested would also be skewed. In 1994, the NPS estimated an $18 \%$ increase of boat launches at Spring Canyon during May through August when compared to 1993. It appears the influx of boats may have amplified the harvest estimates for rainbow trout if the percentage of angler boats versus non anglers were skewed. A more refined analysis pointed at determining the ratio of angler to non-angler boats is required.

### 4.3 Relative Abundance

The relative abundance of kokanee increased substantially ranging from $<1$ to $3 \%$ during 1989 through 1994 to $20 \%$ in 1995 (Table 4.4). The catch per unit of effort increased from as low as 0.08 catch per unit effort (CPUE) in past years to 0.98 CPUE (Table 4.5). The relative abundance, catch per unit of effort and harvest estimates all suggest that the kokanee populations have increased substantially. The hatchery program appears to be the cause for the increased population suggesting that hatchery stocking strategies have benefited the fishery. Before hatchery releases, relative abundance of kokanee salmon and rainbow trout in gillnet surveys was $<1 \%$ (Beckman et al. 1985). In 1995, gill net relative abundance of kokanee salmon and rainbow trout increased to 4 and 5\% respectively. Other

Table 4.4 Comparison of relative abundance (\%) of fish collected during the 1989 through 1995 sampling periods via electroshocking and gillnetting.

## 1989199019911992199319941995

| Effort (hrs) | 482 | 581 | 366 | 436 | 100 | 643 | 2,099 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| bridgelip sucker | 1 | $<1$ | $<1$ | $<1$ |  | $<1$ | $<$ |
| brook trout | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ |
| brown bullhead | 0 | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ |
| brown trout | $<1$ | $<1$ | $<1$ | Cl | $<1$ | $<1$ | $<1$ |
| bull trout | $<1$ | 0 | 0 | 0 | 0 | $<1$ | $<1$ |
| burbot | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | 1 | $<1$ |
| carp | 2 | 2 | $<1$ | 2 | 1 | 1 | 2 |
| chinook salmon | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ |
| chiselmouth | 0 | $<1$ | 0 | 0 | 0 | 0 | 0 |
| Cottus spp. | 2 | 2 | c 1 | 2 | 3 | 16 | 6 |
| crappie | $<1$ | $<1$ | $<1$ | $<1$ | 0 | $<1$ | $<1$ |
| kokanee salmon | 2 | $<1$ | $<1$ | 3 | $<1$ | 3 | 20 |
| lake whitefish | 4 | 3 | $<1$ | 1 | c 1 | 2 | 5 |
| largemouth bass | $<1$ | $<1$ | $<1$ | 0 | 0 | 0 | 0 |
| largescale sucker | 12 | 19 | 34 | 43 | 45 | 35 | 27 |
| longnose sucker | $<1$ | 2 | $<1$ | $<1$ | 0 | 2 | $<1$ |
| mountain whitefish | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ | $<1$ |
| peamouth | $<1$ | 0 | $<1$ | $<1$ | 0 | 0 | 0 |
| pumpkinseed | Cl | $<1$ | 0 | 0 | 0 | 2 | 0 |
| rainbow trout | 5 | 3 | 5 | 6 | 9 | $<1$ | 5 |
| redside shiner | 0 | $<1$ | 0 | 0 | 0 | $<1$ | $<1$ |
| smallmouth bass | 2 | 3 | 15 | 11 | 9 | 8 | 10 |
| squawfish | 4 | 5 | 3 | 2 | 8 | 4 | 2 |
| sturgeon | c 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus spp. | 7 | 0 | 0 | 0 | 0 | 0 | $<1$ |
| tench | $<1$ | $<1$ | $<1$ | $<1$ | 0 | $<1$ | $<1$ |
| walleye | 18 | 13 | 12 | 11 | 11 | 7 | 12 |
| yellow bullhead | $<1$ | 0 | 0 | 0 | 0 | 0 | 0 |
| yellow perch | 40 | 45 | 29 | 17 | 11 | 12 | 7 |

Table 4.5 Comparison of catch per unit effort (No. fish per hour) for fish collected during the 1989 through 1995 sampling periods via electroshocking and gillnetting.

| Effort (hrs) | 482 | 581 | 366 | 436 | 100 | 643 | 2,099 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| bridgelip sucker | 0.21 | 0.01 | 0.03 | C0.01 | 0.00 | x0.01 | 0.06 |
| brook trout | 0.01 | C0. 01 | C0. 01 | 0.01 | 0.02 | 0.02 | 0.01 |
| brown bullhead | 0.00 | C0. 01 | co. 01 | 0.07 | 0.03 | 0.00 | 0.00 |
| brown trout | 0.04 | 0.03 | 0.04 | 0.04 | 0.16 | 0.03 | 0.02 |
| bull trout | C0. 01 | 0.00 | 0.00 | 0.00 | 0.00 | C0. 01 | 0.01 |
| burbot | 0.06 | 0.02 | 0.05 | 0.02 | 0.03 | 0.14 | 0.04 |
| carp | 0.24 | 0.26 | 0.20 | 0.15 | 0.22 | 0.19 | 0.11 |
| chinook salmon | <0.01 | <0.01 | 0.01 | 0.01 | 0.01 | co. 01 | co. 01 |
| chiselmouth | 0.00 | C0. 01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Cottus spp. | 0.27 | 0.22 | 0.06 | 0.16 | 0.62 | 2.13 | 0.27 |
| crappie | 0.09 | 0.02 | C0. 01 | 0.04 | 0.00 | 0.01 | 0.00 |
| kokanee salmon | 0.27 | 0.10 | 0.08 | 0.28 | 0.15 | 0.46 | 0.98 |
| lake whitefish | 0.56 | 0.38 | 0.20 | 0.10 | 0.15 | 0.26 | 0.25 |
| largemouth bass | 0.10 | 0.05 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| largescale sucker | 1.87 | 2.85 | 7.51 | 3.91 | 10.12 | 4.76 | 1.30 |
| longnose sucker | 0.04 | 0.32 | 0.01 | 0.01 | 0.00 | 0.26 | 0.02 |
| mountain whitefish | 0.03 | 0.03 | 0.08 | 0.01 | 0.02 | 0.01 | 0.01 |
| peamouth | 0.03 | 0.00 | C0. 01 | 0.01 | 0.00 | 0.00 | 0.00 |
| pumpkinseed | 0.01 | 0.10 | 0.00 | 0.00 | 0.00 | 0.20 | 0.00 |
| rainbow trout | 0.82 | 0.43 | 1.02 | 0.56 | 2.03 | 0.88 | 0.24 |
| redside shiner | 0.00 | <0.01 | 0.00 | 0.00 | 0.00 | 0.01 | <0.01 |
| smallmouth bass | 0.24 | 0.46 | 3.22 | 1.01 | 2.08 | 1.12 | 0.46 |
| squawfish | 0.61 | 0.80 | 0.59 | 0.21 | 1.84 | 0.49 | 0.11 |
| sturgeon | co. 01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Ca tostomus spp . | 0.99 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| tench | 0.01 | 0.03 | 0.01 | 0.01 | 0.00 | 0.02 | CO. 01 |
| walleye | 2.70 | 1.96 | 2.60 | 0.99 | 2.34 | 1.00 | 0.58 |
| yellow bullhead | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| yellow perch | 6.02 | 6.65 | 6.40 | 1.55 | 2.48 | 1.63 | 0.34 |
| TOTALS | 15.24 | 14.73 | 22.13 | 9.15 | 22.29 | 13.61 | 4.79 |

fish populations in Lake Roosevelt appear to be relatively stable over the last seven years except for yellow perch. Both relative abundance and catch per unit effort indicated that yellow perch populations were declining. Walleye were documented by'previous Lake Roosevelt Monitoring Program Reports to feed on yellow perch (Underwood and Shields 1996). Walleye may be limiting the yellow perch through predation.

### 4.4 Growth and Feeding.

Peone et al. (1990) examined the growth of Lake Roosevelt kokanee, rainbow trout and walleye, in comparison to growth of these species in area lakes. The comparison was made using back calculated lengths from scales. Peone et al. concluded that fish in Lake Roosevelt grew to a larger size at a young age than fish in area lakes. Their statement still holds true in 1995 for rainbow trout and kokanee. However, their statement does not hold true for walleye. The back calculated length of walleye sampled in 1995 was below the walleye length average of area lakes by approximately five centimeters per year of life . The slower growth rates may be an early indicator that the food base for walleye was limited. One of the main prey items for walleye is yellow perch. The yellow perch population appears to be declining based on relative abundance and catch per unit effort.

The feeding habits of rainbow trout, kokanee and walleye differed slightly from previous years. Kokanee salmon utilized mostly Daphnia spp. and chironomids. Walleye fed primarily on fish. Rainbow trout fed primarily on Daphnia spp., chironnomids and a new food item, yellow perch. In 1995, the summer collection of fish for diet analysis changed from August to July in order to capture more fish. August water temperatures cause the fish to reside in deeper water limiting our ability to collect them. As a result, we observed for the first time rainbow trout with numerous yellow perch in their stomachs and yellow perch were estimated to be the second most important food item. Up to this point walleye were thought to be the cause for the reduction in the yellow perch population. Rainbow trout may have had a significant effect of the yellow perch as well. The availability of food should be considered prior to increasing the number of rainbow trout stocked annually.

Feeding habits and growth analysis both suggest that rainbow trout and kokanee populations had ample food. On the other hand, walleye growth appeared to be slowing and the food availability may be the cause. The condition factor of walleye was 0.84 which was not different than past years. The limitation of food may not be large enough to cause
the walleye to starve, causing a reduced condition factor, but may cause a reduced overall fitness of population. This may be why the walleye harvest in 1995 was less than previous years. In addition, rainbow released from the net pens may be competing with walleye for food, the diet overlap between walleye and rainbow trout was moderate (45\%).
Competition between walleye and rainbow trout may be stronger than diet overlap suggests. This may be why the number of walleye harvested has decreased and the relative abundance of yellow perch has decreased, as the number of rainbow trout released from net pens increased over the few years.

The diet overlap among rainbow and kokanee had a high diet overlap ( $80 \%$ ) meaning kokanee and rainbow trout used similar food types. However, growth of kokanee and rainbow trout appears to be good. Kokanee and walleye did not use similar food types (15\%). Food does not appear to be limited for kokanee or rainbow trout.

### 5.0 RECOMMENDATIONS AND RESEARCH NEEDS

1) Quantify the impact of walleye on newly stocked kokanee. This will give us a better estimate of the actual number of kokanee stocked into the lake after walleye have reduced the population.

2 ) Record origin of every kokanee and rainbow sampled so that comparisons can be made between hatchery origin and wild origin fish. We were unable to determine the number of hatchery and wild origin kokanee harvested with any accuracy in past years.

3 ) Evaluate the scientific design of the creel survey and methods used to compute indices. Question accuracy of ratio estimating the number of boats containing anglers versus non-anglers

4 ) Conduct hydroacoustic surveys monthly to identify spatial and temporal accumulations of fish along the length and width of the lake.
5) Vertical net, beam trawl or purse seine in area of know fish assemblages to determine fish species, age structure, feeding habits and growth rates of fish contained within the assemblage.

## LITERATURE CITATIONS

APHA. 1976. Standard Methods for the Examination of Water and Wastewater, 14th Ed. American Public Health Association. Washington, D.C. 1192 pp.

Beckman, L.G., J.F. Novotny, W.R. Parsons, and T.T. Tarrell. 1985. Assessment of the fisheries and limnology in Lake F.D. Roosevelt 1980-1983. U.S. Fish and Wildlife Service. Final Report to U.S. Bureau of Reclamation. Contract No. WPRS-0-07-10-X0216; FWS-14-06-009-904, May 1985. 168 pp.

Borror, D.J., D.M. Delong, CA. Triplehom. 1976. An introduction to the study of insects. 4th ed. Holt, Rinehart, and Winston. 852 pp.

Bowler, B., B.E. Reiman, and V.L. Ellis. 1979. Pend Oreile Lake fisheries investigations. Idaho Department of Fish and Game, Job Performance Report, Project F-73-R1, Boise.

Brooks, J.L. 1957. The systematics of North America Daphnia. Conn. Acad. Arts and Sci. Vol. 13, New Haven, CT. 180 pp.

Carlander, K.D. 1950. Some considerations in the use of the fish growth data based upon scale studies. Trans. Amer. Fish. Soc. 79: 187-194.

Carlander, K.D. 1981. Caution on the use of the regression method of back-calculating lengths from scale measurements. Fisheries 6:2-4.

Edmonds, G.F., S.L. Jensen, and L. Bemer. 1976. The Mayflies of North and Central America. University of Minnesota Press. Minneapolis, MN. 330 pp.

Everhart, W.H. and W.D. Youngs. 1981. Principles of Fishery Science, 2nd Ed. Cornell University Press. Ithaca, New York. 359 pp.

Fletcher, D.H. 1988. Phase management research, first year's work at Kitsap Lake, Kitsap County, Washington. Washington Department of Wildlife, Fisheries Management Division, Olympia, WA. Report No. 88-6:80.

George, E.L. and W.F. Hadley. 1979. Food habitat partitioning between rock bass (Ambloptites rupestris) and smallmouth bass (Micropterus dolomiceu) young-of-the-year. Trans. Amer. Fish. Soc. 108:253-261.

Griffith, J.R. and A.C. McDowel. 1993. Measuresment of Lake Roosevelt Biota in Relations to Reservoir Operations, Annual Report 1992. Bonneville Power Administration. Portland, OR. Project No. 88-63. 87pp plus appendices

Griffith, J.R., A.C. McDowel and A.T. Scholz. 1992. Measure of Lake Rosevelt Biota in Relation to Reservoir Operations, Annual Report 199 1. Bonneville Power Administration. Portland, OR. Project No. 88-63. 138 plus appendices.

Griffith, J.R. and A.T. Scholz. 1991. Lake Roosveit Fisheries Monitoring Program, Annual Report 1990. Bonneville Power Administration. Portland, OR. Project No. 88-63. 218 pp plus appendices

Hile, R. 1970. Body-scale relation and calculation of growth in fishes. Trans. Amer. Fish. Soc. 99:468-474.

Horn, H.S. 1966. Measurement of "overlap" in comparative ecologically studies. Amer. Nat. 100:419-429.

Hubert, W.A. 1983. Passive Capture Techniques. In: L.A. Nielsen and D.L. Johnson (ed.). Fisheries Techniques. Amer. Fish. Soc. Bethesda, MD. 468 pp.

Jagielo, T. 1984. A comparison of nutrient loading, phytoplankton standing crop, and trophic state in two morphologically and hydraulically different reservoirs. MS thesis. University of Washington. Seattle, WA. 99 pp.

Jearld, A. 1983; Age determination. In: L.A. Nielsen and D.L. Johnson (ed.). Fisheries Techniques. Amer. Fish. Soc., Bethesda, MD. 468 pp.

Lambou, V.W. 1961. Determination of fishing pressure from fishermen of party counts with a discussion of sampling problems. Proc. of the S.E. Association of Game and Fish Commissioners: 1961:380-401.

Lambou, V.W. 1966, Recommended method of reporting creel survey data for reservoirs. Oklahoma Fishery Research Laboratory, University of Oklahoma, Norman, OK. Bulletin No. 4. 39pp.

Lewis, S.L. 1975. Evaluation of the Kokanee fishery and hatchery releases at Ode11 Lake, 1964-1975. Oregon Department of Fish and Wildlife. D.J. Rep. F-71-R.

MacArthur, R.H. 1968. The theory of the niche. In: Lewontin, R.C. (ed.). Population Biology and Evolution. Syracuse University Press, Syracuse, New York: 205 pp.

Malvestuto, S.P. 1983. Sampling the Recreational Fishery. In: LA. Nielsen and P.L. Johnson (ed.). Fisheries Techniques. Amer. Fish. Soc. Bethesda, MD. 468 pp.

Malvestuto, S.P., W.D. Davies, and W.C. Shelton. 1978. An evaluation of the roving creel survey with nonuniform probability sampling. Trans. Amer. Fish. Soc. 107:255-262.

Meiole, Melo 1994. Idaho Department of Fish and Game. Couer d'Alene, Idaho.
Mendel, G. and M. Schuck. 1987. Fall 1985 and Spring 1986 Snake River steelhead creel surveys. Washington Department of Wildlife. D.J. Rep. FRILSR- 87-B.

Merritt, R.W. and K.W. Cummins. 1984. An Introduction to the Aquatic Insects of North America. Kendell-Hunt, Dubuque, IA. 722 pp.

Morisita, M. 1959. Measuring of interspecific association and similarity between communities. Mem. Fac. Sci., Kyushu University Sev. E. Biol. 3:65-80.

Novotany, D.W. and G.R. Prigel. 1974. Electrofishing boats: Improved designs and operation guidelines to increase the effectiveness of boom shockers. Wisconsin Department Natural Resources Technical Bulletin No. 73. 48 pp.

NPPC. 1987. Columbia River Basin Fish and Wildlife Program. Section 900 Resident Fish. Northwest Power Planning Council, Portland, OR. 125-126 pp.

Pennak, R.W. 1978. Freshwater Invertebrates of the United States, 2nd ed. Wiley and sons, New York. 803 pp.

Pennak, R.W. 1989. Freshwater Invertebrates of the United States, 3rd ed. Wiley and sons, New York. 628 pp.

Peone, T., A.T. Scholz, J.R. Griffith, S. Graves, and M.G. Thatcher. 1990. Lake Roosevelt Fisheries Monitoring Program. Annual Report, 1988-98. Bonneville Power Administration. Portland, OR. 234pp plus appendices.

Peterson, R.H. and D.J. Martin-Robichaud. 1982. Food habits of fishes in ten New Brunswick Lakes. Can. Tech. Rep. Fish. Aquat. Sci. 1094:43.

Pratt, K.L. 1985. Pend Oreille trout and char life history study. Idaho Department of Fish and Game. Boise, ID. 105 pp.

Reynolds, J.B. 1983. Electrofishing. In: L. A. Nielsen and D.L. Johnson (ed.). Fisheries Techniques. Amer. Fish. Soc. Bethesda, MD: 468 pp.

Rieman, B.E. 1992. Kokanee salmon population dynamics-kokanee salmon monitoring guidelines. Idaho Department of Fish and Goamne, Job Performance Report, Project F-73-R-14, Subproject II, Sudy II, Boise.

Ruttner-Kolisko, A. 1974. Plankton Rotifers Biology and Taxonomy. Die Binnengewasser, Stutgart. 26/1. 146 pp.

Scholz, A.T., R. J. White, M.B. Tilson, and S.A. Horton. 1993. Measurement of Thyroxine Concentration as an Indicator of the Critical Period for Imprinting in Kokanee Salmon (Oncorhynchus nerka): Implications for Operating Lake Roosevelt Kokanee Hatcheries, Annual Report 1992. Bonneville Power Administration. Portland, OR. Project No. 88-63. 60 pp.

Scholz, A.T., R.J. White, V.A. Koehler and S.A. Horton. 1992. Measurement of Thyroxine Concentration as an Indicator of the Critical Period for Imprinting in Kokanee Salmon (Oncorhynchus nerka): Implications for Operating Lake Roosevelt Kokanee Hatcheries, Annual Report 199 1. Bonneville Power Administration. Portland, OR. Project No. 88-63. 96pp.

Thatcher, M.G., A.C. McDowell, J.R. Griffith, and A.T. Scholz. (In press, submitted 1994). Lake Roosevelt Fisheries Monitoring Program, Annual Report 1992. Bonneville Power Administration. Portland, OR. Project No. 88-63. 237pp plus adendices.

Thatcher, M.G., J.R. Griffith, A.C. McDowell, and A.T. Scholz. ( In press submitted 1993). Lake Roosevelt Fisheries Monitoring Program, Annual Report 1991. Bonneville Power Administration. Portland, OR. Project No. 88-63. 237pp plus adendices.

Tilson, M.B., A.T. Scholz, R.J. White and J. Galloway. 1994. Thyroid Induced Chemical Imprinting in Early Life Stages and Assessment of Smoltification in Kokanee Salmon: Implications for operations Lake Roosevelt Kokanee Salmon Hatcheries, Annual Reprot 1993. Bonneville Power Administration. Portland, OR. Project No. 88-63. 156pp.

Underwood, K.D. and J.P. Shields. (in press). Lake Roosevelt Fisheries Monitoring Program, Annual Report 1993. Bonneville Power Administration. Portland, OR. Project No. 88-63. 46pp plus appendices.

Underwood, K.D., J.P. Shields and M.B. Tilson. (in press). Lake Roosevelt Fisheries and Limnological Research, Annual Reports 1994. Bonneville Power Administration. Portland, OR. Project No. 88-63 and 94-43. 358pp.

Voeller, A. C. 1996. Measurements of Lake Roosevelt Biota in Relations to Reservoir Operations. Annual Report 1993. Bonneville Power Administration. Portland, OR. Project No. 94-43.109pp.

Ward, H.B. and G.C. Whipple. 1966. Freshwater Biology, 2nd Ed. John Whiley and Sons, New York. 1,248pp.

Weber, C.I. (ed.). 1973. Biological field and laboratory methods for measuring the quality of surface waters and effluents. NERC/EPA, Cincinnati, Ohio. 176 pp .

Whitt, CR. 1958. Age and growth characteristics of Lake Pend Oreille Kokanee, 1956. Idaho Department of Fish and Game. D.J. Rep. F3-R-6.

Wiggins, G.B. 1977. Larvae of the North American Caddisfly Genera (Trichoptera). University of Toronto. Toronto, ONT: 568 pp.

Williams, K. and L. Brown. 1984. Mid-Columbia walleye life history and management 1979-1982. Washington Department of Game, Fish Management Division. Olympia, WA. Internal report. 38 pp .

Williams, R.A., A.T. Scholz, and J. Whalen. 1989. The assessment of the developing mixed-species fishery in Sprague Lake, Adams and Lincoln Counties, Washington, following restoration with rotenone. Final report submitted to Washington Department of Wildlife, Olympia, WA. Department of Biology, Eastern Washington University, Cheney, WA. 160 pp.

Wonnacott, T.H. and R.J. Wonnacott. 1977. Introductory Statistics. Third edition. John Wiley and Sons, New York, N.Y. 650 pp.

Wydoski, R.S. and R.R. Whitney. 1979. I nI and Fishes of Washington. University of Washington Press. Seattle, WA. 220 pp.

# APPENDIX A 

## Creel Survey Data

Table A. 1 Boat trailer to boats on the water correction factor split by quarters. Correction factors were establised during 1990 through 1993.

|  | YEAR |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| STRATA | 1990 | 1991 | 1992 | 1993 | MEAN $\pm$ STDEV |
| WINTER Dee-Feb | 3.49 | 1.92 | 2.01 | 2.57 | $2.50 \pm 0.72$ |
| SPRING Mar-May | 3.02 | 3.74 | 1.08 | 1.52 | $2.34 \pm 1.25$ |
| SUMMER | Jun-Aug | 3.71 | 3.17 | 1.10 | 1.01 |
| FALL | Aug-Nov | 1.46 | 3.13 | 1.17 | 1.02 |
| ANNUAL Dee-Nov | 2.92 | 2.99 | 1.34 | 1.53 | $2.19 \pm 1.40$ |
| AN |  |  |  |  |  |

Table A. 2 Boat trailer to boats on the water correction factor split by quarters and by weekday (WD) and weekend (WE) strata. The average among years were applied to pressure estimates in 1993 through 1995.

YEAR
1990-1993

|  |  | YEAR |  |  |  | $1990-1993$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| STRATA |  | 1990 | 1991 | 1992 | 1993 | MEAN $\pm$ STDEV |  |
| WINTER | WD | 3.90 | 1.60 | 1.07 | 2.14 | $\mathbf{2 . 1 8} \pm 1.23$ |  |
|  | WE | 1.84 | 2.24 | 2.49 | 2.85 | $\mathbf{2 . 3 5} \pm 0.42$ |  |
| SPRING | WD | 3.65 | 5.73 | 1.50 | 1.43 | $\mathbf{3 . 0 8} \pm 2.05$ |  |
|  | WE | 2.39 | 1.75 | 0.77 | 1.78 | $\mathbf{1 . 6 7} \pm 0.67$ |  |
| SUMMER | WD | 3.37 | 2.96 | 1.13 | 0.66 | $\mathbf{2 . 0 3} \pm 1.33$ |  |
|  | WE | 4.12 | 3.59 | 1.05 | 1.35 | $\mathbf{2 . 5 3} \pm 1.55$ |  |
|  | WD | 1.53 | 4.07 | 1.27 | 0.87 | $\mathbf{1 . 9 3} \pm 1.45$ |  |
|  | WE | 1.41 | 2.20 | 1.10 | 1.33 | $\mathbf{1 . 5 1} \pm 0.48$ |  |
| ANNUAL | WD | 3.11 | 3.59 | 1.24 | 1.28 | $\mathbf{2 . 3 0} \pm 1.22$ |  |
|  | WE | 2.44 | 2.45 | 1.35 | 1.83 | $\mathbf{2 . 0 2} \pm 0.53$ |  |

Table A. 3 Section 1 pressure estimates in hours for boat anglers in 1995 with intermediate calculations.

| STRATA |  | Correct. factor | ```Mean boat trailers for the day``` | \% of boats fishing | $\begin{gathered} \# \\ \text { angler/ } \\ \text { boat } \\ \hline \end{gathered}$ | $\begin{gathered} \text { \# of } \\ \text { angler/ } \\ \text { boat } \\ \text { S.D. } \end{gathered}$ | $\begin{gathered} \text { Corrected } \\ \text { mean } \\ \text { angler } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Corrected } \\ \text { x angler } \\ \text { sd } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| December | WD | 1.60 | 0.50 | 1.00 | 2.00 | 0.00 | 1.6 | 0.0 |
|  | WE | 2.67 | 0.00 | 1.00 | 0.00 | 0.00 | 0.0 | 0.0 |
| January | WD | 1.60 | 0.17 | 1.00 | 2.00 | 0.00 | 0.5 | 0.0 |
|  | WE | 2.67 | 0.00 | 1.00 | 1.23 | 0.00 | 0.0 | 0.0 |
| February | WD | 1.60 | 0.27 | 1.00 | 2.00 | 0.00 | 0.9 | 0.0 |
|  | WE | 2.67 | 1.00 | 0.75 | 2.46 | 0.78 | 4.7 | 1.5 |
| March | WD | 1.46 | 1.36 | 1.00 | 2.33 | 0.58 | 4.6 | 1.2 |
|  | WE | 1.28 | 4.43 | 1.00 | 3.00 | 0.00 | 17.0 | 0.0 |
| April | WD | 1.46 | 2.07 | 0.90 | 1.91 | 0.70 | 5.2 | 1.9 |
|  | WE | 1.28 | 13.40 | 0.72 | 2.29 | 0.83 | 28.3 | 10.3 |
| May | WD | 1.46 | 2.35 | 0.59 | 1.88 | 0.81 | 3.8 | 1.6 |
|  | WE | 1.28 | 30.40 | 0.58 | 3.00 | 0.82 | 67.7 | 18.5 |
| June | WD | 0.90 | 5.00 | 0.85 | 2.25 | 0.84 | 8.6 | 3.2 |
|  | WE | 1.20 | 16.67 | 0.83 | 2.12 | 0.53 | 35.2 | 8.8 |
| July | WD | 0.90 | 18.75 | 0.71 | 2.61 | 1.12 | 31.3 | 13.4 |
|  | WE | 1.20 | 35.00 | 0.50 | 3.25 | 1.75 | 68.3 | 36.8 |
| August | WD | 0.90 | 7.53 | 0.67 | 2.60 | 1.11 | 11.8 | 5.0 |
|  | WE | 1.20 | 34.33 | 0.61 | 3.33 | 1.15 | 0.0 | 28.9 |
| September | WD | 1.07 | 6.75 | 0.86 | 2.25 | 0.89 | 14.0 | 5.5 |
|  | WE | 1.21 | 28.33 | 0.38 | 2.95 | 1.25 | 38.4 | 16.3 |
| October | WD | 1.07 | 1.36 | 1.00 | 1.62 | 0.52 | 2.4 | 0.8 |
|  | WE | 1.21 | 2.40 | 0.69 | 2.00 | 0.50 | 4.0 | 1.0 |
| November | WD | 1.07 | 0.22 | 1.00 | 1.94 | 0.52 | 0.5 | 0.1 |
|  | WE | 1.21 | 0.00 | 1.00 | 2.47 | 0.50 | 0.0 | 0.0 |
| Annual | WD | 1.26 | 3.86 | 0.88 | 2.12 | 0.59 | 7.1 | 2.7 |
|  | WE | 1.59 | 13.83 | 0.75 | 2.25 | 0.68 | 22.0 | 10.2 |

Table A. 4 Section 2 pressure estimates for boat anglers in 1995 with intermediate calculations.

| STRATA |  | Correct. factor | $\begin{aligned} & \text { Mean } \\ & \text { boat } \\ & \text { trailers } \\ & \text { for } \\ & \text { the day } \\ & \hline \end{aligned}$ | \% of boats fishing | $\begin{gathered} \# \\ \text { angler/ } \\ \mathrm{g} \quad \text { boat } \\ \hline \end{gathered}$ | $\begin{gathered} \text { \# of } \\ \text { angler/ } \\ \text { boat } \\ \text { S.D. } \\ \hline \end{gathered}$ | Corrected mean angler | $\begin{gathered} \text { Corrected } \\ \text { x angler } \\ \text { sd } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| December | WD | 1.60 | 7.18 | 1.00 | 1.67 | 0.58 | 19.2 | 6.7 |
|  | WE | 2.67 | 12.00 | 1.00 | 1.83 | 0.40 | 58.6 | 12.8 |
| January | WD | 1.60 | 4.63 | 1.00 | 2.00 |  | 14.8 | 0.0 |
|  | WE | 2.67 | 8.00 | 1.00 | 1.67 | 0.58 | 35.7 | 12.4 |
| February | WD | 1.60 | 6.92 | 1.00 | 2.00 | 0.00 | 22.1 | 0.0 |
|  | WE | 2.67 | 18.75 | 1.00 | 2.00 |  | 100.1 | 0.0 |
| March | WD | 1.46 | 8.27 | 1.00 | 1.92 | 0.90 | 23.2 | 10.9 |
|  | WE | 1.28 | 15.67 | 1.00 | 2.75 | 1.29 | 55.2 | 25.9 |
| April | WD | 1.46 | 4.29 | 1.00 | 2.00 | 0.00 | 12.5 | 0.0 |
|  | WE | 1.28 | 21.00 | 0.78 | 2.57 | 1.13 | 53.9 | 23.7 |
| May | WD | 1.46 | 10.00 | 0.79 | 2.00 | 0.82 | 23.1 | 9.5 |
|  | WE | 1.28 | 32.00 | 0.64 | 2.75 | 0.96 | 72.1 | 25.2 |
| June | WD | 0.90 | 16.11 | 0.67 | 2.50 | 0.55 | 24.3 | 5.3 |
|  | WE | 1.20 | 30.50 | 0.67 | 2.00 | 0.00 | 49.0 | 0.0 |
| July | WD | 0.90 | 62.00 | 0.04 | 3.19 | 1.28 | 7.1 | 2.9 |
|  | WE | 1.20 | 115.00 | 0.06 | 3.27 | 1.03 | 27.1 | 8.5 |
| August | WD | 0.90 | 74.62 | 0.20 | 2.71 | 1.08 | 36.4 | 14.5 |
|  | WE | 1.20 | 149.50 | 0.28 | 3.21 | 1.19 | 161.2 | 59.8 |
| September | WD | 1.07 | 17.00 | 0.91 | 2.00 | 0.41 | 33.1 | 6.8 |
|  | WE | 1.21 | 37.80 | 1.00 | 2.25 | 0.50 | 102.9 | 22.9 |
| October | WD | 1.07 | 8.67 | 1.00 | 1.75 | 0.46 | 16.2 | 4.3 |
|  | WE | 1.21 | 21.25 | 0.78 | 1.83 | 0.41 | 36.7 | 8.2 |
| November | WD | 1.07 | 4.42 | 1.00 | 2.17 | 0.83 | 10.3 | 3.9 |
|  | WE | 1.21 | 2.33 | 1.00 | 2.04 | 0.40 | 5.8 | 1.1 |
| Annual | WD | 1.26 | 18.68 | 0.80 | 2.16 | 0.63 | 20.2 | 5.4 |
|  | WE | 1.59 | 38.65 | 0.77 | 2.35 | 0.72 | 63.2 | 16.7 |

Table A. 5 Section 3 pressure estimates in hours for boat anglers in 1995 with intermediate calculations.

| STRATA |  | Correct. <br> factor | Mean boat trailers for the day | \%of boats fishing | $\begin{gathered} \text { \# } \\ \text { angler/ } \\ \text { g } \end{gathered}$ | angler/ boat S.D. | $\begin{gathered} \text { Corrected } \\ \text { mean } \\ \text { angler } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Corrected } \\ \mathbf{x} \text { angler } \\ \text { sd } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| December | WD | 1.60 | 4.13 | 1.00 | 1.44 | 0.53 | 9.5 | 3.5 |
|  | WE | 2.67 | 2.00 | 1.00 | 2.00 | 0.00 | 10.7 | 0.0 |
| January | WD | 1.60 | 5.55 | 1.00 | 1.50 | 0.76 | 13.3 | 6.7 |
|  | WE | 2.67 | 14.50 | 1.00 | 2.00 |  | 77.4 | 0.0 |
| February | WD | 1.60 | 13.50 | 1.00 | 1.83 | 0.41 | 39.5 | 8.9 |
|  | WE | 2.67 | 39.00 | 1.00 | 2.00 | 0.71 | 208.3 | 73.9 |
| March | WD | 1.46 | 14.90 | 1.00 | 1.40 | 0.55 | 30.5 | 12.0 |
|  | WE | 1.28 | 23.43 | 1.00 | 2.17 | 1.03 | 65.1 | 30.9 |
| April | WD | 1.46 | 3.91 | 1.00 | 1.67 | 0.52 | 9.5 | 3.0 |
|  | WE | 1.28 | 73.57 | 0.78 | 1.53 | 0.52 | 112.4 | 38.2 |
| May | WD | 1.46 | 19.78 | 0.79 | 1.36 | 0.56 | 31.0 | 12.8 |
|  | WE | 1.28 | 80.75 | 0.64 | 1.36 | 0.49 | 90.0 | 32.4 |
| June | WD | 0.90 | 33.85 | 0.67 | 1.61 | 0.66 | 32.9 | 13.5 |
|  | WE | 1.20 | 82.25 | 0.67 | 1.89 | 0.33 | 125.0 | 21.8 |
| July | WD | 0.90 | 84.86 | 0.04 | 1.50 | 0.76 | 0.0 | 0.0 |
|  | WE | 1.20 | 160.40 | 0.06 | 1.91 | 0.54 | 22.1 | 6.2 |
| August | WD | 0.90 | 105.58 | 0.20 | 2.50 | 0.80 | 47.5 | 15.2 |
|  | WE | 1.20 | 113.00 | 0.28 | 1.90 |  | 72.1 | 0.0 |
| September | WD | 1.07 | 11.42 | 0.91 | 1.17 | 0.41 | 13.0 | 4.6 |
|  | WE | 1.21 | 99.13 | 0.50 | 1.67 | 0.58 | 100.2 | 34.8 |
| October | WD | 1.07 | 14.80 | 1.00 | 1.88 | 0.35 | 29.8 | 5.5 |
|  | WE | 1.21 | 17.43 | 0.78 | 1.88 | 0.35 | 30.9 | 5.8 |
| November | WD | 1.07 | 3.85 | 1.00 | 1.00 | 0.00 | 4.1 | 0.0 |
|  | WE | 1.21 | 6.00 | 1.00 | 1.77 |  | 12.9 | 0.0 |
| Annual | WD | 1.26 | 21.02 | 0.80 | 1.57 | 0.53 | 21.7 | 7.1 |
|  | WE | 1.59 | 59.29 | 0.73 | 1.84 | 0.51 | 77.2 | 20.3 |

Table A. 6 Section 1 angling pressure estimates (hrs) from December, 1994 to November, 1995 with intermediate calculations.


Table A. 6 Continued.

| Strata | Hours per day (naut) Hd | Days per month (cal) D s |  | Hours creeled per month n | Time correction factor NsIn | Angler hours per angler Ha | Mean anglers per day Xd | Mean anglers per month x s | anglers <br> per <br> day <br> Sd | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \\ \hline \end{gathered}$ | ```Pressure estimate per month PE``` | Variance of pressure estlmate per month VPE | $\begin{gathered} 95 \% \\ C \\ \text { per } \\ \text { mont } \mathrm{h} \\ \text { CI } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MARCH WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 11.97 | 23 | 275.31 | 43.00 | 6.40 | 0.68 | 0.5 | 11.5 | 1.0 | 23.5 | S0 | 3,524 | 83 |
| Boat | 11.97 | 23 | 275.31 | 43.00 | 6.40 | 5.21 | 4.6 | 105.8 | 1.2 | 21.6 | 3,529 | 4,811 | 98 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 11.97 | 8 | 95.76 | 37.00 | 2.59 | 3.00 | I. 0 | 0.0 | I. 0 | 8.0 | 0 | 166 | 18 |
| Boat | 11.97 | 8 | 95.76 | 37.00 | 2.59 | 6.63 | 17.0 | 136.0 | 0.0 | 0.0 | 2,334 | 0 | 0 |
| total | 11.97 | 31 | 371.07 | 80.00 |  |  | 23.1 | 253.3 | 3.2 | 59.1 | 5,913 | 8,567 | 199 |
| APRIL |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 13.68 | 20 | 273.60 | 71.50 | 3.83 | 2.60 | 1.0 | 20.0 | 1.8 | 35.4 | 199 | 4,795 | 91 |
| Boat | 13.68 | 20 | 273.60 | 71.50 | 3.83 | 6.20 | 5.2 | 104.0 | 1.9 | 38.0 | 2,467 | 5,526 | 104 |
| Weekend |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 13.68 | 10 | 136.80 | 29. 20 | 4.68 | 3.84 | 2.8 | 28.0 | 3.3 | 32.7 | 504 | 5,010 | 99 |
| Boat | 13.68 | 10 | 136.80 | 29.20 | 4.68 | 5.76 | 28.3 | 283.0 | 10.3 | 103.0 | 1,637 | 49,702 | 312 |
| total | 13.68 | 30 | 410.40 | 100.70 |  |  | 37.3 | 435.0 | 17.2 | 209.1 | 10,807 | 65,033 | 612 |
| M A Y |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.20 | 22 | 334.40 | 84.50 | 3.96 | 6.85 | 0.5 | 10.3 | 0.9 | 20.7 | 280 | 1,692 | 58 |
| Boat | 15.20 | 22 | 334.40 | 84.50 | 3.96 | 6.41 | 3.8 | 83.6 | 1.6 | 35.2 | 2,121 | 4,903 | 98 |
| weekend |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.20 | 9 | 136.80 | 26.00 | 5.26 | 0.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 15.20 | 9 | 136.80 | 26.00 | 5.26 | 5.65 | 67.7 | 609.3 | 18.5 | 166.5 | 18,113 | 145,862 | 535 |
| TOTAL | 15.20 | 31 | 471.20 | 110.50 |  |  | 72.0 | 703.2 | 21.0 | 222.4 | 20,514 | 152,457 | 690 |
| JUNE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 16.02 | 22 | 352.44 | 80.20 | 4.39 | 5.93 | 0.1 | 1.2 | 0.2 | 5.3 | 32 | 123 | 15 |
| Boat | 16.02 | 22 | 352.44 | 80.20 | 4.39 | 5.94 | 8.6 | 189.2 | 3.2 | 70.4 | 4,939 | 21,780 | 207 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 16.02 | 8 | 128.16 | 14.00 | 9.15 | 0.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 16.02 | 8 | 128.16 | 14.00 | 9.15 | 5.81 | 35.2 | 281.6 | 8.8 | 10.4 | 14,971 | 45,370 | 298 |
| TOTAL | 16.02 | 30 | 480.60 | 94.20 |  |  | 43.9 | 472.0 | 12.2 | 146.1 | 19,948 | 67,273 | 520 |

Table A. 6 Continued.

N

| STRATA | Hours per day (naut) Hd | Days per month (cal) D s | Hours per month N s | Hours creeled per month n | TI me correction factor NsIn | Angler <br> hours per angler Ha | $\begin{gathered} \text { Mean } \\ \text { anglers } \\ \text { per day } \\ \text { Xd } \end{gathered}$ | Mean anglers per month x s | ```\pm anglers per day S d``` | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { Ss } \end{gathered}$ | Pressure estimate per month PE | Variance of pressure estimate per month V PE | 95\% C.I. per month CI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| J U L Y |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.67 | 20 | 313.40 | 46.50 | 6.74 | 0.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 15.67 | 20 | 313.40 | 46.50 | 6.74 | 6.71 | 31.5 | 630.0 | 13.4 | 268.0 | 28,491 | 484,078 | 974 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.67 | 11 | 172.37 | 32.70 | 5.27 | 0.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 15.67 | 11 | 172.37 | 32.70 | 5.27 | 6.92 | 68.3 | 751.3 | 36.8 | 404.8 | 27,405 | 863,764 | 1,301 |
| T OTAL | 15.67 | 31 | 485.77 | 79.20 |  |  | 99.8 | 1381.3 | 50.2 | 672.8 | 55,896 | 1,347,842 | 2,275 |
| AUGUST |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 14.38 | 23 | 330.74 | 98.50 | 3.36 | 0.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Uoat | 14.38 | 23 | 330.74 | 98.50 | 3.36 | 5.93 | 11.8 | 271.4 | 5.0 | 115.0 | 5,404 | 44,406 | 295 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 14.38 | 8 | 115.04 | 30.00 | 3.83 | 12.00 | 1.0 | 8.0 | 2.5 | 19.6 | 368 | 1,473 | 54 |
| Boat | 14.38 | 8 | 115.04 | 30.00 | 3.83 | 6.31 | 0.0 | 0.0 | 28.9 | 231.2 | 0 | 204.976 | 634 |
| T OTAL | 14.38 | 31 | 445.78 | 128.50 |  |  | 12.8 | 279.4 | 36.4 | 365.8 | 5,772 | 250,856 | 983 |
| SEPTEMBER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 12.45 | 21 | 261.45 | 45.00 | 5.81 | 6.22 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 12.45 | 21 | 261.45 | 45.00 | 5.81 | 8.17 | 14.0 | 294.0 | 5.5 | 115.5 | 13,956 | 77,507 | 390 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 12.45 | 9 | 112.05 | 30.00 | 3.74 | 12.00 | 0.3 | 3.0 | 0.8 | 7.4 | 133 | 203 | 20 |
| Boat | 12.45 | 9 | 112.05 | 30.00 | 3.74 | 7.07 | 38.4 | 345.6 | 16.3 | 146.7 | 9,126 | 80,381 | 397 |
| T OTAL | 12.45 | 30 | 373.50 | 75.00 |  |  | 52.7 | 642.6 | 22.6 | 269.6 | 23,215 | 158,091 | 807 |
| OCTOBER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10.73 | 20 | 214.60 | 58.00 | 3.70 | 6.68 | 0.3 | 5.4 | 0.9 | 18.0 | 133 | 1,199 | 48 |
| Boat | 10.73 | 20 | 214.60 | 58.00 | 3.70 | 5.63 | 2.4 | 48.0 | 0.8 | 16.0 | 1,000 | 947 | 43 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10.73 | 11 | 118.03 | 15.00 | 7.87 | 6.68 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 10.73 | 11 | 118.03 | 15.00 | 7.87 | 6.71 | 4.0 | 44.0 | 1.0 | 11.0 | 2,323 | 952 | 43 |
| TOTAL | 10.73 | 31 | 332.63 | 73.00 |  |  | 6.7 | 97.4 | 2.7 | 45.0 | 3,456 | 3,098 | 135 |

Table A. 6 Continued.

| STRATA | $\begin{gathered} \text { Hours } \\ \text { per } \\ \text { day } \\ \text { (naut) } \\ \text { Hd } \\ \hline \end{gathered}$ | Days per month (cal) D s | Hours per month N s | Hours creeled per month n | $\begin{gathered} \text { Time } \\ \text { correction } \\ \text { factor } \\ \mathrm{Ns} / \mathrm{n} \\ \hline \end{gathered}$ | Angler hours per angler Ha | Mean anglers per day X d | Mean anglers per month $\times 5$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { day } \\ \text { sd } \\ \hline \end{gathered}$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \\ \hline \end{gathered}$ | Pressure estlmate per month PE | ```Varlance of pressure estlmate per month V PE``` | $\begin{gathered} 95 \% \\ \mathrm{C} .1 . \\ \text { per } \\ \text { mont } \mathrm{h} \\ \mathrm{Cl} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NOVEMBER <br> WEEKSDA ${ }^{\text {r }}$ " <br> Boat | $\begin{aligned} & 9.20 \\ & 9.20 \end{aligned}$ | 20 20 | $\begin{aligned} & 184.00 \\ & 184.00 \end{aligned}$ | $\begin{aligned} & 42.50 \\ & 42.50 \end{aligned}$ | $\begin{aligned} & 4.33 \\ & 4.33 \end{aligned}$ | $\begin{aligned} & 6.00 \\ & 6.90 \end{aligned}$ | $\begin{aligned} & 0.6 \\ & 0.5 \end{aligned}$ | $\begin{aligned} & 11.2 \\ & 10.0 \end{aligned}$ | $\begin{aligned} & 0.9 \\ & 0.1 \end{aligned}$ | $\begin{gathered} 17.6 \\ 2.0 \end{gathered}$ | $\begin{aligned} & 291 \\ & 299 \end{aligned}$ | $\begin{gathered} 1.341 \\ 17 \end{gathered}$ | $\begin{gathered} 51 \\ 6 \end{gathered}$ |
|  | $\begin{aligned} & 9.20 \\ & 9.20 \\ & 9.20 \end{aligned}$ | $\begin{array}{r} 10 \\ 10 \\ 30 \end{array}$ | $\begin{gathered} 92.00 \\ 92.00 \\ 276.00 \end{gathered}$ | $\begin{array}{r} 10.00 \\ 10.00 \\ 52.50 \end{array}$ | 9.20 9.20 | $\begin{aligned} & 6.68 \\ & 6.90 \end{aligned}$ | $\begin{array}{r} 2.5 \\ 0.0 \\ 3.6 \\ \hline \end{array}$ | $\begin{gathered} 25.0 \\ 0.0 \\ 46.2 \\ \hline \end{gathered}$ | $\begin{array}{r} 3.5 \\ 0.0 \\ 4.5 \\ \hline \end{array}$ | $\begin{gathered} 35.3 \\ 0.0 \\ 54.9 \\ \hline \end{gathered}$ | $\begin{gathered} 1.536 \\ 0 \\ 2,126 \\ \hline \end{gathered}$ | $\begin{gathered} 11,464 \\ 0 \\ 12,822 \end{gathered}$ | $\begin{gathered} 150 \\ 0 \\ 207 \\ \hline \end{gathered}$ |
| $A H^{\prime} A^{\text {A }}$ | 146.78 | 362.00 | 4441.59 | 1052.30 |  |  | 382.4 | 4,727.8 | 182.7 | 2,214.0 | 152,773 | 2,082,301 | 6,854 |

Table A. 7 Section 2 angling pressure estimates (hrs) from December, 1994 to November, 1995 with intermediate calculations.

|  | StRATA | Hours per day (naut) Hd | Days per month (cal) D s | Hours per month Ns | Hours creeled per month <br> n $\qquad$ | $\begin{gathered} \text { Time } \\ \text { correction } \\ \text { factor } \\ \mathrm{Ns} / \mathbf{n} \\ \hline \end{gathered}$ | Angler hours per angler Ha | Mean anglers per day Xd | Mean anglers per month x 5 | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { day } \\ \text { sd } \\ \hline \end{gathered}$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { mont } h \\ \text { ss } \\ \hline \end{gathered}$ | Pressure estimate per month PE | $\begin{gathered} \text { Varlance of } \\ \text { pressure } \\ \text { estimate } \\ \text { per month } \\ \text { VPE } \\ \hline \end{gathered}$ | $\begin{gathered} 95 \% \\ c .1 . \\ \text { per } \\ m 0 \text { nt h } \\ \text { CI } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DECEMBER WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 8.40 | 22 | 184.80 | 44.60 | 4.14 | 2.60 | 22.6 | 496.1 | 12.5 | 273.9 | 5,345 | 310.850 | 781 |
|  | $\begin{gathered} \text { Boat } \\ \text { WEEKEND } \end{gathered}$ | 8.40 | 22 | 184.80 | 44.60 | 4.14 | 4.00 | 19.2 | 422.4 | 6.1 | 147.4 | 1,001 | 90,025 | 420 |
|  | Shore | 8.40 | 9 | 75.60 | 10.80 | 1.00 | 4.29 | 36.3 | 326.1 | 6.8 | 61.2 | 9,811 | 26,218 | 227 |
|  | Boat | 8.40 | 9 | 75.60 | 10.80 | 1.00 | 4.00 | 58.6 | 527.4 | 12.8 | 115.2 | 14.767 | 92,897 | 427 |
|  | TOTAL | 8.40 | 31 | 260.40 | 55.40 |  |  | 136.7 | 1772.6 | 38.8 | 597.7 | 36,923 | 519,990 | 1,854 |
|  | JANUARY WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 8.83 | 19 | 167.77 | 34.70 | 4.83 | 4.24 | 10.9 | 206.5 | 8.3 | 157.9 | 4,234 | 120,530 | 486 |
| $\begin{aligned} & N \\ & N \end{aligned}$ | Boat | 8.83 | 19 | 167.77 | 34.70 | 4.83 | 4.50 | 14.8 | 281.2 | 0.0 | 0.0 | 6,118 | 0 | 0 |
|  | WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 8.83 |  | 79.47 | 23.20 | 3.43 | 5.89 | 19.0 | 171.0 | 15.9 | 143.3 | 3,450 | 10,321 | 371 |
|  | Boat | 8.83 | 9 | 19.47 | 23.20 | 3.43 | 5.11 | 35.1 | 321.3 | 12.4 | 111.6 | 6,350 | 42,662 | 289 |
|  | TOTAL | 8.83 | 28 | 247.24 | 57.90 |  |  | 80.4 | 980.0 | 36.6 | 412.8 | 20,152 | 233,513 | 1,146 |
|  | FEBRUARY WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 10.25 | 19 | 194.75 | 27.60 | 1.06 | 5.15 | 8.4 | 160.0 | 10.9 | 207.1 | 5,814 | 302,642 | 170 |
|  | WEEKEND | 10.25 | 19 | 194.75 | 27.60 | 1.06 | 5.38 | 22.1 | 419.9 | 0.0 | 0.0 | 15,940 | 0 | 0 |
|  | Shore | 10.25 | 9 | 92.25 | 7.70 | 11.98 | 1.00 | 30.3 | 272.3 | 22.2 | 199.1 | 3,262 | 471.832 | 968 |
|  | TOTAL Boat | 10.25 10.25 | 28 | 92.25 287250 | 7.70 35.30 | 11.98 | 6.42 | 100.1 | 900.9 | 0.0 | 0.0 | 69,293 | ${ }^{\circ}$ | ${ }^{0} 738$ |
|  | T OTAL | 10.25 | 28 | 287.00 | 35.30 |  |  | 160.9 | 1753.0 | 33.1 | 406.8 | 94,308 | 780,474 | 1,738 |

Table A. 7 Continued.

|  | STRATA | $\begin{gathered} \text { Hours } \\ \text { per } \\ \text { day } \\ \text { (naut) } \\ \text { Hd } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Days } \\ \text { per } \\ \text { month } \\ \text { (cal) } \\ \text { Ds } \end{gathered}$ | Hours per month N s | Hours creeled per month $\qquad$ | $\begin{gathered} \text { Time } \\ \text { correction } \\ \text { factor } \\ \text { Nsin } \\ \hline \end{gathered}$ | Angler hours per angler Ha | Mean anglers per day Xd | Mean anglers per month x 5 | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { day } \\ \text { sd } \\ \hline \end{gathered}$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Pressure } \\ \text { estlmate } \\ \text { per } \\ \text { month } \\ \text { PE } \end{gathered}$ | Variance of pressure estlmate per month VPE | $\begin{gathered} 95 \% \\ \text { C. } 1 . \\ \text { per } \\ \text { mont } \mathrm{h} \\ \mathrm{Cl} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| WEEKDAY | MARCH |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 11.97 | 23 | 275.31 | 24.90 | 11.06 | 4.22 | 6.9 | 157.8 | 7.3 | 167.2 | 7,362 | 309.134 | 778 |
|  | Boat | 11.97 | 23 | 275.31 | 24.90 | 11.06 | 5.05 | 23.2 | 533.6 | 10.9 | 250.7 | 29,794 | 694,914 | 1,167 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 11.97 | 8 | 95.76 | 26.50 | 3.61 | 4.79 | 14.7 | 0.0 | 8.2 | 65.7 |  | 15,589 | 175 |
|  | Boat | 11.97 | 8 | 95.76 | 26.50 | 3.61 | 4.78 | 55.2 | 441.6 | 25.9 | 207.2 | 7.6028 | 155,138 | 551 |
|  | TOTAL | 11.97 | 31 | 371.07 | 51.40 |  |  | 99.9 | 1133.0 | 52.3 | 690.8 | 44,784 | 1,174,775 | 2,672 |
|  | APRIL |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 13.68 | 20 | 273.60 | 16.90 | 16.19 | 4.46 | 2.3 | 45.8 | 1.9 | 37.8 | 3,307 | 23,132 | 213 |
|  | Boat | 13.68 | 20 | 273.60 | 16.90 | 16.19 | 5.03 | 12.5 | 250.0 | 0.0 | 0.0 | 20,358 | 0 | 0 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 13.68 | 10 | 136.80 | 18.00 | 7.60 | 4.46 | 27.7 | 276.7 | 15.6 | 155.7 | 9,379 | 184,243 | 601 |
| N | Boat | 13.68 | 10 | 136.80 | 18.00 | 7.60 | 4.70 | 53.9 | 539.0 | 23.7 | 237.0 | 19,253 | 426,884 | 915 |
|  | total | 13.68 | 30 | 410.40 | 34.90 |  |  | 96.4 | 1111.5 | 41.2 | 430.5 | 52,297 | 634,259 | 1,729 |
| WEEKDAY | M A Y |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 15.20 | 22 | 334.40 | 67.20 | 4.98 | 4.10 | 2.5 | 53.9 | 4.2 | 91.5 | 1,100 | 41,680 | 286 |
|  | Boat | 15.20 | 22 | 334.40 | 67.20 | 4.98 | 6.22 | 23.1 | 508.2 | 9.5 | 209.0 | 15,730 | 217,365 | 653 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 15.20 | 9 | 136.80 | 15.30 | 8.94 | 4.10 | 4.3 | 38.3 | 3.3 | 29.7 | 1,402 | 7,887 | 124 |
|  | Boat | 15.20 | 9 | 136.80 | 15.30 | 8.94 | 7.57 | 72.1 | 648.9 | 25.2 | 226.8 | 43,921 | 459,918 | 949 |
|  | TOTAL | 15. 20 | 31 | 471.20 | 82.50 |  |  | 101.9 | 1249.3 | 42.2 | 557.0 | 62,152 | 726,850 | 2,012 |
|  | J UNE |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 16.02 | 22 | 352.44 | 37.70 | 9.35 | 6.00 | I. 1 | 24.4 | 2.0 | 43.1 | 1,370 | 17,382 | 185 |
|  | Boat | 16.02 | 22 | 352.44 | 37.70 | 9.35 | 5.11 | 24.3 | 534.6 | 5.3 | 116.6 | 25,538 | 127,099 | 499 |
|  | Weekend |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | Shore | 16.02 | 8 | 128.16 | 9.50 | 13.49 | 6.00 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
|  | Boat | 16.02 | 8 | 128.16 | 9.50 | 13.49 | 3.88 | 49.0 | 392.0 | 0.0 | 0.0 | 20,519 | 0 | 0 |
|  | TOTAL | 16.02 | 30 | 480.60 | 47.20 |  |  | 74.4 | 951.0 | 7.3 | 159.7 | 47,427 | 144,481 | 684 |

Table A. 7 Continued.

| S TRATA | Hours per day (naut) Hd | Days per month (cal) D s | Hours per month N s | Hours creeled per month n | Time correction factor Ns/n | Angler hours per angler Ha | Mean anglers per day X d | Mean anglers per month x s | ```士 anglers per day S d``` | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \end{gathered}$ | Pressure estimate per month PE | Variance of pressure estimate per month V PE | $\begin{gathered} 95 \% \\ \text { C.I. } \\ \text { per } \\ \text { mont h } \\ \text { CI } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| JULY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.67 | 20 | 313.40 | 14.70 | 21.32 | 6.00 | 0.7 | 13.2 | 1.2 | 23.0 | 1,689 | 11,278 | 149 |
| Boat | 15.67 | 20 | 313.40 | 14.70 | 21.32 | 4.48 | 7.1 | 142.0 | 2.9 | 58.0 | 13,563 | 71,720 | 375 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.67 | 11 | 172.37 | 13.50 | 12.77 | 6.00 | 1.7 | 18.4 | 2.9 | 31.8 | 1,407 | 12,904 | 159 |
| Boat | 15.67 | 11 | 172.37 | 13.50 | 12.77 | 4.00 | 27.1 | 298.1 | 8.5 | 93.5 | 15,225 | 111.622 | 468 |
| T OTAL | 15.67 | 31 | 485.77 | 28.20 |  |  | 36.5 | 471.7 | 15.4 | 206.3 | 31,883 | 207,524 | 1,150 |
| A U G U S T |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 14.38 | 23 | 330.74 | 60.60 | 5.46 | 6.00 | 3.1 | 70.8 | 4.4 | 101.0 | 2,320 | 55.641 | 330 |
| Boat | 14.38 | 23 | 330.74 | 60.60 | 5.46 | 6.15 | 36.4 | 837.2 | 14.5 | 333.5 | 28,101 | 607,024 | 1.091 |
| Shore | 14.38 | 8 | 115.04 | 21.10 | 5.45 | 6.00 | 4.2 | 33.4 | 5.0 | 40.0 | 1,091 | 8,723 | 131 |
| Boat | 14.38 | 8 | 115.04 | 21.10 | 5.45 | 4.09 | 161.2 | 1289.6 | 59.8 | 478.4 | 28,757 | 1,247,811 | 1,564 |
| T 0 T A L | 14.38 | 31 | 445.78 | 81.70 |  |  | 204.9 | 2231.0 | 83.7 | 952.9 | 60,269 | 1,919,200 | 3,116 |
| SEPTEMBER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 12.45 | 21 | 261.45 | 52.10 | 5.02 | 4.20 | 0.7 | 15.3 | 1.6 | 32.8 | 323 | 5,386 | 103 |
| Boat | 12.45 | 21 | 261.45 | 52.10 | 5.02 | 3.99 | 33.1 | 695.1 | 6.8 | 142.8 | 13,918 | 102,331 | 448 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 12.45 | 9 | 112.05 | 18.30 | 6.12 | 3.00 | 5.8 | 52.2 | 11.4 | 102.3 | 959 | 64,116 | 354 |
| Boat | 12.45 | 9 | 112.05 | 18.30 | 6.12 | 3.75 | 102.9 | 926.1 | 22.9 | 206.1 | 21,264 | 260.086 | 714 |
| T OTAL | 12.45 | 30 | 373.50 | 70.40 |  |  | 142.5 | 1688.7 | 42.6 | 484.0 | 36,464 | 431,919 | 1,619 |
| OCTOBER |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10.73 | 20 | 214.60 | 30.60 | 7.01 | 4.20 | 0.0 | 0.0 | 0.0 | 0.0 | 0 | 0 | 0 |
| Boat | 10.73 | 20 | 214.60 | 30.60 | 7.01 | 4.22 | 16.2 | 324.0 | 4.3 | 86.0 | 9,589 | 51,869 | 319 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10.73 | 11 | 118.03 | 22.90 | 5.15 | 3.00 | 3.0 | 33.0 | 1.4 | 15.5 | 510 | 1,240 | 49 |
| Boat | 10.73 | 11 | 118.03 | 22.90 | 5.15 | 5.81 | 36.7 | 403.7 | 8.2 | 90.2 | 12,089 | 41.934 | 287 |
| TOTAL | 10.73 | 31 | 332.63 | 53.50 |  |  | 55.9 | 760.7 | 13.9 | 191.7 | 22,188 | 95,043 | 655 |

Table A. 7 Continued.

| STRATA | Hours per day (naut) Hd | $\begin{gathered} \text { Days } \\ \text { per } \\ \text { month } \\ \text { (cal) } \\ \text { Ds } \end{gathered}$ | Hours per month N s | Hours creeled per month $\qquad$ n | ```Time correctlon factor Ns/n``` | $\begin{gathered} \text { An g } \\ \text { hours } \\ \text { per } \\ \text { angler } \\ \text { Ha } \end{gathered}$ | $\begin{gathered} \text { Ier } \\ \text { Mean } \\ \text { anglers } \\ \text { per day } \\ X d \\ \hline \end{gathered}$ | Mean anglers per month x 5 | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { day } \\ \text { Sd } \\ \hline \end{gathered}$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \end{gathered}$ | ```Pressure estlmate per month PE``` | $\begin{gathered} \text { Variance of } \\ \text { pressure } \\ \text { estimate } \\ \text { per month } \\ \text { VPE } \end{gathered}$ | $\begin{gathered} 95 \% \\ \text { C.I. } \\ \text { per } \\ \text { mont h } \\ \text { CI } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| november WEEKSh ${ }^{\text {O }}$ Y Boat | 9.20 9.20 | $\begin{aligned} & 20 \\ & 20 \end{aligned}$ | $\begin{aligned} & 184.00 \\ & 184.00 \end{aligned}$ | $\begin{aligned} & 68.70 \\ & 68.70 \end{aligned}$ | $\begin{aligned} & 2.68 \\ & 2.68 \end{aligned}$ | $\begin{aligned} & 5.17 \\ & 4.11 \end{aligned}$ | 1.3 10.3 | 26.6 206.0 | 2.7 3.9 | $\begin{aligned} & 54.8 \\ & 78.0 \end{aligned}$ | 368 2,268 | $\begin{array}{r} 8,043 \\ 16,295 \end{array}$ | 126 179 |
| $\begin{aligned} & \text { WEEK E E NoD D } \\ & \text { TOTAL } \\ & \text { TOTAL } \end{aligned}$ | $\begin{aligned} & 9.20 \\ & 9.20 \\ & 9.20 \end{aligned}$ | 10 10 30 | 92.00 92.00 276.00 | $\begin{aligned} & 0.00 \\ & 0.00 \\ & 68.70 \end{aligned}$ | 5.00 5.00 | 3.50 4.30 | $\begin{gathered} 0.0 \\ 5.8 \\ 17.4 \\ \hline \end{gathered}$ | $\begin{array}{r} 0.0 \\ 58.0 \\ 290.6 \end{array}$ | $\begin{aligned} & 0.0 \\ & 1.1 \\ & 7.7 \end{aligned}$ | $\begin{gathered} 0.0 \\ 11.0 \\ 143.8 \end{gathered}$ | $\begin{gathered} 0 \\ 1,247 \\ 3,883 \end{gathered}$ | $\begin{array}{r} 0 \\ 605 \\ 24,943 \\ \hline \end{array}$ | $\begin{gathered} 0 \\ 34 \\ 339 \end{gathered}$ |
| $\begin{aligned} & \text { ANNUAL } \\ & \text { TOTAL } \end{aligned}$ | 146.8 | 362.00 | 4441.59 | 667.10 |  |  | 1,207.7 | 14393.1 | 414.7 | 5,234.0 | 512,731 | 6,892,970 | 18,713 |

Table A. 8 Section 3 angling pressure estimates (hrs) from December, 1994 to November, 1995 with intermediate calculations.

| STRATA | Hours per day (naut) $\qquad$ | Days per month (cal) D s | Hours per month $\mathrm{Ns}$ $\qquad$ | $\begin{gathered} \text { Hours } \\ \text { creeled } \\ \text { per } \\ \text { month } \\ \text { n } \\ \hline \end{gathered}$ | ```TI me correctlon factor Ns/n``` | Angler hours per angler Ha | $\begin{gathered} \text { Mean } \\ \text { anglers } \\ \text { per day } \\ X d \\ \hline \end{gathered}$ | Mean anglers per month x 5 |  | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \\ \hline \end{gathered}$ | Pressure estimate per month Y E | Variance of pressure estimate per month V PE | $\begin{gathered} 95 \% \\ \text { C. } 1 . \\ \text { per } \\ \text { mont } \mathrm{h} \\ \text { CI } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DECEMBER WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 8.40 | 22 | 184.80 | 50.30 | 3.67 | 3.00 | 2.0 | 44.0 | 2.3 | 51.5 | 485 | 9,737 | 138 |
| Boat | 8.40 | 22 | 184.80 | 50.30 | 3.67 | 4.28 | 9.5 | 209.0 | 3.5 | 77.0 | 3,286 | 21,783 | 207 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 8.40 | 9 | 75.60 | 16.80 | 4.50 | 3.00 | 1.3 | 12.0 | 1.6 | 14.7 | 162 | 968 | 44 |
| Boat | 8.40 | 9 | 75.60 | 16.80 | 4.50 | 7.00 | 10.7 | 96.3 | 0.0 | 0.0 | 3,033 | 0 | 0 |
| TOTAL | 8.40 | 31 | 260.40 | 67.10 |  |  | 23.5 | 361.3 | 7.5 | 143.2 | 6,966 | 32,488 | 388 |
| J ANUARY WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 8.83 | 19 | 167.77 | 43.20 | 3.88 | 6.00 | 2.1 | 39.9 | 2.2 | 41.2 | 930 | 6,602 | 114 |
| Boat | 8.83 | 19 | 167.77 | 43.20 | 3.88 | 4.95 | 13.3 | 252.7 | 6.7 | 127.3 | 4,858 | 62,934 | 351 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 8.83 | 9 | 79.47 | 26.20 | 3.03 | 4.92 | 7.7 | 69.0 | 4.5 | 40.5 | 1.030 | 4,975 | 99 |
| Boat | 8.83 | 9 | 79.47 | 26.20 | 3.03 | 4.95 | 77.4 | 696.6 | 0.0 | 0.0 | 10,459 | 0 | 0 |
| T OTAL | 8.83 | 28 | 247.24 | 69.40 |  |  | 100.5 | 1,058.2 | 13.4 | 209.0 | 17,277 | 74,511 | 564 |
| FEBRUARY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10.25 | 19 | 194.75 | 30.40 | 6.41 | 4.50 | 4.6 | 87.4 | 3.7 | 69.4 | 2,520 | 30,810 | 246 |
| Boat | 10.25 | 19 | 194.75 | 30.40 | 6.41 | 4.66 | 39.5 | 750.5 | 8.9 | 169.1 | 22.405 | 183,186 | 599 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 10. 25 | 9 | 92.25 | 20.00 | 4.61 | 4.00 | 9.7 | 87.0 | 6.7 | 59.9 | 1,606 | 16,572 | 180 |
| Boat | 10. 25 | 9 | 92.25 | 20.00 | 4.61 | 5.31 | 208.5 | 1876.5 | 73.9 | 665.1 | 45.960 | 2,040,376 | 2,000 |
| T OTAL | 10.25 | 28 | 287.00 | 50.40 |  |  | 262.3 | 2801.4 | 93.1 | 963.5 | 72,490 | 2,270,944 | 3,025 |

Table A. 8 Continued.

| STRATA | $\begin{gathered} \text { Hours } \\ \text { per } \\ \text { day } \\ \text { (naut) } \\ \text { Hd } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Days } \\ \text { per } \\ \text { mont } h \\ \text { (cal) } \\ \text { Ds } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Hours } \\ \text { per } \\ \text { month } \\ \text { Ns } \\ \hline \end{gathered}$ | $\qquad$ | $\begin{gathered} \text { Time } \\ \text { correctlon } \\ \text { factor } \\ \text { Nsin } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Angler } \\ \text { hours } \\ \text { per } \\ \text { angler } \\ \text { Ha } \end{gathered}$ | Mean anglers per day $\qquad$ | $\qquad$ |  | $\qquad$ | $\begin{gathered} \text { Pressure } \\ \text { estimate } \\ \text { per } \\ \text { month } \\ \text { PE } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Variance of } \\ \text { pressure } \\ \text { estimate } \\ \text { per month } \\ \text { VPE } \\ \hline \end{gathered}$ | $\begin{gathered} 95 \% \\ \text { C. } 1 . \\ \text { per } \\ \text { mont } \mathrm{h} \\ \text { CI } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MARCH |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 11.97 | 23 | 275.31 | 40.30 | 6.83 | 3.50 | 5.8 | 133.4 | 2.9 | 67.4 | 3,190 | 31,025 | 247 |
| Boat | 11.97 | 23 | 275.31 | 40.30 | 6.83 | 4.58 | 30.5 | 701.5 | 12.0 | 276.0 | 21.949 | 520,397 | 1,010 |
| Weekend |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 11.97 | 8 | 95.76 | 29.50 | 3.25 | 4.44 | 11.6 | 0.0 | 6.3 | 50.3 | 0 | 8,219 | 127 |
| Boat | 11.97 | 8 | 95.76 | 29.50 | 3.25 | 6.00 | 65.1 | 520.8 | 30.9 | 247.2 | 10,143 | 198.362 | 624 |
| TOTAL | 11.97 | 31 | 371.07 | 69.80 |  |  | 113.0 | 1,355.7 | 52.1 | 640.9 | 35,282 | 758,004 | 2,007 |
| APRIL |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 13.68 | 20 | 273.60 | 33.70 | 8.12 | 6.00 | 6.6 | 132.0 | 2.4 | 47.4 | 6,430 | 18,241 | 189 |
| Boat | 13.68 | 20 | 273.60 | 33.70 | 8.12 | 6.88 | 9.5 | 190.0 | 3.0 | 60.0 | 10,613 | 29,227 | 239 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 13.68 | 10 | 136.80 | 33.20 | 4.12 | 6.00 | 11.7 | 117.1 | 6.1 | 61.0 | 2,895 | 15.332 | 173 |
| Boat | 13.68 | 10 | 136.80 | 33.20 | 4.12 | 7.84 | 112.4 | 1124.0 | 38.2 | 382.0 | $36,310$ | 601,277 | $1,086$ |
| TOTAL | 13.68 | 30 | 410.40 | 66.90 |  |  | 140.2 | 1,563.1 | 49.7 | 550.4 | $56,248$ | 664,078 | $1,687$ |
| MAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 15.20 | 22 | 334.40 | 34.20 | 9.78 | 6.00 | 3.4 | 75.7 | 2.1 | 46.9 | 4,440 | 21,471 | 205 |
| Boat <br> WEEKEND | 15.20 | 22 | 334.40 | 34.20 | 9.78 | 7.17 | 31.0 | 682.0 | 12.8 | 281.6 | 47,813 | 775,364 | 1,233 |
| Shore | 15.20 | 9 | 136.80 | 19.00 | 7.20 | 6.00 | 13.8 | 123.8 | 12.9 | 116.0 | 5,346 | 96.900 | 436 |
| Boat | 15.20 | 9 | 136.80 | 19.00 | 7.20 | 6.36 | 90.0 | 810.0 | 32.4 | 291.6 | 37,092 | 612,220 | 1,095 |
| TOTAL | 15.20 | 31 | 471.20 | 53.20 |  |  | 138.2 | 1,691.4 | 60.2 | 736.1 | 94,690 | 1,505,954 | 2,969 |
| JUNE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| WEEKDAY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 16.02 | 22 | 352.44 | 29.70 | 11.87 | 7.08 | 5.0 | 110.0 | 5.3 | 116.4 | 9,242 | 160,726 | 561 |
| Boat | 16.02 | 22 | 352.44 | 29.70 | 11.87 | 9.20 | 32.9 | 723.8 | 13.5 | 297.0 | 79.020 | 1,046,747 | 1,432 |
| WEEKEND |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Shore | 16.02 | 8 | 128.16 | 11.00 | 11.65 | 7.08 | 2.5 | 20.0 | 1.7 | 13.8 | 1,650 | 2,232 | 66 |
| Boat | 16.02 | 8 | 128.16 | 11.00 | 11.65 | 7.43 | 125.0 | 1000.0 | 21.8 | 174.4 | 86,566 | 354,367 | 833 |
| TOTAL | 16.02 | 30 | 480.60 | 40.70 |  |  | 165.4 | 1,853.8 | 42.3 | 601.6 | 176,477 | 1,564,071 | 2,893 |

Table A. 8 Continued.


Table A. 8 Continued.

| STRATA | Hours per day (naut) Hd | Days per month (cal) D s | $\begin{gathered} \text { Hours } \\ \text { per } \\ \text { month } \\ \text { Ns } \\ \hline \end{gathered}$ | Hours creeled per month n | Time correction factor $\mathrm{Ns} / \mathrm{n}$ | Angler hours per angler Ha | Mean anglers per day X d | Mean anglers per month $\times 5$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { day } \\ \text { sd } \\ \hline \end{gathered}$ | $\begin{gathered} \pm \\ \text { anglers } \\ \text { per } \\ \text { month } \\ \text { ss } \end{gathered}$ | Pressure estimate per month PE | $\begin{gathered} \text { Variance of } \\ \text { pressure } \\ \text { estimate } \\ \text { per month } \\ \text { VPE } \\ \hline \end{gathered}$ | $\begin{gathered} 95 \% \\ \text { C.I. } \\ \text { per } \\ \text { mont h } \\ \text { Cl } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NOVEMBER WEEKDAY shore | 9.20 | 20 | 184.00 | 50.80 | 3.62 | 3.00 | 0.2 | 4.6 | 0.4 | 8.8 | 50 | 280 | 23 |
| Boat | 9.20 | 20 | 184.00 | 50.80 | 3.62 | 5.19 | 4.1 | 82.0 | 0.0 | 0.0 | 1,720 | 0 | 0 |
| $\begin{array}{r} \text { WEEKEND } \\ \text { Shore } \\ \text { Boat } \end{array}$ | 9.20 9.20 | 10 10 | $92.00$ | $15.20$ | $6.05$ | 0.00 5.79 | 0.0 | 0.0 | 0.0 | 0.0 | 4.521 | 0 | 0 |
| TOTAL | 9.20 | 30 | 276.00 | 66.00 |  |  | 17.2 | 215.6 | 0.4 | 8.8 | 6,290 | 280 | 23 |
| ANNUAL | 146.78 | 362.00 | 4,442 | 754.60 |  |  | 1,293.8 | 15,196 | 418.4 | 5,252 | 684,028 | 8,460,589 | 18,307 |

Table A. 9 Section 1 catch per unit effort (fish/hour) of the harvest (fish kept) in Lake Roosevelt from December, 1994 through November, 1995.

|  | Species | D E C | J A N | FEB | MAR | APR | MAY | JUN | JUL | A UG | S E P | 0 C T | N O V | Annual Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\underset{\sim}{N}$ | kokanee salmon | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | rainbow trout | 0.073 | 0.084 | 0.011 | 0.023 | 0.030 | 0.021 | 0.004 | 0.028 | 0.063 | 0.035 | 0.050 | 0.143 | 0.033 |
|  | walleye | 0.000 | 0.000 | 0.038 | 0.058 | 0.111 | 0.153 | 0.290 | 0.225 | 0.105 | 0.205 | 0.078 | 0.000 | 0.154 |
|  | smallmouth bass | . 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | sturgeon | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | other species* | 0.000 | 0.000 | 0.005 | 0.000 | 0.003 | 0.009 | 0.000 | 0.003 | 0.005 | 0.005 | 0.000 | 0.000 | 0.003 |
|  | Monthly Mean | 0.012 | 0.014 | 0.009 | 0.014 | 0.024 | 0.0310 | 0.049 | 0.043 | 0.029 | 0.041 | 0.021 | 0.024 | 0.032 |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 10 Section 2 catch per unit effort (fish/hour) of the harvest (fish kept) in Lake Roosevelt from December, 1994 through November, 1995.

|  | Species | D E C | J A N | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | 0 CT | N O V | Annual M e a n |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0,000 | 0.007 | 0.073 | 0.000 | 0.000 | 0.004 |
|  | rainbow trout | 0.060 | 0.153 | 0.099 | 0.052 | 0.111 | 0.030 | 0.000 | 0.000 | 0.057 | 0.036 | 0.063 | 0.040 | 0.064 |
|  | walleye | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 | 0.086 | 0.000 | 0.021 | 0.000 | 0.000 | 0.000 | 0.008 |
| N | smallmouth bass | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | sturgeon | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | other species* | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | Monthly Mean | 0.010 | 0.026 | 0.017 | 0.009 | 0.019 | 0.007 | 0.014 | 0.000 | 0.014 | 0.018 | 0.011 | 0.007 | 0.013 |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 11 Section 3 catch per unit effort (fish/hour) of the harvest (fish kept) in Lake Roosevelt from December, 1994 through November, 1995.

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 12 Section 1 catch per unit effort (fish/hour) of the total catch (harvest and release) in Lake Roosevelt from December, 1994 through November, 1995.


[^3]Table A. 13 Section 2 catch per unit effort (fish/hour) of the total catch (harvest and release) in Lake Roosevelt from December, 1994 through November, 1995.

|  | Species | DEC | JAN | FEB | MAR | APR | MAY | JUN | .JUL | AUG | SEP | OCT | NOV | Annual Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\underset{\infty}{\omega}$ | kokanee | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.007 | 0.073 | 0.000 | 0.000 | 0.004 |
|  | rainbow trout | 0.060 | 0.153 | 0.099 | 0.052 | 0.136 | 0.030 | 0.000 | 0.000 | 0.057 | 0.036 | 0.063 | 0.040 | 0.065 |
|  | walleye | 0.000 | 0.000 | 0.000 | 0.000 | 0.012 | 0.025 | 0.086 | 0.000 | 0.021 | 0.000 | 0.063 | 0.000 | 0.014 |
|  | smallmouth bass | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.014 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
|  | sturgeon | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | other species* | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | Monthly Mean | 0.010 | 0.026 | 0.017 | 0.009 | 0.025 | 0.009 | 0.017 | 0.000 | 0.014 | 0.018 | 0.0210 | 0.007 | 0.014 |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappic, chinook, bullhead, etc...

Table A. 14 Section 3 catch per unit effort (fish/hour) of the total catch (harvest and release) in Lake Roosevelt from December, 1994 through November, 1995.

| Species | DEC | JAN | FEB | MAR | APR | MAY | JUN | J U L | A U G | S E P | 0 CT | N 0 V | Annual Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| kokanee | 0.034 | 0.105 | 0.114 | 0.046 | 0.091 | 0.050 | 0.019 | 0.033 | 0.025 | 0.000 | 0.000 | 0.000 | 0.042 |
| rainbow trout | 0.479 | 0.315 | 0.079 | 0.194 | 0.091 | 0.079 | 0.081 | 0.025 | 0.097 | 0.352 | 0.510 | 0.386 | 0.145 |
| walleye | 0.000 | 0.013 | 0.000 | 0.004 | 0.000 | 0.002 | 0.017 | 0.035 | 0.013 | 0.000 | 0.000 | 0.000 | 0.010 |
| smallmouth bass | 0.000 | 0.000 | 0.000 | 0.000 | 0.020 | 0.187 | 0.141 | 0.115 | 0.034 | 0.742 | 0.000 | 0.000 | 0.100 |
| sturgeon | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| other species* | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Monthly Mean | 0.086 | 0.072 | 0.032 | 0.244 | 0.202 | 0.053 | 0.043 | 0.035 | 0.028 | 0.182 | 0.085 | 0.064 | 0.050 |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 15 Total monthly and annual harvest estimates with $\pm 95 \%$ confidence intervals from fish harvested by anglers on all sections of Lake Roosevelt from December, 1994 through November, 1995.

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 16 Monthly and annual harvest estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 1 of Lake Roosevelt from December, 1994 through November, 1995.


Table A. 17 Monthly and annual harvest estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 2 of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | J A N | F E B | M A R | APR | MAY | U UN | JUL | AUG | S $\mathbf{E}$ P | OCT | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 429 | 2,647 | 0 | 0 | 3,076 |
|  |  | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 22$ | f118 | $\pm 0$ | $\pm 0$ | $\pm 140$ |
|  | rainbow trout | 2,218 | 3,084 | 9,363 | 2,327 | 5,811 | 1,894 | 0 | 0 | 3,434 | 1,324 | 1,402 | 157 | 31,014 |
| $\begin{aligned} & N \\ & N \end{aligned}$ |  | flll | $\pm 175$ | $\pm 173$ | $\pm 139$ | f192 | $\pm 61$ | $\pm 0$ | $\pm 0$ | $\pm 178$ | $\pm 59$ | $\pm 41$ | $\pm 14$ | $\pm 1,143$ |
|  | walleye | 0 | 0 | 0 | 0 | 0 | 631 | 4,089 | 0 | 1,288 | 0 | 0 | 0 | 6,008 |
|  |  | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 20$ | $\pm 59$ | $\pm 0$ | $\pm 67$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 146$ |
|  | smallmouth bass | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ |
|  | sturgeon | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ |
|  | other species* | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ | $\pm 0$ |
|  | Monthly <br> Total | 2,218 | 3,084 | 9,363 | 2,327 | 5,811 | 2,525 | 4,089 | 0 | 5,151 | 3,971 | 1,402 | 157 | 40,098 |
|  |  | $\pm 111$ | $\pm 175$ | f173 | f 139 | $\pm 192$ | $\pm 82$ | $\pm 59$ | $\pm 0$ | $\pm 266$ | $\pm 176$ | $\pm 41$ | $\pm 14$ | $\pm 1,429$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 18 Monthly and annual harvest estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 3 of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | JAN | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | OCT | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | $\begin{aligned} & 238 \\ & \pm 13 \end{aligned}$ | $\begin{gathered} 1,814 \\ \pm 59 \end{gathered}$ | $\begin{array}{r} 8,281 \\ +346 \end{array}$ | $\begin{gathered} 1,635 \\ \pm 93 \end{gathered}$ | $\begin{array}{r} 5,137 \\ \pm 154 \end{array}$ | $\begin{aligned} & 4,705 \\ & \pm 148 \end{aligned}$ | $\begin{gathered} 3,302 \\ \pm 54 \end{gathered}$ | $\begin{array}{r} 737 \\ \pm 41 \end{array}$ | $\begin{gathered} 3,492 \\ \pm 43 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 29,102 \\ \mathbf{9 3 8} \end{gathered}$ |
|  | rainbow trout | $\begin{aligned} & 3,334 \\ & \pm 186 \end{aligned}$ | $\begin{aligned} & 5,442 \\ & \pm 178 \end{aligned}$ | $\begin{aligned} & 5,733 \\ & \pm 239 \end{aligned}$ | $\begin{aligned} & \mathbf{6}, 691 \\ & \pm 381 \end{aligned}$ | $\begin{gathered} 5,137 \\ \pm 0154 \end{gathered}$ | $\begin{aligned} & 7,485 \\ & \pm 235 \end{aligned}$ | $\begin{gathered} 12,839 \\ \pm 210 \end{gathered}$ | $\begin{array}{r} 567 \\ \pm 31 \end{array}$ | $\begin{aligned} & 13,384 \\ & \pm 167 \end{aligned}$ | $\begin{gathered} 11,598 \\ \pm 433 \end{gathered}$ | $\begin{gathered} 12,788 \\ \pm 282 \end{gathered}$ | $\begin{gathered} 2,430 \\ \pm 9 \end{gathered}$ | $\begin{aligned} & 87,428 \\ & \pm 2,505 \end{aligned}$ |
| $\begin{aligned} & N \\ & \pm \end{aligned}$ | walleye | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 227 \\ & \pm 7 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 214 \\ & \pm 7 \end{aligned}$ | $\begin{aligned} & 2,568 \\ & \pm 42 \end{aligned}$ | $\begin{array}{r} 680 \\ \pm 38 \end{array}$ | $\begin{aligned} & 1,746 \\ & \pm 22 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 5,434 \\ & \pm 116 \end{aligned}$ |
|  | smallmouth bass | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 855 \\ & \pm 27 \end{aligned}$ | $\begin{gathered} 4,402 \\ \pm 72 \end{gathered}$ | $\begin{aligned} & 227 \\ & \pm 13 \end{aligned}$ | $\begin{gathered} 4,073 \\ \pm 51 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 9,558 \\ & \pm \mathbf{1 6 2} \end{aligned}$ |
|  | sturgeon | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm \mathbf{0} \end{gathered}$ |
|  | other species* | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \mathbf{~} 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ |
|  | Monthly <br> Total | $\begin{aligned} & 3,572 \\ & \text { f199 } \end{aligned}$ | 7,482 $\pm 244$ | $\begin{gathered} 14,014 \\ \pm 585 \end{gathered}$ | $\begin{array}{r} 8,326 \\ \pm 474 \end{array}$ | $\begin{gathered} 10,274 \\ \pm \mathbf{3 0 8} \end{gathered}$ | $\begin{gathered} 13,258 \\ \pm \mathbf{4 1 6} \end{gathered}$ | $\begin{gathered} 23,111 \\ \mathbf{\pm} 79 \end{gathered}$ | $\begin{gathered} 2,211 \\ \pm 122 \end{gathered}$ | $\begin{gathered} 22,695 \\ \pm \mathbf{2 8 3} \end{gathered}$ | $\begin{gathered} 11,598 \\ \mathbf{4 3 3 3} \end{gathered}$ | $\begin{gathered} 12,788 \\ \pm \mathbf{2 8 2} \end{gathered}$ | $\begin{gathered} 2,430 \\ \pm 9 \end{gathered}$ | $\begin{array}{r} 131,759 \\ \mathbf{3}, 734 \end{array}$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 19 Total monthly and annual catch estimates $\pm 95 \%$ confidence intervals from all fish observed by creel clerks on all sections of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | J A N | FEB | MAR | APR | MAY | JUN | JUL | AUG | SEP | OCT | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | $\begin{aligned} & 238 \\ & \pm 13 \end{aligned}$ | $\begin{aligned} & 1,814 \\ & \pm 59 \end{aligned}$ | $\begin{aligned} & 8,281 \\ & \pm 346 \end{aligned}$ | $\begin{aligned} & 1,635 \\ & \pm 93 \end{aligned}$ | $\begin{aligned} & 5,137 \\ & \pm 154 \end{aligned}$ | $\begin{aligned} & 4,705 \\ & \pm 148 \end{aligned}$ | $\begin{gathered} 3,302 \\ \pm 54 \end{gathered}$ | $\begin{aligned} & 912 \\ & \pm 48 \end{aligned}$ | $\begin{gathered} 3,921 \\ \pm 65 \end{gathered}$ | $\begin{aligned} & 2,647 \\ & \text { fl1 } \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 32,353 \\ & \pm \mathbf{1}, 085 \end{aligned}$ |
|  | rainbow trout | $\begin{aligned} & 5,700 \\ & \pm 303 \end{aligned}$ | $\begin{aligned} & 8,660 \\ & \pm 365 \end{aligned}$ | $\begin{aligned} & 15,291 \\ & \text { f198 } \end{aligned}$ | $\begin{aligned} & 9,304 \\ & \text { f. } 525 \end{aligned}$ | $\begin{gathered} 12,566 \\ \mathbf{~} 408 \end{gathered}$ | $\begin{aligned} & 9,903 \\ & \pm 314 \end{aligned}$ | $\begin{aligned} & 14,380 \\ & \text { f237 } \end{aligned}$ | $\begin{gathered} 2,139 \\ \pm 95 \end{gathered}$ | $\begin{aligned} & 17,181 \\ & \pm 407 \end{aligned}$ | $\begin{gathered} 13,736 \\ \pm 520 \end{gathered}$ | $\begin{gathered} 14,387 \\ \pm 34 \end{gathered}$ | $\begin{gathered} 2,891 \\ \pm 53 \end{gathered}$ | $\begin{array}{r} 125,958 \\ \pm \mathbf{3 , 9 7 8} \end{array}$ |
| $N$ | walleye | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 227 \\ & \pm 7 \end{aligned}$ | $\begin{aligned} & 100 \\ & \pm 13 \end{aligned}$ | $\begin{gathered} 640 \\ \pm 24 \end{gathered}$ | $\begin{aligned} & 2,462 \\ & \pm 124 \end{aligned}$ | $\begin{aligned} & 8,953 \\ & \pm 292 \end{aligned}$ | $\begin{gathered} 21,054 \\ \pm 473 \end{gathered}$ | $\begin{aligned} & 27,694 \\ & \text { il. } 139 \end{aligned}$ | $\begin{aligned} & 3,895 \\ & \text { f236 } \end{aligned}$ | $\begin{aligned} & 6,858 \\ & \pm 238 \end{aligned}$ | $\begin{aligned} & 1,796 \\ & \pm 56 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 73,667 \\ & \pm \mathbf{2 , 6 0 9} \end{aligned}$ |
|  | smallmouth bass | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 1,101 \\ & \pm 33 \end{aligned}$ | $\begin{aligned} & 17,749 \\ & \pm 557 \end{aligned}$ | $\begin{gathered} 25,626 \\ \pm 419 \end{gathered}$ | $\begin{aligned} & 2,608 \\ & \pm 144 \end{aligned}$ | $\begin{gathered} 4,655 \\ \pm 58 \end{gathered}$ | $\begin{gathered} 24,486 \\ \pm 914 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 76,224 \\ & \mathbf{+ 2 , 1 2 5} \end{aligned}$ |
|  | sturgeon | $\begin{gathered} 0 \\ \mathbf{\pm 0} \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{array}{r} 13 \\ \pm 2 \end{array}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 13 \\ \pm 2 \end{gathered}$ |
|  | other species* | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 8 \\ \pm 1 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{array}{r} 36 \\ \pm 2 \\ \hline \end{array}$ | $\begin{aligned} & 262 \\ & \pm 9 \\ & \hline \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 175 \\ & \pm 7 \\ & \hline \end{aligned}$ | $\begin{array}{r} 27 \\ \pm 5 \\ \hline \end{array}$ | $\begin{aligned} & 116 \\ & \pm 4 \\ & \hline \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{array}{r} 624 \\ \mathbf{+ 2 8} \\ \hline \end{array}$ |
|  | Monthly <br> Total | $\begin{array}{r} 5,938 \\ \mathbf{\pm 3 1 6} \end{array}$ | $\begin{gathered} 10,700 \\ \mathbf{4 3 1} \end{gathered}$ | 23,502 $\pm 774$ | $\begin{gathered} 11,569 \\ \pm 651 \end{gathered}$ | $\begin{gathered} 21,301 \\ \pm \mathbf{7 2 0} \end{gathered}$ | $\begin{aligned} & 41,571 \\ & \pm \mathbf{1}, \mathbf{3 2 4} \end{aligned}$ | $\begin{array}{r} 64,361 \\ \pm \mathbf{1 , 1 8 3} \end{array}$ | $\begin{aligned} & 33,526 \\ & \pm \mathbf{1}, \mathbf{4 3 3} \end{aligned}$ | $\begin{gathered} 29,693 \\ \pm 741 \end{gathered}$ | $\begin{aligned} & 47,843 \\ & \pm 1,795 \end{aligned}$ | $\begin{gathered} 16,184 \\ \pm \mathbf{3 8 8} \end{gathered}$ | $\begin{array}{r} 2,891 \\ \pm \mathbf{5 3} \end{array}$ | $\begin{array}{r} 309,079 \\ \pm \mathbf{9 , 8 4 1} \end{array}$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 20 Monthly and annual catch estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 1 of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | J A N | FED | M A R | APR | MAY | JUN | JUL | AUG | SEP | OC' | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 175 \\ & \pm 7 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 175 \\ & \pm 7 \end{aligned}$ |
|  | rainbow trout | $\begin{aligned} & 148 \\ & \pm 6 \end{aligned}$ | $\begin{gathered} 134 \\ \pm 12 \end{gathered}$ | $\begin{array}{r} 17 \\ \pm 2 \end{array}$ | $\begin{aligned} & 138 \\ & \pm 5 \end{aligned}$ | $\begin{aligned} & 327 \\ & \pm 19 \end{aligned}$ | $\begin{aligned} & 524 \\ & \pm 18 \end{aligned}$ | $\begin{aligned} & 73 \\ & \pm 2 \end{aligned}$ | $\begin{aligned} & 1,572 \\ & \pm 64 \end{aligned}$ | $\begin{aligned} & 363 \\ & \pm 62 \end{aligned}$ | $\begin{array}{r} 814 \\ \pm 28 \end{array}$ | $\begin{aligned} & 197 \\ & \pm 8 \end{aligned}$ | $\begin{array}{r} 304 \\ \pm 30 \end{array}$ | $\begin{array}{r} 4,609 \\ \pm 255 \end{array}$ |
| $\stackrel{N}{N}$ | walleye | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{array}{r} 100 \\ \pm 13 \end{array}$ | $\begin{aligned} & 481 \\ & \pm 16 \end{aligned}$ | $\begin{aligned} & 1,816 \\ & \pm 103 \end{aligned}$ | $\begin{aligned} & 7,161 \\ & \pm 241 \end{aligned}$ | $\begin{gathered} 14,030 \\ \text { f366 } \end{gathered}$ | $\begin{aligned} & 26,900 \\ & \pm 1,095 \end{aligned}$ | $\begin{gathered} 861 \\ \pm 147 \end{gathered}$ | $\begin{aligned} & 6,858 \\ & \text { f238 } \end{aligned}$ | $\begin{array}{r} 394 \\ \pm 15 \end{array}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 58,602 \\ & \pm 2,234 \end{aligned}$ |
|  | smallmouth bass | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ |
|  | sturgeon | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{array}{r} 13 \\ \pm 2 \end{array}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 13 \\ \pm 2 \end{gathered}$ |
|  | other species* | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 8 \\ \pm 1 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{array}{r} 36 \\ \pm 2 \\ \hline \end{array}$ | $\begin{aligned} & 262 \\ & \pm 9 \\ & \hline \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 175 \\ & \pm 7 \end{aligned}$ | $\begin{array}{r} 27 \\ +5 \\ \hline \end{array}$ | $\begin{aligned} & 116 \\ & \pm 4 \\ & \hline \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 624 \\ & \pm \mathbf{2 8} \\ & \hline \end{aligned}$ |
|  | Monthly Total | 148 $\pm 6$ | 134 $\pm 12$ | 125 $\pm 16$ | 619 $\pm 21$ | 2,179 f123 | 7,947 $\pm 267$ | 14,103 $\pm 368$ | 28,821 $\pm 1,173$ | 1,265 $\pm 215$ | 7,788 f271 | 592 $\pm 23$ | 304 $\pm \mathbf{3 0}$ | $\begin{aligned} & 64,024 \\ & \pm 2,526 \end{aligned}$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 21 Monthly and annual catch estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 2 of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | JAN | FEB | M A R | APR | MAY | JUN | JUL | AUG | SEP | OCT | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 429 \\ & \pm 22 \end{aligned}$ | $\begin{aligned} & 2,647 \\ & \pm 118 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{array}{r} 3,076 \\ \pm 140 \end{array}$ |
|  | rainbow trout | $\begin{aligned} & 2,218 \\ & \pm 111 \end{aligned}$ | $\begin{aligned} & 3,084 \\ & \pm 175 \end{aligned}$ | $\begin{gathered} 9,363 \\ \mathrm{f} 173 \end{gathered}$ | $\begin{array}{r} 2,327 \\ \pm 139 \end{array}$ | $\begin{aligned} & 7,102 \\ & +735 \end{aligned}$ | $\begin{aligned} & 1,894 \\ & \pm 61 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 3,434 \\ & \mathrm{f} 178 \end{aligned}$ | $\begin{aligned} & 1,324 \\ & \pm 59 \end{aligned}$ | $\begin{aligned} & 1,402 \\ & \pm 41 \end{aligned}$ | $\begin{array}{r} 157 \\ \pm 14 \end{array}$ | $\begin{aligned} & 32,305 \\ & \pm \mathbf{1 , 1 8 6} \end{aligned}$ |
| $\begin{aligned} & N \\ & \text { N } \end{aligned}$ | walleye | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 646 \\ & \pm 21 \end{aligned}$ | $\begin{aligned} & 1,578 \\ & \pm 51 \end{aligned}$ | $\begin{gathered} 4,089 \\ \pm 59 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 1,288 \\ & \pm 67 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 1,402 \\ & \pm 41 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 9,002 \\ & \text { f239 } \end{aligned}$ |
|  | smallmouth bass | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 681 \\ & \pm 10 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 681 \\ & \pm 10 \end{aligned}$ |
|  | sturgeon | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ |
|  | other species* | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \mathbf{+ 0} \\ \hline \end{gathered}$ |
|  | Monthly <br> Total | $\begin{gathered} 2,218 \\ \pm 111 \end{gathered}$ | 3,084 $\pm 175$ | $\begin{array}{r} 9,363 \\ \mathrm{f} 173 \end{array}$ | $\begin{array}{r} 2,327 \\ \pm 139 \end{array}$ | $\begin{array}{r} 7,748 \\ \pm 256 \end{array}$ | $\begin{array}{r} 3,472 \\ \pm 112 \end{array}$ | $\begin{array}{r} 4,770 \\ \pm 69 \end{array}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 5,151 \\ & \pm 266 \end{aligned}$ | $\begin{aligned} & 3,971 \\ & \pm 176 \end{aligned}$ | $\begin{gathered} 2,804 \\ \mathbf{8 8 3} \end{gathered}$ | $\begin{aligned} & 157 \\ & \pm 14 \end{aligned}$ | $\begin{aligned} & 45,065 \\ & \pm 1,575 \end{aligned}$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table A. 22 Monthly and annual catch estimates $\pm 95 \%$ confidence intervals for all fish species surveyed in Section 3 of Lake Roosevelt from December, 1994 through November, 1995.

|  | SPECIES | DEC | JAN | FE] | M A R | APR | MAY | JUN | JUL | AUG | s EP | OCT | NOV | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kokanee | $\begin{array}{r} 238 \\ \pm 13 \end{array}$ | $\begin{gathered} 1,814 \\ \pm 59 \end{gathered}$ | $\begin{aligned} & 8,281 \\ & \text { f346 } \end{aligned}$ | $\begin{gathered} 1,635 \\ \pm 93 \end{gathered}$ | $\begin{aligned} & 5,137 \\ & \pm 154 \end{aligned}$ | $\begin{aligned} & 4,705 \\ & \pm 148 \end{aligned}$ | $\begin{gathered} 3,302 \\ \pm 54 \end{gathered}$ | $\begin{array}{r} 737 \\ \pm 41 \end{array}$ | $\begin{gathered} 3,492 \\ \pm 43 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 29,102 \\ \text { f9 } 38 \end{gathered}$ |
|  | rainbow trout | $\begin{aligned} & 3,334 \\ & \pm 186 \end{aligned}$ | $\begin{aligned} & 5,442 \\ & \pm 178 \end{aligned}$ | $\begin{aligned} & 5,733 \\ & \text { f239 } \end{aligned}$ | $\begin{aligned} & 6,839 \\ & \pm 389 \end{aligned}$ | $\begin{gathered} 5,137 \\ \pm 0154 \end{gathered}$ | $\begin{aligned} & 7,485 \\ & \pm 235 \end{aligned}$ | $\begin{gathered} 14,307 \\ \pm 235 \end{gathered}$ | $\begin{array}{r} 567 \\ \pm 31 \end{array}$ | $\begin{aligned} & 13,384 \\ & \pm 167 \end{aligned}$ | $\begin{gathered} \text { I } 1,598 \\ \pm 433 \end{gathered}$ | $\begin{gathered} 12,788 \\ \pm 282 \end{gathered}$ | $\begin{gathered} 2,430 \\ \mathbf{\pm 9} \end{gathered}$ | $\begin{aligned} & 89,044 \\ & \pm 2,537 \end{aligned}$ |
| $\begin{aligned} & N \\ & \Delta \end{aligned}$ | walleye | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 227 \\ & \pm 7 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 149 \\ & \pm 8 \end{aligned}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 214 \\ & \pm 7 \end{aligned}$ | $\begin{gathered} 2,568 \\ \pm 42 \end{gathered}$ | $\begin{aligned} & 794 \\ & \pm 44 \end{aligned}$ | $\begin{gathered} 1,746 \\ \pm 22 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 6,063 \\ & \pm \mathbf{1 3 6} \end{aligned}$ |
|  | smallmouth bass | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 1,101 \\ & \pm 33 \end{aligned}$ | $\begin{gathered} 17,749 \\ \pm 557 \end{gathered}$ | $\begin{gathered} 24,945 \\ \pm 409 \end{gathered}$ | $\begin{aligned} & 2,608 \\ & \pm 144 \end{aligned}$ | $\begin{gathered} 4,655 \\ \pm 58 \end{gathered}$ | $\begin{gathered} 24,486 \\ \pm 914 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{aligned} & 75,543 \\ & \pm \mathbf{2 , 1 1 5} \end{aligned}$ |
|  | sturgeon | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \end{gathered}$ |
|  | other species* | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ | $\begin{gathered} 0 \\ \pm 0 \\ \hline \end{gathered}$ |
|  | Monthly <br> Total | 3,572 f 199 | 7,482 $\pm \mathbf{2 4 4}$ | $\begin{gathered} 14,01 \\ 4 \\ \pm 585 \end{gathered}$ | 8,623 $\pm 491$ | 11,374 $\pm \mathbf{3 4 1}$ | 30,152 $\pm 945$ | 45,488 $\pm 746$ | 4,705 $\pm 260$ | 23,277 f290 | 36,084 $\pm 1,348$ | 12,788 $\mathbf{+ 2 8 2}$ | 2,430 $\pm 9$ | $\begin{array}{r} 199,990 \\ \pm 5,740 \end{array}$ |

*Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

Table B.l Annual electrofishing results for 1995 split by month including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).

| Effort (hrs) | May |  |  | July |  |  | October |  |  | $\frac{\text { Total }}{117.9}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | \% | CPUE | No. |  | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 83 | 2 | 1.43 | 83 | <1 | 0.70 |
| brook trout | 2 | <1 | 0.06 | 0 | 0 | 0.00 | 19 | <1 | 0.33 | 21 | $<1$ | 0.18 |
| brown bullhead | 4 | <1 | 0.11 | 2 | <1 | 0.08 | 0 | 0 | 0.00 | 6 | <1 | 0.05 |
| brown trout | 0 | 0 | 0.00 | 10 | cl | 0.41 | 27 | <1 | 0.47 | 37 | <1 | 0.31 |
| bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 0 | 0.02 | 1 | <1 | 0.01 |
| burbot | 0 | 0 | 0.00 | 15 | <1 | 0.61 | 16 | <1 | 0.28 | 31 | <1 | 0.26 |
| carp | 126 | 4 | 3.56 | 49 | 2 | 2.00 | 31 | <1 | 0.53 | 206 | 2 | 1.75 |
| chinook salmon | 0 | <1 | 0.00 | 0 | <1 | 0.00 | 2 | <1 | 0.03 | 2 | <1 | 0.02 |
| Cottus spp. | 181 | 6 | 5.11 | 84 | 4 | 3.43 | 293 | 8 | 5.05 | 558 | 6 | 4.73 |
| crappie | 9 | <1 | 0.25 | 0 | 0 | 0.00 | 1 | c 1 | 0.02 | 10 | <1 | 0.08 |
| kokanee salmon | 95 | 3 | 2.68 | 82 | 4 | 3.35 | 1,829 | 48 | 31.53 | 2,006 | 22 | 17.01 |
| lake whitefish | 31 | 1 | 0.88 | 0 | 0 | 0.00 | 9 | <1 | 0.16 | 40 | <1 | 0.34 |
| largescale sucker | 1,046 | 36 | 29.55 | 1,057 | 47 | 43.14 | 577 | 15 | 9.95 | 2,680 | 30 | 22.73 |
| longnose sucker | 12 | <1 | 0.34 | 1 | <1 | 0.04 | 12 | $<1$ | 0.21 | 25 | <1 | 0.21 |
| mountain whitefish | 4 | <1 | 0.11 | 1 | <1 | 0.04 | 5 | <1 | 0.09 | 10 | <1 | 0.08 |
| rainbow trout | 144 | 5 | 4.07 | 110 | 5 | 4.49 | 204 | 5 | 3.52 | 458 | 5 | 3.88 |
| redside shiner | 1 | <1 | 0.03 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | <1 | 0.01 |
| smallmouth bass | 506 | 17 | 14.29 | 416 | 18 | 16.98 | 17 | <1 | 0.29 | 939 | 10 | 7.96 |
| squawfish | 113 | 4 | 3.19 | 72 | 3 | 2.94 | 28 | <1 | 0.48 | 213 | 2 | 1.81 |
| Catostomus spp. | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 6 | <1 | 0.10 | 6 | <1 | 0.05 |
| walleye | 557 | 19 | 15.73 | 275 | 12 | 11.22 | 198 | 5 | 3.41 | 1,030 | 11 | 8.74 |
| yellow perch | 93 | 3 | 2.63 | 97 | 4 | 3.96 | 453 | 12 | 7.81 | 643 | 7 | 5.45 |
| Totals | 2,938 |  | 82.99 | 2,271 |  | 92.69 | 3,811 |  | 65.71 | 9,006 |  | 76.39 |

Table B. 2 May electrofishing results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).

|  | Kettle Falls |  |  | Gifford |  |  | Hunters |  |  | Porcupine Bay |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Effort (hrs) <br> Species | No. | 2.2 $\%$ | CPUE | No. | $5.2$ | CPUE | No. | $3.6$ | CPUE |  | $\begin{aligned} & 4,1 \\ & 0 \end{aligned}$ | CPUE |
| Specidgelip sucker | No. | 0 | CPUE | No. | 0 | CPUE | $\xrightarrow{\text { No. }}$ | 0 | - 0.00 | No. | 0 | CPO |
| brook trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | <1 | 0.28 | 1 | <1 | 0.25 |
| brown bullhead | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brown trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 2 | <1 | 0.49 |
| bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| burbot | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| carp | 1 | <1 | 0.45 | 4 | 3 | 0.11 | 7 | 6 | 1.94 | 6 | 3 | 1.48 |
| chinook salmon | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| Cottus spp. | 2 | 3 | 0.90 | 1 | <1 | 0.19 | 1 | <1 | 0.28 | 5 | 3 | 1.23 |
| crappie | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| kokaneesal mon | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| lake whitefish | 1 | 10 | 3.13 | 1 | <1 | 0.19 | 3 | 3 | 0.83 | 0 | 0 | 0.00 |
| largescale sucker | 6 | 9 | 2.69 | 22 | 15 | 4.22 | 9 | 8 | 2.49 | 5 | 3 | 1.23 |
| longnose sucker | 1 | <1 | 0.45 | 4 | 3 | 0.11 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 | 2 | <1 | 0.38 | 1 | <1 | 0.28 | 0 | 0 | 0.00 |
| rainbow trout | 4 | 6 | 1.79 | 12 | 8 | 2.30 | 3 | 3 | 0.83 | 4 | 2 | 0.98 |
| redside shiner | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| smallmouth bass | 0 | 0 | 0.00 | 2 | <1 | 0.38 | 1 | $<1$ | 0.28 | 40 | 20 | 9.84 |
| squawfish | 2 | 3 | 0.90 | 16 | 11 | 3.07 | 1 | <1 | 0.28 | 3 | 2 | 0.74 |
| Catostomus spp. | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| walleye | 44 | 66 | 19.70 | 75 | 52 | 14.38 | 76 | 65 | 21.01 | 87 | 44 | 21.39 |
| yellow perch | 0 | 0 | 0.00 | 4 | 3 | 0.11 | 14 | 12 | 3.87 | 44 | 22 | 10.82 |
| TOTALS | 67 |  | 30.00 | 143 |  | 27.41 | 117 |  | 32.35 | 197 |  | 48.44 |

Table B. 2 Continued.

|  | Effort (hrs) | Little Falls |  |  | Seven Bays |  |  | Keller Ferry |  |  | Sanpoil |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 3.5 |  |  | 3.8 |  |  | 2.4 |  |  | 3.8 |  |  |
|  | Species | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
|  | bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | brook trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | brown bullhead | 1 | <1 | 0.29 | 0 | 0 | 0.00 | 2 | 2 | 0.83 | 0 | 0 | 0.00 |
|  | brown trout | 12 | 8 | 3.48 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | burbot | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | carp | 0 | 0 | 0.00 | 1 | <1 | 0.27 | 1 | <1 | 0.41 | 15 | 15 | 3.95 |
|  | chinook salmon | 0 | 0 | 0.00 | 1 | <1 | 0.27 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | Cottus spp. | 4 | 3 | 1.16 | 4 | 4 | 1.07 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| N | crappie | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | kokanee salmon | 92 | 58 | 26.67 | 1 | <1 | 0.27 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | lake whitefish | 0 | 0 | 0.00 | 12 | 13 | 3.20 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | largescale sucker | 8 | 5 | 2.32 | 0 | 0 | 0.00 |  | 2 | 0.83 | 7 | 7 | 1.84 |
|  | longnose sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | mountain whitefish | 1 | <1 | 0.29 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | rainbow |  | cl | 0.29 | 4 | 4 | 1.07 | 17 | 20 | 7.03 | 28 | 27 | 7.37 |
|  | redside shiner | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | <1 | 0.26 |
|  | smallmouth bass | 6 | 4 | 1.74 | 18 | 19 | 4.80 | 51 | 59 | 21.10 | 29 | 28 | 7.63 |
|  | squawfish | 1 | <1 | 0.29 | 2 | 2 | - 0.53 | 2 | 2 | 0.83 | 7 | 7 | 1.84 |
|  | Catostomus spp. | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | walleye | 33 | 21 | 9.57 | 49 | 52 | 13.07 | 11 | 13 | 4.55 | 16 | 16 | 4.21 |
|  | yellow perch | 0 | 0 | 0.00 | 1 | <1 | 0.27 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | TOTALS | 159 |  | 46.09 | 94 |  | 25.07 | 86 |  | 35.59 | 103 |  | 27.11 |

Table B. 2 Continued.

|  | Spring Canvon |  |  |
| :--- | ---: | ---: | ---: |
| Effort (hrs) | 6.8 |  |  |
| Species | No. | $\boldsymbol{\%}$ | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 |
| brook trout | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| brown bullhead | 1 | $<1$ | 0.15 |
| brown trout | 0 | 0 | 0.00 |
| bull trout | 0 | 0 | 0.00 |
| burbot | 0 | 0 | 0.00 |
| carp | 13 | 5 | 1.91 |
| chinook salmon | 0 | 0 | 0.00 |
| Cottus spp. | 3 | $<1$ | 0.44 |
| crappie | 0 | 0 | 0.00 |
| kokanee salmon | 2 | $<1$ | 0.29 |
| lake whitefish | 0 | 0 | 0.00 |
| largescale sucker | 9 | 4 | 1.32 |
| longnose sucker | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 |
| rainbow | 44 | 18 | 6.47 |
| redside shiner | 0 | 0 | 0.00 |
| smallmouth bass | 137 | 55 | 20.15 |
| squawfish | 1 | 0 | 0.15 |
| Catostomus spp. | 0 | 0 | 0.00 |
| walleye | 37 | 15 | 5.44 |
| yellow perch | 2 | $<1$ | 0.29 |
| TOTALS | 249 |  | 36.62 |

Table B. 3 July electrofishing results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).

| Effort (hrs)Species | Kettle Falls |  |  | Gifford |  |  | Hunters |  |  | Porcunine Bay |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2.3 |  |  | 3.0 |  |  | 3.0 |  |  | 2.8 |  |
|  | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brook trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brown bullhead | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brown trout | 1 | 2 | 0.43 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| burbot | 2 | 4 | 0.86 | 3 | 7 | 0.99 | 2 | 3 | 0.67 | 0 | 0 | 0.00 |
| carp | 0 | 0 | 0.00 | 3 | 7 | 0.99 | 4 | 5 | 1.35 | 1 | <1 | 0.35 |
| chinook salmon | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| Cottus spp. | 0 | 0 | 0.00 | 3 | 7 | 0.99 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| crappie | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| kokanee salmon | 10 | 18 | 4.29 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| lake whitefish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| largescale sucker | 0 | 0 | 0.00 | 6 | 15 | 1.99 | 7 | 9 | 2.36 | 6 | 7 | 2.12 |
| longnose sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 | 1 | 2 | 0.33 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| rainbow trout | 1 | 2 | 0.43 | 2 | 5 | 0.66 | 3 | 4 | 1.01 | 6 | 7 | 2.12 |
| redside shiner | 0 | 0 | 0.00 | 0 | 0 | 0.00 . | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| smallmouth bass | 3 | 5 | 1.29 | 5 | 12 | 1.66 | 3 | 4 | 1.01 | 22 | 27 | 7.76 |
| squawfish | 8 | 14 | 3.43 | 2 | 5 | 0.66 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| Catostomus spp. | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| walleye | 30 | 53 | 12.86 | 15 | 37 | 4.97 | 35 | 47 | 11.80 | 30 | 37 | 10.59 |
| yellow perch | 2 | 4 | 0.86 | 1 | 2 | 0.33 | 21 | 28 | 7.08 | 16 | 20 | 5.65 |
| TOTALS | 57 |  | 24.43 | 41 |  | 13.59 | 75 |  | 25.28 | 81 |  | 28.59 |

Table B. 3 Continued.

| Effort (hrs) | Little Falls |  |  | Seven Bays |  |  | Keller Ferry |  |  | Sanpoil R. |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 2.5 |  |  | 2.8 |  |  |  |  |  | 4.2 |  |
| Species | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| tr-idgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brook trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brown bullhead | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 2 | 0.50 | 0 | 0 | 0.00 |
| brown trout | 9 | 17 | 3.67 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| burbot | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 2 | <1 | 0.48 |
| carp | 0 | 0 | 0.00 | 4 | 3 | 1.41 | 0 | 0 | 0.00 | 1 | cl | 0.24 |
| chinook salmon | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| Cottus spp. | 1 | 2 | 0.41 | 1 | <1 | 0.35 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| crappie | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| kokanee salmon | 11 | 21 | 4.49 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| lake whitefish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| largescale sucker | 8 | 15 | 3.27 | 5 | 4 | 1.76 | 4 | 6 | 2.00 | 9 | 7 | 2.15 |
| longnose sucker | 1 | 2 | 0.41 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| rainbow trout | 4 | 8 | 1.63 | 35 | 27 | 12.35 | 3 | 5 | 1.50 | 28 | 20 | 6.69 |
| redside shiner | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| smallmouth bass | 7 | 13 | 2.86 | 39 | 30 | 13.76 | 26 | 42 | 13.00 | 48 | 35 | 11.47 |
| squawfish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 7 | 5 | 1.67 |
| Catostomus spp. | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| walleye | 10 | 19 | 4.08 | 42 | 32 | 14.82 | 28 | 45 | 14.00 | 30 | 22 | 7.17 |
| yellow perch | 1 | 2 | 0.41 | 6 | 5 | 2.12 | 0 | 0 | 0.00 | 12 | 9 | 2.87 |
| TOTALS | 52 |  | 21.22 | 132 |  | 46.59 | 62 |  | 31.00 | 137 |  | 32.75 |

## Table B. 3 Continued.

|  | Sbring Canvon |  |  |
| :--- | ---: | ---: | :---: |
| Effort (hrs) | $\mathbf{1 . 9}$ |  |  |
| Species | No. | \% | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 |
| brook trout | $\mathbf{0}$ | $\mathbf{0}$ | 0.00 |
| brown bullhead | $\mathbf{0}$ | $\mathbf{0}$ | 0.00 |
| brown trout | $\mathbf{0}$ | $\mathbf{0}$ | 0.00 |
| bull trout | $\mathbf{0}$ | $\mathbf{0}$ | 0.00 |
| burbot | $\mathbf{0}$ | $\mathbf{0}$ | 0.00 |
| carp | 5 | 7 | 2.68 |
| chinook salmon | 0 | 0 | 0.00 |
| Cottus spp. | 0 | 0 | 0.00 |
| crappie | 0 | 0 | 0.00 |
| kokanee salmon | 0 | 0 | 0.00 |
| lake whitefish | 0 | 0 | 0.00 |
| largescale sucker | 4 | 6 | 2.14 |
| longnose sucker | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 |
| rainbow trout | 3 | 4 | 1.61 |
| redside shiner | 0 | 0 | 0.00 |
| smallmou th bass | 47 | 70 | 25.18 |
| squawfish | 1 | $<1$ | 0.54 |
| Catostomus spp. | 0 | 0 | 0.00 |
| walleye | 6 | 9 | 3.21 |
| yellow perch | 1 | $<1$ | 0.54 |
| TOTALS | 67 |  | 35.89 |

Table B. 4 October electrofishing results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).


Table B. 4 Continued.


Table B. 4 Continued.

|  | SpringCanvon |  |  |
| :--- | ---: | ---: | ---: |
| Effort (hrs) | $\mathbf{1 . 1}$ |  |  |
| Species | No. | \% | CPUE |
| bridgelip sucker | $\mathbf{2}$ | 29 | 1.82 |
| brook trout | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| brown bullhead | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| brown trout | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| bull trout | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| burbot | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| carp | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| chinook salmon | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| Cottus spp. | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| crappie | $\mathbf{0}$ | $\mathbf{0}$ | $\mathbf{0 . 0 0}$ |
| kokanee salmon | 1 | 14 | 0.91 |
| lake whitefish | 0 | 0 | 0.00 |
| largescale sucker | 0 | 0 | 0.00 |
| longnose sucker | 0 | 0 | 0.00 |
| mountain whitefish | 0 | 0 | 0.00 |
| rainbow | 3 | 43 | 2.73 |
| redside shiner | 0 | 0 | 0.00 |
| smallmouth bass | 0 | 0 | 0.00 |
| squawfish | 0 | 0 | 0.00 |
| Catostomus spp. | 0 | 0 | 0.00 |
| walleye | 0 | 0 | 0.00 |
| yellow perch | 1 | 14 | 0.91 |
| TOTALS | 7 |  | 6.36 |

Table B.5. Annual gillnet set results for 1995 split by sampling period including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).

| Effort (min) | $\frac{\text { May }}{704}$ |  |  | $\frac{\text { July }}{730}$ |  |  | $\frac{\text { October }}{547}$ |  |  | $\frac{\text { Total }}{1,981}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 33 | 9 | 0.05 | 33 | 3 | 0.05 |
| brown trout | 0 | 0 | 0.00 | 1 | <1 | 0.00 | 0 | 0 | 0.00 | 1 | cl | co. 01 |
| burbot | 20 | 8 | 0.03 | 12 | 3 | 0.02 | 14 | 4 | 0.02 | 46 | 4 | 0.07 |
| carp | 11 | 4 | 0.02 | 0 | 0 | 0.00 | 9 | 2 | 0.01 | 20 | 2 | 0.03 |
| kokanee salmon | 5 | 2 | 0.01 | 33 | 8 | 0.05 | 9 | 2 | 0.01 | 47 | 5 | 0.07 |
| lake whitefish | 131 | 51 | 0.19 | 223 | 55 | 0.32 | 128 | 34 | 0.18 | 482 | 46 | 0.68 |
| largescale sucker | 20 | 8 | 0.03 | 5 | 1 | 0.01 | 16 | 4 | 0.02 | 41 | 4 | 0.06 |
| longnose sucker | 4 | 2 | 0.01 | 14 | 3 | 0.02 | 1 | <1 | 0.00 | 19 | 2 | 0.03 |
| mountain whitefish | 0 | 0 | 0.00 | 1 | <1 | 0.00 | 0 | 0 | 0.00 | 1 | <1 | co. 01 |
| rainbow trout | 4 | 2 | 0.01 | 39 | 10 | 0.06 | 1 | <1 | 0.00 | 44 | 4 | 0.06 |
| smallmouth bass | 2 | <1 | 0.00 | 2 | <1 | 0.00 | 26 | 7 | 0.04 | 30 | 3 | 0.04 |
| squawfish | 5 | 2 | 0.01 | 5 | 1 | 0.01 | 10 | 3 | 0.01 | 20 | 2 | 0.03 |
| tench | 1 | <1 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | cl | $<0.01$ |
| walleye | 48 | 19 | 0.07 | 52 | 13 | 0.07 | 82 | 22 | 0.12 | 182 | 18 | 0.26 |
| yellow perch | 7 | 3 | 0.01 | 19 | 5 | 0.03 | 45 | 12 | 0.06 | 71 | 7 | 0.10 |
| TOTALS | 258 |  | 0.37 | 406 |  | 0.58 | 374 |  | 0.53 | 1038 |  | 1.47 |

Table B. 6 May gillnet results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).


## Table B. 6 Continued.

|  | Effort (hrs) | $\frac{\text { Seven Bays }}{190.6}$ |  |  | $\frac{\text { Keller Ferrv }}{52.9}$ |  |  | $\frac{\text { Sanpoil R. }}{54.2}$ |  |  | Spring Canyon |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 79.6 |  |  |  |  |
|  | Species | No. | \% | CPUE |  |  |  | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| N | bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | brown trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | burbot | 5 | 17 | 0.03 | 0 | 0 | 0.00 | 6 | 12 | 0.11 | 2 | 29 | 0.03 |
|  | carp | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 10 | 19 | 0.18 | 0 | 0 | 0.00 |
|  | kokanee salmon | 2 | 7 | 0.01 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 3 | 43 | 0.04 |
|  | lake whitefish | 15 | 50 | 0.08 | 1 | 50 | 0.02 | 19 | 37 | 0.35 | 0 | 0 | 0.00 |
|  | largescale sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 2 | 29 | 0.03 |
|  | longnose sucker | 1 | 3 | 0.01 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | mountain whitefish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | rainbow trout | 0 | 0 | 0.00 | 1 | 50 | 0.02 | 2 | 4 | 0.04 | 0 | 0 | 0.00 |
|  | smallmouth bass | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 2 | 0.02 | 0 | 0 | 0.00 |
|  | squawfish | 2 | 7 | 0.01 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | tench | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 2 | 0.02 | 0 | 0 | 0.00 |
|  | walleye | 5 | 17 | 0.03 | 0 | 0 | 0.00 | 8 | 15 | 0.15 | 0 | 0 | 0.00 |
|  | yellow perch | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 5 | 10 | 0.09 | 0 | 0 | 0.00 |
|  | TOTALS | 30 |  | 0.16 | 2 |  | 0.04 | 52 |  | 0.96 | 7 |  | 009. |

Table B. 7 July gillnet results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).

| Effort (hrs) | Kettle Falls |  |  | Cifford |  |  | Hunters |  |  | Porcupine Bay |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 94 |  |  | 119 |  |  | 79 |  |  | 78 |  |  |
| Species | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
| bridgelip sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| brown trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 2 | 0.01 |
| burbot | 2 | 3 | 0.02 | 3 | 4 | 0.03 | 4 | 6 | 0.05 | 1 | 2 | 0.01 |
| carp | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| kokanee salmon | 15 | 26 | 0.16 | 2 | 3 | 0.02 | 0 | 0 | 0.00 | 2 | 3 | 0.03 |
| lake whitefish | 24 | 41 | 0.26 | 52 | 77 | 0.44 | 53 | 76 | 0.68 | 42 | 72 | 0.54 |
| largescale sucker | 4 | 7 | 0.04 | 2 | 3 | 0.02 | 0 | 0 | 0.00 | 2 | 3 | 0.03 |
| longnose sucker | 2 | 3 | 0.02 | 2 | 3 | 0.02 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| mountain whitefi | sh 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| rainbow trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 1 | 0.01 | 2 | 3 | 0.03 |
| smallmouth bass | 1 | 2 | 0.01 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| squawfish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| tench | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| walleye | 8 | 14 | 0.09 | 6 | 9 | 0.05 | 4 | 6 | 0.05 | 8 | 14 | 0.10 |
| yellow perch | 2 | 3 | 0.02 | 1 | 2 | 0.01 | 8 | 11 | 0.10 | 0 | 0 | 0.00 |
| TOTALS | 58 |  | 0.62 | 68 |  | 0.57 | 70 |  | 0.89 | 58 |  | 0.74 |

## Table B. 7 Continued.



Table B. 8 October gillnet results for 1995 split by location including number of fish collected (No.), relative abundance (\%) and catch per unit effort (CPUE) based on time (hrs).


Table B. 8 Continued.

|  |  |  | ven |  |  | Iler | errv |  | anpo |  | Spr | ng C | nvon |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort (hrs) |  | 59 |  |  | 56 |  |  | 82 |  |  | 55 |  |
|  | Species | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE | No. | \% | CPUE |
|  | bridgelip sucker | 0 | 0 | 0.00 | I | 4 | 0.02 | 0 | 0 | 0.00 | 1 | 14 | 0.02 |
|  | brown trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | bull trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | burbot | 5 | 16 | 0.09 | 1 | 4 | 0.02 | 2 | 3 | 0.02 | 0 | 0 | 0.00 |
|  | carp | 2 | 7 | 0.03 | 1 | 4 | 0.02 | 3 | 4 | 0.04 | 1 | 14 | 0.02 |
|  | kokanee salmon | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | lake whitefish | 21 | 68 | 0.36 | 1 | 4 | 0.02 | 12 | 16 | 0.15 | 0 | 0 | 0.00 |
|  | largescale sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 1 | 0.01 | 0 | 0 | 0.00 |
|  | longnose sucker | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| N | mountain whitefish | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
| $\cdots$ | rainbow trout | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 1 | 1 | 0.01 | 0 | 0 | 0.00 |
|  | smallmouth bass | 0 | 0 | 0.00 | 14 | 50 | 0.25 | 8 | 11 | 0.10 | 4 | 57 | 0.07 |
|  | squawfish | 0 | 0 | 0.00 | 1 | 4 | 0.02 | 2 | 3 | 0.02 | 0 | 0 | 0.00 |
|  | tench | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 | 0 | 0 | 0.00 |
|  | walleye | 2 | 7 | 0.03 |  | 32 | 0.16 | 22 | 30 | 0.27 | 1 | 14 | 0.02 |
|  | yellow perch | 1 | 3 ' | 0.02 | 0 | 0 | 0.00 | 23 | 31 | 0.28 | 0 | 0 | 0.00 |
|  | TOTALS | 31 |  | 0.53 | 28 |  | 0.50 | 74 |  | 0.91 | 7 |  | 0.13 |

## APPENDIX C

Feeding Habits

Table C. 1 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for all kokanee ( $\mathrm{n}=50$ ) sampled in 1995.

| PREY ITEM | $\%$ by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Unidentified fish | 0.00 | 8.71 | 2.00 | 3.06 |
| Cladocera |  |  |  |  |
| Daphnia spp. | 97.41 | 80.33 | 68.00 | 70.21 |
| Eucopepoda |  |  |  |  |
| L. Ashlandi | 0.14 | 0.03 | 2.00 | 0.62 |
| Copepoda | 0.02 | 0.05 | 2.00 | 0.59 |
| Diptera |  |  |  |  |
| Chironimade pupa | 2.25 | 9.59 | 34.00 | 13.10 |
| Chironomidae larvae | 0.10 | 0.48 | 18.00 | 5.31 |
| Simuliidae larvae | 0.01 | 0.00 | 2.00 | 0.57 |
| Trichoptera |  |  |  |  |
| Hydrop yschidae | 0.00 | 0.22 | 2.00 | 0.64 |
| Hemiptera |  |  |  |  |
| Terrestrial 0.00 |  |  |  |  |
| Insects | 0.07 | 0.57 | 18.00 | 5.33 |

Table C. 2 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $1+$ year old kokanee $(\mathrm{n}=2)$ sampled in 1995.

| PREY ITEM | \% by <br> Number | \% by <br> Weight | Frequency of <br> Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| Cladocera <br> Daphnia spp. | 63.16 | 16.67 | 50.00 | 43.27 |
| Diptera <br> Chironimade pupa | 36.84 | 83.33 | 50.00 | 56.73 |

Table C. 3 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $2+$ year old kokanee $(\mathrm{n}=22)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera Daphnia spp. | 69.61 | 21.89 | 31.82 | 30.48 |
| Eucopepoda L. Ashland Copepoda | $\begin{aligned} & 1.65 \\ & 0.27 \end{aligned}$ | 0.19 0.37 | 4.55 4.55 | $\begin{aligned} & 1.58 \\ & 1.28 \end{aligned}$ |
| Diptera Chironimade pupa Chironomidae larvae Simuliidae larvae | 26.24 1.20 0.09 | 68.18 3.45 0.03 | 68.18 40.91 4.55 | 40.19 11.26 1.15 |
| Trichoptera Hydropyschidae | 0.04 | 1.62 | 4.55 | 1.54 |
| Hemiptera Corixidae Terrestrial | 0.04 0.85 | 0.13 | 4.55 | 1.17 |
| Insects | 0.85 | 4.14 | 40.91 | 11.35 |

Table C. 4 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $3+$ year old kokanee $(\mathrm{n}=26)$ sampled in 1995.

| PREY ITEM | \% by <br> Number | \% by <br> Weight | Frequency of <br> Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| Osteichthyes <br> Unidentified fish | 0.00 | 10.12 | 3.85 | 4.54 |
| Cladocera <br> Daphnia spp. | 99.98 | 89.85 | 100.00 | 94.19 |
| Diptera <br> Chironimade pupa | 0.02 | 0.03 | 3.85 | 1.26 |

Table C. 5 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for all rainbow trout $(\mathbf{n}=90)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Cyprinidae | 0.11 | 6.62 | 2.22 | 2.30 |
| Percidae | 0.87 | 51.81 | 1.11 | 13.83 |
| Unidentified fish | 0.02 | 27.50 | 2.22 | 7.65 |
| Fish eggs | 2.18 | 0.34 | 1.11 | 0.93 |
| Cladocera |  |  |  |  |
| L. kindtii | 2.43 | 0.19 | 4.44 | 1.82 |
| Daphnia spp. | 85.39 | 3.96 | 53.3 | 36.69 |
| Eucopepoda |  |  |  |  |
| E. nevadensis | 0.01 | 0.00 | 1.11 | 0.29 |
| L. ashlandi | 0.01 | 0.00 | 1.11 | 0.29 |
| Basommatophora 0.00 0.02 |  |  |  |  |
| Planorbidae | 0.01 | 0.02 | 1.11 | 0.29 |
| Physidae | 0.01 | 0.88 | 1.11 | 0.51 |
| Diptera |  |  |  |  |
| Chironomidae pupa | 6.20 | 1.70 | 33.33 | 10.60 |
| Chironomidae larvae | 0.28 | 0.07 | 16.67 | 4.37 |
| Sciomyzidae | 0.01 | 0.02 | 1.11 | 0.29 |
| Trichoptera |  |  |  |  |
| Hydropyschidae | 0.01 | 0.00 | 1.11 | 0.29 |
| Hydroptilidae | 0.02 | 0.10 | 2.22 | 0.60 |
| Lepidostomatidae | 0.01 | 0.01 | 2.22 | 0.58 |
| Hemiptera 0.22 |  |  |  |  |
| Corixidae | 0.17 | 0.09 | 6.67 | 1.78 |
| Plecoptera |  |  |  |  |
| Nemouridae | 0.01 | 0.01 | 1.11 | 0.29 |
| Perlodidae | 0.01 | 0.02 | 1.11 | 0.29 |
| Ephemeroptera 0.00 |  |  |  |  |
| Baetidae | 0.02 | 0.00 | 1.11 | 0.29 |
| Leptophlebiidae | 0.35 | 0.10 | 2.22 | 0.69 |
| Odonata |  |  |  |  |
| Anisoptera | 0.04 | 0.02 | 4.44 | 1.16 |
| Zygoptera | 0.11 | 0.05 | 1.11 | 0.33 |
| Coleoptera |  |  |  |  |
| Elmidae | 0.01 | 0.02 | 1.11 | 0.29 |
| Oligochaeta 0.02 |  |  |  |  |
| Lumbriculidae | 0.05 | 3.03 | 5.56 | 2.22 |
| Hydrachnellae |  |  |  |  |
| Hydracharina Terrestrial | 0.07 | 0.01 | 3.33 | 0.88 |
| Terrestrial |  |  |  |  |
| Other | 0.03 | 0.03 | 34.11 | 0.30 |

Table C. 6 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $0+$ year old rainbow trout $(\mathrm{n}=5)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Cladocera |  |  |  |  |
| Daphnia spp. | 68.73 | 27.95 | 40.00 | 24.41 |
| Diptera |  |  |  |  |
| Chironomidae pupa | 7.75 | 3.84 | 60.00 | 12.78 |
| Chironomidae larvae | 0.52 | 0.17 | 40.00 | 7.27 |
| Trichoptera |  |  |  |  |
| Lepidostomatidae | 0.26 | 0.11 | 20.00 | 3.64 |
| Hemiptera |  |  |  |  |
| Corixidae | 1.55 | 3.84 | 40.00 | 8.11 |
| Ephemeroptera 0.11 |  |  |  |  |
| Baetidae | 0.78 | 0.40 | 20.00 | 3.78 |
| Leptophlebiidae | 12.66 | 10.78 | 20.00 | 7.76 |
| Odonata |  |  |  |  |
| Anisoptera | 0.78 | 1.64 | 40.00 | 7.57 |
| Terrestrial |  |  |  |  |
| Insects | 6.98 | 51.27 | 80.00 | 24.69 |

Table C. 7 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $1+$ year old rainbow trout $(\mathrm{n}=31)$ sampled in 1995.

|  | \% by <br> Pumber | \% by <br> Weight | Frequency <br> of | Occurrence |
| :--- | :--- | :--- | :--- | :--- | IRI

Table C. 8 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $2+$ year old rainbow trout $(\mathrm{n}=21)$ sampled in 1995.

|  | \% by <br> Number | \% by <br> Weight | Frequency <br> of | Occurrence |
| :--- | :---: | :---: | :---: | :---: | IRI | PREY ITEM | 10.78 | 16.28 | 4.76 | 9.03 |
| :--- | :---: | :---: | :---: | :---: |
| Osteichthyes <br> Fish eggs | 7.19 | 5.64 | 9.52 | 6.34 |
| Cladocera <br> L. kindtii | 80.23 | 45.42 | 71.43 | 55.93 |
| Daphnia spp . | 0.39 | 1.59 | 23.81 | 7.32 |
| D i t e r a |  |  |  |  |
| Chironomidae pupa <br> Chironomidae larvae | 0.12 | 0.19 | 9.52 | 2.79 |
| Hydrachnellae <br> $\quad$ Hydracharina | 0.27 | 0.19 | 4.76 | 1.48 |
| Terrestrial <br> Insects | 1.02 | 30.70 | 28.57 | 17.11 |

Table C. 9 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $3+$ year old rainbow trout $(\mathrm{n}=24)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Percidae | 2.99 | 61.84 | 4.17 | 16.56 |
| Unidentified fish | 0.04 | 32.51 | 4.17 | 8.81 |
| Cladocera |  |  |  |  |
| L. kindtii | 1.53 | 0.05 | 4.17 | 1.38 |
| Daphnia spp. | 91.98 | 1.26 | 50.00 | 34.38 |
| Diptera |  |  |  |  |
| Chironomidae pupa | 1.72 | 0.12 | 37.50 | 9.44 |
| Chironomidae larvae | 0.29 | 0.03 | 16.67 | 4.08 |
| Sciomyzidae | 0.02 | 0.02 | 4.17 | 1.01 |
| Trichoptera |  |  |  |  |
| Hydropyschidae | 0.02 | 0.00 | 4.17 | 1.01 |
| Hydroptilidae | 0.06 | 0.12 | 8.33 | 2.04 |
| Hemiptera |  |  |  |  |
| Corixidae | 0.04 | 0.01 | 8.33 | 2.01 |
| Plecoptera |  |  |  |  |
| Nemouridae | 0.02 | 0.01 | 4.17 | 1.01 |
| Perlodidae | 0.02 | 0.02 | 4.17 | 1.01 |
| Ephemeroptera |  |  |  |  |
| Leptophlebiidae | 0.17 | 0.02 | 4.17 | 1.05 |
| Odonata |  |  |  |  |
| Oligochaeta |  |  |  |  |
| Lurnbriculidae | 0.15 | 2.81 | 16.67 | 4.71 |
| Hydrachnellae |  |  |  |  |
| Hydracharina | 0.02 | 0.00 | 4.17 | 1.01 |
| Terrestrial |  |  |  |  |
| Insects | 0.71 | 1.12 | 33.33 | 8.46 |
| Other | 0.10 | 0.04 | 4.17 | 1.03 |

Table C. 10 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for 4+ year old rainbow trout $(\mathrm{n}=9)$ sampled in 1995.

| PREY ITEM | \% by <br> Number | \% by <br> Weight | Frequency <br> of Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| Osteichthyes <br> Cyprinidae | 0.120 .41 | 51.896 .86 | 11.1111 .11 | 20.105 .35 |
| Unidentified fish | 19.47 | 0.89 | 11.11 | 9.14 |
| Cladocera <br> $\quad$ L. kindti | 76.11 | 3.95 | 33.33 | 32.92 |
| Daphnia spp. | 0.44 | 22.57 | 11.11 | 9.91 |
| Basommatophora <br> Physidae <br> Diptera <br> $\quad$ Chironomidae pupa <br> Terrestrial <br> Insects 1.77 | 0.37 | 44.44 | 13.53 |  |

Table C. 11 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for all walleye $(\mathrm{n}=153)$ sampled in 1995.

| PREY ITEM | \% by Number | $\begin{gathered} \text { \% by } \\ \text { Weight of } \end{gathered}$ | Frequency Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Catostomidae | 0.38 | 1.46 | 1.96 | 1.17 |
| Cottidae | 10.12 | 14.58 | 26.14 | 15.62 |
| Cyprinidae | 2.58 | 4.61 | 5.88 | 4.02 |
| Percidae | 10.22 | 25.53 | 17.65 | 16.40 |
| Salmonidae | 0.67 | 31.85 | 3.92 | 11.19 |
| Unidentified fish | 7.45 | 19.67 | 34.64 | 18.98 |
| Amphipoda 0.00 0.0.0.0 |  |  |  |  |
| Gammeras | 0.57 | 0.00 | 0.65 | 0.38 |
| Cladocera 0.00 ( ${ }^{\text {a }}$ |  |  |  |  |
| L. kindtii | 30.95 | 0.02 | 3.92 | 10.72 |
| Daphnia spp. | 10.03 | 0.01 | 4.58 | 4.49 |
| Eucopepoda |  |  |  |  |
|  | 2.39 | 0.00 | 0.65 | 0.94 |
| Decapoda 0.65 |  |  |  |  |
| Astacidae | 0.10 | 0.45 | 0.65 | 0.37 |
| Diptera |  |  |  |  |
| Chironomidae pupa | 13.56 | 0.06 | 9.15 | 7.00 |
| Chironomidae larvae | 1.91 | 0.01 | 3.92 | 1.79 |
| Hemiptera 0.00 |  |  |  |  |
| Corixidae | 0.19 | 0.01 | 0.65 | 0.26 |
| Plecoptera |  |  |  |  |
| Perlodidae | 0.10 | 0.00 | 0.65 | 0.23 |
| Ephemeroptera |  |  |  |  |
| Leptophlebiidae | 0.10 | 0.00 | 0.65 | 0.23 |
| Odonata 0.65 |  |  |  |  |
| Anisoptera | 0.19 | 0.05 | 1.31 | 0.48 |
| Zygoptera | 0.19 | 0.00 | 1.31 | 0.46 |
| Oligochaeta |  |  |  |  |
| Lumbriculidae | 0.67 | 1.12 | 1.96 | 1.15 |
| Terrestrial |  |  |  |  |
| Insects | 7.55 | 0.04 | 4.58 | 3.74 |
| Other | 0.10 | 0.53 | 0.65 | 0.39 |

Table C. 12 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $0+$ year old walleye $(\mathrm{n}=8)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Cotticlae | 7.14 | 99.76 | 37.50 | 42.79 |
| Unidentified fish | 2.38 | 0.01 | 12.50 | 4.41 |
| Cladocera |  |  |  |  |
| Daphnia spp. | 14.29 | 0.06 | 25.00 | 11.66 |
| Eucopepoda | 59.52 | 0.11 | 12.50 | 21.37 |
| Diptera |  |  |  |  |
| Chironomidae pupa | 14.29 | 0.04 | 37.50 | 15.36 |
| Chironomidae larvae | 2.38 | 0.01 | 12.50 | 4.41 |

Table C. 13 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $1+$ year old walleye $(\mathrm{n}=37)$ sampled in 1995.

| PREY ITEM | \% by Number | \% by Weight | Frequency of Occurrence | IRI |
| :---: | :---: | :---: | :---: | :---: |
| Osteichthyes |  |  |  |  |
| Cottidae | 0.97 | 28.13 | 31.82 | 16.15 |
| Percidae | 0.19 | 17.71 | 4.55 | 5.95 |
| Unidentified fish | 0.97 | 21.47 | 36.36 | 15.59 |
| Amphipoda Gammeras | 0.10 | 0.05 | 4.55 | 1.24 |
| Cladocera <br> L. kindtii | 0.10 | 0.05 | 4.55 | 1.24 |
| Daphnia spp. | 51.40 | 1.91 | 18.18 | 18.95 |
| Diptera Chironornidae pupa | 12.37 | 3.10 | 27.27 | 11.33 |
| Chironomidae larvae | 32.95 | 6.40 | 18.18 | 15.25 |
| Terrestrial Insects | 0.39 | 0.09 | 13.64 | 3.74 |

Table C. 14 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $2+$ year old walleye $(n=42)$ sampled in 1995.

|  | \% by <br> Number | \% by <br> Weight | Frequency <br> of | Occurrence |
| :--- | :---: | :---: | :---: | :---: |$\quad$ IRI

Table C. 16 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $4+$ year old walleye $(\mathrm{n}=21)$ sampled in 1995.

|  | \% by <br> Number | \% by <br> Weight | Frequency <br> of Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| OREY ITEM | 13.33 | 10.59 | 23.81 | 14.12 |
| Osteichthyes | Cotticlae | 30.00 | 4.89 | 9.52 |
| Cyprinidae | 21.67 | 13.76 | 28.57 | 13.14 |
| Percidae | 3.33 | 48.82 | 9.52 | 18.93 |
| $\quad$ Salmonidae | 21.67 | 20.25 | 47.62 | 26.48 |
| Unidentified fish <br> Decapoda <br> Astacidae | 1.67 | 1.69 | 4.76 | 2.40 |
| Diptera <br> Chironomidae pupa | 6.67 | 0.00 | 9.52 | 4.79 |
| Odonata <br> zygoptera | 1.67 | 0.00 | 4.76 | 1.90 |

Table C. 17 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for $5+$ year old walleye $(\mathrm{n}=13)$ sampled in 1995.

|  | \% by <br> Number | \% by <br> Weight | Frequency <br> of Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| PREY ITEM |  |  |  |  |
| Osteichthyes | 3.45 | 0.96 | 7.69 | 3.74 |
| Catostomidae | 3.45 | 0.03 | 7.69 | 3.46 |
| Cottidae | 17.24 | 8.09 | 23.08 | 14.98 |
| Cyprinidae | 31.03 | 31.48 | 15.38 | 24.11 |
| Percidae | 10.34 | 35.02 | 15.38 | 18.80 |
| Salmonidae | 31.03 | 24.29 | 46.15 | 31.41 |
| Unidentified fish |  | 0.13 | 7.69 | 3.49 |
| Odonata |  |  |  |  |

Table C. 18 Percentage by number, percentage by weight, frequency of occurrence and index of relative importance (IRI) of food items for 6+ year old walleye $(n=4)$ sampled in 1994.

| PREY ITEM | \% by <br> Number | \% by <br> Weight | Frequency <br> of Occurrence | IRI |
| :--- | :---: | :---: | :---: | :---: |
| Osteichthyes <br> Unidentified fish | 100.00 | 100.00 | 100.00 | 100.00 |

SECTION 3

# ARTIFICIAL IMPRINTING OF JUVENILE KOKANEE SALMON (Oncorhynchus nerka): <br> IMPLICATIONS FOR OPERATING LAKE ROOSEVELT KOKANEE SALMON HATCHERIES 

## ANNUAL REPORT 1995

## Prepared by:

Mary Beth Tilson<br>Allan T. Scholz<br>Jennifer L. Miller<br>Upper Columbia United Tribes Fisheries Research Center Eastern Washington University<br>Department of Biology<br>Cheney, Washington

## Funded by:

U.S. Department of Energy

Bonneville Power Administration
Division of Fish and Wildlife
P.O. Box 3621

Portland, OR 97283-3621
Project Number 88-63
Contract Number DE-BI79-88BP91819

SEPTEMBER 1996

## EXECUTIVE SUMMARY

The Lake Roosevelt Monitoring Program was designed in 1988 to evaluate the effectiveness of Lake Roosevelt reservoir kokanee salmon hatcheries which were operational in 1991 and 1992. From 1991 to 1994, most of the kokanee released into Lake Roosevelt were released as fry. Poor adult returns to egg collection and release sites prompted investigations to determine if there was a critical period for thyroxine induced olfactory imprinting in kokanee salmon. These imprinting investigations have shown that there are two critical periods for imprinting, at the alevin/swimup stage and the smolt stage. Additionally, past investigations indicated that kokanee salmon undergo smoltification and residualization similar to anadromous salmonids. Kokanee have been observed in past years emigrating out of Lake Roosevelt as far downstream as Rock Island Dam. Therefore, hatchery managers changed their release strategies from primarily fry releases to smolt releases. This improved the number of adult recoveries substantially ( $99 \%$ of the fish recovered were released as residualized smolts).

The objectives of the present investigation were to: (1) determine the critical period(s) for olfactory imprinting and (2) assess the best times and locations to release kokanee in order to prevent entrainment, and improve returns to creel and egg collection sites. From 1992 to 1995, coded wire tagged (CWT) fish were released as residualized smolts into Lake Roosevelt. These fish were imprinted at different life stages and were given an adipose clip and a distinctive coded wire tag. These returning adults would enable us to determine (1) the number entrained below Grand Coulee Dam, (2) the number harvested by anglers; (2) the number homing to egg collection sites, and (4) the number straying to other locations.

Results of the present investigation continued to show that kokanee can be successfully imprinted to artificial odors - morpholine and phenethyl alcohol - as juveniles from hatch to swimup and again as smolts. Fish double exposed to synthetic chemicals at alevin/swimup and smolt stages had the highest rate of homing to egg collection sites ( $85 \%$ of the morpholine exposed fish returned to morpholine scented streams and $88 \%$ of the phenethyl alcohol exposed fish returned to phenethyl alcohol streams). Additionally, fish exposed to synthetic chemicals were recovered in greater numbers and displayed higher homing ability to egg collection sties than fish that were not exposed to synthetic chemicals. Results also indicated that smolt releases continued to provide better adult recoveries than fry releases. Although there were approximately equal percentages of fry and smolts released ( $42 \%$ and $58 \%$ respectively) from 1992 to 1994, almost all recoveries in 1995 (99\%) came from fish released as smolts.

Based on the results of this investigation, we recommend the following measures for Lake Roosevelt kokanee hatcheries:

1) Make an effort to release more fish into the reservoir. This should be done by (a) drilling a new well at the Spokane Tribal Hatchery to deliver additional flow needed to raise more smolt sized fish; (b) provide new net pen sites; (c) initiate experiments with induction of spawning by injecting females with inducing hormones.
2) Monitor entrainment from Lake Roosevelt.
3) Modify holding facilities and hatchery ladder system at Sherman Creek Hatchery to attract spawning kokanee.
4) Set up an egg collection site at Hawk Creek since 286 fish were recovered at that site.
5) Locate alternative stocks of kokanee with better genetic adaptations than Lake Whatcom fish for the Lake Roosevelt Program.

## ACKNOWLEDGMENTS

We thank Tim Peone (Manager, Spokane Tribal Hatchery) for advice and coordination in all aspects of this study. Spokane Tribal Hatchery personnel Del Brown, James Andrews and Jayne Abraham assisted with the rearing and collection of fish for this study. Del Brown also assisted with monitoring the Little Falls Dam site for returning kokanee. We thank the Spokane Tribal hatchery crew, including Diane Abrahamson, Eugene Abrahamson, Cyndi Day, Barbara Gunter, Shari Smallwood, Vina Timmons and Jo-Anne Wynne for coded wire tagging and fin clipping kokanee. We also thank the Spokane Tribal Lake Roosevelt Monitoring Program crew-Keith Underwood, Amy Voeller, Rudy Peone, Sam Abrahamson, William Matt Jr., Henry Etue III, Bart Kiefer, Veronica Flett, Jason Wynecoop-- and Lake Roosevelt Creel Clerks-- Bill Matt Sr. (Spokane Tribe), Leroy Williams (Colville Tribe), and Jim Meskin (Washington Department of Fish and Wildlife)-- for recovering coded wire tagged fish. We would also like to thank the Eastern Washington University Environmental Biology Club for donating their time electroshocking. These people included Jeff Jordan, Bethany McCabe, Scott Newton, Amy Laush, Anna Gacek, Lila Parlin and James Chase. We also thank Washington Department of Fish and Wildlife personnel, at the Region I office (John Hisata, Kurt Vail) and Sherman Creek Hatchery (Mitch Combs) for use of their unpublished data regarding walleye predation on stocked kokanee, and for advice and coordination in all aspects of this investigation. Charlie Craig (Bonneville Power Administration, Division of Fish and Wildlife) served as BPA Contracting Officer Technical Representative for this project.

This project was supported by a contract from the U.S. Department of Energy, Bonneville Power Administration (BPA), Division of Fish and Wildlife, Project No. 88-63, Contract No. DE-BI179-88BP918 19. Additional financial support for this project was provided by a grant from the U.S. Department of Interior, Bureau of Indian Affairs to the Upper Columbia United Tribes (UCUT) to fund the operation of the UCUT Fisheries Research Center at Eastern Washington University, Department of Biology. Capital equipment for this project was provided by Eastern Washington University, Department of Biology.

[^4]
## TABLE OF CONTENTS

EXECUTIVE SUMMARY ..... 282
ACKNOWLEDGMENTS ..... 283
TABLE OF CONTENTS ..... 284
1.0 INTRODUCTION. ..... 289
1.1 Study Strategy ..... 291
2.0 METHODS AND MATERIALS ..... 294
2.1 Rearing Conditions ..... 294
2.2 Olfactory Imprinting Investigations. ..... 294
2.3 Coded Wire Tagging Program ..... 294
2.4 Statistical Analysis. ..... 298
3.0 RESULTS. ..... 299
3.1 Kokanee Salmon Recoveries ..... 299
3.1.1 Lake Whatcom Brood Recoveries ..... 299
3.1.2 Captive Brood Fish (Unexposed) Recoveries ..... 307
4.0 DISCUSSION. ..... 309
4.1 Imprinting Investigations ..... 309
4.2 Kokanee Harvest. ..... 316
4.3 Kokanee Egg Production Estimates ..... 321
4.4 Management Recommendations ..... 323
LITERATURE CITED ..... 327
APPENDICES ..... 330

## LIST OF TABLES

Table 1. Summary of coded wire tagged kokanee exposed to synthetic chemicals released from 1992-1995 and recovered from 1993-1995. ..... 290
Table 2. Kokanee salmon recoveries in Lake Roosevelt by electrofishing gill net, trap and creel surveys from January through December 1995 ..... 300
Table 3. Number of recoverable coded wire tagged (1991 to 1993 cohort) kokanee salmon released into Lake Roosevelt from 1992 through 1995. ..... 301
Table 4. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1992. Recoveries are total number recovered from creel and fisheries surveys conducted from 1992 to 1995. ..... 302
Table 5. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1993. Recoveries are total number recovered from creel and fisheries surveys conducted from 1993, 1994 and 1995. ..... 303
Table 6. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1994. Recoveries are total number recovered from creel and fisheries surveys conducted in 1994 and 1995. ..... 304
Table 7. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1995. Recoveries are total number recovered from creel and fisheries surveys conducted in 1995. ..... 305
Table 8. Recoveries by location of kokanee salmon with fin clips. These fish were exposed to synthetic chemicals in 1993, held at the Spokane Tribal Hatchery until July 1993, when they were released into the Spokane Arm of Lake Roosevelt as residualized smolts. Recoveries were made from September to November 1993, 1994, and 1995. ..... 306
Table 9. Statistical comparison of the number of kokanee salmon recovered from 1992 to 1995 which were exposed to either morpholine or phenethyl alcohol and recovered at the morpholine egg collection site, phenethyl alcohol egg collection site or at other sites from 1992 to 1995 ..... 308
Table 10. Summary of chemically exposed -v- unexposed coded wire tagged/fin clipped kokanee salmon homing to release sites from 1993 to 1995. ..... 311
Table 11. Catch per unit effort (CPUE) for kokanee salmon recovered at Kettle Falls and the Spokane River by elecuoshocking from September 1 to November 30, 1989 through 1995. ..... 313
Table 12. Mean length (\&SD), weight ( $\pm$ SD), range of length and number in sample for kokanee salmon collected in 1995. ages based on coded wire tag analysis ..... 318
Table 13 Summary of age, mean length, mean weight of coded wire tagged kokanee caught from 1992 to 1995 ..... 318
Table 14. Mean length ( $\pm$ SD), weight (\&SD), range of length and number in sample for kokanee salmon collected in 1995. Ages based on scale analyses. ..... 3.19
Table 15. Mean length ( $\pm$ SD), weight ( $\pm$ SD), range of length and number in sample for kokanee salmon collected in 1995 from electroshocking. Ages based on length $<371 \mathrm{~mm}=$ age 2, 371-479 mm = age 3, $>470 \mathrm{~mm}=$ age 4 . ..... 319
Table 16 Estimates of kokanee salmon fecundity based on females ( $>370 \mathrm{~mm}$ ) recovered from September 1 to November 30, 1995. ..... 322

## LIST OF FIGURES

Figure 1. Location of Lake Roosevelt kokanee hatcheries operated by Spokane Tribe and WDFW ..... 295
Figure 2. Lake elevation (feet) and water retention time (days) in Lake Roosevelt during 1991 and 1995. ..... 315
Figure 3. Length frequency (mm) for coded wire tagged kokanee collected in 1995. ..... 320

### 1.0 INTRODUCTION

The Lake Roosevelt Monitoring Program was designed in 1988 to evaluate the effectiveness of Lake Roosevelt reservoir kokanee salmon hatcheries. In 1995, one of the objectives of this program was to identify temporal and spatial release sites for hatchery reared kokanee salmon which would minimize entrainment through Grand Coulee Dam, maximize angler harvest, and maximize homing. The specific objectives were to:
(1) Determine the critical period(s) for olfactory imprinting in kokanee salmon via investigations in Lake Roosevelt. Knowing the time of imprinting in kokanee is essential for developing a successful Lake Roosevelt Kokanee Hatchery outplanting program. For the hatcheries to be self-sustaining, sufficient numbers of kokanee need to return to egg collection sites. Successful imprinting could improve returns to egg collection sites.
(2) Assess the best times and locations to release kokanee in terms of preventing entrainment below Grand Coulee Dam, and improving returns to creel and egg collection sites.

Both of these objectives were addressed through experiments that were initiated in 1992, 1993, 1994 and 1995 (Scholz et al. 1992, 1993; Tilson et al. 1994, 1995). Fish were released from 1992 to 1995 and monitoring occurred from 1993 to 1995. Table 1 shows that fish released as fry from 1992-1995 could be recovered from 1993-1995. Also, fish released as smolts from 1992-1995 could be recaptured from 1993-1997. During the release years kokanee were exposed to synthetic chemicals, morpholine and phenethyl alcohol, at different life history stages to determine the critical period for imprinting in kokanee. The life stages were: (1) fertilized egg to eyed egg, (2) eyed egg to hatch, (3) time of hatch, (4) hatch to swimup, (5) time of swimup, (6) fry O-1 month post swimup, (7) fry 1-2 months post swimup, (8) fry 3-6 months post swimup, and (9) smolt (about 16-18 months post fertilization). Thyroid hormone levels monitored in these fish indicated peaks at hatch to swimup and smolt stages (Scholz et al. 1992, 1993; Tilson et al. 1994). A portion of these fish were held until maturity at age 2 and 3. Experiments were conducted with these fish to determine if kokanee could be imprinted to synthetic chemicals at different life stages and then home back to their appropriate chemical as mature adults. In these experiments (Tilson et al. 1994, 1995), fish were exposed to synthetic chemicals in 1992 and 1993 and behavioral tests were conducted in 1994 with 2 year old sexually mature fish (1992 cohort) and in 1995 with 2 year old (1993 cohort) and 3 year old (1992 cohort) sexually mature fish (Tilson et al. 1994, 1995). Sexually mature fish were released into a stream with a natural

Table 1. Summary of coded wire tagged kokanee exposed to synthetic chemicals released from 1992-1995 and recovered from 1993-1995.

| Cohort | Year of Release | Life stage at Release | Life stage Exposed | Year of recapture at age |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 2 | 3 | 4 |
| 1990 | 1992 | Smolt | Smolt | -- | 1993 | 1994 |
| 1991 | 1992 | Fry | Eyed egg | 1993 | 1994 | 1995 |
|  |  |  | Hatch | 1993 | 1994 | 1995 |
|  |  |  | Alevin | 1993 | 1994 | 1995 |
|  |  |  | Swimup | 1993 | 1994 | 1995 |
|  |  |  | Fry (Feb-Jul) | 1993 | 1994 | 1995 |
| 1991 | 1993 | Smolt | Smolt | 1993 | 1994 | 1995 |
| 1992 | 1993 | Fry | Eyed egg | 1994 | 1995 | 1996 |
|  |  |  | Hatch | 1994 | 1995 | 1996 |
|  |  |  | Alevin | 1994 | 1995 | 1996 |
|  |  |  | Swimup | 1994 | 1995 | 1996 |
|  |  |  | Fry (Feb-Jul) | 1994 | 1995 | 1996 |
| 1992 | 1994 | Smolt | Hatch | 1994 | 1995 | 1996 |
|  |  |  | Alevin | 1994 | 1995 | 1996 |
|  |  |  | Swimup | 1994 | 1995 | 1996 |
| 1993 | 1994 | Fry | Hatch-Swimup | 1995 | 1996 | 1997 |
|  |  | Fry | Alevin-Swimup | 1995 | 1996 | 1997 |
| 1993 | 1995 | Smolt | Hatch-Swimup | 1995 | 1996 | 1997 |
|  |  |  | Alevin-Swimup | 1995 | 1996 | 1997 |
|  |  |  | Hatch-Swimup and Smolt | 1995 | 1996 | 1997 |
|  |  |  | Alevin-Swimup and Smolt | 1995 | 1996 | 1997 |
| 1994 | 1995 | Fry | Hatch-Swimup | 1996 | 1997 | 1998 |
|  |  |  | Alevin-Swimup | 1996 | 1997 | 1998 |

Y-maze, and traps located at each arm of the maze at the upstream end. Results suggested that imprinting coincided with elevated thyroxine levels. For example, zero age fish ( 1991 and 1992 cohort) displayed basal whole body thyroxine levels of $6.5 \pm 1.3 \mathrm{ng} / \mathrm{g}$ body weight (BW) and 8.3 $\pm 1.1 \mathrm{ng} / \mathrm{g}$ B W respectively and peak levels at the time of swimup of $22.1 \pm 5.2 \mathrm{ng} / \mathrm{g}$ B W and $15.1 \pm 2.1 \mathrm{ng} / \mathrm{g}$ BW respectively. Fish exposed at hatch and swimup displayed 66 and $81 \%$ homing respectively as 2 year olds (1991 and 1992 cohort) and 68 and $92 \%$ homing respectively as 3 year olds (1991 cohort). When plasma thyroxine was measured in yearling fish (1990 and 1991 cohort) the basal levels were $2.2 \mathrm{ng} / \mathrm{ml}$ and $10.6 \mathrm{ng} / \mathrm{ml}$ respectively with peak levels at the smolt stage (early spring) of $20 \mathrm{ng} / \mathrm{ml}$ and $15.2 \mathrm{ng} / \mathrm{ml}$ respectively (Tilson et al. 1994). When these fish (1990 cohort) were tested as mature adults, they displayed $59 \%$ homing as 3 year olds. Fish exposed at pre-hatch and post-swimup stages had whole body thyroxine levels of $<8.0 \mathrm{ng} / \mathrm{g}$ BW and $<1.0 \mathrm{ng} / \mathrm{g}$ BW respectively. These groups displayed less than $30 \%$ homing when tested as mature adults. Therefore, thyroxine concentration was elevated at both hatch to swimup stage and also at the smolt stage. The highest percentage of fish homing to their exposure odor as mature adults (1990 and 1991 cohort tested in 1993, and 1991 and 1992 cohort tested in 1994) were fish exposed at hatch to swimup and the smolt stage.

To determine if the results in these behavioral experiments could be duplicated in the field (Lake Roosevelt), most of the fish which were imprinted at different life stages were marked with an adipose clip and a distinctive coded wire tag that uniquely identified: (1) exposure chemical, (2) life stage exposed, (3) release location, (4) life stage released, and (5) date released. A total of $1,243,774$ coded wire tagged fish were released into Lake Roosevelt from 1992 to 1995. These fish will become sexually mature spawners from 1994 to 1998. Field tests in Lake Roosevelt were conducted to determine the following information for each group of coded wire tagged fish: (1) number entrained below Grand Coulee Dam, (2) number harvested by anglers in Lake Roosevelt; (3) number homing to egg collection sites scented with the appropriate imprinting chemical, and (4) number straying to other locations.

### 1.1 Study Strategy

The field tests were conducted in 1994 and 1995. During 1994, the following tasks were completed (Tilson et al. 1995):
(1) Kokanee were exposed at Spokane Tribal Hatchery to synthetic chemicals from hatch to swimup stage in January - February 1994. The odor delivery system was set up and monitored throughout this period.
(4) Kokanee entrainment below Grand Coulee Dam was monitored at Rocky Reach Dam and at Rock Island Dam. If any adipose clipped kokanee outmigrants were collected during the season, $4 / 1-8 / 31$, the biologists at these facilities were to freeze the fish so we could check for coded wire tags.

Kokanee returns to egg collection sites were monitored by augmenting the creel survey at Little Falls Dam, and by augmenting the electrofishing surveys at Little Falls Dam and Sherman Creek during the spawning season. These augmented electrofishing surveys were conducted by EWU biologists and by EWU Environmental Biology Club volunteers. Additionally, EWU monitored Big Sheep Creek, Colville River, Blue Creek, Hawk Creek, and Barnaby Creek.
(6) Information about tagged kokanee captured in Lake Roosevelt was augmented by conducting a site specific creel survey. Biologists from EWU interviewed anglers at specific sites on Lake Roosevelt from January to July at times (usually evening) when anglers were likely to be catching kokanee and when the scheduled creel clerk was not at those sites.
(7) Recommendations were made about exposure times and release locations and dates that (a) reduce entrainment below Grand Coulee Dam, and/or (b) increase harvest rates in Lake Roosevelt and/or (c) increase returns of adults to egg collection sites.

### 2.0 METHODS AND MATERIALS

### 2.1 Rearing Conditions

Eggs for producing kokanee to be reared at the Spokane Tribal Hatchery near Wellpinit, WA were obtained from two sources: 1) Lake Whatcom stock eyed eggs transferred from the Lake Whatcom Hatchery in Bellingham WA (WDFW); and 2) spawn take from kokanee captive brood held at the Spokane Tribal Hatchery (Figure 1). Water supply to the raceways was a combination of Metamooteles Springs water and well water at $8-11^{\circ} \mathrm{C}$. After swimup, zero age fry were feed trained on Biodiet semi-moist mash (starter feed). Older fry were fed a combination of Biodiet semi-moist grower feed (1.0-2.5 mm crumbles) and Silvercup size 1-4 mm crumbles. Yearling fish were fed Biodry 1000 pellets ( $3.0-4.0 \mathrm{~mm}$ ) obtained from Bioproducts, Inc. Photoperiod was maintained at natural daylength as each raceway was partially exposed to natural conditions of light and weather.

### 2.2 Olfactory Imprinting Investigations

From 1991 to 1994, fish were exposed to either morpholine $\left(\mathrm{C}_{4} \mathrm{H}_{9} \mathrm{NO}\right)$ or phenethyl alcohol $\left(\mathrm{C}_{8} \mathrm{H}_{10} \mathrm{O}\right)$ at various developmental stages. The life stages were: (1) fertilized egg to eyed egg, (2) eyed egg to hatch, (3) time of hatch, (4) hatch to swimup, (5) time of swimup, (6) fry O-1 month post swimup, (7) fry 1-2 months post swimup, (8) fry 3-6 months post swimup, (9) smolt (about 16-18 months post fertilization) (Appendix A). At each life stage, one group of fish was exposed to morpholine $\left(5 \times 10^{-5} \mathrm{mg} / \mathrm{l}\right)$ and a second to phenethyl alcohol $\left(5 \times 10^{-3} \mathrm{mg} / \mathrm{l}\right)$. In 1994 and 1995, fish were "double imprinted", initially at the hatch through swimup stages (i.e. alevin) in 1994 and a second time at the smolt stage in 1995. These fish received the double imprint because our earlier investigations had demonstrated that both the alevin/swimup and smolt stages were sensitive periods for imprinting. Two imprinting chemicals were employed in this experiment so that one odor could act as a control for the other. Details of the synthetic chemical imprinting procedure and methods for calculating steady state concentration of imprinting chemicals were described in a previous annual report (Scholz et al. 1993).

### 2.3 Coded Wire Tagging Program

From 1992 to 1995, most of the hatchery kokanee were exposed to synthetic chemicals and released into Lake Roosevelt at various locations. A portion of each group released were tagged with distinctive coded wire tags. During the spawning season in autumn 1995, both morpholine (at a steady state concentration of $5 \times 10^{-5} \mathrm{mg} /$ ) and phenethyl alcohol ( $5 \times 10^{-3}$

Figure 1. Location of Lake Roosevelt kokanee hatcheries operated by Spokane Tribe and Washington Department of Fish \& Wildlife.


Inset shows location where fish stocks were obtained (Lake Whatcom Hatchery and Lake Roosevelt at Little Falls Dam).

$\mathrm{mg} / \mathrm{l})$ were metered into the ladder trap at Sherman Creek. This protocol was changed from that of 1992 through 1994 when only morpholine was metered into Sherman Creek. Due to a management decision by the Lake Roosevelt Kokanee Hatchery Steering Committee, it was decided to drip both morpholine and phenethyl alcohol into Sherman Creek in order to attract as many adult spawners as possible at that site. Phenethyl alcohol was also metered into the Spokane River at a site below Little Falls Dam from 1992 through 1995. For marking experiments, kokanee were dipnetted out of hatchery raceways and mildly anesthetized with a 50 $\mathrm{mg} / \mathrm{l}$ concentration of tricaine methanesulfonate (MS-222). Coded wire tags (CWT) were then injected into the rostrum using a model MK4 CWT machine (Northwest Marine Technology, Inc.), equipped with two different nose hoods specially fitted for fry and fingerling-sized fish. Lengths and weights of fry which were marked for release in 1995 ranged from 52 to 78 mm and 1.4 to 4.6 g respectively. Lengths and weights of fingerlings ranged from 116 to 124 mm and 15 to 18.6 g respectively (Appendix A). All hatchery fish were given an adipose fin clip as an external identification mark. Marked fish were counted using a tally counter, then released back into hatchery raceways through a quality control device (QCD) (Northwest Marine Technology, Inc.) equipped with a CWT detector. The fish were retained for approximately three weeks before release to estimate mortality rates and tag retention. In 1995, mortality rates were uniformly low ( $<1 \%$ ) (T. Peone, Spokane Tribal Hatchery, personal communication). The mean percent tag retention after 20 days was $96 \%$, and ranged from 93 to $99.1 \%$ (Appendix A).

Reservoir wide'creel surveys were conducted throughout the year by individuals from the Spokane Tribe, Colville Confederated Tribe, and the Washington Department of Fisheries and Wildlife (See section 2.2 of Underwood et al. 1996). Also, a site specific creel survey at Spring Canyon Campground near Coulee Dam, Washington was conducted from January through July 1995 by individuals from the UCUT Fisheries Center at EWU. These augmented creel surveys were conducted on days and/or times when the regularly scheduled creel clerk would not be at that site and when kokanee were most likely to be observed. The augmented surveys at Spring Canyon were conducted on January 9, 10, 11, 12, 13; April 9, 12, 15, 17, 23, 30; May 7, 13; and July 11, 17. At Little Falls Dam the creel survey was augmented on April 14, 25; May 1 and September 30. At other sites, the creel was conducted on September 13, 20 and October 5, 6. In addition, electrofishing/gill net surveys were conducted throughout the year (See Section 2.3 of Underwood et al. 1996) and the electrofishing surveys were augmented by an EWU crew during the spawning season (September-November, 1995). This EWU crew consisted mainly of vollunteers from the EWU EnvironThese augmented electroshocking surveys were done ten days in September ( $9 / 6$ to $9 / 29$ ), 9 days in October (10/3 to 10/27) and 8 days in November ( $11 / 1$ to $11 / 30$ ) at various locations in the reservoir and concentrated at Little Falls Dam and Sherman

Creek. Additionally, a ladder trap was monitored by WDFW personnel at Sherman Creek Hatchery.

The heads of kokanee with adipose clips were cut off and sent to the Fisheries Research Center at Eastern Washington University, where CWT's were dissected out and examined with a dissecting microscope to determine the lot code. The number of fish from each lot returning to Sherman Creek, Little Falls Dam, and other locations was determined.

Percent error of coded wire tags read was determined by having two individuals read all of the tags. If there was a discrepancy on a tag code, both people re-read those tags until they were in agreement. In 1995, a total of 1,578 adipose clipped fish were examined for coded wire tags. There were 1,243 heads which contained tags. However, $2 \%$ ( $n=23$ tags) were lost in the extraction process. All tags were read by two individuals. After Reader 1 and Reader 2 read the tags, there was a discrepancy on 89 of 1,218 tag codes. After those 89 tags were re-read by both readers and a tag code was agreed upon, Reader 1 was incorrect on $67 / 1,218$ tags for a $5.5 \%$ error. Reader 2 was incorrect on $22 / 1,218$ tags for a $1.8 \%$ error. However, the total percent error was $0 \%$ since all tags were read twice and all were agreed upon.

To monitor kokanee entrainment from Lake Roosevelt, a creel survey was conducted by boat for two days in February on Rufus Woods Reservoir (directly below Grand Coulee Dam). However, this was a trial survey which resulted in very few angler interviews. In addition, budget constraints prohibited a more in-depth angler survey.

Therefore, in order to collect coded wire tagged fish below Grand Coulee Dam, we coordinated efforts with the Fish Passage Center's Smolt Monitoring Program at both Rocky Reach Dam (three dams downstream from Grand Coulee Dam) and at Rock Island Dam (four dams down from Grand Coulee Dam) to collect kokanee migrants from April 1 to August 30, 1995. At Rocky Reach Dam, we had technicians examine adipose clipped kokanee salmon of any size. At Rock Island Dam, we had technicians look for adipose clipped kokanee which were 250 mm or larger. The reason for the size differentiation was because Lake Wenatchee stock sockeye were also adipose clipped and coded wire-tagged. It would not be possible to separate the adipose clipped age 1+ Lake Roosevelt kokanee from adipose clipped age 1+ anadromous sockeye passing Rock Island Dam without sacrificing both species. The larger kokanee ( $>250$ mm ) would have been 2 to 4 year old fish which had remained in one of the reservoirs after release until they migrated downstream or had been entrained as subadults. However, there were no adipose clipped kokanee of any size seen at either of the collection facilities (Robert McDonald, Fish Biologist, Chelan County PUD, personal communication). In addition, there
were no adult coded wire tagged fish turned in at the Washington Department of Fish and Wildlife Head Lab (Ken Johnson, Pacific States Marine Fisheries Commission, personal communication).

### 2.4 Statistical Analysis

Statistical evaluation was determined via a chi-squared test at $\mathrm{p} \leq 0.05$. The null hypothesis $\left(\mathrm{H}_{0}\right)$ stated, "Chemically exposedjish were random in their movements. They were recaptured in equal numbers in scented and unscented streams during the spawning migration." If the calculated probability was greater than 0.05 ( $p>0.05$ ), then we interpreted this to mean that fish were not attracted to their exposure odor. If p 20.05, then we accepted the alternative hypothesis (HA), "Chemically exposed fish moved to the scented stream in greater number than theunscented stream" A statistical significance would imply that the fish were homing to their exposure odor.

### 3.0 RESULTS

### 3.1 Kokanee Salmon Recoveries

Recoveries of kokanee salmon captured in gill net, electrofishing, ladder trap and creel surveys in 1995 are shown in Table 2. A total of 2,815 fish were recovered. Of those, $60 \%$ ( $n=1,689$ ) were fin clipped.

Of the fish recovered in the creel surveys ( $n=467$ ), 185 fish were recovered by the Lake Roosevelt Monitoring Program's year round creel survey and 282 fish ( $60.4 \%$ ) were recovered by EWU's augmented, site specific creel survey. Most fish were recovered in fish surveys ( $n=2,348$ ) including gill netting ( $n=45$ ), ladder traps ( $n=3$ ) and electroshocking ( $n=2,300$ ). Many of these fish were recovered during their fall spawning migration at sites scented with synthetic chemicals, including 1,003 at Sherman Creek and 593 at Little Falls (Table 2).

### 3.1.1 Lake Whatcom Brood Recoveries

From 1992-95 a total of 5,413,270 fish were released into Lake Roosevelt, of which 1,041,355 were tagged with CWT/fin clips. Total coded wire tag releases included 507,319 fry and 534,036 smolts (Table 3 ). However, only 445,778 fry releases were considered recoverable (1990-1993 cohort) in 1995. The other 51,411 fish were released as fry in 1995 and thus would not have been adults in 1995. Kokanee in Lake Roosevelt attain lengths of about 250-350 mm by age 2 and spawn principally at age 2 , age 3 and sometimes age 4 . A portion of the total number of coded wire tagged fish ( $88 \%$ ), including 497,189 fry and 418,432 smolts were Lake Whatcom stock fish that had been exposed to either morpholine or phenethyl alcohol. The remainder ( $12 \%$ ) were not exposed to a chemical and were from captive brood stock fish (Table $3)$.

A total of 1,643 CWT/fin clipped fish were recaptured from 1992 to 1995 including 1,205 recaptured at egg collection sites at Sherman Creek and Little Falls Dam/Spokane River (Tables 4-8). Of the total 1,643 CWT/fin clip recoveries, 1,634 fish ( $99.4 \%$ ) had been stocked as smolts or residualized smolts and 9 fish ( $0.6 \%$ ) were stocked as fry. Therefore, the ratio of smolts to recoverable fry released was approximately 1.2 to 1 respectively, the ratio of smolts to fry recovered as adults was approximately 182.5 to 1 respectively. Clearly, smolt releases produced substantially more adults than fry releases.

From 1992-1995, a total of 339 (76\%) morpholine-exposed fish (released at the smolt stage) were recovered as adult spawners at Sherman Creek compared to 46 (10\%) recovered at

Table 2. Kokanee salmon recoveries in Lake Roosevelt by electrofishing, gill net, trap and creel surveys from January through December 1995 ${ }^{1}$.

| Location | Fish Surveys |  |  | Creel Surveys |  |  | Totals |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total \# | Adipose clip only | Other clips | $\begin{gathered} \text { Total } \\ \# \\ \hline \end{gathered}$ | Adipose <br> Clip only | Other Clips | Total \# recovered | Total \# Clipped |
| 0 Northport | 5 | 1 | 1 | 0 | 0 | 0 | 5 | 2 |
| 1 Kettle Falls | 995 | 874 | 53 | 8 | 0 | 0 | 1,003 | 927 |
| 2 Gifford $^{2}$ | 29 | 20 | 0 | 0 | 0 | 0 | 29 | 20 |
| 3 Hunters | 109 | 90 | 0 | 0 | 0 | 0 | 109 | 90 |
| $\omega 4$ Porcupine Bay | 110 | 76 | 9 | 0 | 0 | 0 | 110 | 85 |
| O 5 Little Falls | 577 | 107 | 70 | 16 | 10 | 2 | 593 | 189 |
| 6 Seven Bays2 | 473 | 331 | 15 | 6 | 2 | 1 | 479 | 349 |
| 7 Kellers Ferry | 1 | 0 | 0 | 22 | 0 | 0 | 23 | 0 |
| 8 San Poil | 30 | 18 | 0 | 1 | 0 | 0 | 31 | 18 |
| 9 Spring Canyon | 19 | 2 | 0 | 414 | 7 | 0 | 433 | 9 |
| TOTALS | 2,348 | 1,519 | 148 | 467 | 19 | 3 | 2,815 | 1,689 |

1 Totals include regularly scheduled electroshocking and creel surveys, and augmented electroshocking and creel surveys conducted by Eastern Washingion
University.
2 Gifford location includes Barnaby Creek; Seven Bays location includes Hawk Creek and Blue Creek.

Table 3. Number of recoverable coded wire tagged (1991 to 1993 cohort) kokanee salmon released into Lake Roosevelt from 1992 through 19951.

Imprinted • Lake Whatcom brood

| Stage at |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
| Release | 1992 | 1993 | 1994 | 1995 |
| fry | 171,452 | 204,328 | 69,998 | $51,411^{2}$ |
| smolt | 0 | 53,979 | 108,602 | 255,851 |

Non-imprinted - Captive brood
Stage at

| Release | 1992 | 1993 | 1994 | 1995 |
| :--- | :---: | :---: | :---: | :---: |
| fry | 0 | 0 | 10,130 | 0 |
| smol t | 0 | 26,489 | 22,584 | 66,531 |

1 These numbers represent kokanee that can be positively identified with coded wirc tag data codes which have not been duplicated.
2 These fish were 94 cohort fry and were only $1+$ in 1995 . Therefore, they were not counted in the total number released.

Table 4. $\quad \begin{aligned} & \text { Recoveries by location of coded wire tagged kokanee salmon from releases made in 1992. Recoveries are total } \\ & \text { number recovered from creel and fisheries surveys conducted from } 1992 \text { to } 1995 .\end{aligned}$


[^5]Table 5. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1993. Recoveries are total number recovered from creel and fisheries surveys conducted in 1993, 1994 and 1995.

| Cohort | Stage | Exposure | Release | Life Stage | Total \# CWT | \#recovered at |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Sherman Creek | Spokane Rlver | Other |
|  | Exposed | Odor | Location | At Release | Released | ( MOR') | ( PEA) |  |
| 1992 | Eyed egg | M 0 R | Sherman Creek | Fry | 14.355 | 0 | 0 | 0 |
|  |  | MOR | Spokane River | Fry | 10,903 | 0 | 0 | 0 |
|  |  | PEA | Sherman Creek | Fry | 10,721 | 0 | 0 | 0 |
|  |  | PEA | Spokane River | Fry | 10,960 | 0 | 0 | 0 |
| 1992 | Hatch | MOR | Sherman Creek | Fry | 1,988 | 0 | $1{ }^{1}$ | 0 |
|  |  | MOR | Spokane River | Fry | 31,416 | 0 | 0 | 0 |
|  |  | MOR | Bamaby Creek | Fry | 21,784 | 0 | 0 | 0 |
|  |  | PEA | Sherman Creek | Fry | 1,988 | 0 | 0 | 0 |
|  |  | PEA | Spokane River | Fry | 21,993 | 0 | 0 | 0 |
| 1992 | Alevin | MOR | Sherman Creek | Fry | 10,938 | 0 | 0 | 0 |
|  |  | PEA | Sherman Creek | Fry | 11,791 | 0 | 0 | 0 |
| 1992 | Swimup | MOR | Sherman Creek | Fry | 10,908 | 0 | 0 | 0 |
|  |  | PEA | Sherman Creek | Fry | 10,885 | 0 | 0 | 0 |
| 1992 | Fry (Feb) | MOR | Sherman Creek | Fry | 10,802 | 0 | 0 | 0 |
|  |  | PEA | ShermanCreck | Fry | 10,896 | 0 | 0 | 0 |
| 1991 | Smolt | MOR | Sherman Creek | Smolt | 38,030 | $36^{i}$ | 5 | 0 |
|  |  | PEA | Sherman Creek | Smolt | 1,153 | 20 | 25 | 0 |
|  |  | PEA | Blue Creek | Smolt | 8,196 | 0 | 92 | 8 |
|  |  | NONE | Sherman Creek | Smolt | 26,489 | $1^{i}$ | 0 | 0 |

[^6]Table 6. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1994, Recoveries are total number recovered from creel and fisheries surveys conducted in 1994 and $1995 .{ }^{1}$


[^7]Table 7. Recoveries by location of coded wire tagged kokanee salmon from releases made in 1995. Recoveries are total number recovered from creel and fisheries surveys conducted in $1995^{1}$.

|  |  |  |  |  |  |  |  | CWT recoveries |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Cohort | $\begin{aligned} & \text { Stages } \\ & \text { Exposed } \end{aligned}$ | Exposure Odor | Release Location | Life Stage At Release | $\qquad$ | $\qquad$ | Sherman Cr (MOR/PEA) (n) | Spokane (PEA) (n) | R Hawk Cr (n) | Barnaby <br> (n) | Other <br> (n) |
|  | 1994 | Hatch-swimup | MOR | Sherman Creek | Fry | 40,468 | 3,708 | 0 | 0 | 0 | 0 | 0 |
|  | 1994 | Alevin-swimup | MOR | Chamokane Creek | Fry | 10,943 | 386 | 0 | 0 | 0 | 0 | 0 |
|  | 1993 | -- -- | NONE NONE NONE | Bamaby Creek ${ }^{2}$ <br> Spokane River Kettle Falls Net Pen ${ }^{3}$ | Smolt <br> Smoh <br> Smolt | $\begin{gathered} 21,534 \\ 37,654 \\ 7,343 \end{gathered}$ | $\begin{aligned} & 625 \\ & 987 \\ & 429 \end{aligned}$ | $\begin{aligned} & 1 \\ & 5 \\ & 5 \end{aligned}$ | $\begin{gathered} 56 \\ 80 \\ 0 \end{gathered}$ | $\begin{gathered} 96 \\ 133 \\ 1 \end{gathered}$ | $\begin{gathered} 43 \\ 26 \\ 0 \end{gathered}$ | $\begin{aligned} & 1 \\ & 0 \\ & 0 \end{aligned}$ |
|  | 1993 | Hatch-swimup | $\begin{gathered} \text { MOR } \\ \text { PEA } \end{gathered}$ | Kettle Falls Net Pen Kettle Falls Net Pen | Smolt Smolt | $\begin{aligned} & 17,103 \\ & 23,183 \end{aligned}$ | $\begin{aligned} & 1,097 \\ & 1,283 \end{aligned}$ | $\begin{gathered} 11 \\ 6 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0 \\ & 1 \end{aligned}$ | $\begin{aligned} & 2 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ |
| $\boldsymbol{O}$ | 1993 | Alevin-swimup | MOR | Kettle Falls Net Pen | Smolt | 21,068 | 912 | 4 | 0 | 0 | 0 | 1 |
|  | 1993 | Hatch-swimup and smolt | $\begin{gathered} \text { MOR } \\ \text { PEA } \end{gathered}$ | Sherman Creek Sherman Creek | Smolt Smolt | $\begin{gathered} 16,576 \\ 124,906 \end{gathered}$ | $\begin{gathered} 280 \\ 6,398 \end{gathered}$ | $\begin{aligned} & 110 \\ & 388 \end{aligned}$ | $\begin{aligned} & 2 \\ & 6 \end{aligned}$ | $\begin{aligned} & 12 \\ & 28 \end{aligned}$ | $\begin{gathered} 4 \\ 13 \end{gathered}$ | $\begin{gathered} 2 \\ 12 \end{gathered}$ |
|  | 1993 | Alevin-swimup and smolt | $\begin{gathered} \text { MOR } \\ \text { PEA } \\ \hline \end{gathered}$ | Sherman Creek Sherman Creek | Smoh Smolt | $\begin{array}{r} 51,455 \\ 1,560 \\ \hline \end{array}$ | $\begin{gathered} 2,088 \\ 20 \\ \hline \end{gathered}$ | $\begin{gathered} 109 \\ 4 \\ \hline \end{gathered}$ | $\begin{aligned} & 3 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 3 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 5 \\ & 0 \\ & \hline \end{aligned}$ | 3 0 |

[^8]Table 8. Recoveries by location of kokanee salmon with fin clips. These fish were exposed to synthetic chemicals in 1993, held at the Spokane Tribal Hatchery until July 1993, when they were released into the Spokane Arm of Lake Roosevelt as residualized smolts. Recoveries were made from September to November 1993, 1994 and 1995.


[^9]Little Falls Dam (phenethyl alcohol scented) and 59 (13\%) at other locations (Tables 4-8). In contrast, 626 ( $85 \%$ ) phenethyl alcohol exposed fish were recovered at phenethyl alcohol scented waters (Spokane River and Sherman Creek), 33 (5\%) fish at Sherman Creek when only morpholine was present and 81 (10\%) at other locations (Tables 4-8).

To determine which life stage was most sensitive to synthetic chemical imprinting, different groups of fish were exposed at different life history stages (see Tilson et al. 1994 for a detailed description of the origins of these fish). Although too few fish released as fry were recovered to assess imprinting effectiveness, we were able to assess imprinting effectiveness with fish released as fingerlings (smolts). Table 9 shows that there were no significant differences in distribution between exposed fish returning to their exposure odor and fish returning to other locations ( $p>0.05$ ). The fish which were exposed between hatch and swimup and/or smolt stages showed a significant difference in distribution of chemically exposed fish returning to their exposure odor versus to other locations ( $\mathrm{p}<0.05$ ) (Table 9).

### 3.1.2 Captive Brood Fish (Unexposed) Recoveries

Of the 115,604 CWT fish that were not exposed to any chemical and released in the Spokane River as smolts, a total of 244 were recovered. Of these recoveries, $32.8 \%(n=80)$ were recovered at Little Falls Dam, $54.1 \%(n=133)$ were recovered at Hawk Creek, $10.7 \% \quad(n=26)$ were recovered at Bamaby Creek and $2 \%(\mathrm{n}=5)$ were recovered at Sherman Creek (Table 7). Of the 21,534 fish which were released in Bamaby Creek as smolts, a total of 197 fish were recovered (Table 7). Of these, $21.8 \%(\mathrm{n}=43)$ were recovered at Barnaby Creek, $48.7 \%(\mathrm{n}=96)$ at Hawk Creek, $28.4 \%(\mathrm{n}=56)$ at the Spokane River and $0.5 \%(\mathrm{n}=1)$ at Sherman Creek (Table 7). Of the 29,927 unexposed fish released from the Kettle Falls Net Pen as smolts, a total of 9 fish were recovered. Of these, $88.9 \%(n=8)$ were recovered at Sherman Creek, $11.1 \%(n=1)$ were recovered at Hawk Creek. None were recovered at other locations (Table 6, 7). Of the 26,489 unexposed fish released at Sherman Creek as smolts, 1 of 1 (100\%) was recovered at Sherman Creek (Table 5).

Table 9. Statistical comparison of the number of kokanee salmon (1990-1993 cohort) recovered from 1992 to 1995 which were exposed to either morpholine (MOR) or phenethyl alcohol (PEA) and recovered at the morpholine egg collection site, phenethyl alcohol egg collection site or at other sites from 1992 to 1995. Separate tests were made for each condition. The null hypothesis stated that there was no difference in the distribution of the two set of fish exposed to different odors at a particular life history stage. The null hypothesis was rejected if $\mathbf{p} \leq 0.05$. An asterisk signifies these groups of fish homed to their exposure odor.

| Stage(s) Exposed | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | TOTAL RECOVERED.AT |  |  |  |  |  | Chi <br> Square |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | MOR <br> (\#) | $\begin{gathered} \text { MOR } \\ (\%) \\ \hline \end{gathered}$ | PEA <br> (\#) | PEA <br> (\%) | Other <br> (\#) | Other (\%) |  |
| Eyed egg | MOR | 3 | 50 | 2 | 33 | 1 | 17 | $\mathrm{x}^{2}=0.64$ |
|  | PEA | 0 | 0 | 2 | 67 | 1 | 33 | $\mathrm{p}=0.82{ }^{\text {t }}$ |
| Hatch | MOR | 29 | 52 | 6 | 11 | 21 | 38 | $\mathrm{x}^{2}=3.57$ |
|  | PEA | 1 | 3 | 26 | 70 | 10 | 27 | $\mathrm{p}=0.05{ }^{\text {* }}$ |
| Alevin | MOR | 18 | 47 | 12 | 32 | 8 | 21 | $\mathrm{x}^{2}=9.17$ |
|  | PEA | 4 | 9 | 40 | 85 | 3 | 6 | $\mathrm{p}<0.01{ }^{*}$ |
| Hatch through swimup | MOR | 11 | 85 | 0 | 0 | 2 | 15 | $\mathrm{x}^{2}=7.46$ |
|  | PEA | 0 | 0 | 6 | 75 | 2 | 25 | $\mathrm{p}<0.01{ }^{*}$ |
| Alevin through swimup | $\begin{aligned} & \text { MOR } \\ & \text { PEA } \end{aligned}$ | $\begin{aligned} & 4 \\ & 0 \end{aligned}$ | $\begin{gathered} 80 \\ 0 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1 \\ & 0 \end{aligned}$ | $\begin{gathered} 20 \\ 0 \end{gathered}$ | $\begin{aligned} & \mathrm{n} / \mathrm{a} \\ & \mathrm{n} / \mathrm{a} \end{aligned}$ |
| Swimup | MOR | 19 | 61 | 7 | 23 | 5 | 16 | $\mathrm{x}^{2}=6.53$ |
|  | PEA | 9 | 21 | 30 | 71 | 3 | 7 | $\mathrm{p}<0.05{ }^{\text {* }}$ |
| Fry | $\underset{\text { PEA }}{\text { MOR }}$ | $\begin{aligned} & 4 \\ & 2 \end{aligned}$ | $\begin{aligned} & 40 \\ & 25 \end{aligned}$ | $\begin{aligned} & 5 \\ & 3 \end{aligned}$ | $\begin{aligned} & 50 \\ & 37 \end{aligned}$ | $\begin{aligned} & 1 \\ & 3 \end{aligned}$ | $\begin{aligned} & 10 \\ & 38 \end{aligned}$ | $\begin{gathered} \mathrm{x}^{2}=0.235 \\ \mathrm{p}=0.64 \end{gathered}$ |
| Smolt | MOR | 46 | 82 | 10 | 18 | 0 | 0 | $\mathrm{x}^{2}=66.38$ |
|  | PEA | 22 | 15 | 120 | 80 | 9 | 6 | p<0.01 ${ }^{\text {* }}$ |
| Hatch through swimup and Smolt | MOR PEA | $\begin{gathered} 110 \\ 0 \end{gathered}$ | 85 0 | $\begin{gathered} 2 \\ 394 \end{gathered}$ | $\begin{gathered} 2 \\ 88 \end{gathered}$ | 18 53 | 14 | $\begin{aligned} & x^{2}=263.03 \\ & \mathrm{p}<0.05 t^{*} \end{aligned}$ |
| Alevin through swimup and Smolt | $\begin{gathered} \text { MOR } \\ \text { PEA } \\ \hline \end{gathered}$ | 109 | 89 | $\begin{aligned} & 3 \\ & 4 \\ & \hline \end{aligned}$ | $\begin{gathered} 2 \\ 100 \\ \hline \end{gathered}$ | $\begin{gathered} 11 \\ 0 \\ \hline \end{gathered}$ | $\begin{aligned} & 9 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{gathered} \mathrm{x}^{2}=25.01 \\ \mathrm{p}<0.01 t^{*} \end{gathered}$ |

$t$ Chi square values were calculated for these groups even though the rule staung,"no expected value may be less than 5 " was violated.

### 4.0 DISCUSSION

This study has important implications for management of the Lake Roosevelt kokanee fishery. The results of the CWT investigations continue to show that kokanee can be successfully imprinted to artificial odors - morpholine and phenethyl alcohol - as juveniles from hatch to swimup and again as smolts. Fish double exposed to synthetic chemicals at alevin/swimup and smolt stages had the highest rate of homing to egg collection sites. In addition, the results of the present study continue to support results of previous investigations which show that fish released at the smolt/residualized smolt stage are recovered in greater numbers at egg collection sites than fish released as fry (Tilson et al. 1994, 1995). This information will be helpful in developing management strategies for the Lake Roosevelt kokanee fishery.

### 4.1 Imprinting Investigations

Results of laboratory imprinting investigations reported by Scholz et al. (1993) and Tilson et al. $(1994,1995)$ were similar to the results obtained from field investigations with imprinted, coded wire tagged kokanee released and recovered in Lake Roosevelt. These results show that:

1) Chemical imprinting coincided with elevated thyroxine levels (Scholz et al. 1993, Tilson et al. 1994, 1995). The group that had the highest whole body thyroxine content (swimup stage) also had the highest percentage of fish that were reliably attracted to their exposure odor as sexually mature 2 or 3 year old adults in behavioral tests conducted in a Y-maze in 1993 and 1994 (range 73-93\%). In present field experiments in Lake Roosevelt, fish exposed at swimup exhibited $66 \%$ homing. Recently hatched eggs and alevins also had relatively high thyroxine content and displayed $69 \%$ and $81-87 \%$ homing respectively in the Y-maze test. In the field, these fish displayed 61 and $66 \%$ homing respectively. The other group that displayed accurate homing in the Y-maze test were fish that were exposed to the synthetic chemicals at the smolt stage. As adults in the Y-maze test, this group homed to their exposure odor $67 \%$ of the time. In the field, these fish displayed $61 \%$ homing. Pre-eyed eggs, eyed eggs and four fry stages all had relatively low thyroxine content at the time they were exposed to synthetic chemicals and displayed poor homing ability in laboratory experiments (<32\%) (Scholz er al. 1993, Tilson et al. 1994,
1995). Fish exposed at these stages were all released as fry and were not recovered in field experiments so no comparison could be made between field and laboratory results for these groups of fish.
2) The groups which had the highest percent homing and were recovered in the greatest numbers were groups which were double-exposed to synthetic chemicals. These fish were exposed from hatch to swimup and again at the smolt stage. Morpholine exposed fish displayed $85 \%$ homing and phenethyl alcohol exposed fish displayed $88 \%$ homing.
3) Fish exposed to synthetic chemicals were recovered in greater numbers (\% recovered) and displayed higher homing ability (\% homing) to egg collection sites than fish that were not exposed to synthetic chemicals (Table 10). For example, $80 \%$ of chemically exposed fish released as smolts homed correctly to Sherman Creek with a recovery rate of $0.3 \%$. Even though $100 \%$ of the unexposed fish homed correctly, the recovery rate was only $0.01 \%$. At the Spokane River, $53 \%$ of the chemically exposed fish released as smolts homed correctly, while only $33 \%$ of the unexposed fish homed correctly. In Bamaby Creek, there were no recoveries from chemically exposed fish releases. However, these fish were all released as fry. At Bamaby Creek, $22 \%$ of the unexposed fish released as smolts homed correctly with a $1.1 \%$ recovery rate. However, it is not possible to compare the two groups of fish released at Barnaby Creek because all of the unexposed fish were smolt releases while the chemically exposed fish were fry releases. Both chemically exposed and unexposed fish from the Kettle Falls net pens displayed 0\% homing to the net pen sites and displayed only 0.04 and $0.02 \%$ recovery rates (Table 10). However, these fish were recovered at Sherman Creek cove which is less than 1 mile downstream from the net pen sites.
4) Yearling (smolt) releases continued to provide better adult recoveries than fry releases. Although fish were released at different life stages, and there were approximately equal fry and smolt releases ( 42 and $58 \%$ respectively), almost all of the recoveries $(99 \%)$ have come from fish released into the reservoir as smolts while only $1 \%$ of the recoveries were from fish released as fry.

Table 10. Summary of chemically exposed - v - unexposed coded wire tagged/fin clipped kokanee salmon (1991-1993 cohorts) homing to release sites from 1993 to $1995 .{ }^{1}$

| Release <br> Location | \# Released |  |  | \# Recovered |  | homing ${ }^{1}$ | \% recovered ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | @ home stream | @ other locations |  |  |
|  | Fry | Smolt | Total |  |  |  |  |
| Sherman Creek |  |  |  |  |  |  |  |
| Chem Exposed | 431,722 | 321,699 | 753,421 | 743 | 185 | 80\% | 0.30\% |
| No Chem | 10,130 | 26,489 | 36,619 | 4 | 0 | 100\% | 0.01\% |
| Spokane River |  |  |  |  |  |  |  |
| $\underset{\sim}{\omega} \quad$ Chem Exposed | 75,272 0 | 23,392 37,654 | 98,664 37,654 | $\begin{aligned} & 71 \\ & 80 \end{aligned}$ | $\begin{aligned} & 135 \\ & 243 \end{aligned}$ | $\begin{aligned} & 53 \% \\ & 33 \% \end{aligned}$ | $\begin{aligned} & 0.9 \% \\ & 0.9 \% \end{aligned}$ |
| Blue Creek |  |  |  |  |  |  |  |
| Chem Exposed | 0 | 18,487 | 18,487 | 14 | 24 | 58\% | 0.1\% |
| No Chem | 0 | 0 |  | 0 | 0 | 0\% | 0.0\% |
| Bamaby Creek |  |  |  |  |  |  |  |
| Chem Exposed | 21,784 | ${ }_{0}^{0}$ | 21,784 | ${ }^{0}$ | ${ }^{0}$ | 0\% | 0.0\% |
| No Chem | 0 | 21,534 | 21,534 | 43 | 196 | 22\% | 1.1\% |
| Kettle Falls Net Pens Chem Exposed | 0 | 61,354 | 61,354 | 0 |  | 0\% | 0.04\% |
| No Chem | 0 | 29.927 | 29.927 | 0 | 266 | 0\% | 0.02\% |

1 Percent homing = total number captured in Lake Roosevelt in "home stream" + total number recovered in Lake Roosevelt.
2 Percent recovered $=$ total number recovered + total number smolts released (since $>99 \%$ of fish returning were from fish released as smolts).
5) Table 11 shows that catch per unit effort at egg collection sites (Sherman Creek and Spokane River) increased from 1994 to 1995. This means that although more effort was expended in 1995, the number of kokanee which were recovered was still 6 times higher than it was in 1994.

There are two possible explanations for the low number of recoveries of adult kokanee which were released as fry. One reason may be that fry are not able to avoid predators as well as smolt size fish. Walleye predation is known to occur when kokanee fry are released, as evidenced by observations of CWT kokanee fry in stomachs of walleye collected at release sites (Thatcher et al. 1993). Additionally, walleye collected from the reservoir were occasionally reported containing salmonids, presumably kokanee or rainbow trout in their stomachs (Peone et al. 1990; Griffith and Scholz 1991; Thatcher et al. 1993). In 1995, we observed large walleye (range $500-720 \mathrm{~mm}$ and 4.2 kg ) following schools of mature 2 year old kokanee during their spawning migration. One of those walleye ( $685 \mathrm{~mm}, 4 \mathrm{~kg}$ ), had swallowed a 2 year old adipose clipped kokanee. Although walleye appeared to be preying on kokanee opportunistically during the spawning migration, the numbers seen during these migrations were relatively low ( $\mathrm{n}=8$ at Little Falls and $n=6$ at Sherman Creek). It is our impression that only walleye $>600 \mathrm{~mm}$ could handle a 2 year old kokanee. We never saw walleye where 3 and 4 year old fish were congregating.

The second explanation for the poor adult returns of fish released as fry is that there may be entrainment from Lake Roosevelt when fish are released as fry. Thatcher et al. (1993) found a significant number of kokanee ( 721 fish) at Rock Island Dam between April and August 1991. In addition, Tilson et al. $(1994,1995)$ found that kokanee underwent partial smoltification and exhibited an increased downstream orientation and migratory activity in late winter and early spring. Fish released as $0+$ fry ( 8 months old) in July could undergo smoltification during the following spring, and the higher flows could be sufficient to stimulate downstream displacement causing entrainment through Grand Coulee Dam. In the present study, biologists at both Rocky Reach and Rock Island Dams were looking for adipose clipped kokanee at the time of the regular smolt monitoring program (April 1 to August 31). Although no kokanee were seen at either of these facilities in 1995, it could be that they did not show up at the dams because of turbineinduced mortality. Skaar et al. (1996) investigated fish entrainment through Libby Dam in Montana from December 1990 to June 1994. Skaar et al. found that $81 \%$ of kokanee captured at Libby Dam had turbine-induced injuries, and $49 \%$ of these injuries were termed lethal, soon-to-be-lethal, or prolonged and damaging injuries. Therefore, it is possible that kokanee are being

Table 11. Catch per unit effort (CPUE) for kokanee salmon recovered at Kettle Falls and the Spokane River by electroshocking from September 1 to November 30, 1989 through 1995.

entrained through Grand Coulee Dam but are not making it to either of the collection facilities downstream.

Skaar et al. (1996) found that during a period of continuous monitoring at Libby Dam from January 1992 to January 1993, kokanee entrainment from Lake Koocanusa was estimated to be between 1.12 and 4.36 million kokanee with most fish being lost in December and January during drawdown for power production. The low estimate assumed that the rate of entrainment during unsampled periods was the mean of all measurements for the year. The high estimate assumed that rates from biweekly sampling applied to unsampled days preceding and following the netting session. Losses of $0+$ fish were highest in January, with a maximum of 1,246 fish collected on one date. Another time that entrainment rates were high was in spring between late April and early July when withdrawal depth was shallow, and discharge high for sturgeon/salmon flows. Most of these fish (93\%) were 0+ fish. In 1992, peaks of entrainment of age 1+ fish were from May - June and November - December. Age 2+ and older fish were entrained from June to October with peak numbers lost in September. Skaar et al. estimated the population of kokanee in Libby Reservoir to be about 4.8 million fish from January 1992 to January 1993 and the entrainment to be 1.15 to 4.47 million using the low and high estimates of entrainment respectively. Therefore, the loss of kokanee from the reservoir was estimated at a minimum of $\mathbf{2 3 \%}$, and could have been as high as $92 \%$ in 1992.

Due to lack of funding to effectively monitor entrainment from Lake Roosevelt in the present study, we were unable to estimate the number of kokanee being entrained from the reservoir. However, since $<1 \%$ of fish released as fry were recovered as adults, we have been phasing out fry releases from the hatcheries. Instead we have been releasing more residualized smolts into the reservoir since 1994. This release strategy has probably reduced the amount of entrainment occurring from Lake Roosevelt since these fish are being allowed to residualize at the hatchery.

It is our perception that in years when there is low water retention time (WRT) and low reservoir elevation in Lake Roosevelt, entrainment is higher than in years when there is high WRT and high elevations. In 1991, Thatcher et al. (1993) reported that 721 kokanee were collected at Rock Island Dam's passage facility and an estimate of 25,221 fish were lost over Grand Coulee Dam. During that year the reservoir was drawn down to $1,235 \mathrm{ft}$ with a water retention time of 18 days (Figure 2). In 1995, the reservoir was drawn down $1,259 \mathrm{ft}$ with a water retention time of 40 days and there were no kokanee seen at the fish Passage Facility at Rock Island (Figure 2). Higher flows in 1991 triggered kokanee smoltification and subsequently


Figure 2. Lake elevation (feet) and water retention time (days) in Lake Roosevelt during 1991 and 1995.
entrainment, while the lower flows in 1995 together with the greater number of residualized smolts released, resulted in less entrainment from Lake Roosevelt. This is reflected in the high CPUE and high numbers of kokanee seen in 1995 when WRT and water elevations were low compared with low CPUE and low numbers of kokanee seen in 1991 when WRT and water elevations were high (Figure 2, Table 11).

The Colville Confederated Tribe is currently monitoring entrainment from Grand Coulee Dam by using hydroacoustic sonars and vertical gill nets in the forebay of Grand Coulee Dam. This will provide a more direct method of monitoring entrainment from Lake Roosevelt. Results of their investigations, combined with continued assessment of kokanee counted at Rocky Reach and Rock Island Dams will provide better information to assess both entrainment and fish losses in future years.

### 4.2 Kokanee Harvest

The second objective was to assess the best times and locations to release kokanee in order to improve angler harvest and returns to egg collection sites. Only 19 adipose clipped kokanee were observed in the creel and 8 of these had coded wire tags. Too few tagged kokanee were obtained from anglers to assess harvest. Therefore, we recommend intensifying the creel in order to see more adipose clipped/CWT fish.

The Lake Roosevelt Monitoring Program's year round creel survey was designed to expend approximately the same amount of effort each year [(40 hours/week for 50 weeks/year for each clerk $(n=3)$ ]. This effort was approximately 6,000 hours creeled each year. In 1995, the Monitoring Program's creel expended 9,055 hours. The kokanee harvest was estimated at 32,353 fish. There were 118 kokanee observed by the creel clerks. Out of these, only one fish was fin clipped and was not coded wire tagged.

In 1995, the site specific augmented creel conducted by EWU was designed to observe as many adipose clipped/CWT kokanee as possible. Therefore, we targeted the locations and times when anglers would be most likely to catch kokanee. We recognize that this biased the number of kokanee observed and we did not use this number to discuss trends between years. However, the number of hours of effort in the augmented creel was relatively low in comparison to the number of kokanee observed ( 282 kokanee for 37 hours creeled). Therefore, we recommend using this method in future years to try to collect more adipose clipped/CWT kokanee. Not only would it allow us to collect more heads for information on age and growth using CWT analysis, it would also allow us to collect more scales for age and growth determination.

Table 12 shows the mean length and weight of 2,3 and 4 year old fish captured in 1995. These sizes can be compared to kokanee caught in previous years (Table 13). The mean length of age 2 fish decreased from 350 mm in 1993 to $321 \pm 30 \mathrm{~mm}$ in 1995. However, the mean weight stayed approximately the same. The mean length and weight of age 3 fish in 1994 was 454 mm and $1,181 \mathrm{~g}$ compared to 389 mm and 650 g respectively in 1995 (Tables 12,13 ). However, these data are misleading. The reason for the decrease in size could be because during sampling in autumn 1995, the investigators collected heads from the smaller spawners which they thought to be 2 year olds (less than approximately 370 mm ). When a larger fish was caught, the investigators would transfer it to holding traps at Sherman Creek or Little Falls for future egg collection. These fish, thought to be primarily 3 and 4 year olds, had an average length and weight of 432 mm and $1,141 \mathrm{~g}(\mathrm{n}=385)$, which is close to the size of 3 or 4 year old CWT fish in 1994 (Table 13). Unfortunately, many of these fish (approximately 250) were lost due to otter predation and other factors before coded wire tags could be extracted. Therefore, the information gathered from age 3 and 4 year old CWT fish sizes is not accurate because it represents only the smaller sized fish in those age classes.

Table 14 shows the ages of kokanee collected in 1995 based on scale analyses. This information shows that the age $3+$ fish (which would be 4 year old spawners) averaged 472 mm and $1,180 \mathrm{~g}$. In addition, Table 15 shows that when fish are sorted according to length ( $<371$ $\mathrm{mm}=$ age 2; 371-469 mm =age $3 ;>470 \mathrm{~mm}=$ age 4), the 3 year old fish averaged 416 mm and 867 g and the 4 year old fish averaged 516 mm and $1,403 \mathrm{~g}$. Figure 3 shows a length frequency distribution of age 2,3 , and 4 year old CWT fish. We feel these data (Table 14, 15) portray a more accurate assessment of the sizes of kokanee in Lake Roosevelt than do the information from CWT fish presented in Tables 12 and 13. The distribution of lengths of age 2 and 3 kokanee show some overlap. However, as mentioned above, the data for age 3 fish could be biased toward small fish because of the loss of the larger spawners.

With the CWT kokanee, it is possible to confirm kokanee growth rates in Lake Roosevelt. However, without complete information from all kokanee collected, we cannot assess these growth rates. Lost tags also means other lost information about the effectiveness of the release strategies. We recommend building a new land-based holding system or using inreservoir, predator proof traps to hold these large spawners until egg collection (See Section 4.3).

Table 12. Mean length ( $\pm$ SD), weight ( $\pm$ SD), range of length and number in sample for kokanee salmon collected in 1995. Ages are based on coded wire tag analyses.

|  | Length <br> $(\mathbf{m m})$ | Weight <br> $(\mathbf{g})$ | Range <br> $(\mathbf{m m})$ | Number <br> in sample |
| :--- | :---: | :---: | :---: | :---: |
| Age 2 | $321 \pm 30.4$ | $486 \pm 141$ | $200-434$ | 1,163 |
| Age 3 | $389 \pm 32.2$ | $650 \pm 116.4$ | $310-462$ | 46 |
| Age 4 | $505 \pm 57.2$ | $806 \pm 176.1$ | $445-559$ | 3 |

Table 13. Summary of age, mean length, mean weight of coded wire tagged kokanee caught from 1992 to 1995.

| Age at recovery | $\begin{gathered} \text { Year } \\ \text { recovered } \end{gathered}$ | Total length $(\mathrm{mm})$ | $\begin{gathered} \text { Weight } \\ (\mathrm{g}) \end{gathered}$ | Number in sample |
| :---: | :---: | :---: | :---: | :---: |
| 2 | 1992 | 320 | 347 | 2 |
|  | 1993 | 350 | 490 | 66 |
|  | 1994 | 337 | 418 | 37 |
|  | 1995 | 321 | 486 | i. 163 |
|  | Mean | 331 | 435 |  |
| 3 | 1993 | 908 | 1,592 | 2 24 |
|  | 1995 | 389 | 650 | 46 |
|  | Mean | 449 | 1,124 |  |
| 4 | 1994 | 447 | 1,508 | 5 |
|  | 1995* | 505 | 806 | 3 |
|  | Mean | 465 | 1,060 |  |

[^10]Table 14. Mean length $( \pm \mathbf{S D})$, weight $( \pm$ SD $)$, range of length and number in sample for kokanee salmon collected in 1995. Ages are based on scale analyses.

|  | Length <br> $(\mathbf{m m})$ | Weight* <br> $(\mathbf{g})$ | Range <br> $(\mathbf{m m})$ | Number <br> in sample |
| :--- | :---: | :---: | :---: | :---: |
| Age 1+ | $219 \pm 35$ | $92 \pm 38$ | $169-260$ | 5 |
| Age $2+$ | $385 \pm 81$ | $437 \pm 207$ | $230-520$ | 62 |
| Age $3+$ | $472 \pm 60$ | $1,180 \pm 559$ | $318-570$ | 113 |

Table 15. Mean length ( $\pm$ SD), weight ( $\pm$ SD), range of length and number in sample for kokanee salmon collected in 1995 from electroshocking. Ages are based on length $<371 \mathrm{~mm}=$ age $2,371-469 \mathrm{~mm}=$ age $3,>470 \mathrm{~mm}=$ age 4 .

|  | Length <br> $(\mathbf{m m})$ | Weight <br> $(\mathrm{g})$ | Range <br> $(\mathrm{mm})$ | Number <br> in sample |
| :--- | :---: | :---: | :---: | :---: |
| Age 2 | $318 \pm 28.4$ | $453 \pm 117.6$ | $116-370$ | 1520 |
| Age 3 | $416 \pm 27.7$ | $867 \pm 216.3$ | $371-468$ | 496 |
| Age 4 | $516 \pm 35.2$ | $1,403 \pm 327.4$ | $470-605$ | 235 |



Figure 3. Length frequency (mm) for coded wire tagged kokanee collected in 1995.

### 4.3 Kokanee Egg Production Estimates

One of the goals of the Northwest Power Planning Council's Fish and Wildlife Plan (NPPC 1995) was to release 1 million residualized kokanee smolts into Lake Roosevelt which would provide a kokanee escapement of 1.45 million eggs. This estimate was based on data collected by Peone et al. (1990).

In 1988 and 1989, Peone et al. (1990) estimated fecundity of Lake Roosevelt kokanee. In 1988, the total length/fecundity regression equation was:

$$
y=4.246(x)-283.321
$$

$$
\begin{aligned}
\text { where: } & =\quad \text { number of eggs for individual female, and } \\
x & =\text { length of individual fish captured in } 1988
\end{aligned}
$$

In 1989, the total length/fecundity regression equation was:

$$
\mathrm{y}=2.902(\mathrm{x})+326.013
$$

where: $y=$ number of eggs for individual female, and $x=$ length of individual fish captured in 1988.

In the present report, we measured the potential total egg production to determine if we were on target to achieve the goals mentioned above. We calculated the potential fecundity of individual age 3 and 4 females with lengths $>370 \mathrm{~mm}$. The length of individual females recovered in autumn 1995 was substituted for ( x ) in the above equations and calculated to determine fecundity ( y ) of that individual. The fecundity of each individual was summed to provide a total estimate of the eggs that could have potentially been produced in 1995 using both the 1988 regression equation $\left(r^{2}=0.107\right)$ and the 1989 regression equation $\left(r^{2}=0.416\right)$ (Table 16).

Based on the data collected from Peone et al. (1990), the current estimate was about 523,000 eggs( $n=318$ females) based upon release of approximately 244,995 residualized smolts, so we are on target to achieve this goal (Table 16). A release of 224,995 fish is about $1 / 4$ of our release goal, and it could have provided $1 / 3$ of the egg goal. Therefore, we believe that there is a good probability that we can achieve our egg goal of 1.45 million if 1 million residualized smolt are stocked.

| Table 16. | Estimates of kokanee salmon fecundity based on females ( $>\mathbf{3 7 0} \mathrm{mm}$ ) <br> recovered from September 1 to November 30, 1995. Fecundity was <br> estimated using the 1988 regression equation and the 1989 regression <br> equation (Peone et al. 1990). |
| :---: | :---: | :---: |
| $\frac{\text { Total \#eggs using }}{1989 \text { regression }}$equation | (n) |
| 1988 regression |  |
| equation |  |

This information is encouraging from the standpoint of collecting a sufficient number of eggs to support hatchery operation. However, one of the problems associated with collecting eggs in 1995 was that when the females were electroshocked and moved to holding pens at Sherman Creek cove, the eggs quit developing and didn't ripen. This could have been due to the fact that Sherman Creek water is approximately $10^{\circ} \mathrm{F}$ colder than Lake Roosevelt temperatures. Therefore, one of our recommendations is to induce spawning with pituitary hormones or steroids.

Investigators have successfully induced spawning in a number of species of fish (Hamman 1985a,b 1986; Ako et al. 1994). In addition, several studies have shown that Pacific salmon can be induced to ovulate ahead of normal time using pituitary hormones and/or steroid hormones (Jalabert 1976; Hunter et al. 1978; Jalabert et al. 1978; Donaldson et al. 1981 a,b,c; Sower et al. 1982). Fitzpatrick et al. (1994) have shown that injections of $5 \mu \mathrm{~g}$ luteinizing hormone releasing hormone analog (LHRHa) followed by another injection of $5 \mu \mathrm{~g}$ induces spawning in coho salmon in 3 days or less. Therefore, we recommend inducing ovulation of kokanee salmon with LHRHa in an experimental lot of fish.

### 4.4 Management Recommendations

From information gathered in this report and from the previous investigations of Tilson et al. (1994, 1995), it was concluded that kokanee salmon released as yearlings (smolts) are recovered in greater numbers than fish released as fry. In addition, chemically exposed fish home more accurately and in greater numbers to egg collection sites. Based upon the results of our investigations, we make the following recommendations for managing Lake Roosevelt kokanee:
(1) To achieve the escapement goal of 1.45 million eggs, more fish must be released into the reservoir. For this to be accomplished, the following must be done:
a) A new production well capable of delivering 2-4 CFS of additional flow would need to be drilled at the hatchery. This would allow approximately 500,000 fish to be raised to residualized smolt stage instead of the current 100,000 to 300,000 fish. [Note: At present, the hatchery can carryover about 300,000 fish until late March or April. However that is the height of smoltification. The hatchery can currently retain 130,000 to 150,000 fish for release in June.] See Tilson et al. (1994) for more discussion on this point.
b) Provide a net pen system that is moveable with three anchoring sites so the pens can be moved if there are drawdowns. One of these sites should be as far into Sherman Creek cove as possible for best site imprinting. The second should also be in Sherman Creek cove, but further out toward the reservoir where these fish could be moved in case of drawdowns. The third site should be in the reservoir off and slightly downstream from Sherman Creek in a place deep enough to withstand a minimum pool elevation of 1,208 feet.
c) We recommend placing 50,000 zero-age fish into net pens from October until the following June when they have residualized (Appendix D). Since there were only 32 recoveries out of 91,281 smolts released from the Kettle Falls Net Pens, we recommend building new net pens in Sherman Creek cove and discontinuing releases from Kettle Falls Marina. The hatcheries should hold as many kokanee smolts as possible before transferring any fish to the net pens since the best returns came from hatchery released fish. However, if there is not enough room at the hatcheries to hold these fish until release, we recommend using the Kettle Falls Net Pens until the Sherman Creek net pens have been constructed.
d) Initiate experiments with induction of spawning by injecting females with LHRHa or other inducing hormones.
(2) Make effort to monitor Rufus Woods Reservoir for entrained kokanee.
a) Continue to get information from Rock Island, Rocky Reach and McNary Dams on adipose clipped kokanee.
b) Monitor Rock Island Dam earlier in the year, preferably starting in January, instead of May which is the standard time in which the Fish Passage Center's Smolt Monitoring Program starts. We feel that kokanee are being entrained during the winter months through Grand Coulee and Rock Island Dams as well as in early spring.
c) Conduct electrofishing surveys in Rufus Woods reservoir.
d) Studies on entrainment being conducted by the Colville Confederated Tribes need to continue throughout the year. They have been studying entrainment with hydroacoustics at Grand Coulee Dam in the spring, as
well as initiating gill net surveys in the forebay of Grand Coulee Dam. According to Skaar et al. (1996), most of the kokanee entrainment occurs in the winter months in Lake Koocanusa (See Section 4.2).

Provide more adequate adult holding facilities at Sherman Creek so age 3 and age 4 fish are able to be kept for spawning purposes. In 1995, approximately 250 age 3 and 4 fish were lost owing to river otter predation in traps. We recommend using a land based system of tanks to ensure safety. Additionally, if it became necessary to hold large numbers of fish, we recommend building 4 or 5 otter proof wire holding cages in the reservoir so we can safely hold adults until spawning. That way, we could separate fish that were at different stages of maturity and/or separate sexes in different wire cages.
(4) Modify hatchery ladder entrance to attract spawning kokanee. There are two potential problems with the ladder trap system at Sherman Creek. The hatchery ladder enters Sherman Creek at a right angle instead of parallel to the stream and requires that the fish jump about 8 inches through a narrow slot in order to enter the first pool of the ladder. This construction may cause the fish to avoid the ladder. This is evidenced by the fact that only 3 fish were seen in the ladder. Bell (1986) noted that it is generally more effective to introduce attraction water (a) in parallel rather than perpendicular to the main current, (b) through a bottom diffusing area, so that fish jumping could be reduced to a minimum (kokanee are not known for their jumping). The second problem is that fish do not seem to want to move from the cove at the mouth of the creek, up into the creek. For example, about 70 fish were observed in Sherman Creek near the ladder entrance and about 1,500 fish were observed 100 feet downstream in the cove of the stream. There are two potential reasons these fish are not moving past the cove. One reason may be that the cove provides fish with lots of cover (logs, brush) and is about 12 ft deep whereas the creek is shallow ( $\approx 1 \mathrm{ft}$ ) with no cover. It could be that fish do not want to move from the covered, protected area of the cove to the open creek. The other reason may be that there is a temperature barrier between the cove and the creek. In past years, there has been a $15^{\circ} \mathrm{F}$ difference from water in the cove and water in the main creek (M. Combs, Hatchery Manager, Sherman Creek Hatchery). To correct this, we recommend initiating a feasibility study to determine the most effective method of trapping fish. We have identified the following recommendations:
a) Re-engineer the hatchery ladder so the fish can swim up through the steps instead of jumping by punching holes in the steps. We also recommend submerging the ladder entrance to eliminate the jump at the beginning of the ascent.
b) Construct a submerged culvert system to divert the fish from the cove, up to the ladder. To correct the temperature barrier problem, cove water, which is warmer, could be pumped into the fish ladder so it could mix with the colder creek water. Fish could then be lured to the warmer mix of water coming from the culvert, swim up the culvert directly into the ladder.
c) Construct a Merwin trap in the cove to trap fish before they ascended into the creek.
(5) Set up an egg collection site at Hawk Creek since 286 fish were recovered at that site.
(6) We encourage fish managers to locate alternative stocks of kokanee, with better genetic adaptations than Lake Whatcom fish for the Lake Roosevelt Program. This should be tested via coded wire tag investigations by a paired release strategy. For example, tag 20,000 fish from the Lake Whatcom brood stock and 20,000 fish from an alternative stock. Both groups would be subject to the same rearing conditions and release strategies. That way, we could determine which stock is recovered in greater numbers and exhibits better homing ability. We also recommend conducting smolt physiology tests for each stock to determine the time and degree of smoltification, so that they can be stocked into Lake Roosevelt as residualized smolts.
(7) Intensify efforts to recapture more CWT fish by both creel surveys and fisheries (electrofishing/gill net/trawl) surveys so we can compare augmented creel data between years.
(8) As part of the Lake Roosevelt Monitoring Program, further assess potential impacts of walleye predation at kokanee release sites. Since smallmouth bass are increasing dramatically in Lake Roosevelt, also examine smallmouth bass predation on kokanee.

## LITERATURE CITED

Ako H., C.S. Tamaru, and C. Lee. 1994. Chemical and physical differences in milkfish (Chanos chanos) eggs from naturally and hormonally induced spawns. Aquaculture 127:157-167.

Bell, M.C. 1986. Fisheries Handbook of Engineering Requirements and Biological Criteria. Fish Passage Development and Evaluation Program, U.S. Department of the Army, Corps of Engineers, Portland, OR. 290 pp.

Griffith J.R. and A.T. Scholz. 1991. Lake Roosevelt Fisheries Monitoring Program. Annual Report 1990 Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 218 pp.

Griffith J.R. and A.T. Scholz. 1991. Lake Roosevelt Fisheries Monitoring Program. Appendices to 1990 Annual report. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 358 pp.

Hamman, R.L. 1985a. Induced spawning of hatchery-reared razorback sucker. The Progressive Fish Culturist 47(3): 187-189.

Hamman, R.L. 1985b. Induced spawning of hatchery-reared bonytail. The Progressive Fish Culturist 47(4):239-241.

Hamman, R.L. 1986. Induced spawning of hatchery-reared Colorado squawfish. The Progressive Fish Culturist 48:72-74.

Jalabert B., F.W. Goetz, B. Breton, A. Fostier, and E.M. Donaldson. 1978. Precocious induction of oocyte maturation and ovulation in coho salmon, Oncorhynchus kisutch. Journal of the Fisheries Research Board of Canada 35: 1423-1429.

Peone, T.L., A.T. Scholz, J.R. Griffith, S. Graves and M.G. Thatcher. 1990. Lake Roosevelt Fisheries Monitoring Program. Annual Report 1988-1989. Prepared by Upper Columbia United Tribes Fisheries Center for Bonneville Power Administration, Portland, Oregon. 234 pp.

Peone, T.L., A.T. Scholz, J.R. Griffith, S. Graves and M.G. Thatcher. 1990. Lake Roosevelt Fisheries Monitoring Program. Annual Report Appendices 1988-1989. Prepared by Upper Columbia United Tribes Fisheries Center for Bonneville Power Administration, Portland, Oregon. 431 pp.

Scholz, A.T., R.J. White, V.A. Koehler and S.A. Horton. 1992. Measurement of thyroxine concentration as an indicator of the critical period for imprinting in kokanee salmon (Oncorhynchus nerku): Implications for operating Lake Roosevelt kokanee hatcheries. Supplement to 1991 Annual report. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon.

Scholz, A.T., R.J. White, M.B. Tilson and S.A. Horton. 1993. Artificial imprinting of Lake Roosevelt kokanee salmon (Oncorhynchus nerka) with synthetic chemicals: measurement of thyroxine content as an indicator of the sensitive period for imprinting to olfactory cues. Annual Report 1992. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 60 p.

Skaar, D., J. DeShazer, L. Garrow, T. Ostrowski and B. Thomburg. 1996. Quantification of Libby Reservoir levels needed to maintain or enhance reservoir fisheries: Investigations of fish entrainment through Libby Dam, 1990-1994. Final Report 1996. Prepared by Montana Fish, Wildlife and Parks for Bonneville Power Administration. Portland, Oregon. 110 pp.

Sower, S.A., C.B. Schreck, and E.M. Donaldson. 1982. Hormone-induced ovulation of coho salmon (Oncorhynchus kisutch) held in seawater and freshwater. Canadian Journal of Fisheries and Aquatic Science 39:627-632.

Thatcher, M.G., J.R. Griffith, A.C. McDowell and A.T. Scholz. (1993). Lake Roosevelt Fisheries Monitoring Program. Annual Report 199 1. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 237 pp.

Thatcher, M.G., J.R. Griffith, A.C. McDowell and A.T. Scholz. (1993). Lake Roosevelt Fisheries Monitoring Program. Appendices to 1991 Annual Report. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon.

Tilson, M.B., A.T. Scholz, R.J. White and H. Galloway. 1994. Thyroid-induced chemical imprinting in early life stages and assessment of smoltification in kokanee salmon hatcheries. 1993 Annual Report. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 156 pp.

Tilson, M.B., A.T. Scholz, R.J. White and J.L. Hendrickson. 1995. Artificial imprinting and smoltification in juvenile kokanee salmon: Implications for operating Lake Roosevelt kokanee salmon hatcheries. 1994 Annual Report. Prepared by Upper Columbia United Tribes Fisheries Research Center for Bonneville Power Administration. Portland Oregon. 127 pp.

Appendix A. Summary of kokanee salmon coded wire tagged at the Spokane Tribal Hatchery In 1994 and 1995. Tagging information includes stage at the time of tagging, mean length (mm) and mean weight (g) at time of tagging, number tagged and number released after retention estimate.

| CWT Code | $\begin{gathered} \text { Date } \\ \text { Tagged } \end{gathered}$ | Stage © Tagging | $\begin{gathered} \text { Mean } \\ \mathrm{Ln} \\ (\mathrm{~mm}) \\ \hline \end{gathered}$ | Mean Wt (a) | $\begin{gathered} \# \\ \text { Injected } \end{gathered}$ | $\begin{gathered} \text { \#Taggec } \\ \text { QCD } \end{gathered}$ | $\begin{aligned} & \text { d \% } \\ & \text { Tagged } \end{aligned}$ | $\%$ Retention 1 | \# CWT Released | Year Released | Stage © |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 62.52.21 | Apr. 94 | smolt | 156 | 36.7 | 11,253 | 11,101 | 98.6\% | 98.9 | 10,979 | 94 | smolt |
| 62.52.22 | Apr-94 | smolt | 156 | 36.7 | 9,568 | 9,435 | 98.6\% | 99.0 | 9,341 | 94 | smolt |
| 62.52.23 | Apr. 94 | smolt | 156 | 36.7 | 11,128 | 10,904 | 98.0\% | 97.7 | 10,653 | 94 | smoll |
| 62.52.24 | Apr. 94 | smolt | 156 | 36.7 | 11,098 | 10,672 | 96.2\% | 97.5 | 10,405 | 94 | smoll |
| 62.52.25 | Apr. 94 | smolt | 157 | 37.8 | 11,168 | 11,029 | 98.8\% | 98.7 | 10,886 | 94 | smolt |
| 62.52.26 | Apr. 94 | smolt | 157 | 37.8 | 11,236 | 11,117 | 98.9\% | 97.7 | 10,861 | 94 | smolt |
| 62.52.21 | Apr. 94 | smolt | 157 | 37.8 | 11,497 | 11,439 | 99.5\% | 98.4 | 11,256 | 94 | smolt |
| 62.52.28 | Apr. 94 | smolt | 167 | 45.1 | 11,709 | 11,512 | 98.3\% | 98.4 | 11,328 | 94 | smolt |
| 62.52.29 | May. 94 | smolt | 167 | 45.1 | 11,242 | 11,139 | 99.1\% | 98.0 | 10,919 | 94 | smolt |
| 62.52.30 | May. 94 | smoit | 167 | 45.1 | 10,899 | 10,836 | 99.4\% | 98.0 | 10,613 | 94 | smolt |
| 62.52.31 | May. 94 | smolt | 168 | 46.2 | 11,255 | 11,169 | 99.2\% | 98.3 | 10,291 | 94 | smolt |
| 62.52.32 | May. 94 | smolt | 168 | 46.2 | 14,786 | 13,671 | 92.5\% | 98.3 | 11,140 | 94 | $s \mathrm{molt}$ |
| 62.52.33 | May. 94 | smolt | 184 | 60.7 | 8,484 | 1,445 | 87.8\% | 98.9 | 1,303 | 94 | smolt |
| 111.2.a | Jun. 94 | fry | 58 | 1.9 | 12,750 | 11,643 | 91.3\% | 84.0 | 9,780 | 94 | fry |
| 111.2 .9 | Jun.94 | fiy | 58 | 1.9 | 11,018 | 10,132 | 92.0\% | 88.0 | 8,916 | 94 | fry |
| 62.52.34 | Jun.94 | fry | 58 | 1.9 | 10,935 | 10,813 | 98.9\% | 91.3 | 10,099 | 94 | fry |
| 62.52.35 | Jun. 94 | fry | 54 | 1.5 | 11,252 | 11,078 | 98.5\% | 91.3 | 10,114 | 94 | fry |
| 62.52.36 | Jun. 94 | fry | 54 | 1.5 | 11,197 | 11,072 | 98.9\% | 91.6 | 10,147 | 94 | fry |
| $62 \cdot 52 \cdot 37$ | Jun. $j$ un. a | fry | 57 | 1.8 | 11,206 | 11, 050 | 98.6\% | 94.8 | 10,475 | 94 | fry |
| 62.52.38 | Jun.94 | fry | 57 | 1.8 | 11,218 | 11,041 | 98.4\% | 94.8 | 10,467 | 94 | fry |

(1) Percent retention is estimated by randomly capturing 500 fish after $10-20$ days and counting the number of fish with tags.

## Appendix A. Continued.

| CWT Code | $\begin{gathered} \text { Date } \\ \text { Tagged } \end{gathered}$ | Stage © Tagging | $\begin{gathered} \text { Mean } \\ \text { Ln } \\ (\mathrm{mm}) \end{gathered}$ | $\begin{gathered} \hline \text { Mean } \\ \text { Wt } \\ \text { (g) } \\ \hline \end{gathered}$ | $\begin{gathered} \# \\ \text { Injected } \end{gathered}$ | $\begin{gathered} \hline \text { \# Tagged } \\ \text { QCD } \\ 1 \\ \hline \end{gathered}$ | Tagged | $\%$ Retention 2 | \# CWT Released 3 | $\begin{gathered} \text { Year } \\ \text { Released } \end{gathered}$ | $\begin{gathered} \text { Stage © } \\ \text { Release } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 22.34.31 | 701095 | fry | 70 | 3.4 | 10,855 | 10,670 | 98.3 | 93.0 | 9,923 | 95 | fry |
| 62.54.38 | Jun.95 | fry | 70 | 3.4 | 11,152 | 10,004 | 89.7 | 93.0 | 10,271 | 95 | fry |
| 62.54.39 | Jun. 95 | fry | 70 | 3.4 | 11,397 | 11,223 | 98.5 | 93.0 | 10,437 | 95 | fry |
| 62.54.40 | Jun. 95 | fry | 70 | 3.4 | 10,172 | 10,571 | 98.2 | 93.0 | 9, 837 | 95 | fry |
| 62.54.48 | Ju1.95 | fry | 75 | 4.1 | 11,329 | 11,281 | 99.6 | 97.0 | 10,943 | 95 | fry |
| 62.52.39 | Jun. 94 | fry | 57 | 1.8 | 1,896 | 1,763 | 98.3 | 94.8 | 1,507 | 95 | smolt |
| 62.52.40 | Jun.94 | fry | 60 | 2.1 | 10,982 | 10,919 | 99.4 | 95.2 | 5,682 | 95 | smolt |
| 62.52 .41 | Jun.94 | fry | 60 | 2.1 | 11,181 | 11,030 | 98.6 | 95.2 | 10,501 | 95 | smolt |
| 62.53.35 | Ju1.94 | fry | 61 | 2.2 | 11,189 | 11,052 | 98.8 | 95.2 | 5,704 | 95 | smolt |
| 62.53.36 | Ju1.94 | fry | 61 | 2.2 | 11,208 | 11, 070 | 98.8 | 95.2 | 5,713 | 95 | smolt |
| ${\underset{\sim}{\omega}}_{\sim}^{\text {W2.53.37 }}$ | Ju1.94 | fry | 61 | 2.2 | 11,218 | 11,144 | 99.3 | 95.2 | 5,789 | 95 | smolt |
| 62.53.38 | Ju1.94 | fry | 61 | 2.2 | 11,114 | 11, 052 | 99.4 | 95.2 | 5, 752 | 95 | smolt |
| 62.53.39 | Ju1. 94 | fry | 57 | 1.8 | 11,187 | 11,154 | 99.7 | 95.2 | 5,806 | 95 | smoll |
| 62.53.40 | Ju1.94 | fry | 57 | 1.8 | 11,194 | 11,151 | 99,6 | 95.2 | 5,836 | 95 | smoll |
| 62.53.41 | Ju1.94 | fry | 52 | 1.4 | 11,180 | 11,102 | 99.3 | 94.3 | 10,293 | 95 | smoll |
| 62.53.42 | Ju1.94 | fry | 52 | 1.4 | 11,228 | 11,131 | 99.1 | 94.3 | 10,197 | 95 | smolt |
| 62.53.43 | Ju1.94 | fry | 52 | 1.4 | 11,243 | 11,172 | 99.4 | 94.3 | 10,244 | 95 | smoll |
| 62.53.44 | Ju1.94 | fry | 52 | 1.4 | 11,238 | 11,151 | 99.2 | 94.3 | 10,228 | 95 | smolt |
| 62.53.45 | J 41.94 | fry | 52 | 1.4 | 11,354 | 11,210 | 98.7 | 94.3 | 10,301 | 95 | smoll |
| 62.53.46 | Ju1.94 | fry | 52 | 1.4 | 11,199 | 11,155 | 99.6 | 94.3 | 10,204 | 95 | smolt |
| 62-53-47 | Jul. 94 | fry | 55 | 1.6 | 11,239 | 11,173 | 99.4 | 95.7 | 10,224 | 95 | smolt |
| 62.53.48 | Jul. 94 | fry | 55 | 1.6 | 11,308 | 11,268 | 99.6 | 96.6 | 10,410 | 95 | smolt |
| 62.53 .49 | Ju1.94 | fry | 55 | 1.6 | 11,286 | 11,215 | 99.4 | 96.6 | 10,363 | 95 | smolt |
| 62.53. 51 | Ju1.94 | fry | 55 | 1.6 | 11,169 | 11,113 | 99.5 | 96.6 | 10,173 | 95 | smolt |
| 62.53.50 | Ju1.94 | fry | 55 | 1.6 | 11,190 | 11,128 | 99.4 | 96.6 | 10,285 | 95 | smolt |
| 62.51.25 | Aug. 94 | fry | 66 | 2.8 | 2,995 | 2,959 | 98.8 | 98.1 | 1,560 | 95 | smoll |
| 62.53.52 | Aug. 94 | fry | 66 | 2.8 | 11,203 | 11,147 | 99.5 | 96.6 | 10,457 | 95 | smoll |
| 62.53.53 | Aug. 94 | fry | 66 | 2.8 | 11,192 | 11,143 | 99.6 | 96.6 | 10,590 | 95 | smolt |
| 62.53.54 | Aug. 94 | fry | 66 | 2.8 | 11,171 | 11,129 | 99.6 | 96.6 | 10,576 | 95 | smoll |

Appendix A. Continued.

| CWT Code | $\begin{gathered} \text { Date } \\ \text { Tagged } \end{gathered}$ | Stage © Tagging | $\begin{gathered} \text { Mean } \\ \text { Ln } \\ (\mathrm{mm}) \\ \hline \end{gathered}$ | Mean Wt (g) | Injected | $\begin{gathered} \text { \# Tagge } \\ \text { QCD } \\ 1 \\ \hline \end{gathered}$ | Tagged | Retention 2 | \# CWT Released 3 | Year Released | Stage © Release |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 62.53.55 | Aug. 94 | 'firy | 66 | 2.8 | 11,197 | 11,128 | 99.4 | 96.6 | 10,380 | 95 | smoll |
| 62.53.56 | Aug-94 | fry | 78 | 4.6 | 11,192 | 11,081 | 99.0 | 96.6 | 10,542 | 95 | smolt |
| 62.53.57 | Aug. 94 | fry | 78 | 4.6 | 11,195 | 11,089 | 99.1 | 96.6 | 10,526 | 95 | smolt |
| 62.53.58 | Aug. 94 | fry | 66 | 2.8 | 11,183 | 11,136 | 99.6 | 97.9 | 10,725 | 95 | smolt |
| 62.53.59 | Aug. 94 | fry | 66 | 2.8 | 11,158 | 11,126 | 99.7 | 97.9 | 10,711 | 95 | smolt |
| 62.51.63 | Nov. 94 | fingerling | 116 | 15 | 11,658 | 11,574 | 99.3 | 95.4 | 10,939 | 95 | smolt |
| $62.51 \cdot 26$ | Dee. 94 | fingerling | 120 | 16.8 | 5,594 | 5,522 | 98.7 | 98.2 | 5,423 | 95 | smolt |
| 62.51.28 | Dee. 94 | fingerling | 120 | 16.8 | 8,375 | 8,301 | 99.1 | 98.9 | 8,210 | 95 | $s \mathrm{molt}$ |
| 62.51.34 | Dee. 94 | fingerling | 120 | 16.8 | 5,483 | 5,443 | 99.3 | 98.9 | 5,383 | 95 | smolt |
| 62.51 .44 | Dee. 94 | fingerling | 116 | 15 | 6,007 | 5,960 | 99.2 | 95.4 | 5,637 | 95 | $s \mathrm{molt}$ |
| 62.51.24 | Jan. 95 | fingerling | 120 | 16.8 | 4,932 | 4,913 | 99.6 | 98.3 | 5,430 | 95 | smolt |
| 62.51.42 | Jan. 95 | fingerling | 120 | 16.8 | 4,712 | 4,669 | 99.1 | 97.5 | 4, 552 | 95 | $s \mathrm{molt}$ |
| 62.51.48 | Jan. 95 | fingerling | 124 | 18.6 | 11,030 | 11 | 00.1 | 94.5 | 10,681 | 95 | smolt |
| 62.51 .49 | Jan. 95 | fingerling | 124 | 18.6 | 8,081 | 8,056 | 99.7 | 98.0 | 1,895 | 95 | smolt |
| 62.51.50 | Jan. 95 | fingerling | 124 | 18.6 | 1,993 | 1,975 | 99.8 | 99.0 | 1,736 | 95 | smolt |
| 62.51.53 | Feb. 95 | fingerling | 124 | 18.6 | 452 | 448 | 99.1 | 99.1 | 444 | 95 | $s \mathrm{molt}$ |
| 62.51.54 | Feb. 95 | fingerling | 124 | 18.6 | 3,230 | 3,215 | 99.5 | 97.0 | 3,119 | 95 | smolt |

(1) Number actually tagged after running fish through quality control device.
(2) Percent retention is estimated by randomly capturing 500 fish $10-20$ days after tagging and counting the number still tagged.
(3) Number cwt released is the number of fish released after mortality.

Appendix B. Total number of kokanee salmon released into Lake Roosevelt from 1992 to 1995. Numbers taken from Appendix C.

| STAGE © RELEASE | $\begin{gathered} \text { CWT } \\ (n) \\ \hline \end{gathered}$ | $\begin{gathered} 1992 \\ \text { AD ONLY } \\ (n) \\ \hline \end{gathered}$ | TOTAL <br> (n) | $\begin{gathered} \text { CWT } \\ (\mathrm{n}) \\ \hline \end{gathered}$ | $\begin{gathered} 1993 \\ \text { AD ONL } \\ (\mathrm{n}) \\ \hline \end{gathered}$ | TOTAL <br> (n) | $\begin{aligned} & \text { C WT } \\ & (\mathrm{n}) \\ & \hline \end{aligned}$ | $\begin{gathered} 1994 \\ \text { AD ONLY } \\ (\mathrm{n}) \\ \hline \end{gathered}$ | TOTAL <br> (n) | $\begin{aligned} & \text { CWT } \\ & (\mathrm{n}) \\ & \hline \end{aligned}$ | $\begin{gathered} 1995 \\ \text { AD ONLY } \\ (n) \\ \hline \end{gathered}$ | TOTAL (n) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FRY | 171,452 | 21,983 | 193,435 | 241,952 | 3,105 | 245,057 | 59,899 | 8,174 | 68,073 | 51,411 | 4,094 | 55,505 |
| SMOLT | 132,029 | 0 | 132,029 | 80,468 | 1,845 | 82,313 | 137,457 | 5,225 | 142,682 | 369,106 | 16,944 | 386,050 |

Appendix C. Summary of marked kokanee salmon released into Lake Roosevelt from 1992 to 1995.

334

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | Exposure Stage | $\begin{aligned} & \hline \text { CWT } \\ & \text { Code } \end{aligned}$ | Number Tagged ( n ) | Adipose clipped only (n) | Release Location | Stage © <br> Release | Year Released | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 90 | M 0 R | Smolt | 62.51 .12 | 1,501 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | 5 molt | 62.51.13 | 2,525 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | Smolt | 62.51.1 4 | 5,392 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | 5 molt | 62.51.15 | 1,796 |  | Sherman Cr | Smolt | 92 |  |
| 90 | MOR | Smolt | 62.51.16 | 3,734 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | Smolt | 62.51.17 | 5,691 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | Smolt | 62.51.18 | 4,491 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | MOR | Smoit | 62.51.19 | 3,492 |  | Sherman Cr | 5 molt | 92 |  |
| '90 | PEA | Smolt | 62.51.13 | 4,855 |  | Sherman Cr | Smolt | 92 |  |
| 90 | PEA | Smolt | 62.51.14 | 1,665 |  | Sherman Cr | Smolt | 92 |  |
| 90 | PEA | Smolt | 62.51.1 5 | 1,717 |  | Sherman Cr | Smolt | 92 |  |
| 90 | PEA | Smoit | 62.51.16 | 6,769 |  | Sherman Cr | Smolt | 92 |  |
| 90 | PEA | Smolt | 62.51.17 | 5,471 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | PEA | 5 molt | 62.51.18 | 1,535 |  | Sherman Cr | Smolt | 92 |  |
| 90 | PEA | Smolt | 62.51.19 | 9, 215 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | PEA | Smolt | 62.51.21 | 5,143 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | PEA | Smolt | 62.51.22 | 3,211 |  | Sherman Cr | 5 molt | 92 |  |
| 90 | NONE |  | 62.51.12 | 9, 756 |  | Blue Cr | smoit | 92 | captive brood |
| 90 | NONE | - | 62.51.20 | 1,382 |  | Sherman Cr | Smolt | 92 | captive brood |
| 90 | NONE | - | 62.51 .20 | 3,153 |  | Lit falls | 5 molt | 92 | captive brood |
| 90 | NONE | - | 62.51.21 | 6,299 |  | Sherman Cr | smolt | 92 | captive brood |
| 90 | NONE | - | 62.51.22 | 4,124 |  | Sherman Cr | 5 molt | 92 | captive brood |
| 90 | NONE | - | 62.51.22 | 4,075 |  | Lit Falls | 5 molt | 92 | captive brood |
| 90 | NONE | - | 62.51.23 | 1,872 |  | Sherman Cr | 5 molt | 92 | captive brood |
| 90 | NONE | - | 62.51.23 | 9,159 |  | Lit Falls | 5 molt | 92 | captive brood |
|  |  |  | TOTAL | 132,029 | 0 |  |  |  |  |

Appendix C. Continued.

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | $\begin{gathered} \text { Exposure } \\ \text { Stage } \end{gathered}$ | $\begin{aligned} & \hline \text { CWT } \\ & \text { Code } \end{aligned}$ | Number Tagged ( n ) | Adipose clipped only ( n ) | Release Location | Stage (a) <br> Release | Year Released | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 91 | M 0 R | eye.hatch | 62.51.28 | 2,967 | 225 | Sherman Cr | fry | 92 |  |
| 91 | MOR | eye.hatch | 62.51.44 | 3,507 | 668 | Sherman Cr | fry | 92 |  |
| 91 | PEA | eye.hatch | 62.51.21 | 10,595 | 798 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | MOR | hatch | 62.51.30 | 10,169 | 1,006 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | MOR | hatch | 62.51.32 | 10,053 | 994 | Sherman Cr | fry | 92 | Ad,RV2,000 |
| 91 | PEA | hatch | 62.51.29 | 10,665 | 803 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | PEA | hatch | 62.51.31 | 10,599 | 1,048 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | MOR | h. su | 62.51.37 | 10,411 | 1,030 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | PEA | h. su | 62.51.33 | 9, 455 | 1,413 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | M 0 R | $s$ wl mup | 62.51.36 | 7,617 | 753 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | PEA | swimup | 62.51.35 | 9,323 | 1,393 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | M 0 R | Feb.fry | 62.51.24 | 4,627 | 881 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | M 0 R | Feb.fry | 62.51.25 | 6,247 | 1,190 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | M 0 R | Feb.fry | 62.51.26 | 6,089 | 1,160 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | PEA | Feb-fry | 62.51.34 | 5,242 | 783 | Sherman Cr | fry | 92 | Ad, RV 2,000 |
| 91 | M 0 R | Mar-fry | 62.51.38 | 8,916 | 882 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | PEA | Mar-fry | 62.51.39 | 9, 520 | 1,298 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | MOR | Apr.fry | 62.51 .40 | 10,072 | 1,373 | Sherman Cr | fry | 92 | Ad,RV 2, 000 |
| 91 | PEA | Apr.fry | 62.51.41 | 10,142 | 1,383 | Sherman Cr | fry | 92 | Ad,RV 2,000 |
| 91 | MOR | May.fry | 62.51.42 | 5,744 | 1,094 | Sherman Cr | fry | 92 |  |
| 91 | PEA | May.fry | 62.51.43 | 9,492 | 1,808 | Sherman Cr | fry | 92 |  |
|  |  |  | TOTAL | 171,452 | 21,983 |  |  |  |  |
| 91 | PEA | 5 molt | 62.51.54 | 8,196 | 184 | Blue Cr | $s \mathrm{molt}$ | 93 |  |
| 91 | PEA | 5 molt | 62.51 .48 | 732 | 19 | Sherman Cr | $s \mathrm{molt}$ | 93 |  |
| 91 | PEA | Smolt | 62.51 .49 | 3,454 | 89 | Sherman Cr | smolt | 93 |  |
| 91 | PEA | Smoit | 62.51.50 | 3,567 | 91 | Sherman Cr | $s \mathrm{molt}$ | 93 |  |
| 91 | MOR | 5 molt | 62.51.45 | 12,396 | 318 | Shermancr | $s \mathrm{molt}$ | 93 |  |
| 91 | M 0 R | 5 molt | 62.51 .46 | 12,664 | 325 | Sherman Cr | smolt | 93 |  |
| 91 | M OR | 5 molt | 62.51.47 | 12,970 | 333 | Shermancr | $s \mathrm{molt}$ | 93 |  |
| 91 | NONE | - | 62.51.51 | 9, 751 | 179 | Sherman Cr | smolt | 93 |  |
| 91 | NONE | - | 62.51.52 | 9, 800 | 180 | Sherman Cr | smolt | 93 |  |
| 91 | NONE | - | 62.51 .53 | 6,938 | 127 | Sherman Cr | smoll | 93 |  |

Appendix C. Continued.

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | Exposure Stage | CWT Code | Number Tagged (n) | Adipose clipped only ( n ) | Release Location | Stage © Release | Year Released Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 92 | PEA | hatch | 62.52.32 | 325 |  | Chamokane Cr 2 | 2 year old | 94 |
| 92 | PEA | eye.hatch | 62.52.31 | 325 |  | Chamokane Cr 2 | year old | 94 |
| 92 | MOR | surel | 62.52.07 | 10,870 | 121 | Sherman Cr | fry | 93 |
| 92 | MOR | suret | 62.51.55 | 10,802 | 266 | Sherman Cr | fry | 93 |
| 92 | PEA | surel | 62.52.06 | 10,896 | 121 | Sherman Cr | fry | 93 |
| 92 | M 0 R | eye.hatch | 62.51 .56 | 10,961 | 269 | Sherman Cr | fry | 93 |
| 92 | MOR | eye.hatch | 62.52.09 | 3,394 | 38 | Sherman Cr | fry | 93 |
| 92 | MOR | eye.hatch | 62.52 .09 | 7,509 | 53 | Sherman Cr | fry | 93 |
| 92 | PEA | eye.hatch | 62.51.57 | 10,721 | 264 | Sherman Cr | fry | 93 |
| 92 | PEA | eye.hatch | 62.52.10 | 10,960 | 11 | Sherman Cr | fry | 93 |
| 92 | PEA | eye.hatch | 62.52.1 6 | 10,863 | 121 | Sherman Cr | fry | 93 |
| 92 | MOR | hatch | 62.52 .13 | 11,001 | 78 | Sherman Cr | fry | 93 |
| 92 | MOR | hatch | 62.52.1 7 | 10,863 | 121 | Sherman Cr | fry | 93 |
| 92 | MOR | hatch | 62.52 .14 | 10,916 | 11 | Sherman Cr | fry | 93 |
| 92 | MOR | hatch | 62.52.1 5 | 9, 499 | 67 | Sherman Cr | fry | 93 |
| 92 | MOR | hatch | 62.51.59 | 10,086 | 221 | Sherman Cr | fry | 93 |
| 92 | PEA | hatch | 62.51.58 | 10,767 | 265 | Sherman Cr | fry | 93 |
| 92 | PEA | hatch | 62.52.11 | 10,971 | 78 | Sherman Cr | fry | 93 |
| 92 | PEA | hatch | 62.52 .12 | 11,022 | 78 | Sherman Cr | fry | 93 |
| 92 | MOR | h.su | 62.51 .60 | 10,938 | 122 | Sherman Cr | fry | 93 |
| 92 | PEA | h.su | 62.51 .61 | 11,791 | 144 | Sherman Cr | fry | 93 |
| 92 | MOR | swimup | 62.52 .03 | 10,908 | 121 | Sherman Cr | fry | 93 |
| 92 | PEA | swimup | 62.52 .05 | 10,885 | 121 | Sherman Cr | fry | 93 |
| 92 | MOR | swimup | 62.52.18 | 2,712 | 31 | Barnaby Cr | fry | 93 |
| 92 | MOR | h.su | 62.52 .18 | 2,712 | 30 | Barnaby Cr | fry | 93 |
| 92 | PEA | eye.hatch | 62.52 .18 | 2,712 | 30 | Barnaby Cr | fry | 93 |
| 92 | MOR | eye.hatch | 62.52.18 | 2,712' | 30 | Barnaby Cr | fry | 93 |
| 92 | PEA | swimup | 62.52 .19 | 3,637 | 41 | Barnaby Cr | fry | 93 |
| 92 | M 0 R | hatch | 62.52 .19 | 3,637 | 40 | Barnaby Cr | fry | 93 |

Appendix C. Continued.

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | Exposure Stage | CWT Code | Number Tagged (n) | Adipose clipped only ( n ) | Release Location | Stage © <br> Release | Year Released | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 92 | M 0 R | h.su | 62.52.19 | 3,637 | 40 | Barnaby Cr | fry | 93 |  |
| 92 | PEA | swimup | 62.52.20 | 1,190 | 13 | Barnaby Cr | fry | 93 |  |
| 92 | M 0 R | hatch | 62.52.20 | 1,190 | 13 | Barnaby Cr | fry | 93 |  |
| 92 | M 0 R | h. su | 62.52.20 | 1,190 | 14 | Barnaby Cr | fry | 93 |  |
|  |  |  | TOTAL | 241,952 | 3,105 | 246,057 |  |  |  |
| 92 | M 0 R | eye.hatch | 62.52.31 | 10,291 | 975 | ${ }^{\text {Blue }} \mathrm{Cr}$ | smolt | 94 |  |
| 92 | MOR | eye.hatch | 62.52.32 | 3,334 | 91 | Blue Cr | smolt | 94 |  |
| 92 | MOR | hatch | 62.52.30 | 10,613 | 338 | Sherman Cr | smolt | 94 |  |
| 92 | MOR | hatch | 62.52.22 | 2,822 | 46 | Sherman Cr | smolt | 94 |  |
| 92 | PEA | hatch | 62.52.32 | 1,806 | 130 | A.Frame | $s \mathrm{molt}$ | 94 |  |
| 92 | PEA | eye.hatch | 62.52.33 | 8,352 | 132 | A.Frame | smolt | 94 |  |
| 92 | MOR | h.su | 62.52.26 | 4,604 | 232 | Sherman Cr | smolt | 94 |  |
| 92 | MOR | h. su | 62.52.29 | 10,919 | 546 | Sherman Cr | smolt | 94 |  |
| 92 | MOR | swimup | 62.52.21 | 10,979 | 274 | Sherman Cr | smolt | 94 |  |
| 92 | MOR | $s$ wl mup | 62.52.22 | 6,938 | 190 | Sherman Cr | smolt | 94 |  |
| 92 | PEA | swimup | 62.52.23 | 10,653 | 475 | Sherman Cr | smoll | 94 |  |
| 92 | PEA | swimup | 62.52.24 | 10,405 | 694 | Sherman Cr | smolt | 94 |  |
| 92 | PEA | su-fry | 62.52.25 | 10,886 | 282 | Sherman Cr | smoll | 94 |  |
| 92 | PEA | su.fry | 62.52.26 | 6,271 | 198 | Sherman Cr | smolt | 94 |  |
| $\begin{aligned} & 92 \\ & 92 \end{aligned}$ | $\begin{aligned} & \text { NONE } \\ & \text { NONE } \end{aligned}$ | . | $\begin{aligned} & 62 \cdot 52 \cdot 27 \\ & 62.52 .28 \end{aligned}$ | $\begin{aligned} & 11,256 \\ & 11,328 \end{aligned}$ | $\begin{aligned} & 241 \\ & 381 \end{aligned}$ | KF Net Pen <br> KF Net Pen | smolt smolt |  | captive brood captive brood |
|  |  |  | TOTAL | 137,457 | 5,225 | 142,662 |  |  |  |
| 93 | MOR | al.sul | 111.2 .8 | 9,780 | 2,970 | Sherman Cr | fir | 94 |  |
| 93 | M 0 R | al.su | 11.1 .2 .9 | 8,916 | 1,216 | Sherman Cr | fry | 94 |  |
| 93 | MOR | h. su | 62.52.35 | 10,114 | 1,312 | Sherman Cr | fry | 94 |  |
| 93 | M 0 R | h. su | 62.52.36 | 10,147 | 1,151 | Sherman Cr | fry | 94 |  |
| 93 | PEA | al.su | 62.52.37 | 10,475 | 174 | Sherman Cr | fry | 94 |  |
| 93 | PEA | al.su | 62.52.38 | 10,467 | 751 | Sherman Cr | fry | 94 |  |
|  |  |  | TOTAL | 59,699 | 6,174 | 66,073 |  |  |  |

## Appendix C. Cqntlnued.

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | $\begin{gathered} \text { Exposure } \\ \text { Stage } \end{gathered}$ | $\begin{aligned} & \hline \text { CWT } \\ & \text { Code } \end{aligned}$ | Number Tagged <br> ( n ) | Adipose clipped only(n) | Release Location | Stage © <br> Release | Year Released | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 93 | PEA | al-sw/smoll | 62.51.25 | 1,560 | 20 | Sherman Cr | smoll | 95 |  |
| 93 | PEA | h -su/smoll | 62.53.41 | 10,293 | 704 | Sherman Cr | smoll | 95 |  |
| 93 | PEA | h -su/smoll | 62.53.42 | 10,197 | 697 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h-su/smoll | 62.53.43 | 10,244 | 671 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h-su/smoll | 62.53.44 | 10,228 | 688 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h -su/smoll | 62.53.45 | 10,301 | 728 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h -su/smoll | 62.53.46 | 10,204 | 674 | Sherman Cr | smoll | 95 |  |
| 93 | PEA | h-su/smolt | 62.53.52 | 10,457 | 424 | Sherman Cr | smoll | 95 |  |
| 93 | PEA | h -su/smolt | 62.53.53 | 10,590 | 418 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h -su/smolt | 62.53.54 | 10,576 | 418 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h-su/smolt | 62.53.55 | 10,380 | 432 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h-su/smoll | 62.53.58 | 10,725 | 275 | Sherman Cr | smolt | 95 |  |
| 93 | PEA | h-su/smolt | 62.53.59 | 10,711 | 263 | Sherman Cr | smolt | 95 |  |
| 93 | NONE |  | 62.51.28 | 8,210 | 165 | Spokane R | smolt | 95 | captive brood |
| 93 | NONE | - | 62.51.34 | 5,383 | 100 | Spokane R | smolt | 95 | captive brood |
| 93 | NONE | - | 62.51.42 | 4,552 | 160 | Spokane R | smolt | 95 | captive brood |
| 93 | NONE |  | 62.51 .49 | 8,210 | 165 | Spokane R | smolt | 95 | captive brood |
| 93 | none | - | 62.51.50 | 1,736 | 257 | Spokane R | smolt | 95 | captive brood |
| 93 | NONE |  | 62.51.53 | 444 | 8 | Spokane R | smolt | 95 | captive brood |
| 93 | none |  | 62.51.54 | 3,119 | 111 | Spokane R | smoit | 95 | captive brood |
| 93 | NONE | - | 62.51.24 | 5,430 | 102 | Barnaby Cr | smolt | 95 | captive brood |
| 93 | NONE | - | 62.51.26 | 5,423 | 174 | Barnaby Cr | smolt | 95 | captive brood |
| 93 | none | - | 62.51.48 | 10,681 | 349 | Barnaby Cr | smolt | 95 | captive brood |
| 93 | NONE |  | 62.52 .40 | 10,295 | 573 | KF Net Pen | smolt | 95 | captive brood |
| 93 | M OR | h.su | 62.52.41 | 10,501 | 680 | KF Net Pen | smoit | 95 |  |
| 93 | MOR | h.su | 62.53.35 | 10,522 | 667 | KF Net Pen | smolt | 95 |  |
| 93 | M 0 R | h. su | 62.53.36 | 10,539 | 669 | KF Net Pen | smolt | 95 |  |
| 93 | PEA | h. su | 62.53.37 | 10,609 | 609 | KF Net Pen | smolt | 95 |  |
| 93 | PEA | h. su | 62.53.38 | 10,522 | 574 | KF Net Pen | smolt | 95 |  |
| 93 | PEA | h. su | 62.53.39 | 10,619 | 568 | KF Net Pen | $s \mathrm{molt}$ | 95 |  |
| 93 | PEA | h.su | 62.53.40 | 10,616 | 579 | KF Net Pen | smolt | 95 |  |
| 93 | M 0 R | al.su | $62 \cdot 53 \cdot 56$ | 10,542 | 439 | KF Net Pen | smolt | 95 |  |

Appendix C. Continued.

| Cohort | $\begin{aligned} & \text { Exposure } \\ & \text { Odor } \end{aligned}$ | Exposure Stage | CWT Code | Number Tagged (n) | Adipose clipped only ( n ) | Release Location | Stage © <br> Release | Year Released | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 93 | NONE | - | 62.52.39 | 10,130 | 736 | KF Net Pen | smolt | 95 | captive brood |
| 93 | M 0 R | h-su/smolt | 62.51.44 | 5,637 | 68 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | h-su/smolt | 62.51.63 | 10,939 | 212 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | al-su/smolt | 62.53.47 | 10,224 | 426 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | al-su/smolt | 62.53.48 | 10,410 | 400 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | al-su/smolt | 62.53.49 | 10,363 | 432 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | al-su/smolt | 62.53.50 | 10,285 | 417 | Sherman Cr | smolt | 95 |  |
| 93 | M 0 R | al-su/smolt | 62.53.51 | 10,173 | 413 | Sherman Cr | smolt | 95 |  |
|  |  |  | TOTAL | 369,106 | 16,944 | 366,050 |  |  |  |
| 94 | M 0 R | h.su | 62.54.37 | 9,923 | 932 | Sherman Cr | fry | 95 |  |
| 94 | M 0 R | h.su | 62.54.38 | 10,271 | 881 | Sherman Cr | fry | 95 |  |
| 94 | M 0 R | h.su | 62.54.39 | 10,437 | 960 | Sherman Cr | fry | 95 |  |
| 94 | M 0 R | h.su | 62.54.40 | 9, 837 | 935 | Sherman Cr | fry | 95 |  |
| 94 | M 0 R | h. su | 62.54.48 | 10,943 | 386 | Chamokane Cr | fry | 95 |  |
|  |  |  | TOTAL | 51,411 | 4,094 | 55,505 |  |  |  |

Appendix D. Recommended release strategy for 1995 cohort kokanee to be released as residualized smolts in 1997.

| Stock | Exposure <br> Chemical | Number <br> Released | Release <br> Site |
| :---: | :---: | :---: | :---: |
| Captive brood | None | 97,000 | Little Falls |
| Lake Whatcom | MOR | 241,000 | Sherman Creek |
| Lake Whatcom | MOR | 50,000 | Net Pens |


[^0]:    * Indicates that this species was observed in 1995.

[^1]:    '.-' Indicates no data collected or no value.

[^2]:    * Indicates suspect pH readings due to sensor malfunction.

[^3]:    *Includes yellow perch, largemouth bass, suckers, squawfish, black crappie, chinook, bullhead, etc...

[^4]:    Disclaimer: Use of brand names in this report does not imply endorsement by BPA, Eastern Washington University, Upper Columbia United Tribes or the authors.

[^5]:    In 1995, both morpholine (MOR) and phenethyl alcohol (PEA) were dripped at Sherman Creek, PEA was also dripped into the Spokane River at Little Falls.
    One of these fish was recovered as a 4 year old in 1995 .
    These fish were recovered in 1992 to 1994 as 2 to 4 year old fish respectively.

[^6]:    In 1995, both MOR and PEA were dripped into Sherman Creek.
    $t$ This fish was recovered as a three year old in 1995.
    i One of these fish was recovered as a four year old in 1995.

[^7]:    11992 cohort fish were recovered at age 2 and 3 in 1994 and 1995 respectively. 1993 cohort fish were recovered as age 2 in 1995.
    2 In 1995, both MOR and PEA were dripped at Sherman Creek, PEA was also dripped into the Spokane River at Little Falls Dam.

[^8]:    1 An additional 160,000 1994 cohort fry were released into Sherman Creek. These fish were unmarked and unexposed. Fish from the 1994 cohort are not expected to return until 1996. 1997 and 1998 as 2, 3 and 4 year olds respectively. Fish from the 1993 cohort returned as age 2 in 1995.
    2 Bamaby Creek is located approximately 11 miles south of Sherman Creek.
    3 The Kettle Falls net pen is located at the Kettle Falls Marina.

[^9]:    1 In 1995, both morpholine and phenethyl alcohol were dripped at Sherman Creek, PEA was also dripped in the Spokane River at Little Falls Dam.

[^10]:    * Weight is based on only one fish.

